Draft Report Evaluation of Dredged Material Proposed for Open-Water Placement at Site 104, Chesapeake Bay, Maryland

Prepared for

Department of the Army Corps of Engineers Baltimore District 10 South Howard Street Baltimore, Maryland 21201



Prepared by EA Engineering, Science, and Technology, Inc. 15 Loveton Circle Sparks, Maryland 21152



EA Engineering, Science, and Technology, Inc. October 2000

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25 October 2000

Christina E. Correale Chief - Operations Division USACE Baltimore District P.O. Box 1715 10 North Howard Street - Room 8620 Baltimore, MD 21203-1715

RE: Draft Sediment Report - Site 104, Chapters 1-7

Dear Ms. Correale:

Accompanying this letter please find the draft of Chapters 1-7 of the above referenced report. These have been carefully reviewed by senior technical scientists and by our senior editor. Chapter 7 has been slightly modified from that which we gave to Jeff McKee on Monday following an additional final review. We are currently in the process of completing the remaining chapters and will forward them to you as soon as they are completed. Please feel free to call either Peggy Derrick or me with any questions.

Sincerely,

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Peggy Derrick, Project Manager/Scientist

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Frank W. Pine, Ph.D. Senior Scientist/Project Director

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CHAPTER 4 -FIELD SAMPLING PROGRAM

4. FIELD SAMPLING PROGRAM

4.1 OVERVIEW OF FIELD SAMPLING ACTIVITIES

The field effort for the Site 104 dredged material evaluation consisted of two separate rounds of sampling and subsequent chemical and biological testing. Mobilization for the first round of approach channel and Site 104 sediment sampling commenced on 13 September 1999. Sample collection was initiated on 15 September 1999 and continued through 28 September 1999. A total of 56 stations in the upper Chesapeake Bay were successfully sampled (Figures 4-1 through 4-13). Grab samples were collected at 26 locations in six reaches using a stainless steel Van Veen grab sampler, and sediment cores were collected at 30 locations in eight reaches with targeted lengths ranging from 2.5 to 15 ft, using a vibracoring system. A combined total of 30 grabs and 141 cores (one from each coring reach) were submitted to the U.S. Army Engineer Research and Development Center–Waterways Experiment Station (WES) for particle settling tests.



In response to recommendations from USEPA Region III–Philadelphia for additional chemical and biological testing, a second round of sampling was conducted during the period of 8-15 December 1999. Sampling locations included those locations sampled in September 1999 in addition to an Outside Site 104 reference area. A total of 55 stations were successfully sampled in the second round of sampling. Grab samples were collected at 25 locations using a stainless steel Van Veen grab sampler, and sediment cores from 30 locations with targeted lengths of 2.5 to 15 ft, using a vibracoring system. A combined total of 29 grabs and 92 cores were collected from 15 sampling reaches in the second round of sampling.

Sediments from each station were submitted for chemical analyses, and a sediment composite from each sampling reach was prepared and submitted for biological testing. For both rounds of sampling, site water from each sampling reach was collected and submitted for the preparation of reach elutriates that were targeted for chemical and biological testing. In addition, receiving water samples, targeted for chemical analysis, were collected from Inside Site 104 during both rounds of sampling, and from Outside Site 104 during the second round of sampling.

Sediment and site water from the NODS reference area were collected on 01 February 2000 in conjunction with the Woodrow Wilson Bridge sediment sampling and testing program (EA 2000b).

The project Field Sampling Plan (FSP) (Appendix A) describes the field sampling and datagathering methods for the Approach Channels and Site 104 project. The FSP was prepared following guidance provided by the USACE Engineer Manual (EM) 200-1-3 *Requirements for the Preparation of Sampling and Analysis Plans* (1994). Sampling methodologies for sediment and water collection at an Atlantic Ocean reference site are documented in the FSP for the Woodrow Wilson Bridge Dredged Material Evaluation (EA 2000b).

4.2 SAMPLING OBJECTIVES

The field investigations consisted of obtaining sediment cores and surficial sediments from within the Baltimore Harbor approach channels, from inside and outside Placement Site 104, and from the NODS reference area. In addition to collecting sediment, site water was collected for chemical analysis, elutriate preparation, and bioassay testing. Samples were submitted to Severn Trent Laboratories–Baltimore (STL–Baltimore) for physical and chemical analysis, to EA's Ecotoxicology Laboratory for ecotoxicological testing, and to the U.S. Army Waterways Experiment Station (USACE–WES) for specialized physical testing.

The overall objective of the sampling effort was to obtain and analyze sediment and water samples representative of the areas proposed for dredging during the period that Placement Site 104 would be used for placement.

The specific objectives of the field sampling and sample processing were:

Field Sampling

- Collect sediment cores from within channels proposed for upcoming maintenance dredging and new work dredging to depths ranging from -2.5 ft. to -15 ft. below the sediment surface (30 locations).
- Collect surficial sediments from within channels that have been recently dredged (17 locations), in new work areas (2 locations), from within Placement Site 104 (5 locations), and from areas immediately surrounding Placement Site 104 (outside reference 4 locations).
- Obtain the required sediment volume necessary for physical, chemical, and biological testing.
- Collect one additional core in each channel reach (eight coring reaches only) for particle settling tests to be conducted by the USACE–WES.
- Sample 4 additional stations in the vicinity (north, south, east, and west) of station KI-7 in the southern portion of Placement Site 104 to roughly delineate the geographic extent of contamination (if possible), due to previous placement activity. Submit samples for physical and chemical analysis only.
- Collect site water for chemical analysis, elutriate preparation, and bioassay testing.
- Submit equipment blanks for analytical testing.

- Transport sediment cores to EA's office in Sparks, Maryland under temperaturecontrolled conditions (4°C) and according to the requirements of chain-of-custody protocols.
- Complete appropriate chain-of-custody documentation.

Sample Processing

- Extrude sediment from core liners.
- Composite and homogenize core sediments and grab (surficial) sediments according to protocols that ensure sample integrity.
- Distribute homogenized and/or composited sediment samples into appropriate containers for submittal to appropriate laboratories (either EA's Ecotoxicology Laboratory, STL-Baltimore, or USACE-WES).
- Complete appropriate chain-of-custody documentation.

4.3 STATION LOCATION DETERMINATION

4.3.1 Station Locations

Stations were located in channels proposed for maintenance dredging (28 locations), in channels recently dredged (17 locations), in new work areas (2 locations), inside Placement Site 104 (5 locations) and outside Placement Site 104 (4 locations). Channels proposed for dredging in FY00 and FY01 include Tolchester (Figure 4-9), Craighill Entrance (Figure 4-5), Craighill Angle (Figure 4-4), Swan Point (Figure 4-8), and northern C&D approaches (Figure 4-2). The number of stations sampled in each channel was determined by the volume of material proposed for removal. For this project, one station was sampled for every 100,000 cubic yards (cy) of material proposed for removal. The channels proposed for dredging in FY00 and FY01, the estimated volume of material to be dredged, and the number of stations to be sampled in each FY00 and FY01 channel are provided in Table 4-1. Station locations within the channels were chosen in consultation with USACE–Baltimore, and locations were targeted to specific channel areas proposed for dredging (i.e., shoaled locations).

The Craighill Angle (Figure 4-4) and the Tolchester Channel (Figure 4-9) were each broken into two sampling/testing reaches. The material proposed for dredging in these channels is geographically separated (east and west sides of the Craighill Angle; northern and southern areas of the Tolchester Channel). Dredging in these channels would likely result in separate placement events for the east/west (Craighill Angle) and north/south (Tolchester) regions; therefore, the biological testing results represent the potential impacts that could be associated with the separate placement events for each geographic region of the channels.



For channels that have been recently dredged (Craighill (Figure 4-3), Craighill Upper Range (Figure 4-6), Cutoff Angle (Figure 4-7), Brewerton Eastern Extension (Figure 4-1), and the southern C&D approach channel (Figure 4-2), the sampling locations corresponded to previously sampled locations in the FY95 and FY98 sediment characterization studies (EA 1996a and 2000c). Sampling locations in the C&D approach channels were chosen by USACE–Baltimore in consultation with USACE–Philadelphia (CENAP).

The C&D approach channel was divided into two sampling/testing reaches. The northern reach is proposed for dredging in FY00 or FY01, and the southern reach has been dredged recently. Biological testing was conducted separately for the northern and southern areas.

New work dredging was represented by two sampling locations adjacent to the Tolchester Channel where straightening of the channel S-Turn is proposed (Figure 4-10). These stations were targeted for vibracoring to a depth of 15 ft below the sediment surface.

Sampling locations inside Site 104 (Figure 4-11) included stations KI-3, KI-5, and KI-7 from the previous sediment characterization study (EA 1998a) and two additional stations located in the southern portion of the site (KI-S-1 and KI-S-2) where placement of material is proposed. Five additional samples from the southern portion of Site 104 in the vicinity of KI-7 (Figure 4-12) were collected and analyzed in an attempt to roughly delineate the extent of the previously documented contamination from KI-7 (KI-7REF, KI-7-N, KI-7-S, KI-7-E, and KI-7-W). Sampling locations outside Site 104 (Figure 4-13) corresponded to stations KI-11, KI-14, KI-15, and KI-16 that were sampled in a previous sediment investigation (EA 1998a).

Sampling in the Ocean Reference area consisted of grab sediment collection at each of four locations within the reference area (Figure 4-14). The sediment from the four locations was combined into one composite sample for physical, chemical, and biological testing.

Station locations for each sampling reach are depicted in alphabetical order in Figures 4-1 through 4-10. Station locations for Inside Site 104 and Outside Site 104 are depicted in Figures 4-11 and 4-13, respectively. Sampling station coordinates are provided in Table 4-2.

4.3.2 Global Positioning System Equipment

Core sampling was conducted from two vessels: the USACE vessel *Reynolds* and the University of Maryland (UMD) vessel *Aquarius*. For the first round of core sampling conducted from the USACE vessel *Reynolds*, stations were located using a Differential Global Positioning System (DGPS) on the survey vessel *Linthicum*. The station locations were marked in the field by the *Linthicum* crew using the DGPS to locate the target position and by placing a buoy at the target location. The work platform (*Reynolds*) was positioned and maintained adjacent to the buoy to facilitate collection of cores at the target location. During the second round of sampling on the *Reynolds*, a Northstar DGPS that was located directly on the sampling vessel was utilized to locate target stations. The UMD *R/V Aquarius* utilized a Northstar 941XD DGPS for both the September and December sampling events.

4.4 SAMPLE VOLUME REQUIREMENTS

Approximately 20-25 gallons of sediment were required per sampling reach for the toxicity testing, bioaccumulation studies, sediment chemistry, and elutriate preparation. Approximately 137 cores and 22 grabs were required to meet the sediment volume requirements in the first round of sampling, and approximately 92 cores and 26 grabs were required to meet the sediment volume requirements in the second round of sampling. Eight additional cores were required in the first round of sampling for the WES particle settling tests. Fewer cores were required in the second round of sampling than in the first round of sampling due to greater sediment recovery during the December sampling effort. A station-by-station breakdown of the targeted core depth and quantity is provided in Table 4-3. The actual number of cores collected is provided in Tables 4-4A (September 1999) and 4-4B (December 1999). If sediment recovery was less than the target penetration depth, additional cores were collected to obtain the necessary sediment volume.

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4.5 VIBRACORING PROCEDURES



In channels proposed for upcoming maintenance dredging and new work dredging (8 sampling reaches), sediment samples were collected using a Rossfelder-P3 vibracoring system capable of collecting sediment cores ranging from 2.5 to 15 ft in length. The targeted core depth was representative of the material to be removed, varied between stations, and was dependent upon the proposed dredging depth (Table 4-3). Core penetration depths and recovery lengths for each station are provided in Tables 4-4A and 4-4B for the September and December sampling, respectively. Cellulose Acetate Butylrate (CAB) core liners with an outer diameter of 4 inches and an inner diameter of 3.75 inches were used within the coring device. The core liners were fitted with a stainless steel catcher at the bottom to retain sediment during retrieval. The core barrel was fitted with a steel cutter head to facilitate sediment penetration. Sampling equipment coming into direct contact with the sediment was decontaminated prior to sampling using the protocol described in Section 4.10.

Coring equipment was provided by Aqua Survey, Inc. (Flemington, New Jersey) for both the September and December sampling events. Coring operations were conducted from the USACE vessel *Reynolds* in September 1999 and from the University of UMD *R/V Aquarius* in December 1999. Both vessels were outfitted with lifting equipment and electrical hook-ups to facilitate coring operations. The project staging areas for the vibracoring activities were located at Fort McHenry (Baltimore, Maryland) and Sandy Point State Park (Whitehall, Maryland) for the September and December coring efforts, respectively.

The following procedure was followed for core collection at each station:

- A clean, decontaminated section of cellulose acetate butyrate plastic liner was fitted with a clean, decontaminated stainless steel core catcher at the bottom.
- The liner was placed inside the outer steel core barrel.

- A clean stainless steel core cutter, or nose cone, was placed at the bottom of the core barrel.
- The boat was positioned at the target location and anchored.
- The core barrel was lowered until the bottom of the barrel was just above the sediment surface.
- The vibracoring motor was engaged and the core barrel was penetrated into the sediment to the target depth.
- The approximate water depth and the position were recorded.
- After the position was recorded, the core barrel was brought up on deck.
- The core liner was removed from the corer, the core catcher was removed from the bottom of the core, a core cap was placed on the bottom of the core and taped in place. The core was moved into a vertical position and excess liner above the sediment-water interface was cut off using a hacksaw. The top of the core was capped and taped. The depth of actual sediment recovery was measured and recorded.
- The liner and both caps were labeled. Labeling included the following:
 - Station Location/Site
 - Core Number out of Total Number of Cores for Station
 - Reference to top or bottom
- The process was repeated at the site until the required sediment volume was attained.
- The boat was relocated to the next station, and the process repeated.

In addition to collecting cores for analytical and ecotoxicological testing, one additional core was collected within each sampling reach (coring reach) for USACE–WES. These cores were used for particle settling tests that are not part of this evaluation.

Cores were chilled with bags of ice and kept on-board the work vessel until the end of each workday. A summary of coring activities is provided in Tables 4-4A (September) and 4-4B (December). Copies of the field logbooks and data sheets are provided in Attachment I.

4.6 GRAB SAMPLING PROCEDURES

Surficial sediments were collected from channels that have recently been dredged, from inside Placement Site 104, and from several locations outside Placement Site 104 (7 sampling reaches).



Surficial sediments in the channel locations are most representative of materials to be dredged in the future, and surficial sediments inside Placement Site 104 are representative of the existing surficial conditions that will be disturbed as the result of placement. The surficial sediments in the Outside Site 104 area locations represent substrate used by biota and substrate to which biota are exposed. The Outside Site 104 sediments have not been influenced by previous dredged material placement activities.

The surficial sediments were collected using a stainless steel Van Veen grab sampler with a 10.5-gallon capacity. Grab sampling operations were conducted from the UMD's R/V Aquarius in September and from both the USACE vessel Reynolds (2 days) and the R/V Aquarius (1 day) in December. One grab was collected and composited from each station to obtain the necessary sample volume for physical, chemical, and biological analysis. Sediment targeted for analytical testing was thoroughly homogenized with a stainless steel spoon in a stainless steel mixing bowl on the work vessel. The homogenized sample was transferred to appropriate sample containers using a stainless steel spoon, samples were labeled, and chilled on ice on the work vessel. Sediment targeted for ecotoxicological testing was transferred directly from the sampling device to pre-cleaned, 5-gallon, polypropylene buckets with sealed lids. These samples were homogenized (composited) in the laboratory prior to initiation of ecotoxicological testing. The buckets were labeled and chilled with ice. In the second round of sampling, for samples targeted for volatiles analysis, a subsample from each station was directly removed from the undisturbed grab sampler and was placed immediately into the appropriate sample container prior to homogenization. To avoid the loss of volatiles from the sample, the volatiles sample container was filled to the top and capped, leaving minimal headspace and trapped air.

During the grab sampling effort inside Placement Site 104, an attempt was made to determine the extent of surficial contamination at Site 104 in the vicinity of station KI-7. Surficial grabs were collected at approximate 5-10 m increments to the north, south, east, and west of KI-7. Previous sampling (in 1997) at this location indicated that the sediments were visibly different than at other locations within the site (EA 1998a). Grabs were inspected visually and stirred to reveal signs of visual contamination (i.e., oily residue or sheen, petroleum-like odors, unusual textures, foreign objects, etc).

No visual signs of contamination were evident in the surficial sediments collected at KI-7 during the September sampling effort (unlike the 1997 sampling event). Visual signs of contamination were observed, however, in surficial sediment that was collected approximately 10-15 m south of KI-7. Material from this location (KI-7REF) was archived and submitted for analytical testing. In addition, sub samples of sediment from KI-7REF were included in the sediment composite for the ecotoxicological testing and bioaccumulation studies.

Grabs were then collected at the 5-10 m distances (north, south, east, and west of KI-7REF) until no visual signs of surficial contamination existed. When the apparent boundary of surficial contamination was delineated, four grabs were collected (one to the north, south, east, and west) immediately outside the boundary for chemical analysis. These grabs were not utilized in the Inside Site 104 sediment composite for biological testing. In addition, site water was collected and an elutriate sample was created using sediment from KI-7REF. The intent of the elutriate



test was to provide information regarding the potential release of contaminants from resuspension of bottom sediments during material placement.

Summaries of the grab sampling activities are provided in Tables 4-5A and 4-5B for the September and December sampling events, respectively.

4.7 OCEAN REFERENCE SITE SAMPLING

Surficial sediments were collected the Norfolk Ocean Disposal Site reference area on 01 February 2000 (Figure 4-14). The sampling was conducted from a 45-ft vessel supplied and operated by Sea Search of Norfolk, Virginia. The surficial sediment was collected using a stainless steel Van Veen grab sampler with a 10.5-gallon capacity. The sampler was decontaminated prior to use following protocols described in Section 4.10. Four grabs were collected and composited into one sample for the analytical and ecotoxicological testing. The sample submitted for analytical testing consisted of an equal volume of surficial sediment subsampled from four grabs. The sample was composited and thoroughly homogenized with a stainless steel spoon in a stainless steel mixing bowl. The sample for volatiles analysis was removed from one undisturbed grab (prior to homogenizing) and was placed immediately into the appropriate sample container. To avoid the loss of volatiles from the sample, the volatiles sample container was filled to the top and capped leaving minimal headspace and trapped air. The containers were labeled, chilled on ice on the work vessel, and then hand delivered to STL-Baltimore for analytical testing. Holding times for the grab samples began when the samples were collected.

For the ecotoxicological testing, 5 gallons of sediment from each of the four grabs were archived individually in pre-cleaned, polypropylene buckets with sealed lids. The buckets were labeled and chilled on ice on the work vessel. The sediment was then transported to EA's ecotoxicological facility, stored at 4°C, and later composited and homogenized prior to ecotoxicological testing.

In addition to collecting ocean reference sediment, USEPA Region III requested that a lower Chesapeake Bay control sediment be collected for use as the laboratory control in the *Leptocheirus plumulosus* amphipod testing. Four grabs were also collected at a control site located approximately 5 miles northwest of Fisherman's Island in the southern Chesapeake Bay (Figure 4-14). The sediments were composited into one sample and were tested as the laboratory control in the whole-sediment bioassays only. A summary of station information for the ocean reference and lower Chesapeake Bay control site sampling is provided in Table 4-6.

4.8 SAMPLE STORAGE AND TRANSPORT

In the field, cores were capped at both ends, sealed, labeled, and bagged on the work vessel. Cores were chilled on the work vessel using bagged ice until the end of the workday. Cores were hand-delivered each evening to EA's facility in Sparks, Maryland where they were stored at 4°C until processing. A chain-of-custody form accompanied the cores during transport to Sparks, Maryland. The chain-of-custody form documented core name and date and time of collection. Copies of the chain-of-custody forms for the cores are provided in Attachment I.

For surficial sediments, one grab sample from each station was composited on-board the work vessel using pre-cleaned stainless steel bowls and spoons. These samples were placed in appropriate holding containers, labeled, chilled on ice, and hand-delivered to Severn Trent Laboratories (STL-Baltimore) each evening for analytical testing. The sediment for ecotoxicological testing was transported to EA's Ecotoxicological Laboratory each evening, stored at 4°C, and later composited for ecotoxicological testing. A chain-of-custody form accompanied the sediment samples during transport to STL and the Ecotoxicology Laboratory.

4.9 SITE WATER SAMPLING

Approximately 10 gallons of site water were collected from one specified station in each sampling reach for elutriate preparation and and bioassay testing. In addition, site water samples were collected from inside and outside Placement Site 104 for chemical analysis. One site water sample was collected from station KI-3, one from station KI-7, and one from station KI-14. In addition, a duplicate water sample was collected at KI-7. Approximately 6 gallons of water were required for the chemical analyses. The water was collected from mid-depth of the water column using a peristaltic pump with Tygon tubing. Water targeted for chemical analysis was pumped directly into laboratory-prepared sample containers containing laboratory-specified preservatives. Elutriate preparation water was stored at 4° C in 1-gallon, pre-cleaned, amber glass bottles. Water for the bioassay testing was stored at 4° C in 5-gallon, pre-cleaned, high-density polyethylene containers. A summary of water sampling activities is provided in Table 4-7A (September) and 4-7B (December).

Holding times for water samples began when the water was collected. Site water targeted for chemical analysis was hand delivered to STL-Baltimore on the evening of the day of collection. Elutriate preparation water was hand delivered to the analytical laboratory with the sediment samples. A chain-of-custody form was submitted to STL-Baltimore with each delivery of samples. Required sample containers, preservation techniques, and holding time requirements for water samples are provided in Table 4-8.

4.10 EQUIPMENT DECONTAMINATION AND WASTE HANDLING PROCEDURES

Equipment that would come into direct contact with sediment was decontaminated prior to deployment in the field to minimize cross-contamination. This included CAB core liners, core caps, stainless steel cutters, stainless steel catchers, and stainless steel processing equipment (spoons, knives, bowls, extruder, etc.). Nose cones and core catchers that were re-used in the field were decontaminated on-board the work vessel between stations. While performing the decontamination procedure, phthalate-free nitrile gloves were used to prevent phthalate contamination of the sampling equipment or the samples.

The decontamination procedure described below was utilized:

- Rinse equipment using clean tap or site water;
- Wash and scrub with non-phosphate detergent (Alconox or other laboratory-grade detergent);
- Rinse with tap water;
- Rinse with 10 percent nitric acid (HNO₃);
- Rinse with distilled or de-ionized water;
- Rinse with methanol followed by hexane;
- Rinse with distilled or de-ionized water;
- Air dry (in area not adjacent to the decontamination area); and
- Wrap equipment in aluminum foil, shiny side out.

Waste liquids were contained during decontamination procedures and transferred to a 55-gallon drum for characterization and disposal at the end of the field effort. Waste liquids were disposed from EA's warehouse facility (in Hunt Valley, Maryland) using standard disposal procedures and contractors.

4.11 FIELD QC SAMPLES

4.11.1 Equipment Blanks

Equipment blanks were collected for the Site 104 project for each round of sampling. Equipment blanks were collected by pouring deionized (DI) water over sampling equipment that had been decontaminated using the procedure outlined in Section 4.10. The rinsate water was placed in laboratory-prepared containers, submitted to the analytical laboratory, and tested for the same chemical parameters as the sediments. One equipment blank was collected for the vibracoring equipment (core catchers, liners, etc.), one equipment blank was collected for the grab sampling equipment (grab, stainless steel bowls and spoons), and one equipment blank was collected for the same state collection device (peristaltic pump tubing). Equipment blanks were hand-delivered to STL-Baltimore in the evening on the day of collection. Chain-of-custody documentation was submitted with the equipment blanks.

4.11.2 Field Duplicates

Three field duplicate sediment samples were collected for the Site 104 project. Field duplicates are samples collected simultaneously from the same sampling location and are used as measures

of matrix homogeneity and sampling precision for the analytical testing. Duplicate samples were collected as individual, co-located samples (i.e., separate grabs from the same station location). The samples were homogenized separately and placed in separate containers. Field duplicate sediment samples were collected at stations CRU2, BE3, and KI-7 during both sampling efforts. A field duplicate water sample was collected at station KI-7 during the first round of sampling. Field duplicate samples were differentiated from other samples with an "FD" as the last two characters of the sample ID. Field duplicates were hand-delivered to the analytical laboratory with their corresponding co-located samples.

4.11.3 Trip Blanks

Trip blanks (also called transport blanks) were analyzed to evaluate the effect of ambient site conditions and sample transport on sample integrity and to ensure proper container preparation and handling techniques. Trip blanks consist of analyte-free water placed in organic vials (preserved with HCl) in the laboratory. Trip blanks are analyzed for VOCs. Trip blanks were only analyzed for the second round of sampling, because VOC analyses were only conducted on samples in the second round of sampling. One trip blank was analyzed per sampling day or sample processing day during the second round of sampling. All volatiles samples were stored in the same cooler as the trip blank.

4.12 SAMPLE PROCESSING

Sediments were processed in a designated area at EA's warehouse facility on 30 September and 01 October 1999, and 17 December 1999. Prior to processing, cores and surficial sediments were sorted and checked against the chain-of-custody forms.

4.12.1 Core Processing

Cores were processed in a designated area at EA's warehouse facility. Sediments were extracted from each core using a stainless-steel extrusion rod and were homogenized using pre-cleaned stainless-steel spoons in stainless-steel bowls. Multiple cores were homogenized for each station. Core samples from each station were composited and homogenized until the sediment was thoroughly mixed and of uniform consistency. For each station, samples for volatile analysis were removed from a longitudinal section of one core and immediately placed into sample containers to avoid loss of volatiles. These samples were not stirred or homogenized. When compositing and homogenization of sediment from each station was completed, sub-samples of sediment were placed in appropriate sample jars and submitted to STL–Baltimore for chemical analysis.

Holding times for the core sediment samples began when the sediment was removed from the core liner, composited, homogenized, and placed in the appropriate sample containers. Holding times for the surficial grabs began when the sediment was collected. Sample containers, preservation techniques, and analytical holding requirements for sediment samples are provided in Table 4-9.



Sediment cores collected for USACE-WES were shipped directly to USACE-WES via overnight delivery.

4.12.2 Reach Composites

After a subsample of sediment from each station was removed for analytical testing, 20-gallon composite samples from each reach were created for ecotoxicological testing using an equal volume of sediment from each station within the channel reach. Twenty-gallon composite samples for ecotoxicological testing were also created for reaches where surficial sediments were collected using an equal volume of sediment from each station within the reach. The composites for each reach were individually homogenized in 55-gallon fiberglass holding containers using large stainless-steel spoons. Sample processing equipment that came into direct contact with the sediment was decontaminated according to the protocols specified in the Section 4.10. The reach sediment composites were placed in pre-cleaned, decontaminated, 5-gallon polypropylene buckets and submitted to EA's Ecotoxicology Laboratory for toxicity testing and bioaccumulation studies. Toxicological holding times for the sediment composites began when the sediment was composited. A list of the composite samples for ecotoxicological testing is provided in Table 4-10.

4.13 SAMPLE LABELING, CHAIN-OF-CUSTODY, AND DOCUMENTATION

4.13.1 Field Logbook

A log of coring activities, station locations, water depths, weather conditions, and core recoveries was recorded in permanently bound logbooks or datasheets. In addition to sampling information, personnel names, local weather conditions, and other information that impacted the field sampling program was recorded. Each page of the logbook (field and sample processing) was numbered, dated, and signed by the personnel entering information. Full copies of the project logbooks are contained in Attachment I.

4.13.2 Numbering System

For the coring program, two separate, but related sample numbering systems were used. One applied to the cores, the other to the sediment samples. The core numbering system was used to communicate between the field crew and the sampling processing crew, and indicated which cores were collected from each station. The sample numbering system provided communication between the sample processing operation and the laboratories performing the desired analyses.



Core Numbering

Cores were numbered as follows:

| Example: | CRE01-CORE1 |
|----------|-------------|
| | CRE01-CORE1 |

where the first 2-3 letters denote the site designation and the following numbers denote the station number. CORE1, CORE2, etc., represented the multiple cores collected from each site.

Sample Numbering

Sample IDs for sediment, site water, and elutriate samples followed the numbering system similar to the system utilized in the FY95 and FY98 sampling programs. A breakdown of each sample type and the corresponding sample ID is provided in Table 4-11 (approach channels) and Table 4-12 (Inside Site 104, Outside Site 104, and Ocean Reference). Sediment samples from the homogenized and composited sediment cores were designated with a "VC" as the last 2 characters of the sample ID.

Blank Numbering System

Equipment blanks were labeled, respectively, as follows:

EQBCORE-mmddyy (rinsate from coring equipment) EQBGRAB-mmddyy (rinsate from grab sampling equipment) EQBWAT-mmddyy (rinsate from water collection equipment) TB-mmddyy (trip blanks)

where the 2-digit month, day, and year of collection were designated within each sample ID.

4.13.3 Sample Labeling

Both cores and processed sediment were labeled. Sediment cores collected in the field were labeled with the site location, station number, core orientation (top and bottom), and date of collection. Sample containers for the surficial grabs, processed sediment, and water samples were labeled with the following information:

- Client name
- Project number
- Sample ID
- Station location
- Date and time of collection
- Sampler's initials
- Type of analyses required

4.13.4 Chain-of-Custody Records

Sediment cores collected in the field were documented on a core-specific chain-of-custody form. This chain-of-custody accompanied the cores to the sample processing facility at EA's facility in Sparks, Maryland. Sample processing personnel prepared a separate chain-of-custody for sample submittal to EA's Ecotoxicology Laboratory, STL-Baltimore, and USACE-WES. Copies of the chain-of-custody forms for the cores and WES samples are provided in Attachment I. Copies of the chain-of-custody forms for bulk sediment, water and elutriates, and ecotoxicological testing are provided in Attachments II, III, and V, respectively.





Figure 4-1. Grab sampling locations in the Brewerton Eastern Extension.











Figure 4-4. Vibracoring locations in the Craighill Angle (East and West).



Figure 4-5. Vibracoring locations in the Craighill Entrance.







Figure 4-7. Grab sampling locations in the Cutoff Angle.



Figure 4-8. Vibracoring locations in the Swan Point Channel.



Figure 4-9. Vibracoring locations in the Tochester Channel (North and South).







Figure 4-11. Grab sampling locations Inside Site 104.







Figure 4-13. Grab sampling locations Outside Site 104.

TABLE 4-1 CHANNELS PROPOSED FOR DREDGING IN FY00 AND FY01, VOLUME OF MATERIAL TO BE REMOVED, AND NUMBER OF SAMPLING STATIONS

| Channel | Required Depth (ft) | Volume (cubic yards) | No. of Sampling Stations |
|--------------------|------------------------|-------------------------|--------------------------------|
| Craighill Entrance | 51 | 172,000 | 2 |
| | | 170,000 | 2 |
| Craighill Angle | 51 | 290,000 | 3 |
| | | 285,000 | 3 |
| Tolchester Channel | 37 | 360,000 | 4 |
| | | 170,000 | 2 |
| | | 75,000 | 1 |
| | | 250,000 | 3 |
| Swan Point | 37 | 725,000* | 6 |

* Assumes whole channel will be dredged. Only 6 stations chosen for sampling because actual sediment volume removed will probably be less than 700,000cy.

| | | (NA | | | |
|-----------------------------|-------------|------------|------------|---------------------|--|
| Sampling Reach | Station | Northing | Easting | ing Sampling Method | |
| Brewerton Eastern Extension | BEI | 546054.6 | 1484173.9 | Grab | |
| | BE2 | 544207.3 | 1490512.6 | Grab | |
| | BE3 | 545890.3 | 1485066.2 | Grab | |
| | BE4 | 548684.8 | 1479219.6 | Grab | |
| C&D Approach Channels | CD-001VC | 608800 | 1541400 | Core | |
| - F F | CD-002VC | 612175 | 1545400 | Core | |
| | CD003 | 573600 | 1528000 | Grab | |
| | CD004 | 581300 | 1528650 | Grab | |
| | CD005 | 601300 | 1534250 | Grab | |
| | CD006 | 607525 | 1540450 | Grab | |
| Craighill | CR1 | 515128.6 | 1484264.4 | Grab | |
| - | CR2 | 521097.4 | 1484206.8 | Grab | |
| | CR3 | 526054.6 | 1484173.9 | Grab | |
| Craighill Angle-East | CRA-E-001VC | 532042.07 | 1484445.93 | Core | |
| | CRA-E-002VC | 531405.32 | 1484484.88 | Core | |
| | CRA-E-003VC | 535877.91 | 1482174.72 | Core | |
| Craighill Angle-West | CRA-W-001VC | 532237.22 | 1482683.56 | Core | |
| | CRA-W-002VC | 530098.27 | 1483263.04 | Core | |
| | CRA-W-003VC | 533112.97 | 1482200.47 | Core | |
| Craighill Entrance | CRE-001VC | 498692.06 | 1487831.57 | Core | |
| _ | CRE-002VC | 499552.33 | 1487570.33 | Core | |
| | CRE-003VC | 511593.04 | 1484861.36 | Core | |
| | CRE-004VC | 509853.57 | 1484454.57 | Core | |
| Craighill Upper Range | CRU1 | 538976.2 | 1479833.7 | Grab | |
| | CRU2 | 543109.9 | 1477601.3 | Grab | |
| | CRU3 | 547444.2 | 1475053.4 | Grab | |
| Cutoff Angle | CUTI | 549257.3 | 1473781.9 | Grab | |
| - | CUT2 | 550463 | 1472435.6 | Grab | |
| | CUT3 | 551262.7 | 1470855.7 | Grab | |
| Inside Placement Site 104 | KI-3 | 497783.9 | 1498316.6 | Grab | |
| | KI-5 | 492612.6 | 1496696.3 | Grab | |
| | K1-7 | 487234.6 | 1494445.2 | Grab | |
| | K1-S-1 | 486219.1 | 1493899.7 | Grab | |
| | K1-S-2 | 486641.5 | 1496423 | Grab | |
| Outside Placement Site 104 | K1-11 | 510575.7 | 1504376.8 | Grab | |
| | K1-14 | 478014.2 | 1492456.7 | Grab | |
| | K1-15 | 491442.1 | 1488417.1 | Grab | |
| | K1-16 | 499865 | 1492147.3 | Grab | |
| Swan Point | SWP-001VC | 516863.07 | 1508075.38 | Core | |
| | SWP-002VC | 521804.29 | 1510170.87 | Core | |
| | SWP-003VC | 525645.48 | 1509913.08 | Core | |
| | SWP-004VC | 520350 | 1509900 | Core | |
| | SWP-005VC | 520325 | 1511100 | Core | |
| | SWP-006VC | 521075 | 1511275 | Core | |
| Tolchester Channel – North | TLC-005VC | 544160.25 | 1507775.94 | Core | |
| | TLC-006VC | 54 /953.08 | 1511035.91 | Core | |
| | TLC-007VC | 555257.61 | 1518689.51 | Core | |
| | | 558000.20 | 152152/.01 | Core | |
| | | 561672.92 | 1322833.33 | Core | |
| Talahastar Channal Cauth | TLC-010VC | 52013/2.83 | 15250/0.55 | Core | |
| roicnester Channel – South | TLC-001VC | 520002.92 | 1504030.41 | Core | |
| | TLC-002VC | 530111 04 | 1504007.52 | Core | |
| | TLC-003VC | 530059 27 | 1504093.93 | Core | |
| Talahastar Straightaning | | 557840 5 | 1510614 6 | Core | |
| rotenester straightening | TLS-002VC | 560593.6 | 1572269 | Core | |

TABLE 4-2 SAMPLING STATION COORDINATES



TABLE 4-3DEPTH OF CORES AND NUMBER OF
CORES REQUIRED PER STATION

| | | | Number of Cores |
|----------------------------|-------------|-------------------|------------------|
| Sampling Reach | Station | Core Depth | Required* |
| C&D Approach Channels | CD-001VC | 3 | 9 |
| | CD-002VC | 3 | 9 |
| Craighill Angle-East | CRA-E-001VC | 4.5 | 4 |
| | CRA-E-002VC | 2.5 | 7 |
| | CRA-E-003VC | 2.7 | 6 |
| Craighill Angle-West | CRA-W-001VC | 4 | 4 |
| | CRA-W-002VC | 4 | 4 |
| | CRA-W-003VC | 4 | 4 |
| Craighill Entrance | CRE-001VC | 5 | 3 |
| E . | CRE-002VC | 5 | 3 |
| | CRE-003VC | 4 | 4 |
| | CRE-004VC | 2.5 | 5 |
| Swan Point | SWP-001VC | 4 | 4 |
| | SWP-002VC | 5 | 3 |
| | SWP-003VC | 4 | 4 |
| | SWP-004VC | 4 | 3 |
| | SWP-005VC | 3 | 5 |
| | SWP-006VC | 3 | 5 |
| Tolchester Channel – North | TLC-005VC | 4.7 | 2 |
| | TLC-006VC | 4.8 | 2 |
| | TLC-007VC | 4.5 | 2 |
| | TLC-008VC | 5 | 2 |
| | TLC-009VC | 5 | 2 |
| | TLC-010VC | 4.5 | 2 |
| Tolchester Channel – South | TLC-001VC | 6 | 2 |
| | TLC-002VC | 7 | 2 |
| | TLC-003VC | 7 | 2 |
| | TLC-004VC | 7.5 | 2 |
| Tolchester Straightening | TLS-001VC | 15 | 2 |
| | TLS-002VC | 15 | 2 |

* Pre-sampling estimate based on 100% recovery.



TABLE 4-5ASUMMARY OF GRAB SAMPLING ACTIVITIESSEPTEMBER 1999

| | | Sediment Sample | | | Compositing |
|-----------------------------|------------|-----------------|-------------|------------|---------------------|
| Sampling Reach | Station ID | ID | Sample Date | Time (EDT) | Date ^(a) |
| Brewerton Eastern Extension | BE1 | BE1SED | 09/27/1999 | 1615 | 10/01/1999 |
| | BE2 | BE2SED | 09/27/1999 | 1443 | |
| | BE3 | BE3SED | 09/27/1999 | 1715 | |
| | BE3 | BE3SEDFD | 09/27/1999 | 1720 | |
| | BE4 | BE4SED | 09/27/1999 | 1745 | 1 1 |
| C&D Approach Channels | CD003 | CD003SED | 09/27/1999 | 1455 | 10/01/1999 |
| | CD004 | CD004SED | 09/27/1999 | 1410 | |
| | CD005 | CD005SED | 09/27/1999 | 1340 | |
| | CD006 | CD006SED | 09/27/1999 | 1300 | |
| Craighill | CR1 | CR1SED | 09/28/1999 | 1246 | 10/01/1999 |
| | CR2 | CR2SED | 09/28/1999 | 1209 | |
| | CR3 | CR3SED | 09/28/1999 | 1140 | 1 [|
| Craighill Upper Range | CRU1 | CRU1SED | 09/28/1999 | 1057 | 10/01/1999 |
| | CRU2 | CRU2SED | 09/28/1999 | 1025 | |
| | CRU2 | CRU2SEDFD | 09/28/1999 | 1035 | |
| | CRU3 | CRU3SED | 09/28/1999 | 0955 | |
| Cutoff Angle | CUT1 | CUT1SED | 09/28/1999 | 0925 | 10/01/1999 |
| | CUT2 | CUT2SED | 09/28/1999 | 0840 | |
| | CUT3 | CUT3SED | 09/28/1999 | 0810 | |
| Inside Placement Site 104 | KI-7N | KI-7N-SED | 09/29/1999 | 0818 | NC |
| | KI-7S | KI-7S-SED | 09/29/1999 | 1124 | |
| | KI-7E | KI-7E-SED | 09/29/1999 | 1323 | |
| | KI-7W | KI-7W-SED | 09/29/1999 | 1412 | |
| | KI-7 | KI-7-SED | 09/28/1999 | 1510 | |
| | KI-3 | KI-3-SED | 09/28/1999 | 1410 | 10/01/1999 |
| | KI-5 | KI-5-SED | 09/28/1999 | 1448 | |
| | KI-7-REF | KI-7-REFSED | 09/29/1999 | 0934 | |
| | KI-7-REFFD | KI-7-REFSEDFD | 09/29/1999 | 1100 | |
| | KI-S-1 | KI-S-1-SED | 09/28/1999 | 1700 | |
| | KI-S-2 | KI-S-2-SED | 09/28/1999 | 1725 | |

(a) = Date that reach composite was created for ecotoxicological testing.

not composited for ecotoxicological testing; chemical analysis of individual samples only.
TABLE 4-5BSUMMARY OF GRAB SAMPLING ACTIVITIESDECEMBER 1999

| | | Sediment | Collection | Collection | Compositing |
|-----------------------------|-------------------|--------------|------------|------------|---------------------|
| Sampling Reach | Station ID | Sample ID | Date | Time (EST) | Date ^(a) |
| Brewerton Eastern Extension | BE1 | BE1SED | 12/14/1999 | 1540 | 12/17/1999 |
| | BE2 | BE2SED | 12/14/1999 | 1415 | |
| | BE3 | BE3SED | 12/14/1999 | 1500 | |
| | BE3 | BE3SEDFD | 12/14/1999 | 1500 | |
| | BE4 | BE4SED | 12/14/1999 | 1555 | |
| C&D Approach Channels | CD003 | CD003SED | 12/14/1999 | 1235 | 12/17/1999 |
| | CD004 | CD004SED | 12/14/1999 | 1155 | |
| | CD005 | CD005SED | 12/14/1999 | 1115 | |
| | CD006 | CD006SED | 12/14/1999 | 1055 | |
| Craighill | CR1 | CR1SED | 12/13/1999 | 1535 | 12/17/1999 |
| | CR2 | CR2SED | 12/13/1999 | 1555 | Ī |
| | CR3 | CR3SED | 12/13/1999 | 1630 | |
| Craighill Upper Range | CRU1 | CRU1SED | 12/15/1999 | 835 | 12/17/1999 |
| | CRU2 | CRU2SED | 12/15/1999 | 915 | |
| | CRU2 | CRU2SEDFD | 12/15/1999 | 915 | |
| | CRU3 | CRU3SED | 12/15/1999 | 940 | |
| Cutoff Angle | CUT1 | CUTISED | 12/15/1999 | 1005 | 12/17/1999 |
| | CUT2 | CUT2SED | 12/15/1999 | 1030 | |
| | CUT3 | CUT3SED | 12/15/1999 | 1105 | |
| Inside Placement Site 104 | KI-3 | KI-3-SED | 12/13/1999 | 1340 | 12/17/1999 |
| | KI-5 | KI-5-SED | 12/13/1999 | 1350 | |
| | KI-7-REF | KI-7-REF-SED | 12/13/1999 | 1425 | |
| | KI-S-1 | KI-S-1-SED | 12/13/1999 | 1405 | |
| | KI-S-2 | KI-S-2-SED | 12/13/1999 | 1505 | |
| Outside Placement Site 104 | KI-11 | KI-11-SED | 12/13/1999 | 1115 | 12/17/1999 |
| | KI-14 | KI-14-SED | 12/13/1999 | 1225 | |
| | KI-15 | KI-15-SED | 12/13/1999 | 1155 | |
| | KI-16 | KI-16-SED | 12/13/1999 | 1130 | |

(a) = Date that reach composite was created for ecotoxicological testing.

TABLE 4-6SAMPLING LOCATION COORDINATES FOR NORFOLKOCEAN DISPOSAL SITE REFERENCE AREA AND LOWERCHESAPEAKE BAY CONTROL SITE

| Station | Collection Date | Latitude | Longitude | Station Depth (ft) |
|-----------------|------------------------|-------------|-------------|--------------------|
| Ocean Reference | Feb. 1, 2000 | 36° 52.6' N | 75° 41.5' W | 71 |
| Control | Jan. 12, 2000 | 37° 15.0' N | 76° 5.7' W | 51 |

4-7A SUMMARY OF WATER SAMPLES COLLECTED IN SEPTEMBER 1999

| | | Water Sample | Sample | Sample | |
|-----------------------------|-------------|--------------|------------|--------|-----------------------------|
| Sampling Reach | Station ID | ID | Date | Time | Sample Type |
| Brewerton Eastern Extension | BE2 | BE-SW | 09/27/1999 | 1645 | Elutriate Preparation Water |
| C&D Approach Channels | CD-002VC | CD-VC-SW | 09/21/1999 | 1230 | Elutriate Preparation Water |
| | CD004 | CD-SW | 09/27/1999 | 1420 | Elutriate Preparation Water |
| Craighill | CR2 | CR-SW | 09/28/1999 | 1205 | Elutriate Preparation Water |
| Craighill Angle-East | CRA-E-002VC | CRA-E-SW | 09/18/1999 | 1315 | Elutriate Preparation Water |
| Craighill Angle-West | CRA-W-002VC | CRA-W-SW | 09/18/1999 | 1645 | Elutriate Preparation Water |
| Craighill Entrance | CRE-002VC | CRE-SW | 09/19/1999 | 1052 | Elutriate Preparation Water |
| Craighill Upper Range | CRU2 | CRU-SW | 09/28/1999 | 1020 | Elutriate Preparation Water |
| Cutoff Angle | CUT2 | CUT-SW | 09/28/1999 | 835 | Elutriate Preparation Water |
| Inside Placement Site 104 | KI-3 | KI-3WAT | 09/28/1999 | 1415 | Receiving Water |
| | KI-7-REF | KI-SW | 09/28/1999 | 1540 | Elutriate Preparation Water |
| | KI-7-REF | KI-7WAT | 09/28/1999 | 1600 | Receiving Water |
| | KI-7-REF | KI-7WATFD | 09/28/1999 | 1615 | Receiving Water |
| | KI-7-REF | KI-7WATMS | 09/28/1999 | 1630 | Receiving Water |
| | KI-7-REF | KI-7WATMSD | 09/28/1999 | 1645 | Receiving Water |
| Swan Point | SWP-002VC | SWP-SW | 09/19/1999 | 1530 | Elutriate Preparation Water |
| Tolchester Channel - North | TLC-N-007VC | TLC-N-SW | 09/20/1999 | 1130 | Elutriate Preparation Water |
| Tolchester Channel - South | TLC-S-002VC | TLC-S-SW | 09/20/1999 | 850 | Elutriate Preparation Water |
| Tolchester Straightening | TLS-001VC | TLS-SW | 09/27/1999 | 1520 | Elutriate Preparation Water |

TABLE 4-7BSUMMARY OF WATER SAMPLES COLLECTED IN
DECEMBER 1999

| | | Water Sample | Sample | Sample | |
|-----------------------------|-------------|--------------|------------|--------|------------------------------------|
| Sampling Reach | Station ID | ID | Date | Time | Sample Type |
| Brewerton Eastern Extension | BE2 | BE-SW | 12/14/1999 | 1420 | Elutriate Preparation Water |
| C&D Approach Channels | CD-002VC | CD-VC-SW | 12/08/1999 | 845 | Elutriate Preparation Water |
| | CD004 | CD-SW | 12/14/1999 | 1210 | Elutriate Preparation Water |
| Craighill | CR2 | CR-SW | 12/13/1999 | 1600 | Elutriate Preparation Water |
| Craighill Angle-East | CRA-E-002VC | CRA-E-SW | 12/15/1999 | 930 | Elutriate Preparation Water |
| Craighill Angle-West | CRA-W-002VC | CRA-W-SW | 12/14/1999 | 950 | Elutriate Preparation Water |
| Craighill Entrance | CRE-002VC | CRE-SW | 12/15/1999 | 1300 | Elutriate Preparation Water |
| Craighill Upper Range | CRU2 | CRU-SW | 12/15/1999 | 915 | Elutriate Preparation Water |
| Cutoff Angle | CUT2 | CUT-SW | 12/15/1999 | 1035 | Elutriate Preparation Water |
| Inside Placement Site 104 | KI-7-REF | KI-SW | 12/13/1999 | 1430 | Elutriate Preparation Water |
| | KI-7-REF | KI-7WAT | 12/13/1999 | 1430 | Receiving Water |
| | KI-7-REF | KI-7WATMS | 12/13/1999 | 1430 | Receiving Water |
| | KI-7-REF | KI-7WATMSD | 12/13/1999 | 1430 | Receiving Water |
| Outside Placement Site 104 | KI-14 | KI-OUT-SW | 12/13/1999 | 1235 | Elutriate Preparation Water |
| | KI-14 | KI-14WAT | 12/13/1999 | 1235 | Receiving Water |
| Swan Point | SWP-002VC | SWP-SW | 12/09/1999 | 1145 | Elutriate Preparation Water |
| Tolchester Channel - North | TLC-N-007VC | TLC-N-SW | 12/08/1999 | 1315 | Elutriate Preparation Water |
| Tolchester Channel - South | TLC-S-002VC | TLC-S-SW | 12/09/1999 | 950 | Elutriate Preparation Water |
| Tolchester Straightening | TLS-001VC | TLS-SW | 12/09/1999 | 830 | Elutriate Preparation Water |

| Parameter | Volume Required (mL) | Container ^(b) | Preservative | Holding Time ^(a) |
|---|----------------------------|----------------------------|--|--|
| Inorganics | a 5 - 4 - 18 - | and and and a second | | |
| Mercury | 100 | Р | pH <2 with HNO ₃ Cool, 4°C | 28 days |
| Other Metals | 100 | Р | pH <2 with HNO ₃ Cool, 4°C | 6 months |
| Cyanide | 500 | P,G | NaOH to pH >12 Ascorbic Acid Cool, 4°C | 14 days |
| Sulfide | 500 | P,G | NaOH to pH >9 Zinc Acetate Cool, 4°C | 7 days |
| Ammonia | 500 | P,G | H ₂ SO ₄ to pH <2 Cool, 4°C | 28 days |
| Biological Oxygen Demand | 1000 | P,G | Cool, 4°C | 48 hours |
| Chemical Oxygen Demand | 50 | P,G | H ₂ SO ₄ to pH <2 Cool, 4°C | 28 days |
| Nitrogen (Ammonia, Total Kjeldahl, Nitrate+Nitrite), Total Phosphorus | 1050 | P,G | H ₂ SO ₄ to pH <2 Cool, 4°C | 28 days |
| Organics | | | an a | the has a second of the second |
| Total Organic Carbon | 50 | P,G | H ₂ SO ₄ or HCl to PH <2 Cool, 4°C | 28 days |
| Organotins | 1000 | G, teflon- lined cap | Cool, 4°C | 7 days until extraction, 7 days from extraction to derivatization, 40 days after extraction |
| Volatile Organic Compounds | 80 | G, teflon- lined septum | Cool, 4 C | 14 days |
| Dioxins/Furans | 1,000 | G, teflon- lined cap | Cool, 4 C | 30 days until extraction, 40 days after extraction |

TABLE 4-8 REQUIRED CONTAINERS, PRESERVATION TECHNIQUE, AND HOLDING TIMES FOR AQUEOUS SAMPLES

(a) From time of sample collection.
(b) P = plastic; G = glass.





TABLE 4-9REQUIRED CONTAINERS, PRESERVATION TECHNIQUE, AND
HOLDING TIMES FOR SEDIMENT SAMPLES

| Parameter | Mass Required (g) | Container ^(b) | Preservative | Holding Time ^(a) |
|---|-------------------------|--|--------------|--|
| Inorganics | | 5 | | |
| Mercury | 5 | Р | 4°C | 28 days |
| Other Metals | 5 | Р | 4°C | 6 months |
| Cyanide | 50 | P,G | 4°C | 14 days |
| Sulfide | 10 | P,G | 4°C | 7 days |
| Acid Volatile Sulfides (AVS) | 25 | P, G | 4°C | 14 days |
| Ammonia | 10 | G | 4°C | 28 days |
| Biological Oxygen Demand | 10 | G | 4°C | 48 hours |
| Chemical Oxygen Demand | 50 | P,G | 4°C | 28 days |
| Nitrogen (Ammonia, Total Kjeldahl Nitrate+Nitrite), Total Phosphorus | 150 | P,G | 4°C | 28 days |
| Physical Parameters | | | | |
| Elutriate Preparation | 1500 | G | 4°C | 14 days until elutriate prep. Follow aqueous hold times after prep. |
| Total Moisture, Grain Size, Atterberg Limits, Specific Gravity | 1000 | P,G | 4°C | 6 months |
| Organics | | | | |
| Total Organic Carbon | 5 | Heat treated glass vial with Teflon- lined lid ^(b) | 4°C | 14 days |
| Organotins | 100 | G | 4°C | 14 days until extraction, 7 days from extraction to derivatization, 40 days after extraction |
| Pesticides (Organochlorine and Organophosphate),PCB Congeners, Semivolatile Organics, Polynuclear Aromatic Hydrocarbons | 400 | G | 4°C | 14 days until extraction, 40 days after extraction |
| Volatile Organic Compounds | 50 | Heat-heated glass vial with Teflon- lined lid | 4 <u>°</u> C | 14 days |
| Dioxins/Furans | 40 | G | 4 <u>°</u> C | 30 days until extraction, 40 days after extraction |





(a) From time of sample collection for grab samples; from the time of removal from the liner for cores.
(b) P = plastic; G = glass.

TABLE 4-10RECOMMENDED CONTAINERS, PRESERVATION TECHNIQUE, AND
HOLDING TIMES FOR TOXICITY AND BIOACCUMULATION TESTING

| Parameter | Mass Required (g) | Container ^(b) | Preservative | Holding Time ^(a) |
|------------------------------|-------------------------|---|--------------|---|
| Toxicity and Bioaccumulation | Testing | de la companya da serie da se Serie da serie da seri Serie da serie da ser | | an a |
| Whole Sediment | 30 L | Р | ≤ 4°C/dark | Optimum 14 days, maximum 6 weeks |
| Water Column | 10 L | Р | ≤ 4°C/dark | Elutriate from sediment prepared within 24 hours of test initiation |

(a) From time of sample collection per USEPA 1991, The Green Book, and USEPA 1995, QA/QC Guidance for Sampling and Analysis of Sediments, Water, and Tissues for Dredged Material Evaluations.
(b) Polyethylene (P) or glass (G).





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Nereis virens: mean metal concentrations compared to inside Site 104.

Nereis virens: mean metal concentrations compared to outside Site 104.

Nereis virens: mean metal concentrations compared to ocean reference.

Nereis virens: bioaccumulation factors for metals.

Macoma nasuta: mean metal concentrations compared to inside Site 104.

Macoma nasuta: mean metal concentrations compared to outside Site 104.

Macoma nasuta: mean metal concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for metals.

Nereis virens: mean chlorinated pesticide concentrations compared to inside Site 104.

Nereis virens: mean chlorinated pesticide concentrations compared to outside Site 104.

Nereis virens: mean chlorinated pesticide concentrations compared to ocean reference.

Nereis virens: bioaccumulation factors for chlorinated pesticides.

Macoma nasuta: mean chlorinated pesticide concentrations compared to inside Site 104.

Macoma nasuta: mean chlorinated pesticide concentrations compared to outside Site 104.

Macoma nasuta: mean chlorinated pesticide concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for chlorinated pesticides.

Nereis virens: mean PAH concentrations compared to inside Site 104.

Nereis virens: mean PAH concentrations compared to outside Site 104.

Nereis virens: mean PAH concentrations compared to ocean reference.

Nereis virens: bioaccumulation factors for PAHs.

Macoma nasuta: mean PAH concentrations compared to inside Site 104.

Macoma nasuta: mean PAH concentrations compared to outside Site 104.

Macoma nasuta: mean PAH concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for PAHs.

Nereis virens: mean PCB congener concentrations compared to inside Site 104.

Nereis virens: mean PCB congener concentrations compared to outside Site 104.

Nereis virens: mean PCB congener concentrations compared to ocean reference.

Nereis virens: bioaccumulation factors for PCB congeners.

Macoma nasuta: mean PCB congener concentrations compared to inside Site 104.

Macoma nasuta: mean PCB congener concentrations compared to outside Site 104.

Macoma nasuta: mean PCB congener concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for PCB congeners.

Nereis virens: mean semivolatile organic compound concentrations compared to inside Site 104.

Nereis virens: mean semivolatile organic compound concentrations compared to outside Site 104.

Nereis virens: mean semivolatile organic compound concentrations

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Nereis virens: bioaccumulation factors for semivolatile organic compounds.

Macoma nasuta: mean semivolatile organic compound concentrations compared to inside Site 104.

Macoma nasuta: mean semivolatile organic compound concentrations compared to outside Site 104.

Macoma nasuta: mean semivolatile organic compound concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for semivolatile organic compounds.

Nereis virens: mean dioxin and furan concentrations compared to inside Site 104.

Nereis virens: mean dioxin and furan concentrations compared to outside Site 104.

Nereis virens: mean dioxin and furan concentrations compared to ocean reference.

Nereis virens: bioaccumulation factors for dioxins and furans.

Macoma nasuta: mean dioxin and furan concentrations compared to inside Site 104.

Macoma nasuta: mean dioxin and furan concentrations compared to outside Site 104.

Macoma nasuta: mean dioxin and furan concentrations compared to ocean reference.

Macoma nasuta: bioaccumulation factors for dioxins and furans.

Nereis virens: mean lipid and percent moisture concentrations.

Macoma nasuta: mean lipid and percent moisture concentrations.
LIST OF ABBREVIATIONS, ACRONYMS, AND UNITS

| AAS | Atomic Adsorption Spectrophotometry |
|-------------------------------------|--|
| AET | Apparent Effects Threshold |
| ASA | American Society of Agronomy |
| ASI | Agua Survey, Inc. |
| ASTM | American Society for Testing and Materials |
| AVS | Acid Volatile Sulfides |
| | |
| BE | Brewerton Eastern Extension |
| BOD | Biological Oxygen Demand |
| BSAF | Biota Sediment Accumulation Factor |
| BZ | Ballschmiter and Zell |
| | |
| °C | Celsius |
| C&D | Chesapeake and Delaware Canal (Approach Canal) |
| CAB | Cellulose Acetate Butylrate |
| CBR | Critical Body Residue |
| CENAB | Corps of Engineers North Atlantic – Baltimore |
| CENAP | Corps of Engineers North Atlantic – Philadelphia |
| CdCl | Cadmium Chloride |
| CF | Concentration Factor |
| CFR | Code of Federal Regulations |
| cm/sec | centimeter per second |
| COC | Chain-of-Custody |
| COD | Chemical Oxygen Demand |
| CBP | Chesapeake Bay Program |
| COE | Corps of Engineers |
| COMAR | Code of Maryland |
| COPC | Contaminant of Potential Concern |
| CORMIX | (USACE Corps of Engineers model) |
| CR | Craighill Channel |
| CRA-E | Craighill Angle-East |
| CRA-W | Craighill Angle-West |
| CRE | Craighill Entrance |
| CRU | Craighill Upper Range |
| CuSO ₄ 5H ₂ O | Copper Sulfate |
| CUT | Cutoff Angle |
| CVAAs | Cold Vapor Atomic Absorption Spectrophotometry |
| CWA | Clean Water Act |
| су | cubic yard |
| - | - |
| DGPS | Differential Global Positioning System |

| DI | De-ionized (Water) |
|------------------|--|
| DL | Detection Limit |
| DO | Dissolved Oxygen |
| DQO | Data Quality Objective |
| DVR | Data Validation Report |
| EA | EA Engineering, Science, and Technology, Inc. |
| EC50 | Effective Sub-lethal Concentration |
| ECD | Electron Capture Detector |
| EDM | Estimated Detection Limit |
| EDS | Environmental Data Services, Inc. |
| EDT | Eastern Davlight Time |
| EIS | Environmental Impact Statement |
| EM | Engineer Manual |
| EMAP | Environmental Monitoring and Assessment Program |
| EMPC | Estimated Maximum Possible Concentration |
| ERED | Environmental Residue-Effects Database |
| ERL | Effect Range Low |
| ERM | Effect Range Median |
| EST | Eastern Standard Time |
| FD | Field Duplicate |
| FDA | Food and Drug Administration |
| ft | foot |
| ft/sec | feet per second |
| FSP | Field Sampling Plan |
| FPD | Flame Photometric Detector |
| FY | Fiscal Year |
| GC/MS | Gas Chromatography / Mass Spectrometry |
| GFAAS | Graphite Furnace Atomic Absorption Spectrophotometry |
| GPC | Gel Permeation Chromatography |
| gm/cc | gram per cubic centimeter |
| g/kg | gram per kilogram (ppt) |
| g/L | gram per liter (ppt) |
| hr | hour |
| H ₋ S | Hydrogen Sulfide |
| HCl | Hydrochloric Acid |
| HH | Human Health |
| HMW | High Molecular Weight |
| HNO ₃ | Nitric Acid |
| HPLČ | High Pressure Liquid Chromatography |
| | |

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| HRGC | High Resolution Gas Chromatography |
|----------|--|
| 111(1915 | Tigh Resolution Wass Spectrometry |
| ICP | Inductively Counted Plasma |
| | Instrument Detection Limit |
| ITM | Inland Testing Manual |
| IUPAC | International Union of Pure and Applied Chemistry |
| 101110 | |
| KCl | Potassium Chloride |
| L | liter |
| LC50 | Median Lethal Concentration |
| LCS | Laboratory Control Sample |
| LMW | Low Molecular Weight |
| LOED | Lowest Observed Effect Dose |
| LPC | Limiting Permissible Concentration |
| LSD | Least Significant Difference |
| | |
| m | meter |
| MB | Method Blank |
| mg/kg | milligram per kilogram (ppm) |
| mg/L | milligram per liter (ppm) |
| MDE | Maryland Department of the Environment |
| MDL | Method Detection Limits |
| Ml | milliliter |
| MLW | Mean Low Water |
| MLLW | Mean Lower Low Water |
| MPRSA | Marine Protection Research and Sanctuaries Act of 1972 |
| MS | Matrix Spike |
| MSA | Method of Standard Addition |
| MSD | Matrix Spike Duplicate |
| MSE | Mean Square Error |
| M&N | Moffatt & Nichol Engineers |
| NA | Not Analyzed |
| NAD | North American Datum |
| NAS | Northwest Aquatic Sciences |
| ND | Non-Detect |
| NIST | National Institute of Standards and Technology |
| NOAA | National Oceanic and Atmospheric Administration |
| NODS | Norfolk Ocean Disposal Site |
| NOEC | No Observed Effect Concentration |
| NOED | No Observed Effect Dose |
| | |



| NPD | Nitrogen/Phosphorous Detector |
|-----------|---|
| NPDES | National Pollutant Discharge Elimination System |
| ng/kg | nanogram per kilogram (pptr) |
| ng/l | nanograms per liter (pptr) |
| DAU | Polymuslear Aromatic Hydrogarhons |
| | Porodiam Apolytical Laboratories |
| | Palacingin Analytical Laboratories |
| | Probable Effect I evel |
| rEL | norte per hillion (ug/kg or ug/L) |
| ppo | parts per billion (ug/kg or ug/L) |
| ppin | parts per finnion (fig/kg of fig/L) nexts per thousand (solinity units) $(g/kg \text{ or } g/L)$ |
| ppt | parts per thousand (samity units) (g/kg of g/L) |
| pptr | parts per trillion (ng/kg or ng/L) |
| QA | Quality Assurance |
| QAPP | Quality Assurance Project Plan |
| QC | Quality Control |
| RBC | Risk-Based Concentration/Criteria |
| RIM | Regional Implementation Manual |
| RL | Reporting Limit |
| SDG | Samula Delivery Groun |
| SDG | Sodium Dodecyl Sulfate |
| SDS SF | Standard Error |
| SE | Standard Error Simultaneously Extracted Metals |
| SOP | Standard Operating Procedure |
| SOG | Sediment Quality Guidelines |
| SQU | Sample Quantitation Limit |
| SQL | Standard Reference Material |
| | Standard Reference Material |
| STEATE | (numerical mixing model) |
| STIAL | Severn Trent Laboratories |
| SV | Screening Value |
| SVOA | Semivolatile Organic Analysis |
| SVOC | Semivolatile Organic Compound |
| SVOC | Swap Point Channel |
| SWI | Swan Fornt Channel |
| TBP | Theoretical Bioaccumulation Potential |
| TBT | Tributyltin |
| TDL | Target Detection Limit |
| TEF | Toxicity Equivalency Factor |
| TEQ | Toxicity Equivalency Quotient |

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|-----------|--|
| TEL | Threshold Effect Level |
| TIC | Tentatively Identified Compounds |
| TKN | Total Kjeldahl Nitrogen |
| TLC-N | Tolchester Channel-North |
| TLC-S | Tolchester Channel-South |
| TLS | Tolchester Straightening |
| TOC | Total Organic Carbon |
| TP | Total Phosphorus |
| | |
| UCLM | Upper Confidence Limit of the Mean |
| µg/kg | microgram per kilogram (ppb) |
| μg/l | microgram per liter (ppb) |
| µmol/g | micromoles per gram |
| UMD-CES | University of Maryland, Center for Environmental Studies |
| USACE | U.S. Army Corps of Engineers |
| USACE-WES | U.S. Army Corps of Engineers – Waterways Experiments Station |
| USEPA | U.S. Environmental Protection Agency |
| | |
| VC | Vibracoring (location or core sample) |
| VOA | Volatile Organic Analysis |
| VOC | Volatile Organic Compound |
| | |
| WQC | Water Quality Criteria |
| WQS | Water Quality Standards |
| VOI | V-lloss Covin as Instances ante |
| 1 21 | r enow springs instruments |



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CHAPTER 1 -EVALUATION GUIDELINES

1. EVALUATION OF SEDIMENTS PROPOSED FOR OPEN WATER AND OCEAN PLACEMENT

The objective of this project is to evaluate dredged material proposed for open water placement in the Chesapeake Bay. Thirteen channels proposed for either maintenance or new work dredging are evaluated in this report. In addition, two proposed placement areas (Site 104 and the Norfolk Ocean Disposal Site) and an upper Chesapeake Bay reference area (Outside Site 104) are evaluated using the tiered testing approach as recommended by U.S. Environmental Protection Agency/U.S. Army Core of Engineers (USEPA/USACE) guidance (1998). The dredged material evaluation follows guidelines in the Inland Testing Manual (ITM) (USEPA/USACE 1998) as related to section 404(b)(1) guidelines. Specifically, this project is designed to evaluate dredged material proposed for placement at Site 104.

This report does not provide a definitive decision regarding whether the dredged material is appropriate or inappropriate for open water or ocean placement. The information presented provides weight of evidence to assist decision-makers with a factual determination regarding the potential for short-term or long-term impacts associated with the placement of dredged material in open water. The data and information presented in this report may be used to evaluate the potential impacts associated with other dredged material placement alternatives – both alternative placement methods and locations.

1.1 **REGULATORY OVERVIEW**

The USACE and the USEPA are responsible for regulating placement of dredged material. Open water placement of dredged material in waters of the U.S. is governed by the USEPA's Guidelines for Specification of Disposal Sites for Dredged or Fill Material implemented in Section 404(b)(1) of the Clean Water Act (CWA). The Section 404(b)(1) Guidelines are published in the Code of Federal Regulations (CFR) in 40 CFR Part 230. The technical evaluation of potential contaminant-related impacts that may be associated with open water placement of dredged material is conducted in accordance with 40 CFR 230.60 and 230.61. Evaluation of dredged material proposed for ocean placement is mandated by Section 103 of the Marine Protection, Research and Sanctuaries Act (MPRSA) of 1972. The technical evaluation of potential contaminant-related impacts that may be associated with ocean placement of dredged material is conducted in accordance with 40 CFR 230.60 and 230.61.

The USACE is responsible for issuing 404 permits that specify placement sites for dredged material and for issuing ocean placement permits in accordance with Section 103 of MPRSA. Projects in waters of the U.S. must comply with 40 CFR 230 Guidelines and 33 CFR 320-330 (public interest review) prior to being issued a 404 permit. Projects proposed for ocean placement must comply with 40 CFR 220-228 and 33 CFR 320-330 prior to being issued an ocean permit.

The USEPA is responsible for assisting USACE with the development of environmental guidelines for evaluating permit applications. In addition, the USEPA is responsible for reviewing and commenting on permit applications, regulating the placement of materials that

may adversely impact the aquatic environment, determining authority, and defining exemptions. The USACE and USEPA share the authority to enforce the regulations (USACE/USEPA 1998).

1.2 GUIDANCE FOR TESTING AND EVALUATIONS

The USEPA and USACE have jointly published guidance documents that describe the recommended testing and evaluation procedures for dredged materials proposed for placement in waters of the U.S. and in ocean waters. The *Inland Testing Manual* (USEPA/USACE 1998) describes testing and evaluation procedures for dredged material proposed for placement in either fresh, estuarine, or saline (near coastal) waters of the United States, in accordance with 40 CFR 230.60 and 230.61. The *Ocean Testing Manual* (USEPA/USACE 1991) (also known as "*The Green Book*") describes the testing and evaluation procedures for dredged material proposed for dredged material proposed for placement, in accordance with 40 CFR 220-228.

The ITM serves as the national framework and primary guidance document for evaluating impacts related to open water placement of dredged material in inland waterways. It was created as the counterpart to the Ocean Testing Manual, and the testing and evaluation procedures in the ITM were patterned after the guidance in the Ocean Testing Manual. The ITM serves as the basis for developing and implementing dredged material evaluations for navigation projects and was created with the intent to provide uniformity between dredged material evaluations under the CWA and MPRSA. The testing framework permits regional flexibility with regard to implementation and application to accommodate project-specific, site-specific, or situation-specific issues.

Both the ITM and the Ocean Testing Manual describe a tiered testing approach for evaluating the potential for unacceptable adverse impacts associated with dredged material placement. Either open water or ocean placement may be eliminated as a dredged material placement alternative if there is "reason to believe" that unacceptable adverse effects will occur to either the water column or the benthic environment. Therefore, evaluation of both potential water column and potential benthic effects associated with dredged material placement are required.

Although the ITM and the Ocean Testing Manual serve as the primary guidance manuals for the evaluation of dredged material, various USEPA regions and states have developed Regional Implementation Manuals (RIMs) that describe region-specific testing requirements, testing methodologies, and recommended test species. Currently, neither USEPA Region III nor the State of Maryland has published guidance regarding region-specific requirements for dredged material evaluations. The USEPA Region III office in Philadelphia and the Maryland Department of the Environment (MDE) provide consultation on requirements on a project-specific basis.

Overall, the testing and evaluation procedures and criteria in the ITM and the Ocean Testing Manual are nearly equivalent. The ITM provides the most recent federal guidance for evaluating dredged material and is referenced as the primary information source for the remaining sections of both this chapter and report.

1.3 TIERED TESTING

The "tiered" testing approach to evaluate contaminant-related impacts associated with placement of dredged material is depicted in Figure 1-1 and a Tiered Testing Flow Diagram is provided in Figure 1-2. The process for evaluating potential water column and benthic impacts is detailed in Figures 1-3 and 1-4, respectively.

The initial tier (Tier I) uses readily available existing information to evaluate the impact of placement. If this information is inadequate to support a decision, testing proceeds through subsequent tiers of more extensive and specific testing until sufficient information is generated to support a decision. It is necessary to proceed through the tiers only until information on each topic becomes sufficient to make the required factual determination.

1.3.1 Tier I

Tier I consists of an evaluation of existing information to determine (1) if there is evidence or "reason to believe" that adverse effects could potentially occur, and (2) to identify potential contaminants of concern. If there is sufficient information to determine that the sediments are not contaminated and are similar to sediments in the proposed placement site, then the material may be excluded from additional testing, or only limited additional testing may be required.

1.3.2 Tier II

Tier II involves sediment and water chemistry. Sediments are collected for physical and chemical testing, and water is collected for elutriate testing. Bulk sediment chemistry data are used to calculate Theoretical Bioaccumulation Potential (TBP), the potential for contaminants to accumulate in organism tissue. Potential water column impacts are evaluated by comparing concentrations of contaminants in elutriates to national or state Water Quality Standards (WQS). If TBP exceeds a reference site value, or if dissolved contaminants exceed WQS, then Tier III testing is recommended.

1.3.3 Tier III

Tier III involves toxicity and bioaccumulation studies. Water column bioassays, whole sediment bioassays, and bioaccumulation studies are conducted to determine acute toxicity and to determine the bioavailability of contaminants. Water column tests evaluate the effects of dissolved and suspended particulates on organisms, after allowance for mixing that would occur within the water column. Whole sediment bioassays evaluate the effect of the sediment exposure to benthic organisms. Bioaccumulation tests evaluate the uptake of contaminants from the dredged material into the tissue of benthic organisms. The results of Tier III are usually sufficient to determine if the dredged material will cause adverse impacts.



1.3.4 Tier IV

Tier IV is implemented in situations where earlier tiers do not provided sufficient information to determine the potential effects of the dredged material. Tier IV uses case-specific toxicity and bioaccumulation studies designed to answer case-specific questions. These studies may include risk assessment, calculation of steady-state bioaccumulation, field assessments of resident biological communities, or food web modeling.

1.4 TESTING REQUIREMENTS

Regional testing requirements of the sediments proposed for dredging are dictated by a number of factors, including the history of the dredging location, proximity to point sources of contamination, present use of the dredging location, and the proposed placement location. Types of testing pertinent to open water and ocean placement include physical and chemical analysis of bulk sediment, elutriate testing, and bioassay and bioaccumulation studies. Table 1-1 summarizes the analyses or constituents analyzed in each type of test. Target detection limits, preservation techniques, and holding time requirements for target analytes are provided in *QA/QC Guidance for Sampling and Analysis of Sediments, Water, and Tissues for Dredged Material Evaluations* (USEPA/USACE 1995). Target analytes and benchmark species may be modified based on consultation with regulatory agencies to accommodate regional or project-specific requirements.

1.4.1 Reference Sediment

Reference sediment tests serve as the point of comparison (reference benchmark) to which all benthic test results are compared. Until recently, sediment from the proposed placement site has served as the point of comparison for the test results. Section 404 guidelines have recently been updated to include a "reference sediment" that is located outside the proposed placement site as a point of comparison for benthic evaluations, as opposed to sediments from inside the placement site (USEPA/USACE 1998). USEPA designates a reference location that is similar in grain size composition and subject to similar influences, but has not been impacted by previous placement of dredged material. Comparisons to non-impacted reference sediments are expected to yield more valid results regarding the potential for individual and cumulative impacts. More than one location may serve as a reference comparison, and depending upon the objectives of the project, it may be important to compare test results to both the actual conditions within the placement site and to a non-impacted outside reference area.

Importantly, "reference" sediment should not be confused with "control" sediment, or with "standard reference material" (SRM). "Control" sediment is a natural sediment that is used in the ecotoxicology laboratory to assess the health of the testing organisms and the acceptability of test conditions. SRM is sediment with a certified concentration of a constituent and is used by the analytical laboratory to monitor analytical accuracy.

1.4.2 Physical Analyses

Physical testing represents the minimal testing that would be required prior to any dredging project. Physical sediment characteristics include grain size and moisture content determinations. Since sediment contaminants tend to sorb to organic particles, sediments with a high percentage of sand are likely to contain fewer and lower concentrations of contaminants than sediments with higher percentages of silt and clay particles. Sediments primarily comprised of sand or gravel may qualify for testing exclusions (USEPA/USACE 1998). Specific gravity and Atterberg Limits may also be determined for evaluating compaction and settling of particles at open water placement sites.

1.4.3 Bulk Sediment Chemistry

Chemical analysis of bulk sediments is required for every type of placement alternative, with the exception of sediments that qualify for exclusions. Contaminants in sediment include: bulk organics (hydrocarbons that include oil and grease), halogenated hydrocarbons (persistent organics that degrade slowly), polynuclear aromatic hydrocarbons (PAHs) (organics that include petroleum products and petroleum by-products), metals, and nutrients. Chemical fractions targeted in the ITM are summarized in Table 1-1. Elevated concentrations of the target compounds may indicate a potential for toxicity to living organisms and a potential to bioaccumulate. Detected concentrations of the target compounds are often compared to screening criteria such as Sediment Quality Guidelines (SQG). Concentrations are also used to calculate the Theoretical Bioaccumulation Potential (TBP) of the compound. The TBP is compared to reference site sediment TBP values or literature values.

1.4.4 Elutriate Testing

Elutriate testing is required for open water and ocean placement of dredged material. Elutriates are created by mixing sediments and site water at a known ratio, allowing the particulates to settle, then testing the overlying water for dissolved constituents (USEPA/USACE 1998). The test simulates mixing and release of contaminants that would occur in the water column if the sediments were released or pumped into an aquatic environment. Water column effects are evaluated by measuring dissolved analytes and comparing the concentration to national and state Water Quality Standards (WQS) after allowance for mixing. A numerical mixing model (STFATE) (Johnson and Fong 1995) is used to calculate analyte concentrations in the water column under various mixing scenarios. Standard guidelines for elutriate preparation are described in the ITM.

1.4.5 Bioassays and Bioaccumulation Studies

Depending upon the results of bulk sediment chemistry, biological testing may be required for open water placement. Bioassays include water column toxicity tests, solid phase toxicity tests, and bioaccumulation studies. The ITM contains USEPA-developed and approved bioassay protocols and recommended benchmark species. Benchmark species are either easily cultured in a laboratory or easily collected in the field from undisturbed environments, respond to contaminants, and are relevant from an ecological perspective. Estuarine or marine species that would be appropriate for an estuarine or marine testing program are provided in Table 1-1. Water column tests measure the acute toxicity of elutriates to water column species and simulate impacts that would be expected to occur during a placement event. Solid phase tests measure the toxicity of the sediments to bottom-dwelling species that would be expected to live within the sediment post-placement. Bioassay results are based on percent survival and are statistically compared to the results for a laboratory control (elutriate testing) or to the reference sediment (solid phase testing). LC50 concentrations (the elutriate concentration that is lethal to 50% of the organisms) or EC50 concentrations (the elutriate concentration that is sub-lethal to 50% of the organisms) are calculated.

Bioaccumulation tests are conducted to determine the uptake (or bioavailability) of contaminants into organism tissue when exposed to sediments for 28 days. Tissue concentrations are statistically compared to reference tissue concentrations and to U.S. Food and Drug Administration (USFDA) Action Levels and USEPA Tolerance Values. Tissue-residue concentrations that statistically exceed the reference values may be compared to residue-effects data (if available) to identify potential physiological, morphological, and reproductive impacts to the organism. For constituents where few or no effect data are available, other approaches are used to evaluate the potential for toxic response in the organism.

1.5 EVALUATION CRITERIA

Each tier of testing includes evaluation criteria to determine if the existing information is sufficient to make a factual determination regarding unacceptable adverse impact or to determine if additional testing is required. The evaluation criteria for each tier are presented in the following sections. Criteria presented are specific to guidance in the ITM. Unless otherwise noted, the criteria for both open water and ocean placement are equivalent.

1.5.1 Tier I

According to the ITM (USEPA/USACE 1998), after consideration of all available Tier I information, one of the following conclusions is reached:

- 1) Existing information does not provide a sufficient basis for making factual determinations. In this case, further evaluation in higher tiers (Tier II, Tier III, etc.) is appropriate.
- 2) Existing information provides a sufficient basis for making factual determinations. In this case, one of the following decisions is reached:
 - (a) The material meets the criteria for exclusion from testing and proposed placement proceeds with no additional testing.
 - (b) The material does not meet the criteria for exclusion from testing, but information concerning the potential impact of the material is sufficient to make a factual

determination regarding potential for water column impact, benthic toxicity, and benthic bioaccumulation.

1.5.2 Tier II

1.5.2.1 Water Quality

According to the ITM (USEPA/USACE 1998), after consideration of the Tier II water quality data, one of two possible conclusions is reached regarding the potential water column impact of the proposed dredged material:

- 1) The available water quality requirements are met. Further information on water column toxicity must be evaluated in Tier III when there are contaminants of concern for which applicable water quality criteria or standards are not available or where interactive effects are of concern.
- 2) Concentrations of one or more of the dissolved contaminants of concern, after allowance for mixing that would occur in the water column at the placement site, exceeds applicable water quality criteria or standards beyond the boundaries of the mixing zone. In this case, the proposed discharge of dredged material does not comply with the water quality criteria or standards.

1.5.2.2 Benthos

According to the ITM (USEPA/USACE 1998), after comparison of the Tier II TBP data for nonpolar organic contaminants in the proposed dredged material and reference sediment, one of the following two conclusions is reached:

- The dredged material is predicted to not result in unacceptable adverse effects due to bioaccumulation of the measured non-polar organic compounds (i.e., TBP for dredged material does not exceed TBP for reference). However, further evaluation of *biological effects* in Tier III is necessary to furnish information to make determinations under the Guidelines.
- 2) The available information is not sufficient to predict whether the dredged material will result in unacceptable adverse effects due to bioaccumulation of non-polar organic compounds (i.e., TBP for dredged material exceeds TBP for reference). Further evaluation of *bioaccumulation* in Tier III is necessary to furnish information to make determinations under the Guidelines.



1.5.3 Tier III

1.5.3.1 Water Column Toxicity

According to the ITM (USEPA/USACE 1998), after considering water column test results and mixing at the placement site, one of the following conclusions is reached:

- 1) The 100% dredged material elutriate toxicity is not statistically higher than the laboratory dilution water. Therefore, the dredged material is not predicted to be acutely toxic to water column organisms. However, benthic impact must also be evaluated.
- 2) The concentration of dissolved plus suspended contaminants, after allowance for mixing, does not exceed 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material is not predicted to be acutely toxic to water column organisms.
- 3) The concentration of dissolved plus suspended contaminants, after allowance for mixing in the water column, exceeds 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material is predicted to be acutely toxic to water column organisms.

1.5.3.2 Benthic Toxicity

According to the ITM (USEPA/USACE 1998), benthic toxicity testing of contaminants in the dredged material in Tier III will result in one of the following possible conclusions:

- 1) Mean test organism mortality in the dredged material is not statistically greater than in the reference sediment, or does not exceed mean mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). Therefore, the dredged material is predicted not to be acutely toxic to benthic organisms. However, bioaccumulation of contaminants and water quality effects must also be considered.
- 2) Mean test organism mortality in the dredged material is statistically greater than in the reference sediment *and* exceeds mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). In this case, the dredged material is predicted to be acutely toxic to benthic organisms.

1.5.3.3 Bioaccumulation

According to the ITM (USEPA/USACE 1998), tissue residues are compared to FDA levels and one of the following conclusions is reached:

1) Tissue concentrations of one or more contaminants are not statistically less than the FDA levels. Therefore, the dredged material is predicted to result in unacceptable benthic bioaccumulation of contaminants.

2) Tissue concentrations of all contaminants are either statistically less than FDA levels or there are no FDA levels for the contaminants. In this case, the information is insufficient to make a factual determination with respect to benthic bioaccumulation of contaminants. The dredged material requires further evaluation under Tier III as described below to make a factual determination under the Guidelines.

Contaminant concentrations in tissues exposed to dredged material that are statistically lower than FDA levels, or for which no FDA levels exist, are compared to tissue contaminant concentrations for organisms exposed to reference sediment. One of the following conclusions is reached:

- Tissue concentrations of contaminants of concern in organisms exposed to dredged material do not statistically exceed those of organisms exposed to reference sediment. Therefore, the dredged material is not predicted to result in unacceptable benthic bioaccumulation of contaminants. However, benthic toxicity tests must also be evaluated.
- 2) Tissue concentrations of contaminants of concern in organisms exposed to dredged material statistically exceed those of organisms exposed to reference sediment. The final conclusion regarding benthic bioaccumulation of contaminants requires region-specific technical evaluation. Additional testing (Tier IV) may be required and benthic toxicity must be evaluated.

Region-specific technical evaluation of bioaccumulation data for contaminants in dredged material tests that statistically exceed the reference may include the following:

- 1) What is the toxicological importance of the contaminants that statistically exceed the reference? (Do they biomagnify? Do they have effects at low concentrations?)
- 2) By what magnitude does bioaccumulation from the dredged material exceed bioaccumulation of the reference material?
- 3) What is the propensity for the contaminants with statistically significant bioaccumulation to biomagnify within food webs?
- 4) What is the magnitude by which contaminants, whose bioaccumulation from the dredged material exceeds that from the reference material, also exceed the concentrations found in comparable species living in the vicinity of the proposed placement site?
- 5) For how many contaminants is bioaccumulation from the dredged material statistically greater than bioaccumulation from the reference?

According to the ITM (USEPA/USACE 1998), after considering of all of the above evaluation factors, one of the following conclusions is reached:

- 1) Placement of the dredged material is predicted to not result in above-reference toxicity or unacceptable benthic bioaccumulation of contaminants.
- 2) Placement of the dredged material is predicted to result in above-reference toxicity or unacceptable bioaccumulation of contaminants.
- 3) Further information is needed to make factual determination, specifically in Tier IV.

1.6 DECISION PROCESS

The tiered evaluation process provides decision-makers with the weight of evidence regarding the potential for short-term, long-term, or unacceptable contaminant-related impacts as a result of dredged material placement. The results of the evaluation process may be used to facilitate dredged material management options, implement a proposed open water placement program, or identify alternative placement locations or dredging and discharge methods. Results may demonstrate that open water placement is a viable placement alternative, or results may demonstrate that alternative placement location or dredging and discharge methods are required to comply with 404(b)(1) Guidelines.



Figure 1-1 Inland Testing Manual Tiered Approach for Evaluating Potential Impacts Related to Dredged Material Placement (USEPA/USACE 1998)





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Figure 1-2 Tiered Testing Flow Diagram





CHAPTER 2 -PROJECT DESCRIPTION

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2. PROJECT DESCRIPTION

2.1 PROJECT BACKGROUND

Site 104 is a formerly used open-water dredged material placement site in the upper Chesapeake Bay. The Baltimore District USACE is proposing to reuse the site to meet the short-term placement needs generated by scheduled maintenance dredging and new work dredging of the mainstem Chesapeake Bay navigation channels. Physical, chemical, and biological characterization of the sediment is required to comply with Section 404 of the CWA and to determine if the sediments are appropriate for open-water placement.

EA Engineering, Science, and Technology, Inc. was contracted by the USACE-Baltimore District to conduct an evaluation of dredged material proposed for placement at Site 104. A tiered evaluation of the sediment proposed for dredging was conducted to determine if materials from the Baltimore Harbor approach channels were appropriate for open-water placement. In addition to evaluating the sediments with regard to open-water placement, the sediments were also evaluated with regard to ocean placement at the Norfolk Ocean Disposal Site (NODS) in the Atlantic Ocean. The project consisted of a tiered evaluation that included review of existing information; collection of sediments and site water; chemical analysis of sediment, site water, and elutriates; and water column bioassays, sediment bioassays, and bioaccumulation studies.

2.2 **PROJECT PURPOSE**

The purpose of this project was (1) to document existing sediment quality conditions and potential impacts related to open water placement of dredged material in the Environmental Impact Statement (EIS) at Site 104; (2) to determine if the sediments are appropriate for openwater (Chesapeake Bay) placement; (3) to evaluate feasible placement alternatives (such as ocean placement); and (4) to facilitate dredged material management decisions.

This project provides the data necessary to document the existing physical, chemical, and biological characteristics of sediments and water in the channels, in the proposed placement area, and in two additional reference areas (outside Placement Site 104 and the Ocean Reference site).

2.3 PROJECT OBJECTIVES

The overall objective of the project was to obtain, analyze, and evaluate sediment and water samples that are representative of the areas proposed for dredging during the period that Site 104 would be used for placement. The results of this investigation document the existing physical, chemical, and biological characteristics of sediment and water from the approach channels proposed for maintenance dredging and the areas proposed for new work dredging.



Specific objectives of the project were to:

- Collect the required volume of sediment and site water for physical, chemical, and biological analyses, and for elutriate preparation.
- Collect samples from specified locations with sufficient distribution to characterize dredging sites within positioning accuracy appropriate for the project objectives.
- Collect and transfer sediment to appropriate, laboratory-prepared containers and preserve/hold samples for analysis according to protocols that ensure sample integrity.
- Test and characterize sediments with regard to physical characteristics, chemical contamination, biological toxicity, and the potential for bioaccumulation.
- Test site water, elutriates, and placement site/reference site water with regard to potential chemical contamination.
- Evaluate the results of the sediment chemistry, elutriate, toxicity, and bioaccumulation data with regard to open-water and ocean placement alternatives.

2.4 DESCRIPTION OF PROJECT AREA

The project area is located in the upper Chesapeake Bay, Maryland, as shown in Figure 2-1. A total of 56 locations were targeted for sampling in the testing program. Stations were located in Baltimore Harbor approach channels proposed for maintenance dredging (28 locations), in approach channels recently dredged (17 locations), in new work areas (2 locations), inside Placement Site 104 (5 locations), and outside Placement Site 104 (4 outside reference locations). Baltimore Harbor Approach Channels proposed for dredging in FY00 and FY01 include: Craighill Entrance, Craighill Angle, Swan Point, Tolchester, and the northern Chesapeake and Delaware (C&D) approaches. Channels that have been recently dredged include: Craighill, Craighill Upper Range, Cutoff Angle, Brewerton Eastern Extension, and areas of the southern C&D Approach Channels. The proposed Tolchester Channel straightening represents a new work dredging project.

2.4.1 Baltimore Harbor Approach Channels

This project includes approach channels to Baltimore Harbor located east of Rock Point/North Point. The approach channels maintained by USACE–Baltimore District include: Craighill Entrance, Craighill, Craighill Upper Range, Cutoff Angle, Brewerton Eastern Extension, Tolchester, and Swan Point. In addition, several locations in the C&D Canal approaches have been included in this study. The channels are maintained to a depth of -51 ft Mean Lower Low Water (MLLW), with the exception of the Tolchester Channel, C&D Canal approaches, and Swan Point Channel, which are maintained to a depth of -37 ft MLLW. Shoaled areas within the channels are typically dredged every 3-4 years, depending upon the shoaling rate as indicated by periodic USACE bathymetric surveys.

Previous investigations (EA 1996a and 2000c) have characterized the physical and chemical characteristics of surficial sediments in the approach channels. The channel sediments generally consist of fine-grained silt and clay materials. Detected chemical constituents include metals, PAHs, and chlorinated organic compounds [pesticides and polychlorinated biphenyls (PCBs)]. The detected concentrations are generally similar to background values that are typical of the upper Chesapeake Bay (Eskin et al. 1996). Results from previous investigations are described in Chapter 3.

2.4.2 Placement Site 104

Site 104 is a previously used, 1800-acre site located approximately 0.37 mile north of the Bay Bridge and one mile west of Kent Island. The site was established as a placement area in 1924 by USACE and was used for dredged material placement through 1975. The site is approximately 4.2 miles long and 0.65 mile wide. Depths range from -42 to -78 ft MLLW.

The physical and chemical characteristics of surficial sediments within and directly outside of the Site 104 boundaries were characterized by a previous study in 1997 (EA 1998a). Values of metals, PAHs, and chlorinated organic compounds in most sediment were similar to values reported for other areas of the upper Chesapeake Bay. One station, KI-7, located in the southern end of Site 104, exhibited concentrations of PAHs and metals that were elevated above sediment quality guideline values for aquatic life. Results from previous investigations are described in Chapter 3.

2.4.3 Reference Areas

Originally, the proposed placement site (Inside Site 104) was chosen as the reference area for channel comparisons of chemical and biological data. An upper Chesapeake Bay reference area (Outside Site 104) was added to the evaluation process at the request of USEPA Region III. The Outside Site 104 reference area represents an area that is physically and hydrologically similar to Inside Site 104, but has not been impacted by historical placement activities. The outside reference sampling locations corresponded to stations KI-11, KI-14, KI-15, and KI-16 that were sampled in previous sediment investigations (EA 1998a) (see Figure 4-13). Depths in the Outside Site 104 reference area ranged from -35 to -90 ft MLLW. Channel comparisons to Inside Site 104 are important for determining how sediment quality at the actual placement site will be impacted by the proposed placement activities. Channel comparisons to Outside Site 104 are important for determining how sediment quality at the actual placement site will be impacted by the proposed placement activities. Channel comparisons to Outside Site 104 are important for determining how sediment quality at the actual placement site will be impacted by the proposed placement activities. Channel comparisons to Outside Site 104 are important for determining if the materials proposed for placement are of a quality that is similar to areas where placement has not previously occurred and for assessing individual and cumulative impacts.

In addition to the inside and outside Site 104 reference areas, an Atlantic Ocean reference area (specified by USEPA Region III) was chosen as the point of comparison for the ocean placement alternative. The Norfolk Ocean Disposal Site reference area was located approximately 14 miles

southwest of the NODS and 2 miles southeast of the Chesapeake Light Tower in the Atlantic Ocean. Water depths in the area ranged from -71 to -75 ft. Sampling and analysis of sediment from the NODS reference area were conducted in conjunction with a sediment evaluation for the Woodrow Wilson Bridge Project (EA 2000b) and was conducted subsequent to the approach channel, Inside Site 104, and Outside Site 104 testing.

2.5 EXPERIMENTAL DESIGN

The analytical and ecotoxicological components of the Approach Channel and Placement Site 104 dredged material evaluation followed the tiered testing guidance described in the following documents:

- USEPA/USACE, 1998 (EPA-823-B-98-004). Evaluation of Dredged Material Proposed for Discharge in Waters of the U.S.–Testing Manual (Inland Testing Manual).
- USEPA/USACE, 1991 (EPA-503/8-91/001) Evaluation of Dredged Material Proposed for Ocean Disposal, Testing Manual (The Green Book).
- USEPA/USACE, 1995 (EPA-823-B-95-001). QA/QC Guidance for Sampling and Analysis of Sediments, Water, and Tissues for Dredged Material Evaluations.

In addition, regional requirements/recommendations were provided by USEPA Region III during the collection, testing, and evaluation process.

The analytical testing program included the following components:

- Physical analyses of bulk sediment (grain size, Atterberg Limits, specific gravity, and total solids determinations).
- Chemical analysis of bulk sediment, site water, and elutriates for project-specific target analytes: volatile organic compounds (VOCs), semivolatile organic compounds (SVOCs), metals, chlorinated and organophosphorus pesticides, PAHs, PCB congeners, dioxin and furan congeners, butyltins, cyanide, total sulfides, simultaneously extracted metals (SEM)/acid volatile sulfides (AVS), ammonia, total Kjeldahl nitrogen (TKN), total organic carbon (TOC), total phosphorus (TP), and nitrate + nitrite.

The ecotoxicological testing program included the following components:

- Water column bioassays with *Mysidopsis bahia* (opossum shrimp), *Cyprinodon* variegatus (sheepshead minnow), *Mytilus* sp. (blue mussel), and *Menidia beryllina* (inland silverside).
- 10-day whole sediment bioassays with *Leptocheirus plumulosus* (estuarine amphipod) and *Neanthes arenaceodentata* (estuarine polychaete).

• 28-day whole sediment bioaccumulation studies with *Nereis virens* (sand worm) and *Macoma nasuta* (blunt-nose clam).

Data analysis and evaluation included the following components:

- Chemical data collected in 1997 and 1998 for Site 104 (EA 1998a) and the approach channels (EA 2000c), respectively, were reviewed to identify contaminants of concern and to determine if testing in subsequent tiers was necessary.
- Chemical concentrations in bulk sediment were compared to reference sediment concentrations and to published Sediment Quality Guidelines (SQGs) (Buchman 1999; MacDonald 1994; MacDonald et al. 1996).
- TBP was calculated for pesticides, PAHs, PCB congeners, and dioxin and furan congeners.
- Chemical concentrations in elutriate samples were compared to concentrations in receiving waters (Site 104 and the Ocean Reference site) and to USEPA saltwater acute and chronic Water Quality Criteria (WQC) for protection of aquatic life and to USEPA WQC for human health (consumption of aquatic organisms).
- Water quality modeling was conducted using the STFATE model to determine the maximum expected dilution (mixing) within the specified mixing zone and to determine water quality compliance. Placement events at both Site 104 and the NODS were modeled.
- For the water column bioassays, LC50 (lethal concentration) and EC50 (effect concentration) values were calculated for survival and effect data, respectively. In addition, results were statistically analyzed to determine whether survival in the channel, placement site, or reference sediments was significantly lower than the laboratory control survival.
- For the whole sediment bioassays, survival data were statistically compared to the reference site survival data to determine whether channel sediments were acutely toxic.
- 28-day bioaccumulation survival in the channel sediment was statistically compared to survival in the reference sediments to determine if organism survival in the channel sediments was significantly lower than organism survival in the reference sediment.
- Chemical concentrations in worms and clams exposed to the channel sediments were statistically compared to chemical concentrations in organisms exposed to the reference sediments (Inside Site 104, Outside Site 104, and the Ocean Reference) to determine if contaminant tissue-residues were significantly higher in the channel-exposed organisms.

- Concentration Factors (CFs) (comparing day 0 to day 28 tissue residues) were calculated to quantify the magnitude of contaminant accumulation in tissues.
- Concentrations of pesticides and PCBs in the worm and clam tissue were statistically compared against USFDA Action Levels and USEPA Tolerance Values (USFDA 1998 and USEPA 1998) to determine if analyte concentrations in tissues were significantly lower than FDA Action Levels, USEPA Tolerance Values, or USEPA Guidance Levels (for food consumption) at the 95% confidence level.
- Tissue residue concentrations that were significantly higher than reference tissue residue concentrations were compared to residue-effects data derived from the USACE-Waterways Experiment Station (WES) Environmental Residue Effects Database (ERED) and from a tissue residue-effects database compiled by Jarvinen and Ankley (1999).
- Critical Body Residue (CBR) was calculated to assess the potential impact of PAHs and pesticide body burden in aquatic organisms.
- Tissue residue concentrations that were significantly higher than reference tissue residue concentrations for which FDA Action Levels and EPA Tolerance Values or Guidance Levels did not exist were compared to EPA Fish Tissue Advisory Screening Values (SVs) (USEPA 1995) and USEPA Region III Risk Based Criteria (RBCs) for fish tissue (USEPA 2000a).

2.6 **REPORT ORGANIZATION**

This report contains a comprehensive summary of historical data and field activities; results of bulk sediment testing, elutriate testing, ecotoxicological testing, and bioaccumulation studies; and evaluation of the data with respect to open water and ocean placement alternatives. Bulk sediment and elutriate data collected during previous investigations in 1997 and 1998 are summarized in Chapter 3. The field sampling program for the project is described in Chapter 4. Analytical methodologies for the testing of bulk sediment, site water, elutriates, and tissue are provided in Chapter 5. Results for the bulk sediment testing, site water and elutriate testing, ecotoxicological testing, and bioaccumulation studies are provided in Chapters 6, 7, 8, and 9, respectively. Chapter 10 provides a summary of significant findings. A list of cited references is provided in Chapter 11. Chapter 12 provides a list of persons who assisted with the preparation and review of this document.

The project Field Sampling Plan (FSP), Quality Assurance Project Plans (QAPPs) for analytical and ecotoxicological testing, supplemental reports, and raw data tables are appended to this report. Field logbooks, field and laboratory data sheets, chain-of-custody documentation, and analytical narratives are included as attachments in subsequent volumes of this report.

2.7 **DEFINITIONS OF TERMS**

The following words and terms are used throughout this document. Definitions are provided as follows:

- *sampling reach* refers to a channel, placement site, or reference area where samples were collected. Sediment from each channel reach was tested and evaluated separately to determine if material from a particular reach was suitable for open water or ocean placement.
- *placement site* refers to either the proposed Placement Site 104 (Inside Site 104), or the Norfolk Ocean Placement Site (NODS).
- *reference sites* refers collectively to the Inside Site 104 (Figure 4-11), Outside Site 104, (Figure 4-13) and the Ocean Reference areas (Figure 4-14), unless individually specified.
- *maintenance dredging area* refers to an area that is routinely dredged every 3-5 years; typically 1-4 ft of deposited unconsolidated material or shoaled material is removed to maintain channel depths.
- *new work dredging* refers to dredging required for new projects where dredging has not previously occurred (i.e., new channels, channel widening, or channel straightening); consolidated, native or historical sediments are removed to depths dictated by project requirements.
- *reference sediment* refers to sediment that the channel sediments are compared to (Inside Site 104, Outside Site 104, or Ocean Reference).
- *control sediment or control* refers to a natural sediment or control media that is used in the biological laboratory to assess the health of the test organisms and acceptability of a test.





Figure 2-1. Location Map: Baltimore Harbor Approach Channels and Site 104.

CHAPTER 3 -AVAILABLE INFORMATION

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3. AVAILABLE TIER I INFORMATION

3.1 EXISTING HISTORICAL DATA SETS

A Tier I evaluation requires the review of available data that are representative of the areas proposed for dredging and/or placement. The objective of the review is to identify contaminants of potential concern (COPCs) and to determine if additional testing in subsequent tiers is necessary. Two previous studies by USACE–Baltimore District have documented sediment quality/chemistry in the Baltimore Harbor approach channels (EA 1996a and EA 2000c). One previous study conducted in 1997 has also documented sediment quality/chemistry inside Site 104 and in the vicinity immediately outside Site 104 (EA 1998a). Data from these studies is reviewed in the following section.

In 1995 (EA 1996a) and 1998 (EA 2000c), USACE–Baltimore District conducted routine sediment chemistry and elutriate studies in the Chesapeake Bay Baltimore Harbor approach channels. Because maintenance dredging continuously occurs in the approach channels, USEPA Region III considers data collected no longer than within the past 3 years prior to dredging to be most representative of existing conditions within proposed dredging areas. In addition, the analytical methodology for sediment testing has improved substantially within the past 5 years. Analytical testing laboratories are able to identify and report analyte concentrations to lower detection limits, and they have a better understanding of the complexities of matrix and moisture interferences associated with sediment testing. Based on this information, the 1998 channel data most accurately identify and represent the existing conditions within the areas proposed for dredging and placement at Site 104, and the 1997 data best represent the conditions inside Site 104 and the vicinity outside Site 104. Therefore, only the 1998 channel data are discussed in this chapter; the 1995 channel data are appended to this report for historical reference (Appendix E). These data provide chemical information that is sufficient to make a factual determination regarding the need for additional testing.

3.2 DATA ANALYSIS

3.2.1 Bulk Sediment Data

For sediments, data were evaluated based on mean concentrations of constituents detected within each channel. The mean concentrations best represent the concentrations that would be expected when the material is dredged, mixed together, and placed in large volumes.

3.2.1.1 Mean Calculations

Mean concentrations of detected analytes were calculated for each of the eight sampling reaches tested by USACE-Baltimore in 1998 (Brewerton Eastern Extension, Craighill Entrance, Craighill, Craighill Angle, Craighill Upper Range, Cutoff Angle, Swan Point, and Tolchester) (EA 2000c) and for the Inside Site 104 and Outside Site 104 areas tested in 1997 (EA 1998b). The detection limit was substituted for non-detected analytes in the calculations of the mean. If

an analyte was not detected in any sample tested within a sampling reach, the mean detection limit was reported and qualified with a "U". One field duplicate sample was included in the means calculations for the Tolchester Channel.

For individual samples, PCB concentrations were determined by summing the 18 summation congeners (as specified in Table 9-3 of the ITM). The total summed concentration was then multiplied by a factor of 2 following the National Oceanic and Atmospheric Administration (NOAA) (1993) approach for total PCB determinations. Total PAHs were determined by summing the concentrations of PAHs in each sample. In the summation calculations for both Total PCBs and Total PAHs, three total values are presented in the data tables:

- Non-detects = zero (ND=0);
- Non-detects = $\frac{1}{2}$ of the detection limit (ND= $\frac{1}{2}$ DL); and
- Non-detects = the detection limit (ND=DL).

The substitution of the detection limit (ND=DL) provides the most conservative approach to calculating and evaluating the data. However, in cases where few PCB congeners or PAHs are detected, the detection limit drives the total value and overestimates the actual expected concentration.

3.2.1.2 Comparisons to Sediment Quality Guidelines

Mean concentrations of detected analytes in sediment samples were compared against SQG for marine sediments. Among the most commonly used of the methods that attempt to provide sediment contaminant concentration values that differentiate sediments of little concern from those predicted to have adverse biological effects are the Threshold Effect Level (TEL) and Probable Effect Level (PEL) (Buchman 1999; MacDonald 1994; MacDonald et al. 1996).

TELs represent the contaminant concentration below which adverse biological effects rarely occur. PELs represent the contaminant concentration above which adverse biological effects frequently occur. Values that fall between the TEL and PEL represent the concentrations at which adverse biological effects occasionally occur. TEL and PEL screening values are provided in Table 3-1.

O'Connor et al. (1998) and O'Connor and Paul (1999) quantitatively evaluated the reliability of sediment toxicity predictions based on Effect Range Low (ERL) / Effect Range Median (ERM) values, which are derived by a process very similar to the TEL/PEL process and have similar values. Both papers attempt to validate the values using large independent datasets that contain both sediment chemistry and sediment toxicity data for each sample. Both papers define a toxic sediment as one that produces less than 80 % survival of amphipods after a 10-day exposure to whole sediment, the same criterion of whole sediment toxicity used in the ITM. Both papers reach the same conclusions. O'Connor and Paul (1999) found that within a 2,475-sample dataset, 2,087 (84%) of the samples were not toxic. A total of 730 samples did not exceed any ERL (comparable to TEL), and 697 (95.5%) of these were not toxic. This indicates that not

exceeding an ERL is a reliable predictor of non-toxicity, and the same should be true for the closely related TEL. Of the 453 samples that exceeded at least one ERM, 186 (41%) actually produced toxicity. Therefore, exceeding an ERM (much less an ERL) is less than 50% accurate as a predictor of sediment toxicity, and the same is expected to be true for the closely related PEL (and TEL). This independent evaluation indicates that:

- not exceeding a TEL should reliably predict the absence of whole sediment toxicity,
- exceeding a PEL (much less a TEL) does not reliably indicate toxicity, and
- many, perhaps even most, sediments that exceed one or more PELs are not toxic.

Because TEL/PEL have been widely used despite their recently demonstrated low reliability in predicting toxicity, the mean concentrations of contaminants in the sediments proposed for placement at Site 104 were compared to the TEL and PEL values for all analytes for which TEL/PEL values have been developed. Comparison of sediment chemistry to SQGs is not a part of the tiered testing evaluations in the ITM (USEPA/USACE 1998) or the Green Book (USEPA/USACE 1991). For dredged material evaluations, SQG are used as a tool to assist with identification of COPCs and to provide additional weight of evidence in the evaluation (USACE–WES 1998). Comparisons to TEL/PEL values were used only for these purposes in this evaluation of the sediments proposed for placement at Site 104.

3.2.2 Elutriate Data

Analytes detected in the elutriates were compared to Maryland Department of the Environment (MDE) proposed water quality criteria [Maryland Register 27(17):1628-1636] and to USEPA saltwater acute and chronic aquatic life water quality criteria and water quality criteria for the protection of human health from the consumption of aquatic organisms (USEPA 1998 National Recommended Water Quality Criteria [63 Federal Register 68354 – 68364; 10 December 1998]). MDE's proposed criteria are more extensive than the State of Maryland's current standards, and the proposed criteria are identical/equivalent to the USEPA criteria. Applicable USEPA/ proposed MDE water quality criteria values (for detected analytes only) are provided in Table 3-2. Human health criteria are based on MDE's risk-factor of 10^{-5} .

Total PCB concentrations in the elutriates were determined by summing the 18 summation congeners (as specified in Table 9-3 of the ITM). The total summed concentration was then multiplied by a factor of 2 following the NOAA (1993) approach for total PCB determinations. Total PAHs were determined by summing the concentrations of PAHs in each sample. In the summation calculations for both Total PCBs and Total PAHs, three total values are presented in the data tables:

- Non-detects = zero (ND=0);
- Non-detects = $\frac{1}{2}$ of the detection limit (ND= $\frac{1}{2}$ DL); and
- Non-detects = the detection limit (ND=DL).

The substitution of the detection limit (ND=DL) provides the most conservative approach to calculating and evaluating the data. However, in cases where few PCB congeners or PAHs are detected, the detection limit drives the total value and overestimates the actual expected concentration.

3.3 BULK SEDIMENT RESULTS

Results of the bulk sediment chemistry analyses are presented in the following subsections. Sample weights were adjusted for percent moisture (up to 50% moisture) prior to analysis to achieve the lowest possible detection limits. Because sediments contain a large proportion of moisture, each analyte has a sample-specific detection limit. The detection limit range within each reach is provided within each analytical summary table. Analytical results are reported on a dry weight basis. Definitions of organic and inorganic data qualifiers are presented in Tables 3-3 and 3-4, respectively. Station locations and analytical methodology utilized in the 1997 and 1998 sampling efforts are provided in EA (1998a) and EA (2000c), respectively. Analytical data for individual stations within each channel reach are provided in Appendix E.

Analytical results (mean concentrations) are provided in Tables 3-5 through 3-15. Results of TEL and PEL comparisons for mean concentrations are provided in Tables 3-16 and 3-17, respectively. Frequency of detection (number of detected analytes / total number of tested analytes) by analytical fraction is provided in Table 3-18.

3.2.2 Physical Analysis

Results of the physical analyses are provided in Table 3-5. Grain size determinations indicated that the channel sediments and sediment Inside Site 104 and Outside Site 104 were primarily comprised of silt and clay. Sediments in the Craighill Channel, Craighill Entrance, and Craighill Angle contained the highest proportions of sand (36%, 24%, and 14%, respectively). Sand comprised less than 8% of the sediments in the other tested channels.

3.2.3 Inorganic Non-Metals and Nutrients

Results of inorganic (non-metal) analyses and nutrients are provided in Table 3-6. Mean TOC in the channels ranged from 3.4% (Craighill Channel) to 14.1% (Cutoff Angle). Mean TOC concentrations were 6.8% and 8.9% for Inside and Outside Site 104, respectively. Overall, the highest ammonia-nitrogen concentration was reported for Outside Site 104 (137.7 mg/kg). Mean ammonia-nitrogen concentrations in the channels varied widely, ranging from 2.47 to 127.9 mg/kg, with the highest mean value at Swan Point. Nitrate + nitrite concentrations in the channels ranged from 0.12 to 4.7 mg/kg, with the values in the shallowest channels on the eastern side of the Upper Bay (Swan Point and Tolchester) ranging from 14 to 40 times higher than the other channels. The highest channel TKN (organic nitrogen + nitrogen- ammonia) concentrations were reported for Swan Point (1,430.0 mg/kg), Tolchester (2,407.5 mg/kg), and Craighill Angle (1565.0 mg/kg). Total phosphorus in the channels ranged from 161.8 mg/kg to (Craighill) to 320.3 mg/kg (Brewerton Eastern Extension). Total sulfides were substantially

higher in Inside and Outside Site 104 sediments (1,804 and 1,692 mg/kg, respectively) than in the channel sediments (range of 39.17 to 612.3 mg/kg).

3.3.3 Volatile Organic Compounds

Mean concentrations of VOCs are provided in Table 3-7. Only 3 of 34 tested VOCs were detected in the channel sediments (carbon disulfide, chloromethane, and dichloromethane). In the 8 channel reaches combined, VOCs were detected in 20 of 850 analyses (2.4%) in the channel sediments (Table 3-18). VOCs were not detected Inside or Outside Site 104 (0 of 136 cases for each location). None of the tested VOCs was detected in sediments from the Cutoff Angle, Swan Point, or Inside and Outside Site 104. Mean concentrations of the VOCs that were detected were generally low within the channels. Although none of the compounds was detected in the laboratory method blanks, both carbon disulfide and dichloromethane (methylene chloride) can be laboratory contaminants.

3.3.4 Semivolatile Organic Compounds

Mean concentrations of SVOCs are provided in Table 3-8. Only 7 of 51 of the tested SVOCs were detected in the channel sediments. Only 3 of 49 tested SVOCs were detected within the Inside and Outside Site 104 sediments. In the 8 channel reaches, SVOCs were detected in 28 of 1,275 cases (2.1%) (Table 3-18). In the Inside and Outside Site 104 sediments, SVOCs were detected in 2 of 208 possible cases (1%) in each area, respectively.

The detection limit for 2-methylnaphalene exceeded the TEL value (Table 3-16); therefore, it is not possible to determine whether concentrations of 2-methylnaphalene exceeded the TEL in channels where it was not detected. 2-Methylnaphalene was detected above the TEL in Brewerton Eastern Extension, Craighill Angle, and Tolchester Channel, but the mean concentrations did not exceed the PEL value (Table 3-17). Mean concentrations of bis(2-ethylhexyl) phthalate exceeded the TEL value (Table 3-16) in each reach where it was detected (Craighill, Craighill Entrance, and Tolchester) except Inside Site 104.

3.3.5 Chlorinated Pesticides

Mean concentrations of chlorinated pesticides are provided in Table 3-9. 73 % of the tested chlorinated pesticides were not detected in the sediments. Six of 22 tested chlorinated pesticides were detected in sediments from the channels and from Inside Site 104. Three of 22 tested chlorinated pesticides were detected in sediments from Outside Site 104. In the eight channel reaches, chlorinated pesticides were detected in 20 of 550 cases (3.6%). Chlorinated pesticides were detected in 7 of 88 possible cases Inside Site 104 (8%) and in 4 of 88 cases (4.5%) Outside Site 104 (Table 3-18). The Brewerton Eastern Extension and Inside Site 104 had the highest number of detected pesticides (six). Mean concentrations of 4,4'DDD and gamma-BHC in the Craighill Entrance were the only detected pesticides that exceeded TEL values (Table 3-16). The mean detection limit for chlordane was above the TEL and PEL values in all channels and Inside and Outside Site 104, so it is not possible to determine whether chlordane exceeded the TEL or PEL in any sample.
3.3.6 Organophosphorus Pesticides

Mean concentrations of organophosphorus pesticides are provided in Table 3-10. Organophosphorus pesticides were not detected in either the channel, Inside Site 104, or Outside Site 104 sediments (0 of 165 cases).

3.3.7 PCB Aroclors and Congeners

Four of seven PCB aroclors (aroclor 1254 and aroclor 1260) were detected in the channel sediments, and two PCB aroclors were detected in sediments from Inside and Outside Site 104 (Table 3-11). Aroclors were detected in 27 of 174 cases (15.5%) in the channel sediments, in 3 of 28 cases (11%) Inside Site 104 and in 2 of 28 cases (7%) Outside Site 104 (Table 3-18). Of the eight channels, PCB aroclors were not detected in Craighill Channel or Craighill Entrance.

Mean concentrations of PCB congeners are provided in Table 3-12. Congeners were not tested in Craighill Channel, Craighill Entrance, Swan Point, or Tolchester Channel in the 1998 study. In the four channels that were tested, PCB congeners were detected in only 11 of 182 cases (6%). Eight of the 26 target PCB congeners were detected in the channel sediments (BZ#101, BZ#138, BZ#153, BZ#170, BZ#180, BZ#187, BZ#206, and BZ#209). None of the 26 tested congeners was detected in sediments from Inside Site 104 or Outside Site 104 (0 of 50 cases). Total PCB calculations indicated that if the detection limit was substituted for non-detected analytes (ND=DL), the total PCB concentration in 4 of the 8 channels, exceeded the TEL, including one of the channels that had no detected congeners (Cutoff Angle). Total PCB calculations using ND=1/2 DL and ND=0 indicate that only Brewerton Eastern Externsion exceeded to TEL. Brewerton Eastern Extension had the greatest number of detected congeners, and the mean total PCB concentration (ND=1/2 DL and ND=DL) for Brewerton Eastern Extension exceeded the TEL value of 21.55 μ g/kg (Table 3-16).

3.3.8 PAHs

Mean concentrations of PAHs are provided in Table 3-13. In the eight channels, PAHs were detected 282 of 400 cases (71%). PAHs were detected in 55 of 64 cases (86%) Inside Site 104 and in 43 of 64 cases (67%) Outside Site 104 (Table 3-18). Of the 16 PAHs tested, acenaphthylene was the only PAH not detected in any of the channel sediments. The detection limits for acenaphthene and acenaphthylene exceeded the TEL values, but did not exceed the PEL values (Tables 3-16 and 3-17, respectively). Although flourene, naphthalene, and phenanthrene exceeded the TEL in several channels (Brewerton Eastern Extension, Craighill Angle, and Tolchester), none of the concentrations exceeded the PEL. The highest mean total PAH concentrations were reported in sediments from Brewerton Eastern Extension, Tolchester Channel, and Inside Site 104, but all total PAH concentrations were substantially lower than the TEL. The total PAH concentration (Inside Site 104) was approximately one-half the TEL value.

3.3.9 Metals

Mean concentrations of metals are provided in Table 3-14. Fourteen of the 15 tested metals were detected in the channel sediments. Cadmium was the only tested metals that was not detected in the channel sediments. In the eight tested channels, metals were detected in 337 of 400 cases (84%) (Table 3-18). Metals were detected in 57 of 60 cases (95%) Inside Site 104 and in 56 of 60 cases (93%) Outside Site 104 (Table 3-18). Seven metals that were detected in the channels and Outside Site 104 exceeded TEL values, and eight metals that were detected Inside Site 104 exceeded TEL values, and eight metals that were detected Inside Site 104 exceeded TEL values, and nickel exceeded the TEL in all eight of the channels, and Inside Site 104 and Outside Site 104. Arsenic, lead, mercury, and zinc exceeded the TEL in seven of the eight channel sediments (not exceeded in Craighill Channel) and Inside and Outside Site 104. Mean cadmium concentrations Inside and Outside Site 104 exceeded the TEL, but cadmium was not detected in the channels. Chromium exceeded the TEL only Inside Site 104.

Only two metals (nickel and zinc) exceeded PEL values in the channel sediments (Table 3-17). Mean concentrations of six metals (arsenic, cadmium, copper, lead, mercury, and zinc) exceeded the PEL Inside Site 104 (Table 3-16). Cadmium was the only metal that exceeded the PEL in the Outside Site 104 sediment.

3.3.10 Butyltins

Results of butyltin analyses are provided in Table 3-15. Tributyltin (TBT) was detected in six of the seven tested channel sediments. Butyltins were not tested in the Brewerton Eastern Extension or Inside and Outside Site 104. Mean concentrations in the channels ranged from $3.12 \mu g/kg$ (Cutoff Angle) to $94.27 \mu g/kg$ (Tolchester Channel). TBT is a common component of anti-fouling paint. There are no TEL/PEL values for TBT.

3.3.11 Summary of Bulk Sediment Results

Overall, only 61 of 189 tested organic and inorganic constituents (32%) (VOCs, SVOCs, PAHs, chlorinated and organophosphorus pesticides, PCB aroclors and congeners, metals, and butyltins) were detected at least once in sediment from the approach channels.

Forty-three of the 186 tested organic constituents (23%) were detected at least once in sediments from the vicinity of Site 104.

For the approach channels, 81.6% of the total number of tests conducted did not yield detectable concentrations of organic or inorganic constituents (Table 3-18). Similarly, 80.3% and 83% of the total number of tests conducted for Inside Side 104 and Outside Site 104, respectively, yielded detectable concentrations of organic or inorganic constituents (Table 3-18). Metals and PAHs were the frequently detected constituents in the sediments from both the approach channels and the vicinity of Site 104.

Evaluation of the 1997 and 1998 sediment data set indicated that at least one analyte exceeded TEL and PEL values in every tested channel reach. The number of exceedances for TELs and PELs are summarized by reach in Tables 3-18 and 3-19, respectively. Sediments from Inside Site 104 exhibited more TEL and PEL exceedances than Outside Site 104 and the channel reaches. Inside Site 104 sediments had nearly twice as many PEL exceedances as any of the channel reaches. Overall, arsenic, copper, nickel, and zinc exceeded the TELs in every channel reach and both Inside and Outside Site 104. Of all of the channel reaches, Brewerton Eastern Extension exhibited the highest number of TEL and PEL exceedances, followed by Tolchester Channel and Craighill Entrance.

Of the 189 tested organic and inorganic constituents, a total of 15 constituents (8% of total tested) exceeded TELs in the channel sediments (7 metals, 3 PAHs, 2 pesticides, total PCBs, and 2 SVOCs). Of the 189 organic and inorganic constituents tested, only 2 constituents (1% of total tested), nickel and zinc, exceeded the PEL value in the channel sediments. The detection limit for one PAH (acenaphthene) exceeded the TEL, so is it not known whether concentrations of this constituent in the channel sediments exceeded the TEL. The detection limit for one pesticide (chlordane) exceeded the TEL and PEL, so it is not known whether concentrations of this constituent in the channel sediments exceeded the TEL or PEL.

Although toxicity cannot be predicted by the number of TEL or PEL exceedances within a given channel reach, none of the tested sediments may be ruled out as non-toxic without additional evaluation (O'Connor et al. 1998).

3.4 ELUTRIATE RESULTS

Results of the 1998 elutriate analyses are presented in the following subsections. A total of six elutriates were tested for the approach channels. No elutriates were prepared for the Inside and Outside Site 104 sediments. Definitions of organic and inorganic data qualifiers are presented in Tables 3-3 and 3-4, respectively. Station locations, analytical methodologies, and elutriate preparation procedures for the 1998 sampling effort are provided in EA (2000d). Analytical results for elutriate analyses are provided in Tables 3-21 through 3-28. Frequency of detection for each analytical fraction is provided in Tables 3-30A (acute WQC), 3-230A (chronic WQC), and 3-31A (human health). Mixing factors that would be required to comply with WQC are provided in Tables 3-30B (acute WQC), 3-31B (chronic WQC), and 3-32B (human health).

3.4.1 Inorganic Non-Metals/Nutrients

Results of inorganic analyses are provided in Table 3-21. Overall, the elutriate from the Craighill Angle had the highest ammonia-nitrogen, TKN, and total phosphorus (TP) concentrations. Ammonia-nitrogen concentrations in the elutriates ranged from 1.7 mg/L (Craighill Entrance/Craighill) to 8.8 mg/L (Craighill Angle). None of the ammonia-nitrogen concentrations exceeded the USEPA acute saltwater criterion of 43 mg/L (Table 3-30A). Ammonia-nitrogen in the Craighill Angle and Swan Point elutriates exceeded the chronic saltwater criterion of 6.4 mg/L (based on salinity=10 ppt, water temperature =10°C, and pH=7.4)



(Table 3-31A). However, mixing factor of only one-fold of the 100% elutriate would be required to achieve compliance with the saltwater chronic criterion (which is based on a 4-day average exposure concentration).

Detected concentrations of TKN ranged from 0.26 mg/L (Brewerton Eastern Extension) to 7.3 mg/L (Craighill Angle). TP ranged from 0.09 mg/L (Brewerton Eastern Extension) to 0.38 mgL (Craighill Angle). The results for the TKN analyses are below the reported ammonia values in every reach except Craighill Entrance/Craighill. The TKN results may be biased low due to chloride interferences or interference from inorganic salts in the elutriates. Cyanide and total sulfide were not detected in any of the channel elutriates.

3.4.2 Volatile Organic Compounds

Results of Volatile Organic Analyses (VOAs) are provided in Table 3-22. VOCs were detected in 4 of 210 cases (2%) in the channel elutriates. Of the 35 tested VOCs, dichloromethane (methylene chloride) was the only analyte detected in the elutriates. Methylene chloride is a common laboratory contaminant, and the detected concentrations did not exceed the human health criterion (Table 3-32A).

3.4.3 Semivolatile Organic Compounds

Results of Semivolatile Organic Analyses (SVOAs) are provided in Table 3-23. None of the 54 tested SVOCs was detected in the channel elutriates (0 of 324 cases).

3.4.4 Chlorinated Pesticides

Results of chlorinated pesticides and PCB aroclors are provided in Table 3-24. Eight of the 22 tested chlorinated pesticides were detected in the channel elutriates (alpha-BHC, beta-BHC, chlorbenside, delta-BHC, endosulfan I, gamma-BHC, heptachlor, heptachlor epoxide). Overall, chlorinated pesticides were detected in 30 of 132 cases (23%). Beta-BHC, gamma-BHC, heptachlor and heptachlor epoxide were detected in method blanks run with the samples; therefore the reported concentrations were flagged with a "B" indicating that the concentrations may have originated from laboratory contamination. Detected concentrations of alpha-BHC, beta-BHC, gamma-BHC, and endosulfan I in the 100% elutriate did not exceed applicable water quality criteria (Tables 3-30A, 3-31A, and 3-32A).

Both heptachlor and heptachlor epoxide in the 100% elutriate exceeded the saltwater chronic criteria for aquatic life and the human health criteria for consumption of aquatic organisms (Tables 3-31A and 3-32A). These constituents were detected in the laboratory method blank, and therefore, the concentrations do not accurately represent release of these constituents in the field during open-water placement. If the detected concentrations were representative of in-situ release, a maximum mixing factor of 10 for heptachlor and 7 for heptachlor epoxide would be required to achieve compliance with the most stringent criteria.

3.4.5 Organophosphorus Pesticides

Results for organophosphorus pesticides are provided in Table 3-25. None of the tested organophosphorus pesticides was detected in the channel elutriates (0 of 30 cases).

3.4.6 PCB Aroclors and Congeners

Results for PCB aroclors are provided in Table 3-26. None of the seven tested PCB aroclors was detected in the elutriate samples (0 of 42 cases).

PCB congeners were not tested in the 1998 channel elutriate samples.

3.4.7 Polynuclear Aromatic Hydrocarbons

Results for PAHs are provided in Table 3-27. None of the 15 tested PAH compounds was detected in the channel elutriates (0 of 90 cases).

3.4.8 Metals

Results of the metals analyses are provided in Table 3-28. Ten of the 16 tested metals were detected in the channel elutriates. Metals were detected in the channel elutriates in 43 of 96 cases (45%). Aluminum, antimony, arsenic, and manganese were detected in each of the channel elutriates. Selenium was detected in 5 of the 6 channel elutriates, and chromium and silver were detected in 4 of the 6 channel elutriates. Neither antimony, chromium, nor selenium concentrations in the 100% elutriate exceeded applicable water quality criteria (Tables 3-30A, 3-31A, and 3-32A). Silver concentrations in the 100% elutriate from Craighill Upper Range/Craighill and Tolchester were slightly higher than the acute saltwater criteria. Mercury and nickel concentrations in the 100% elutriate from Craighill Entrance/Craighill and Craighill Angle, respectively, exceeded the chronic saltwater criteria. Compliance with the acute and chronic criteria would require a mixing factor of 2 (Tables 3-30B and 3-31B).

Manganese and arsenic exceeded the human health criteria for consumption of aquatic organisms in 100% elutriates from all of the channel reaches. A maximum mixing factor of 112 (Swan Point) would be required to achieve compliance with human health criterion for consumption of aquatic organisms (3-32B).

3.4.9 Summary of Elutriate Results

Only 77 of 924 total analyses (8%) conducted on the channel elutriate samples contained measurable concentrations of organic or inorganic constituents (VOCs, SVOCs, chlorinated and organophosphorus pesticides, PCB aroclors, PAHs, and metals) (Table 3-27). The majority of detected constituents were metals and chlorinated pesticides.

Evaluation of the 1998 elutriate data set indicates that at least several chemical constituents detected in the 100% elutriate from each channel reach exceed applicable WQC (Table 3-33). Overall, the human health criteria for consumption of aquatic organisms were the criteria that was most frequently exceeded. The 100% elutriate from Craighill Angle exhibited the greatest number of WQC exceedances (5). Manganese, a naturally occurring trace metal was the constituent that required the highest mixing factor in 5 of the 6 tested elutriates (Brewerton Eastern Extension, Craighill Angle, Craighill Upper Range/Cutoff Angle, Swan Point, and Tolchester). Mercury required the highest mixing factor in Craighill Entrance/Craighill Upper Range.

Of the 152 organic and inorganic constituents tested, only 8 constituents (5%) in the 100% elutriate exceeded WQC. Only 1 constituent (silver) exceeded acute WQC; 5 constituents (ammonia-nitrogen, heptachlor, heptachlor epoxide, mercury, and nickel) exceeded chronic WQC; and 6 constituents (heptachlor, heptachlor epoxide, arsenic, manganese, mercury, and nickel) exceeded human health criteria for the consumption of aquatic organisms. Both heptachlor and heptachlor epoxide were detected in the laboratory method blank; thus, the detected concentrations do not accurately represent release of these constituents from sediments to the water column during open-water placement.

Importantly, the acute, chronic, and human health criteria are based on the following assumptions:

- acute criteria: 1-hr average exposure concentrations;
- chronic criteria: 4-day average exposure concentrations; and
- human health criteria: daily lifetime (70 years) consumption of aquatic organisms (10⁻⁵ risk-level).

The maximum mixing factors required to achieve the acute, chronic, and human health criteria would be 2 (silver), 10 (heptachlor) (although detected in laboratory blank, and 112 (manganese), respectively.

Overall, during open-water placement, the majority of constituents would be expected to meet WQC at the point of release. The few constituents that would not meet compliance at the point of release would be expected to reach the WQC quickly in relation to the times implicit in the WQC (i.e., 1-hr average, 4-day average, daily lifetime consumption of aquatic organisms). Based on the assumptions associated with the WQC, the nature of the detected constituents, and the concentrations of the detected constituents, none of the constituents detected in the elutriates would be expected to cause "significant degradation" [33 CFR 230.1(c)] in the aquatic environment.

3.5 TIER I EVALUATION

According to the ITM (USEPA/USACE 1998), after consideration of all available Tier I information, one of the following conclusions is reached.

- 1. Existing information does not provide a sufficient basis for making factual determinations. In this case, further evaluation in higher tiers is appropriate.
- 2. Existing information provides a sufficient basis for making factual determinations. In this case, one of the following decisions is reached:
 - (a) The material meets the criteria for exclusion from testing.
 - (b) The material does not meet the criteria for exclusion from testing, but information concerning the potential impact of the material is sufficient to make factual determinations.

Review and evaluation of the 1997 and 1998 data sets indicates that the channel sediments do not qualify for exclusion from additional testing. None of the sediments is exclusively comprised of sand, and several chemical constituents were detected above TEL or PEL values in each tested channel. Although toxicity cannot be predicted from TEL and PEL comparisons, none of the sediments can be ruled out as non-toxic without additional information.

Detected constituents in the 100 percent elutriate samples exceeded one or more applicable water quality criteria for each of the tested channel reaches. Although the majority of detected constituents meet WQC at the point of release, water quality modeling in subsequent tiers will determine the expected water column mixing at the placement site and will determine the distance and time required to achieve applicable water quality criteria.

Evaluation in higher tiers is appropriate because the existing information does not provide a sufficient basis for making factual determinations regarding potential water column and benthic impacts.

TABLE 3-1 MARINE SEDIMENT QUALITY GUIDELINES (SQGs)

| | | Threshold | Probable |
|--------------------------------|-------|---------------|---------------|
| | | Effects Level | Effects Level |
| Chemical Name | Units | (TEL) | (PEL) |
| METALS | | | |
| ARSENIC | MG/KG | 7.24 | 41.6 |
| CADMIUM | MG/KG | 0.676 | 4.21 |
| CHROMIUM | MG/KG | 52.3 | 160.4 |
| COPPER | MG/KG | 18.7 | 108.2 |
| LEAD | MG/KG | 30.24 | 112.18 |
| MERCURY | MG/KG | 0.13 | 0.696 |
| NICKEL | MG/KG | 15.9 | 42.8 |
| SILVER | MG/KG | 0.73 | 1.77 |
| ZINC | MG/KG | 124 | 271 |
| PAHs | | | |
| ACENAPHTHENE | UG/KG | 6.71 | 88.9 |
| ACENAPHTHYLENE | UG/KG | 5.87 | 127.87 |
| ANTHRACENE | UG/KG | 46.85 | 245 |
| BENZO(A)PYRENE | UG/KG | 88.81 | 763.22 |
| BENZO[A]ANTHRACENE | UG/KG | 74.83 | 692.53 |
| CHRYSENE | UG/KG | 107.77 | 845.98 |
| DIBENZ(A,H)ANTHRACENE | UG/KG | 6.22 | 134.61 |
| FLUORANTHENE | UG/KG | 112.82 | 1493.54 |
| FLUORENE | UG/KG | 21.17 | 144.35 |
| NAPHTHALENE | UG/KG | 34.57 | 390.64 |
| PHENANTHRENE | UG/KG | 86.68 | 543.53 |
| PYRENE | UG/KG | 152.66 | 1397.6 |
| PAHs, TOTAL | UG/KG | 1684.06 | 16770.4 |
| PESTICIDES | h | | |
| 4,4'-DDT | UG/KG | 1.19 | 4.77 |
| CHLORDANE | UG/KG | 2.26 | 4.79 |
| DDD | UG/KG | 1.22 | 7.81 |
| DDE | UG/KG | 2.07 | 374.17 |
| DIELDRIN | UG/KG | 0.715 | 4.3 |
| GAMMA-BHC | UG/KG | 0.32 | 0.99 |
| PCBs | | | <u> </u> |
| PCBs,TOTAL | UG/KG | 21.55 | 188.79 |
| SEMIVOLATILE ORGANIC COMPOUNDS | | | |
| 2-METHYLNAPHTHALENE | UG/KG | 20.21 | 201.28 |
| BIS(2-ETHYLHEXYL)PHTHALATE | UG/KG | 182.16 | 2646.51 |

Source: Buchman 1999

| ſ | | SALTWATER CRITERIA | | | | | |
|-----------------------------|----------|--------------------|----------------------|---------------------------|--|--|--|
| ANALYTE | UNITS | ACUTE * | CHRONIC ^b | HUMAN HEALTH [•] | | | |
| NON-METALS | | | | | | | |
| CYANIDE | UG/L | 1 ^d | lq | | | | |
| AMMONIA-NITROGEN | MG N/ | 43 ^s | 6.4 ^s | | | | |
| SULFIDE. TOTAL | UG/L | | 2 ^v | | | | |
| VOCs | | | | | | | |
| DICHLOROMETHANE | UG/L | | | 16000 ^y | | | |
| SVOCs | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | | | 59 ^y | | | |
| PESTICIDES | | | | | | | |
| ALPHA-BHC | UG/L | | | 0.13 ^y | | | |
| BETA-BHC | UG/L | | | 0.46 ^y | | | |
| ENDOSULFAN I | UG/L | 0.034 ⁿ | 0.0087 ° | 240 | | | |
| GAMMA-BHC | UG/L | 0.16 | | 0.63 ^y | | | |
| HEPTACHLOR | UG/L | 0.053 | 0.0036 | 0.0021 ^y | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | 0.0036 | 0.0011 ^y | | | |
| PCBs | · | · · · · · · | | | | | |
| TOTAL PCBs | UG/L | | 0.03 ^e | 0.0017 ^{fy} | | | |
| PAHs | <u> </u> | | | | | | |
| FLUORENE | UG/L | | | 14,000 | | | |
| PYRENE | UG/L | | | 11000 | | | |
| METALS | | | | | | | |
| ANTIMONY | UG/L | | | 4300 | | | |
| ARSENIC | UG/L | <u>69 °</u> | 36 * | 0.14 * | | | |
| BERYLLIUM | UG/L | | | 0.117' | | | |
| CHROMIUM | UG/L | 1100 * | 50 * | | | | |
| COPPER | UG/L | 4.8 | 3.1 ' | | | | |
| LEAD | UG/L | 210° | 8.1 ° | | | | |
| MANGANESE | UG/L | | | 100 | | | |
| MERCURY | UG/L | 1.8 ^p | 0.94 ^p | 0.051 | | | |
| NICKEL | UG/L | 74 ^r | 8.2 ^r | 4600 | | | |
| SELENIUM | UG/L | 290 ¹ | 71 ^v | | | | |
| SILVER | UG/L | 1.9 ^u | | | | | |
| ZINC | UG/L | 90 * | 81 ** | | | | |
| ORGANOTINS | | | | | | | |
| TRIBUTYLTIN | UG/L | 0.37 ^q | 0.01 ^q | | | | |

TABLE 3-2 USEPA/ (PROPOSED) MDE WATER QUALITY CRITERIA*

* = Applicable to detected analytes in 1998 elutriate samples.

Source: USEPA 1998 [63 Federal Register 68354-68364] and Maryland Register 2000 [27(17):1628-1636].

Superscripts:

.

a = acute aquatic life criteria based on 1-hr average exposure concentrations.

b = chronic aquatic life criteria based on 4-day average exposure concentrations.

c = human health criteria based on daily lifetime (70-ycar) average consumption of aquatic organisms; criteria based on 10-5 risk for carcinogens.

d = ug free cyanide as CN/L.

c = applies to aroclors 1242, 1254, 1221, 1232, 1248, 1260, and 1016.

f = applies to Total PCBs (sum of all congeners or isomer analyses).

g = total dissolved arsenic.

h = inorganic arsenic only.

j = from EPA 1986 Gold Book; no EPA 1998 number.

k = dissolved chromium; hexavalent.

l = dissolved copper.

n = most appropriately applied to sum of alpha (1) and beta (11) endosulfan.

o = dissolvcd lcad.

p = dissolved total mcrcury.

q = proposed criteria.

r = dissolvcd nickel.

s = total ammonia as nitrogen; criterion assumes cold weather conditions: salinity = 10 ppt, water temperature = 10 C, and pH=7.4

t = dissolvcd selenium.

u = dissolved silver.

v = undissociated hydrogen sulfide (H2S).

w = dissolved zine.

y = carcinogen

TABLE 3-3ORGANIC DATA QUALIFIERS

Qualifiers other than those listed below may be required to properly define the results. If used, they are given an alphabetic designation not already specified in this table or in a project/program document such as a Quality Assurance Project Plan or a contract Statement of Work. Each additional qualifier is fully described in the Analytical Narrative section of the laboratory report.

- U Indicates a target compound was analyzed for but not detected. The sample Reporting Limit (RL) is corrected for dilution and, if a soil sample, for percent moisture, if reported on a dry weight basis.
- J Indicates an estimated value. This qualifier is used under the following circumstances:
 - 1) when estimating a concentration for tentatively identified compounds (TICs) in GC/MS analyses, where a 1:1 response is assumed,
 - 2) when the mass spectral and retention time data indicate the presence of a compound that meets the volatile and semivolatile GC/MS identification criteria, and the result is less than the RL but greater than the method detection limit (MDL).
- **B** This qualifier is used when the analyte is found in the associated method blank as well as in the sample. It indicates possible/probable blank contamination and warns the data user to take appropriate action. For GC/MS analyses, this qualifier is used for a TIC, as well as, for a positively identified target compound.
- **E** This qualifier identifies compounds whose concentrations exceed the calibration range of the instrument for that specific analysis.
- **D** When applied, this qualifier identifies all compound concentrations reported from a secondary dilution analysis.
- A This qualifier indicates that a TIC is a suspected aldol-condensation product.
- N Indicates presumptive evidence of a compound. This qualifier is only used for GC/MS TICs, where the identification is based on a mass spectral library search. For generic characterization of a TIC, such as chlorinated hydrocarbon, the N qualifier is not used.
- **P** When applied, this qualifier indicates a reported value from a GC analysis when there is greater than 25% difference for detected concentrations between the two GC columns.

C (Concentration) qualifiers:

- **B** Reported value is less than the project-specified Reporting Limit (RL), but greater than the method-specified Instrument Detection Limit (IDL) or Method Detection Limit (MDL).
- U Analyte analyzed for but not detected (concentration is less than the methodspecified Instrument Detection Limit (IDL) or Method Detection Limit (MDL).

Q (Quality control) qualifiers:

- **E** Reported value is estimated because of presence of interference.
- M Duplicate injection precision not met.
- N Spiked sample recovery is not within control limits.
- S Reported value is determined by the method of standard additions (MSA).
- W Postdigestion spike for furnace Atomic Absorption Spectrophotometric (AAS) AAS analysis is out of control limits (85-115%) and sample absorbance is less than 50% of spike absorbance.
- * Duplicate analyses is not within control limits.
- + Correlation coefficient for MSA is less than 0.995.

M (Method) qualifiers:

- P Inductively Coupled Plasma (ICP)
- A Flame AAS
- F Furnace AAS
- CV Cold Vapor AAS
- AV Automated Cold Vapor AAS
- AS Semiautomated Spectrophotometric
- C Manual Spectrophotometric
- T Titrimetric
- NR Analyte is not required to be determined.



Figure 4-14. Grab sampling locations at the Norfolk Ocean Disposal Reference Area and lower Chesapeake Bay control site





| | | | Total # | | | | | | Processing / |
|-----------------------|-------------|--------------------|-----------|----------|------------|------------|-------------|----------|--------------|
| | | | Cores | | Collection | Collection | Penetration | Recovery | Compositing |
| Sampling Reach | Station ID | Sediment Sample ID | Collected | Core No. | Date | Time (EDT) | Depth (ft) | (ft) | Date |
| C&D Approach Channels | CD-001VC | CD-001VCSED | 10 | 1 | 09/21/1999 | 1219 | 6 | 3.8 | 09/30/1999 |
| | | | | 2 | 09/21/1999 | 1227 | 6 | 4.4 | |
| | | | | 3 | 09/21/1999 | 1235 | 6 | 3* | |
| | | | | 4 | 09/21/1999 | 1244 | 6 | 5.25 | |
| | | | | 5 | 09/21/1999 | 1252 | 6 | 3.4 | |
| | | | | 6 | 09/21/1999 | 1300 | 6 | 5.25 | |
| | | | | 7 | 09/21/1999 | 1307 | 6 | 4.1 | |
| | | | | 8 | 09/21/1999 | 1315 | 6 | 3.8 | |
| | | | | 9 | 09/21/1999 | 1324 | 6 | 3 | |
| | | | | 10 | 09/21/1999 | 1333 | 6 | 4.5 | |
| | CD-002VC | CD-002VCSED | 9 | 1 | 09/21/1999 | 1104 | 6 | 5.2 | 09/30/1999 |
| | | | | 2 | 09/21/1999 | 1111 | 6 | 3.6 | |
| | | 1 | | 3 | 09/21/1999 | 1119 | 6 | 5.2 | |
| | | | | 4 | 09/21/1999 | 1126 | 6 | 4.1 | |
| | | | | 5 | 09/21/1999 | 1134 | 6 | 4.3 | |
| 1 | | | | 6 | 09/21/1999 | 1143 | 6 | 3.4 | |
| | | | | 7 | 09/21/1999 | 1153 | 6 | 5.2 | |
| | | | | 8 | 09/21/1999 | 1200 | 6 | 4.6 | |
| | | | | 9 | 09/21/1999 | 1207 | 6 | 5.3 | |
| Craighill Angle-East | CRA-E-001VC | CRA-E-001VCSED | 5 | 1 | 09/18/1999 | 1130 | 6 | 3.25 | 10/01/1999 |
| | | | | 2 | 09/18/1999 | 1141 | 6 | 3.6 | |
| | | | | 3 | 09/18/1999 | 1150 | 6 | 2.6 | |
| | | | | 4 | 09/18/1999 | 1159 | 6 | 3.5 | |
| | | | | 5 | 09/18/1999 | 1210 | 6 | 2.5 | |
| | CRA-E-002VC | CRA-E-002VCSED | 9 | 1 | 09/19/1999 | 1316 | 6 | 3 | 10/01/1999 |
| | | | | 2 | 09/19/1999 | 1320 | 6 | 3 | |
| | | | 1 | 3 | 09/19/1999 | 1340 | 6 | 3.5 | |
| | | 1 | | 4 | 09/19/1999 | 1346 | 6 | 3 | |
| | | | | 5 | 09/19/1999 | 1401 | 6 | 2 | |
| | 1 | | | 6 | 09/19/1999 | 1412 | 6 | 2.5 | |
| | | | | 7 | 09/19/1999 | 1425 | 6 | 2.25 | |
| | | | | 8 | 09/19/1999 | 1433 | 6 | 2.0 | |
| | | | | 9 | 09/19/1999 | 1450 | 6 | 1.75* | |
| | CRA-E-003VC | CRA-E-003VCSED | 6 | 1 | 09/15/1999 | 1120 | 10 | 3.25 | 10/01/1999 |
| | | | | 2 | 09/15/1999 | 1155 | 10 | 5 |] |
| | | | | 3 | 09/15/1999 | 1220 | 10 | 5 | |
| | | | | 4 | 09/15/1999 | 1240 | 10 | 5 | |
| | | | | 5 | 09/18/1999 | 1010 | 10 | 5 | |
| | | | | | 09/18/1999 | 1030 | 10 | 5 | L |
| | | | | | | | | | |

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TABLE 4-4A SUMMARY OF VIBRACORING ACTIVITIES (SEPTEMBER 1999)





TABLE 4-4A (CONTINUED)

| | | | Total # | | | | | | Processing / |
|----------------------|-------------|--------------------|-----------|----------|------------|------------|-------------|----------|--------------|
| | | - | Cores | | Collection | Collection | Penetration | Recovery | Compositing |
| Sampling Reach | Station ID | Sediment Sample ID | Collected | Core No. | Date | Time (EDT) | Depth (ft) | (ft) | Date |
| Craighill Angle-West | CRA-W-001VC | CRA-W-001VCSED | 7 | 1 | 09/18/1999 | 1520 | 6 | 2.6 | 09/30/1999 |
| | | | | 2 | 09/18/1999 | 1532 | 6 | 2.7 | |
| | | | | 3 | 09/18/1999 | 1545 | 6 | 2.4 | |
| | | | | 4 | 09/18/1999 | 1556 | 6 | 2.8 | |
| | | | | 5 | 09/18/1999 | 1603 | 6 | 1.9 | |
| | | | | 6 | 09/18/1999 | 1612 | 6 | 2.6* | |
| | | | | 7 | 09/18/1999 | 1623 | 6 | 2 | |
| | CRA-W-002VC | CRA-W-002VCSED | 7 | 1 | 09/18/1999 | 1640 | 6 | 2.4 | 09/30/1999 |
| | | | | 2 | 09/18/1999 | 1648 | 6 | 1.8 | |
| | | | | 3 | 09/18/1999 | 1656 | 6 | 2.3 | |
| | | | | 4 | 09/18/1999 | 1706 | 6 | 1.8 | |
| | | | | 5 | 09/18/1999 | 1717 | 6 | 2.7 | |
| | | | | 6 | 09/18/1999 | 1726 | 6 | 2 | |
| | | | | 7 | 09/18/1999 | 1747 | 6 | 1.9 | |
| | CRA-W-003VC | CRA-W-003VCSED | 5 | 1 | 09/19/1999 | 840 | 6 | 3 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 845 | 6 | 2.7 | |
| | | | | 3 | 09/19/1999 | 852 | 6 | 3.5 | |
| | | | | 4 | 09/19/1999 | 902 | 6 | 3.7 | |
| | | | | 5 | 09/19/1999 | 911 | 6 | 3.4 | |
| Craighill Entrance | CRE-001VC | CRE-001VCSED | 5 | 1 | 09/19/1999 | 955 | 6 | 2.7 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 1003 | 6 | 3.2 | |
| | | | | 3 | 09/19/1999 | 1014 | 6 | 3.4 | |
| | | | | 4 | 09/19/1999 | 1025 | 6 | 2.8 | |
| | | | | 5 | 09/19/1999 | 1036 | 6 | 1.3 | |
| | CRE-002VC | CRE-002VCSED | 6 | 1 | 09/19/1999 | 1053 | 6 | 3.3 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 1059 | 6 | 2.4 | |
| | | | | 3 | 09/19/1999 | 1109 | 6 | 3.7 | |
| | | | | 4 | 09/19/1999 | 1116 | 6 | 3.8 | |
| | | | | 5 | 09/19/1999 | 1125 | 6 | 3.5* | |
| | | | | 6 | 09/19/1999 | 1134 | 6 | 3.1 | |
| | CRE-003VC | CRE-003VCSED | 5 | 1 | 09/19/1999 | 1337 | 6 | 3.8 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 1342 | 6 | 5 | |
| 1 | | | | 3 | 09/19/1999 | 1350 | 6 | 4.8 | |
| | | | | 4 | 09/19/1999 | 1358 | 6 | 4.8 | |
| | | | | 5 | 09/19/1999 | 1406 | 6 | 5.2 | |
| | CRE-004VC | CRE-004VCSED | 5 | 1 | 09/19/1999 | 1157 | 6 | 2.5 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 1215 | 6 | 3 | |
| | | | | 3 | 09/19/1999 | 1222 | 6 | 2.8 | |
| | | | | 4 | 09/19/1999 | 1229 | 6 | 4.2 | |
| | | | | 5 | 09/19/1999 | 1236 | 6 | 3.8 | |

TABLE 4-4A (CONTINUED)

| Sampling Reach | Station ID | Sediment Sample ID | Total # Cores Collected | Core No. | Collection Date | Collection Time (EDT) | Penetration Depth (ft) | Recovery (ft) | Processing / Compositing Date | | | | | | | | | | | | | | | | |
|----------------|------------|--------------------|-------------------------------|----------|--------------------|--------------------------|---------------------------|------------------|-------------------------------------|--|--|--|--|--|--|--|--|--|--|---|------------|------|---|-----|---|
| Swan Point | SWP-001VC | SWP-001VCSED | 4 | 1 | 09/19/1999 | 1605 | 6 | 5.1 | 09/30/1999 | | | | | | | | | | | | | | | | |
| | | | | 2 | 09/19/1999 | 1612 | 6 | 3.7 | | | | | | | | | | | | | | | | | |
| | | | | 3 | 09/19/1999 | 1623 | 6 | 4.2 | | | | | | | | | | | | | | | | | |
| | | | | 4 | 09/19/1999 | 1630 | 6 | 4.0 | | | | | | | | | | | | | | | | | |
| | SWP-002VC | SWP-002VCSED | 4 | 1 | 09/19/1999 | 1525 | 6 | 3.5 | 09/30/1999 | | | | | | | | | | | | | | | | |
| 1 | | | | 2 | 09/19/1999 | 1533 | 6 | 3.7 | | | | | | | | | | | | | | | | | |
| | | | | 3 | 09/19/1999 | 1544 | 6 | 5.2 | | | | | | | | | | | | | | | | | |
| | | | | 4 | 09/19/1999 | 1553 | 6 | 4.5* | | | | | | | | | | | | | | | | | |
| | SWP-003VC | SWP-003VCSED | 4 | 1 | 09/19/1999 | 1446 | 6 | 4.6 | 09/30/1999 | | | | | | | | | | | | | | | | |
| | | | | 2 | 09/19/1999 | 1456 | 6 | 3.3 | | | | | | | | | | | | | | | | | |
| | | | | 3 | 09/19/1999 | 1505 | 6 | 5.3 | | | | | | | | | | | | | | | | | |
| | | | | 4 | 09/19/1999 | 1512 | 6 | 4.5 | ļ | | | | | | | | | | | | | | | | |
| | SWP-004VC | SWP-004VCSED | 5 | 1 | 09/23/1999 | 1529 | 6 | 3.2 | 09/30/1999 | | | | | | | | | | | | | | | | |
| | | | | 2 | 09/23/1999 | 1538 | 6 | 2.8 | | | | | | | | | | | | | | | | | |
| | | | | 3 | 09/23/1999 | 1549 | 6 | 4.0 | | | | | | | | | | | | | | | | | |
| | | | | 4 | 09/23/1999 | 1601 | 6 | 3.0 | | | | | | | | | | | | | | | | | |
| | | | | 5 | 09/23/1999 | 1611 | 6 | 3.8 | | | | | | | | | | | | | | | | | |
| | SWP-005VC | SWP-005VCSED | 5 | 1 | 09/23/1999 | 1411 | 6 | 3.1 | 09/30/1999 | | | | | | | | | | | | | | | | |
| | | | | 2 | 09/23/1999 | 1435 | 6 | 3.9 | 1 | | | | | | | | | | | | | | | | |
| | | | 1 | 3 | 09/23/1999 | 1447 | 6 | 3.8 | 1 | | | | | | | | | | | | | | | | |
| | | | | 4 | 09/23/1999 | 1458 | 6 | 2.5 |] | | | | | | | | | | | | | | | | |
| | SWP-006VC | SWP-006VCSED | 5 | 1 | 09/23/1999 | 1308 | 6 | 4.7 | 09/30/1999 | | | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | | | | | | | 2 | 09/23/1999 | 1319 | 6 | 3.2 | |
| | | | | | | | | | | | | | | | | | | | | | | | | | 3 |
| | | | | 4 | 09/23/1999 | 1341 | 6 | 3.4 | 1 | | | | | | | | | | | | | | | | |
| | | | | 5 | 09/23/1999 | 1354 | 6 | 3.8 | 1 | | | | | | | | | | | | | | | | |

Page 3 of 4



TABLE 4-4A (CONTINUED)

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| | | | Total # | | | | - | | Processing / |
|----------------------------|-------------|--------------------|--------------------|----------|-----------------|------------|---------------------------|------------------|---------------------|
| Sampling Reach | Station ID | Sediment Sample ID | Cores Collected | Core No. | Date Collection | Collection | Penetration Depth (ft) | Recovery (ft) | Compositing Date |
| Tolchester Channel - North | TLC-N-005VC | TLC-N-005VCSED | 2 | 1 1 | 09/20/1999 | 1022 | 6 | 5.2 | 10/01/1999 |
| | | | | 2 | 09/20/1999 | 1031 | 6 | 4.7 | 10,01,1775 |
| | TLC-N-006VC | TLC-N-006VCSED | 3 | 1 | 09/20/1999 | 1046 | 6 | 4.3 | 10/01/1999 |
| | | | | 2 | 09/20/1999 | 1054 | 6 | 4.2 | |
| | | | | 3 | 09/20/1999 | 1107 | 6 | 5.2 | |
| | TLC-N-007VC | TLC-N-007VCSED | 4 | 1 | 09/20/1999 | 1129 | 6 | 4.2 | 10/01/1999 |
|] | | | | 2 | 09/20/1999 | 1138 | 6 | 3.5 | |
| • | | | | 3 | 09/20/1999 | 1149 | 6 | 3.3* | |
| t | | | | 4 | 09/20/1999 | 1158 | 6 | 5.2 | |
| | TLC-N-008VC | TLC-N-008VCSED | 2 | 1 | 09/20/1999 | 1307 | 6 | 4.6 | 10/01/1999 |
| | | | | 2 | 09/20/1999 | 1316 | 6 | 5.1 | |
| | TLC-N-009VC | TLC-N-009VCSED | 3 | 1 | 09/20/1999 | 1329 | 6 | 3.2 | 10/01/1999 |
| | | | | 2 | 09/20/1999 | 1338 | 6 | 5.5 | |
| | | | | 3 | 09/20/1999 | 1347 | 6 | 4.9 | |
| | TLC-N-010VC | TLC-N-010VCSED | 3 | 1 | 09/20/1999 | 1404 | 6 | 3.1 | 10/01/1999 |
| | | | | 2 | 09/20/1999 | 1412 | 6 | 5.0 | |
| | | | | 3 | 09/20/1999 | 1424 | 6 | 5.1 | |
| Tolchester Channel - South | TLC-S-001VC | TLC-S-001VCSED | 3 | 1 | 09/19/1999 | 1708 | 6 | 4.8 | 09/30/1999 |
| | | | | 2 | 09/19/1999 | 1717 | 6 | 5.3 | |
| | | | | 3 | 09/19/1999 | 1725 | 6 | 4.8 | |
| | TLC-S-002VC | TLC-S-002VCSED | 3 | 1 | 09/20/1999 | 855 | 10 | 6.9 | 09/30/1999 |
| | | | | 2 | 09/20/1999 | 910 | 10 | 6.1 | |
| | | | | 3 | 09/20/1999 | 928 | 6 | 2.7* | |
| | TLC-S-003VC | TLC-S-003VCSED | 3 | 1 | 09/23/1999 | 1109 | 12 | 8.1 | 09/30/1999 |
| | | | | 2 | 09/23/1999 | 1126 | 12 | 4.7 | |
| | | | | 3 | 09/23/1999 | 1138 | 12 | 2.3 | |
| | TLC-S-004VC | TLC-S-004VCSED | 2 | 1 | 09/20/1999 | 941 | 10 | 7.8 | 09/30/1999 |
| | | | | 2 | 09/20/1999 | 949 | 10 | 7.2 | |
| Tolchester Straightening | TLS-001VC | TLS-001VCSED | 4 | 1 | 09/23/1999 | 906 | 12 | 4.5* | 09/30/1999 |
| | | | | 2 | 09/23/1999 | 922 | 12 | 9.3 | |
| | | | | 3 | 09/23/1999 | 936 | 12 | 9.4 | |
| | | | | 4 | 09/23/1999 | 1022 | 12 | 8.1 | |
| | TLS-002VC | TLS-002VCSED | 3 | 1 | 09/20/1999 | 1519 | 16 | 12 | 09/30/1999 |
| | | | | 2 | 09/20/1999 | 1636 | 12 | 7.8 | |
| | | | | 3 | 09/20/1999 | 1651 | 12 | 9.5 | |

(a) = Core from individual stations composited first; sub-sample removed for analyical testing; then reach composite created for ecotoxicological testing.

* = Core submitted to WES for particle settling tests.

| | | 1 | Totai # | | | | | | Processing / |
|-----------------------------------|-------------|--------------------|-----------|----------|------------------------|------------|-------------|----------|--------------|
| | | | Cores | | | Collection | Penetration | Recovery | Compositing |
| Sampling Reach | Station ID | Sediment Sample ID | Collected | Core No. | Collection Date | TIme (EST) | Depth (ft) | (ft) | Date |
| C&D Approach Channels | CD-001VC | CD-001VCSED | 4 | 1 | 12/08/1999 | 935 | 10.5 | 5.83 | 12/17/1999 |
| · · · · · · · · · · · · · · · · · | | | | 2 | 12/08/1999 | 950 | 10 | 7.08 | 12/17/1999 |
| | | | | 3 | 12/08/1999 | 1004 | 10 | 6.5 | 12/17/1999 |
| | | | | 4 . | 12/09/1999 | 1518 | 10 | 4 | 12/17/1999 |
| | CD-002VC | CD-002VCSED | 4 | 1 | 12/08/1999 | 829 | 8 | 4.67 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 841 | 8 | 4.625 | 12/17/1999 |
| | | | | 3 | 12/08/1999 | 855 | 8 | 4.5 | 12/17/1999 |
| | | | | 4 | 12/09/1999 | 1540 | 10 | 4 | 12/17/1999 |
| Craighill Angle-Fast | CRA-F-001VC | CRA-E-001VCSED | 4 | 1 | 12/15/1999 | 1105 | 10 | 4 | 12/17/1999 |
| Claight Angle-Last | | | | 2 | 12/15/1999 | 1120 | 10 | 3.17 | 12/17/1999 |
| | | | | 3 | 12/15/1999 | 1135 | 10 | 3.5 | 12/17/1999 |
| | | | | 4 | 12/15/1999 | 1150 | 10 | 4 | 12/17/1999 |
| | CRA-F-002VC | CRA-E-002VCSED | 7 | 1 | 12/15/1999 | 918 | 10 | 4.75 | 12/17/1999 |
| | | | | 2 | 12/15/1999 | 930 | 10 | 2.08 | 12/17/1999 |
| | | | | 3 | 12/15/1999 | 950 | 10 | 4.25 | 12/17/1999 |
| | | | | 4 | 12/15/1999 | 1005 | 10 | 3.92 | 12/17/1999 |
| • | | | | 5 | 12/15/1999 | 1020 | 10 | 6.17 | 12/17/1999 |
| 1 | | | | 6 | 12/15/1999 | 1028 | 10 | 4.33 | 12/17/1999 |
| | | | | 7 | 12/15/1999 | 1045 | 10 | 2.67 | 12/17/1999 |
| | CRA-F-003VC | CRA-E-003VCSED | 7 | 1 | 12/15/1999 | 743 | 8 | 4.67 | 12/17/1999 |
| 1 | CIUTE UNITO | | | 2 | 12/15/1999 | 755 | 8 | 4.5 | 12/17/1999 |
| | | | | 3 | 12/15/1999 | 810 | 8 | 4.67 | 12/17/1999 |
| | | | 1 | 4 | 12/15/1999 | 820 | 8 | 3 | 12/17/1999 |
| | | | | 5 | 12/15/1999 | 830 | 8 | 4.5 | 12/17/1999 |
| | | | | 6 | 12/15/1999 | 845 | 8 | 4.92 | 12/17/1999 |
| | | | | 7 | 12/15/1999 | 856 | 8 | 3.58 | 12/17/1999 |
| Craighill Angle-West | CRA-W-001VC | CRA-W-001VCSED | 4 | 1 | 12/14/1999 | 755 | 8 | 3.5 | 12/17/1999 |
| Craighth Angle-West | | | | 2 | 12/14/1999 | 816 | 10 | 4.83 | 12/17/1999 |
| | | | | 3 | 12/14/1999 | 842 | 10 | 4.5 | 12/17/1999 |
| | | | | 4 | 12/14/1999 | 908 | 10 | 3.67 | 12/17/1999 |
| | CRA-W-002VC | CRA-W-002VCSED | 5 | 1 | 12/14/1999 | 945 | 10 | 4.5 | 12/17/1999 |
| 1 | | | | 2 | 12/14/1999 | 1035 | 10 | 3.83 | 12/17/1999 |
| | | | | 3 | 12/14/1999 | 1115 | 10 | 3.42 | 12/17/1999 |
| | | | | 4 | 12/14/1999 | 1125 | 10 | 3 | 12/17/1999 |
| | | | | 5 | 12/14/1999 | 1145 | 10 | 4.33 | 12/17/1999 |
| | CRA-W-003VC | CRA-W-003VCSED | 4 | | 12/14/1999 | 1250 | 10 | 4 | 12/17/1999 |
| | | | · · | 2 | 12/14/1999 | 1315 | 10 | 2.17 | 12/17/1999 |
| | | | | 3 | 12/14/1999 | 1345 | 10 | 2.5 | 12/17/1999 |
| | | | | 4 | 12/14/1999 | 1420 | 10 | 2.58 | 12/17/1999 |

TABLE 4-4B SUMMARY OF VIBRACORING ACTIVITIES (DECEMBER 1999)





TABLE 4-4B(CONTINUED)

| | | | Total # | | | | | | Processing / |
|----------------------------|-------------|--------------------|-----------|----------|------------------------|------------|-------------|----------|--------------|
| | | | Cores | | | Collection | Penetration | Recovery | Compositing |
| Sampling Reach | Station ID | Sediment Sample ID | Collected | Core No. | Collection Date | Time (EST) | Depth (ft) | (ft) | Date |
| Craighill Entrance | CRE-001VC | CRE-001VCSED | 5 | 1 | 12/15/1999 | 1334 | 10 | 3.75 | 12/17/1999 |
| | | | | 2 | 12/15/1999 | 1345 | 10 | 3.67 | 12/17/1999 |
| | | | | 3 | 12/15/1999 | 1412 | 10 | 3.5 | 12/17/1999 |
| | | | | 4 | 12/15/1999 | 1426 | 10 | 3.08 | 12/17/1999 |
| | | | | 5 | 12/15/1999 | 1440 | 10 | 3.25 | 12/17/1999 |
| | CRE-002VC | CRE-002VCSED | 3 | i | 12/15/1999 | 1245 | 7 | 3 | 12/17/1999 |
| | | | | 2 | 12/15/1999 | 1305 | 9 | 2.25 | 12/17/1999 |
| | | | | 3 | 12/15/1999 | 1315 | 10 | 3 | 12/17/1999 |
| | CRE-003VC | CRE-003VCSED | 2 | 1 | 12/14/1999 | 1623 | 10 | 2.67 | 12/17/1999 |
| | | | | 2 | 12/14/1999 | 1640 | 10 | 3.33 | 12/17/1999 |
| | CRE-004VC | CRE-004VCSED | 4 | 1 | 12/14/1999 | 1 500 | 10 | 1.92 | 12/17/1999 |
| | | | | 2 | 12/14/1999 | 1535 | 10 | 5.17 | 12/17/1999 |
| | | | | 3 | 12/14/1999 | 1550 | 10 | 5.58 | 12/17/1999 |
| | | | | 4 | 12/14/1999 | 1603 | 10 | 5.58 | 12/17/1999 |
| Swan Point | SWP-001VC | SWP-001VCSED | 2 | 1 | 12/10/1999 | 718 | 10 | 4 | 12/17/1999 |
| | | | | 2 | 12/10/1999 | 736 | 10 | 4 | 12/17/1999 |
| | SWP-002VC | SWP-002VCSED | 2 | 1 | 12/09/1999 | 1129 | 10 | 5 | 12/17/1999 |
| | | | 1 | 2 | 12/09/1999 | 1140 | 10 | 6.58 | 12/17/1999 |
| | SWP-003VC | SWP-003VCSED | 2 | i | 12/09/1999 | 1056 | 10 | 5.75 | 12/17/1999 |
| 1 | | | | 2 | 12/09/1999 | 1109 | 10 | 6.75 | 12/17/1999 |
| | SWP-004VC | SWP-004VCSED | 2 | i | 12/10/1999 | 802 | 10 | 4.83 | 12/17/1999 |
| | | | | 2 | 12/10/1999 | 815 | 10 | 4 | 12/17/1999 |
| | SWP-005VC | SWP-005VCSED | 3 | 1 | 12/10/1999 | 850 | 10 | 4.67 | 12/17/1999 |
| | | | | 2 | 12/10/1999 | 900 | 10 | 4.42 | 12/17/1999 |
| | | | | 3 | 12/10/1999 | 934 | 10 | 4.25 | 12/17/1999 |
| | SWP-006VC | SWP-006VCSED | 3 | 1 | 12/10/1999 | 1014 | 10 | 5 | 12/17/1999 |
| | | | | 2 | 12/10/1999 | 1026 | 10 | 4.25 | 12/17/1999 |
| | | | | 3 | 12/10/1999 | 1037 | 10 | 4.67 | 12/17/1999 |
| Tolchester Channel - North | TLC-N-005VC | TLC-N-005VCSED | 2 | 1 | 12/08/1999 | 1435 | 10 | 6.42 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1450 | 10 | 6.67 | 12/17/1999 |
| | TLC-N-006VC | TLC-N-006VCSED | 2 | 1 | 12/08/1999 | 1358 | 10 | 4.67 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1414 | 10 | 7 | 12/17/1999 |
| | TLC-N-007VC | TLC-N-007VCSED | 2 | 1 | 12/08/1999 | 1307 | 10 | 5.58 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1320 | 10 | 5.83 | 12/17/1999 |
| 1 | TLC-N-008VC | TLC-N-008VCSED | 2 | 1 | 12/08/1999 | 1157 | 10 | 5.04 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1211 | 10 | 5.67 | 12/17/1999 |
| | TLC-N-009VC | TLC-N-009VCSED | 2 | 1 | 12/08/1999 | 1126 | 10 | 6.42 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1140 | 10 | 5.83 | 12/17/1999 |
| | TLC-N-010VC | TLC-N-010VCSED | 2 | 1 | 12/08/1999 | 1056 | 10 | 5.83 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1109 | 10 | 5.625 | 12/17/1999 |

TABLE 4-4B (CONTINUED)

| Sampling Reach | Station ID | Sediment Sample ID | Total # Cores Collected | Core No. | Collection Date | Collection Time (EST) | Penetration Depth (ft) | Recovery (ft) | Processing / Compositing Date |
|----------------------------|-------------|--------------------|-------------------------------|----------|-----------------|--------------------------|---------------------------|------------------|-------------------------------------|
| Tolohester Channel - South | TLC-S-001VC | TI C-S-001VCSED | 2 | 1 1 | 12/09/1999 | 1015 | 10 | 7.67 | 12/17/1999 |
| Tolchester Channer - South | TEC-5-0017C | | | 2 | 12/09/1999 | 1025 | 10 | 7.42 | 12/17/1999 |
| | TLC-S-002VC | TLC-S-002VCSED | 2 | 1 | 12/09/1999 | 940 | 10 | 6.5 | 12/17/1999 |
| | | | | 2 | 12/09/1999 | 952 | 10 | 7.17 | 12/17/1999 |
| | TLC-S-003VC | TLC-S-003VCSED | 2 | i | 12/09/1999 | 910 | 10 | 5 | 12/17/1999 |
| | | | | 2 | 12/09/1999 | 920 | 10 | 5.71 | 12/17/1999 |
| | TLC-S-004VC | TLC-S-004VCSED | 3 | 1 | 12/08/1999 | 1523 | 10 | 5.67 | 12/17/1999 |
| | | | | 2 | 12/08/1999 | 1553 | 10 | 5.67 | 12/17/1999 |
| 1 | | | | 3 | 12/08/1999 | 1605 | 10 | 6.33 | 12/17/1999 |
| Tolchester Straightening | TLS-001VC | TLS-001VCSED | 2 | 1 | 12/09/1999 | 820 | 18 | 12.08 | 12/17/1999 |
| | | | | 2 | 12/09/1999 | 1420 | 18 | 13 | 12/17/1999 |
| | TLS-002VC | TLS-002VCSED | 2 | 1 | 12/09/1999 | 744 | 18 | 12.42 | 12/17/1999 |
| | | | | 2 | 12/09/1999 | 1340 | 18 | 12.875 | 12/17/1999 |





TABLE 4-11 BALTIMORE HARBOR APPROACH CHANNELS: SUMMARY OF SAMPLE IDs AND SAMPLE TYPES

| Sampling Reach | Station ID | Sediment Sample ID | Ecotox Sample ID | Elutriate Site Water ID | Elutriate Sample ID | |
|-----------------------------|-------------|--------------------|------------------|-------------------------|---------------------|--|
| Brewerton Eastern Extension | BEI | BEISED | BE-TOX | BE-SW | BE-EL | |
| | BE2 | BE2SED | | | | |
| | BE3 | BE3SED | | | | |
| | BE4 | BE4SED | | | | |
| C&D Approach Channels | CD-001VC | CD-001VCSED | CD-VC-TOX | CD-VC-SW | CD-VC-EL | |
| | CD-002VC | CD-002VCSED | | | | |
| | CD003 | CD003SED | CD-TOX | CD-SW | CD-EL | |
| | CD004 | CD004SED | | | | |
| | CD005 | CD005SED | | | | |
| | CD006 | CD006SED | | | | |
| Craighill | CR1 | CRISED | CR-TOX | CR-SW | CR-EL | |
| | CR2 | CR2SED | | | | |
| | CR3 | CR3SED | | | | |
| Craighill Angle-East | CRA-E-001VC | CRA-E-001VCSED | CRA-E-TOX | CRA-E-SW | CRA-E-VC-EL | |
| | CRA-E-002VC | CRA-E-002VCSED | | | | |
| | CRA-E-003VC | CRA-E-003VCSED | | | | |
| Craighill Angle-West | CRA-W-001VC | CRA-W-001VCSED | CRA-W-TOX | CRA-W-SW | CRA-W-VC-EL | |
| | CRA-W-002VC | CRA-W-002VCSED | | | | |
| | CRA-W-003VC | CRA-W-003VCSED | | | | |
| Craighill Entrance | CRE-001VC | CRE-001VCSED | CRE-TOX | CRE-SW | CRE-VC-EL | |
| | CRE-002VC | CRE-002VCSED | | | | |
| 1 | CRE-003VC | CRE-003VCSED | | | | |
| | CRE-004VC | CRE-004VCSED | | | | |
| Craighill Upper Range | CRUI | CRUISED | CRU-TOX | CRU-SW | CRU-EL | |
| | CRU2 | CRU2SED | | | | |
| | CRU3 | CRU3SED | | | | |
| Cutoff Angle | CUTI | CUTISED | CUT-TOX | CUT-SW | CUT-EL | |
| | CUT2 | CUT2SED | | | | |
| | CUT3 | CUT3SED | | | | |
| Swan Point | SWP-001VC | SWP-001VCSED | SWP-TOX | SWP-SW | SWP-VC-EL | |
| | SWP-002VC | SWP-002VCSED | | | | |
| | SWP-003VC | SWP-003VCSED | | | | |
| | SWP-004VC | SWP-004VCSED | | | | |
| | SWP-005VC | SWP-005VCSED | | | | |
| | SWP-006VC | SWP-006VCSED | | | | |
| Tolchester Channel - North | TLC-N-005VC | TLC-N-005VCSED | TLC-N-TOX | TLC-N-SW | TLC-N-VC-EL | |
| | TLC-N-006VC | TLC-N-006VCSED | | | | |
| | TLC-N-007VC | TLC-N-007VCSED | | | | |
| 1 | TLC-N-008∨C | TLC-N-008VCSED | | | | |
| | TLC-N-009VC | TLC-N-009VCSED | | | | |
| | TLC-N-010VC | TLC-N-010VCSED | | | | |
| Tolchester Channel - South | TLC-S-001VC | TLC-S-00IVCSED | TLC-S-TOX | TLC-S-SW | TLC-S-VC-EL | |
| | TLC-S-002∨C | TLC-S-002VCSED | | | | |
| | TLC-S-003VC | TLC-S-003VCSED | | | | |
| | TLC-S-004VC | TLC-S-004VCSED | | | | |
| Tolchester Straightening | TLS-001VC | TLS-00IVCSED | TLS-TOX | TLS-SW | TLS-VC-EL | |
| | TLS-002VC | TLS-002VCSED | | | | |

| | T | | | | Elutriate Site | Elutraite Sample |
|----------------------------|----------------------|--------------------|--------------------|---------------------|----------------|------------------|
| Sampling Reach | Station ID | Sediment Sample ID | Water Sample ID | Ecotox Sample ID | Water ID | ID |
| Inside Placement Site 104 | KI-3 | KI-3SED | KI-3WAT | KI-TOX | KI-SW | KI-EL |
| | KI-5 | KI-5SED | • | | | |
| | KI-7 | KI-7REFSED | KI-7WAT | | | |
| | | | KI-7WATFD | | | |
| | KI-S-1 | KI-S-1SED | - | | | |
| | KI-S-2 | KI-S-2SED | - | | | |
| | KI-7 | KI-7REFSED | - | - | KI-SW | KI-7-EL |
| | KI-7 | KI-7-REFSEDFD | - | | | |
| | KI-7 | KI-7SED* | - | | | |
| | | KI-7SEDFD* | - | | | |
| | | KI-7N** | - | | | |
| | | KI-7S** | - | | | |
| | | KI-7E** | - | | | |
| | | KI-7W* | - | | | |
| Outside Placement Site 103 | KI-11 | KI-11SED | - | KI-OUT-TOX | KI-OUT-SW | KI-OUT-EL |
| | KI-14 | KI-14SED | KI-14WAT | | | |
| | KI-15 | KI-15SED | - | | | |
| | KI-16 | KI-16SED | - | | | |
| Ocean Reference | Ocean | Ocean Reference | Ocean Reference-SW | Ocean Reference-TOX | Reference-SW | Reference-EL |
| | Lower Chesapeake Bay | Control | - | | | |

TABLE 4-12 PLACEMENT SITE 104: SUMMARY OF SAMPLE IDs AND SAMPLE TYPES

* not included in tox or elutriate composite.

** for delineation of KI-7.

TABLE 1-1SUMMARY OF TARGET ANALYTES FOR PHYSICAL AND CHEMICAL TESTING AND
BENCHMARK SPECIES FOR BIOLOGICAL TESTING^(a)

| Physical Analysis | Bulk Sediment Chemistry | Elutriate Testing | Bioassays |
|--|---|---|--|
| Grain size Moisture content Atterberg Limits Specific Gravity | Volatile Organic Compounds (VOCs) Semivolatile Organic Compounds (SVOCs) Chlorinated Pesticides Organophosphorus Pesticides Polychlorinated Biphenyls (PCBs) Congeners Polynuclear Aromatic Hydrocarbons (PAHs) Metals Butyltins Dioxin and Furan Congeners Nutrients (Ammonia, nitrate, nitrite, inorganic nitrogen, phosphorus) Cyanide Simultaneously Extracted Metals (SEM) / Acid Volatile Sulfides (AVS) Biological Oxygen Demand (BOD) Chemical Oxygen Demand (COD) Total Organic Carbon (TOC) | Same as bulk sediment or dependent upon federal or regional/state water quality criteria | <u>Water column bioassays</u> Mysidposis bahia (opossum shrimp) Cyrpinodon variegatus (sheepshead minnow) Menidia beryllina (inland silverside) Mytilus sp. (blue mussel) Solid phase bioassays Leptocheirus plumulosus (estuarine amphipod) Mysidposis bahia (opposum shrimp) Neanthes arenaceodenata (estuarine polychaete) Bioaccumulation Nereis virens (sand worm) Macoma nasuta (blunt-nose clam) |

(a) From USEPA/USACE 1998; benchmark species appropriate for estuarine or marine placement alternatives

TABLE 3-5MEAN VALUES FOR PHYSICAL CHARACTERISTICS IN SEDIMENT FROMBALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | Inside Site | Outside | Brewerton Eastern | | Craighill | Craighill | Craighill Upper | Cutoff | Swan | |
|------------------|---------|-------------|----------|----------------------|-----------|-----------|-----------|--------------------|--------|-------|------------|
| REA | ACH ID: | 104 | Site 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| SAMPLE | E TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| SAMPLE SIZE (N) | | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | | | | | | | | | | |
| MOISTURE CONTENT | % | 67.5 | 71.9 | 74.0 | 46.33 | 69.6 | 70.67 | 69.4 | 72.67 | 71.9 | 69.3 |
| SPECIFIC GRAVITY | T/4C | 1.26 | 1.24 | 1.2 | 2.61 | 1.2 | 2.7 | 1.33 | 1.2 | 1.23 | 1.4 |
| CLAY | % | 43.88 | 33.85 | 35.95 | 28.6 | 21.3 | 27.83 | 28.9 | 21.47 | 26.93 | 46.8 |
| GRAVEL | % | 0.0 | 0.0 | 0.47 | 0.33 | 0.0 | 1.9 | 0.0 | 0.0 | 0.0 | 0.0 |
| SAND | % | 10.7 | 5.18 | 4.15 | 35.83 | 14.35 | 24.37 | 7.77 | 4.03 | 0.42 | 1.17 |
| SILT | % | 45.17 | 60.97 | 59.43 | 35.23 | 64.35 | 45.9 | 63.34 | 74.5 | 72.63 | 52.02 |

TABLE 3-6 MEAN CONCENTRATIONS OF INORGANIC NON-METALS (MG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | S | REACH ID: AMPLE TYPE: | Inside Site 104 Grab | Outside Site 104 Grab | Brewerton Eastern Extension Grab | Craighiii Grab | Craighiii Angle Grab | Craighili Entrance Grab | Craighiii Upper Range Grab | Cutoff Angie Grab | Swan Point Grab | Toichester Grab |
|---|-------|---------------------------------------|----------------------------|-----------------------------|---|---|--|--|-------------------------------------|-------------------------|-----------------------|--------------------|
| | SAN | IPLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| CHROMIUM (HEXAVALENT) | MG/KG | 0.13 - 0.56 | 20. 0,44 | 0.31 U | 0.4 U | 0.19 U | 0.34 U | 0.37 U | 0.33 U | 0.36 U | 0.37 11 | 03111 |
| CYANIDE | MG/KG | 0.21 - 2.1 | 0.83 | 0.78 U | 0.22 U | 0.23 U | | 0.28 | 0.22 | 0.21 U | 0.21 U | 0.22 |
| NITROGEN, AMMONIA | MG/KG | 0.05 - 2.8 | 51.23 | \$137.68 | F 10 43.9 | 12019.13 | ASE 7.35 | 1. 178.43 | 64.53 | | 127.9 | 16.57 |
| NITROGEN, NITRATE AND NITRITE | MG/KG | 0.01 - 2.3 | 1.02 | S.S. 0.97 | 12 0:33 | 0.01 U | Sec. 10.23 | 0.08 | | 0.12 | ANTE 344.7 | 38: 318: 205 |
| NITROGEN, TOTAL KJELDAHL | MG/KG | 23 - 30.8 | \$1240.5 | \$1000.75 | 未到於1413.5 | 307.0 | 1565.0 | 963.67 | 1 667.67 | 700.67 | 1430.0 | 3 2 7207'5 |
| OXYGEN DEMAND, BIOCHEMICAL | MG/KG | 120 - 1200 | #10042.5 | 5245.0 | Kanin 754:75 | 315.0 | WW 995.5 | | 1516.0 | 964.0 | 1218.33 | 1323 75 |
| OXYGEN DEMAND, CHEMICAL | MG/KG | 990 - 2070 | \$84375:0 | A73925:0 | MH#51925:0 | 32526.67 | 80300.0 | \$45036.67 | 90966.67 | 186900.0 | 67466.67 | 51550 0 |
| PHOSPHORUS, TOTAL | MG/KG | 4.2 - 5.6 | NA | NA | 320.32 | 51161.83 | 312.5 | 5 227.0 | 280.33 | 14 311233 | 2516 | 301 75 |
| SULFIDE, TOTAL | MG/KG | 0.1 - 10.6 | 1803.75 | 41692.0 | Act 1376.55 | 此清秋39.17 | 414,45 | 527.33 | 612:33 | 239.0 | 392.37 | 11 |
| TOTAL ORGANIC CARBON | MG/KG | 6920 - 19500 | 67975.0 | \$ 89825.0 | ¥¥123500.0 | 33813.33 | \$ 107800.0 | 113100.0 | 97533.33 | 140666.67 | 103166.67 | 107675 0 |
| II - not detected in any comptentiation and built | | 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 | | | | the second of the second se | And the second s | And the second sec | | | | 10-010-0 |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NA = not analyzed.

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

•



| | SA | REACH ID: MPLE TYPE: | Inside Site 104 Grab | Outside Site 104 Grab | Brewerton Eastern Extension Grab | Craighill Grab | Cralghill Angle Grab | Craighill Entrance Grab | Craighlll Upper Range Grab | Cutoff Angle Grab | Swan Polnt Grab | Tolchester Grab |
|-----------------------------|-------|-------------------------|----------------------------|-----------------------------|---|-------------------|----------------------------|-------------------------------|-------------------------------------|-------------------------|-----------------------|--------------------|
| | SAM | PLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL(RANGE) | | | | | | | | | | |
| 1,1,1-TRICHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 1,1,2,2-TETRACHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 1,1,2-TRICHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 1,1-DICHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 1,1-DICHLOROETHYLENE | UG/KG | 0.8 - 3 | 0.8 U | 1.1 U | 0.83 U | 0.8 U | 0.8 U | 0.8 U | 0.8 U | 3.0 U | 0.8 U | 0.8 U |
| 1,2-DICHLOROBENZENE | UG/KG | 0.9 - 3 | 0.9 U | 1.18 U | 0.93 U | 0.9 U | 0.9 U | 0.9 U | 0.9 U | 3.0 U | 0.9 U | 0.9 U |
| 1,2-DICHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 1,4-DICHLOROBENZENE | UG/KG | 1-4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| 2-CHLOROETHYL VINYL ETHER | UG/KG | 49 - 190 | 50.0 U | 62.5 U | 51.25 U | 49.0 U | 50.0 U | 49.67 U | 50.33 U | 180.0 U | 50.0 U | 50.0 U |
| ACROLEIN | UG/KG | 12 - 46 | 12.0 U | 15.0 U | 12.25 U | 12.0 U | 12.0 U | 12.0 U | 12.0 U | 43.33 U | 12.0 U | 12.0 U |
| ACRYLONITRILE | UG/KG | 14 - 54 | 14.0 U | 17.5 U | 14.25 U | 14.0 U | 14.0 U | 14.0 U | 14.0 U | 50.67 U | 14.0 U | 14.0 U |
| BENZENE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| BROMODICHLOROMETHANE | UG/KG | 0.9 - 3 | 0.9 U | 1.18 U | 0.93 U | 0.9 U | 0.9 U | 0.9 U | 0.9 U | 3.0 U | 0.9 U | 0.9 U |
| BROMOMETHANE | UG/KG | 3 - 11 | 3.0 U | 3.75 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 10.67 U | 3.0 U | 3.0 U |
| CARBON DISULFIDE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | .4.0 | 2.0 U | 2.33 | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| CARBON TETRACHLORIDE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| CFC-11 | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| CFC-12 | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| CHLOROBENZENE | UG/KG | 0.9 - 3 | 0.9 U | 1.18 U | 0.93 U | 0.9 U | 0.9 U | 0.9 U | 0.9 U | 3.0 U | 0.9 U | 0.9 U |
| CHLORODIBROMOMETHANE | UG/KG | 0.8 - 3 | 0.8 U | 1.1 U | 0.83 U | 0.8 U | 0.8 U | 0.8 U | 0.8 U | 3.0 U | 0.8 U | 0.8 U |
| CHLOROETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| CHLOROFORM | UG/KG | 1-4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| CHLOROMETHANE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 3+ Est 8.25 | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| CIS-1,3-DICHLOROPROPENE | UG/KG | 0.8 - 3 | 0.8 U | 1.1 U | 0.83 U | 0.8 U | 0.8 U | 0.8 U | 0.8 U | 3.0 U | 0.8 U | 0.8 U |
| DICHLOROMETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 11.75 | Alimate 1.33 | See 1.5 | 1.33 | . 1.33 | 3.67 U | 1.0 U | 15.75 |
| ETHYLBENZENE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| M-DICHLOROBENZENE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| METHYLBENZENE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| TETRACHLOROETHENE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| TRANS-1,2-DICHLOROETHENE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |
| TRANS-1,3-DICHLOROPROPENE | UG/KG | 0.8 - 3 | 0.8 U | 1.1 U | 0.83 U | 0.8 U | 0.8 U | 0.8 U | 0.8 U | 3.0 U | 0.8 U | 0.8 U |
| TRIBOMOMETHANE | UG/KG | 1 - 4 | 1.0 U | 1.25 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 3.67 U | 1.0 U | 1.0 U |
| TRICHLOROETHYLENE | UG/KG | 0.9 - 3 | 0.9 U | 1.18 U | 0.93 U | 0.9 U | 0.9 U | 0.9 U | 0.9 U | 3.0 U | 0.9 U | 0.9 U |
| VINYL CHLORIDE | UG/KG | 2 - 8 | 2.0 U | 2.5 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 7.33 U | 2.0 U | 2.0 U |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

TABLE 3-8 MEAN CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | REACH ID: | Inside Site 104 | Outside Site 104 | Brewerton Eastern Extension | Craighill | Cralghill Angle | Craighlll Entrance | Cralghlll Upper Range | Cutoff Angle | Swan Point | Tolchester |
|------------------------------|-------|----------------|--------------------|---------------------|-----------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|---------------|
| | S. | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | IPLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| 1,2,4-TRICHLOROBENZENE | UG/KG | 98 - 100 | 100.0 U | 100.0 U | 99.0 U | 99.67 U | 99.5 U | 99.67 U | 100.0 U | 99.3311 | 99 33 11 | 99.5.11 |
| 1,2-DIPHENYLHYDRAZINE | UG/KG | 91 - 95 | 93.0 U | 93.0 U | 92.5 U | 93.0 U | 93.5 U | 92.67 U | 93.67 U | 93.011 | 93.011 | 93 25 11 |
| 1-METHYLNAPHTHALENE | UG/KG | 25 - 27 | NA | NA | 29.0 | 26.0 U | 26.0 U | 26.0 U | 26.0 U | 25.67 11 | 25.67 11 | 100 100 10075 |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | UG/KG | 220 - 230 | 230.0 U | 230.0 U | 227.5 U | 230.0 U | 230.0 U | 230.0 U | 230.0 U | 230.0 U | 230.0 11 | 230.0 [] |
| 2,4,5-TRICHLOROPHENOL | UG/KG | 97 - 100 | 99.0 U | 99.0 U | 98.25 U | 99.0 U | 99.0 U | 98.67 U | 99.67 U | 99.0 11 | 99.01 | 99.011 |
| 2,4,6-TRICHLOROPHENOL | UG/KG | 90 - 94 | 92.0 U | 92.0 U | 91.5 U | 92.0 U | 92.5 U | 91.67 U | 92.67 U | 92.0 U | 92.01 | 92 25 11 |
| 2,4-DICHLOROPHENOL | UG/KG | 80 - 84 | 82.0 U | 82.0 U | 81.5 U | 82.0 U | 82.0 U | 81.67 U | 82.67 U | 82.33 U | 82.011 | 82 25 11 |
| 2,4-DIMETHYLPHENOL | UG/KG | 90 - 94 | 92.0 U | 92.0 U | 91.5 U | 92.0 U | 92.5 U | 91.67 U | 92.67 U | 92.0 U | 92.011 | 92 25 11 |
| 2,4-DINITROPHENOL | UG/KG | 190 - 190 | 190.0 U | 190.0 U | 190.0 U | 190.0 U | 190.0 U | 190.0 U | 190.0 U | 190.011 | 190.011 | 190.011 |
| 2,4-DINITROTOLUENE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 11 | 110.011 | 110.011 |
| 2,6-DINITROTOLUENE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 11 | 110.0 U |
| 2-CHLORONAPHTHALENE | UG/KG | 120 - 120 | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.011 | 120.011 |
| 2-CHLOROPHENOL | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.011 | 110.0 11 | 110.011 |
| 2-METHYL-4,6-DINITROPHENOL | UG/KG | 230 - 240 | 240.0 U | 240.0 U | 237.5 U | 240.0 U | 240.0 U | 240.0 U | 240.0 U | 240.011 | 240.0 11 | 240.011 |
| 2-METHYLNAPHTHALENE | UG/KG | 28 - 30 | NA | NA | 184 677.25 | 29.0 U | Sec. 38.5 | 29.0 U | 29.0 11 | 28.67 11 | 29011 | 240.0 0 |
| 2-METHYLPHENOL | UG/KG | 89 - 93 | 91.0 U | 93.25 | 90.5 U | 91.0 U | 91.5 U | 90.67 U | 91.6711 | 91 33 11 | 91.011 | 01 25 11 |
| 2-NITROANILINE | UG/KG | 94 - 98 | 96.0 U | 96.0 U | 95.5 U | 96.0 U | 96.5 U | 95.67 U | 96.67 11 | 96.011 | 96.011 | 96 25 11 |
| 2-NITROPHENOL | UG/KG | 98 - 650 | 100.0 U | 100.0 U | 99.0 U | 99.67 U | 99.5 U | 99.67 U | 100.0 U | 99.3311 | 99 33 11 | 237.5 [] |
| 3,3'-DICHLOROBENZIDINE | UG/KG | 69 - 72 | 71.0 U | 71.0 U | 70.5 U | 71.0 U | 71.0 U | 70.67 U | 71.67 U | 71.3311 | 71 33 11 | 71.011 |
| 3,4-METHYLPHENOL | UG/KG | 98 - 100 | 192.5 | 137.5 | June 114.25 | 99.67 U | 14 735.0 | 1 Re - 160.0 | | 176.67 | 11.550 | 11.0 0 |
| 3-NITROANILINE | UG/KG | 89 - 93 | 91.0 U | 91.0 U | 90.5 U | 91.0 U | 91.5 U | 90.67 U | 91.67 U | 91 33 11 | 91.0.11 | 01 25 11 |
| 4-BROMOPHENYL PHENYL ETHER | UG/KG | 120 - 120 | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 U | 120.0 11 | 120.011 | 120 0 11 | 120.011 |
| 4-CHLORO-3-METHYLPHENOL | UG/KG | 94 - 98 | 96.0 U | 96.0 U | 95.5 U | 96.0 U | 96.5 U | 95.67 U | 96.67 U | 96.011 | 96.011 | 96 25 11 |
| 4-CHLOROPHENYL PHENYL ETHER | UG/KG | 130 - 130 | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.011 | 130.011 | 130.011 | 130.011 |
| 4-NITROPHENOL | UG/KG | 85 - 89 | 87.0 U | 87.0 U | 86.5 U | 87.0 U | 87.0 U | 86.67 U | 87.67 U | 87 33 11 | 87.011 | 87 25 11 |
| BENZIDINE | UG/KG | 58 - 60 | 59.0 U | 59.0 U | 58.75 U | 59.0 U | 59.5 U | 58.67 U | 59.67 11 | 59.011 | 59 33 11 | 59.011 |
| BENZYL ALCOHOL | UG/KG | 80 - 84 | 82.0 U | 82.0 U | 81.5 U | 82.0 U | 82.0 U | 81.67 U | 82.67 11 | 82.33 11 | 82 0 11 | 82 25 11 |
| BENZYL BUTYL PHTHALATE | UG/KG | 87 - 91 | 89.0 U | 89.0 U | 88.5 U | 459.67 | 89.0 U | 88.67 U | 89.67 U | 89.33 U | 89.0 U | 89.25 U |









TABLE 3-8 (CONTINUED)

| | | | | | Brewerton | | | | Craighill | | | |
|---------------------------------|--------|------------------|----------|----------|-----------|-----------|-----------|-----------|-----------|---------|---------|------------|
| | | | Inside | Outside | Eastern | | Cralghill | Craighill | Upper | Cutoff | Swan | |
| | | REACH ID: | Site 104 | Site 104 | Extension | Cralghill | Angle | Entrance | Range | Angle | Point | Tolchester |
| | S. | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| # | STATIO | NS SAMPLED: | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| BIS(2-CHLOROETHOXY)METHANE | UG/KG | 98 - 100 | 100.0 U | 100.0 U | 99.0 U | 99.67 U | 99.5 U | 99.67 U | 100.0 U | 99.33 U | 99.33 U | 99.5 U |
| BIS(2-CHLOROETHYL) ETHER | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/KG | 92 - 96 | 115.5 | 94.0 U | 93.5 U | 203.0 | 94.5 U | 212.67 | 94.67 U | 94.0 U | 94.0 U | 235.5 |
| CARBAZOLE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U |
| CYCLOHEXANONE | UG/KG | 86 - 90 | 88.0 U | 88.0 U | 87.5 U | 88.0 U | 88.0 U | 87.67 U | 88.67 U | 88.33 U | 88.0 U | 88.25 U |
| DI-N-BUTYL PHTHALATE | UG/KG | 150 - 150 | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U | 150.0 U |
| DI-N-OCTYL PHTHALATE | UG/KG | 83 - 87 | 85.0 U | 85.0 U | 84.5 U | 223.67 | 85.0 U | 84.67 U | 85.67 U | 85.33 U | 85.0 U | 85.25 U |
| DIBENZOFURAN | UG/KG | 130 - 130 | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U | 130.0 U |
| DIETHYL PHTHALATE | UG/KG | 86 - 90 | 88.0 U | 88.0 U | 87.5 U | 88.0 U | 88.0 U | 87.67 U | 88.67 U | 88.33 U | 88.0 U | 88.25 U |
| DIMETHYL PHTHALATE | UG/KG | 82 - 86 | 84.0 U | 84.0 U | 83.5 U | 84.0 U | 84.0 U | 83.67 U | 84.67 U | 84.33 U | 84.0 U | 84.25 U |
| HEXACHLORO-1,3-BUTADIENE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U |
| HEXACHLOROBENZENE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U |
| HEXACHLOROCYCLOPENTADIENE | UG/KG | 93 - 97 | 95.0 U | 95.0 U | 94.5 U | 95.0 U | 95.5 U | 94.67 U | 95.67 U | 95.0 U | 95.0 U | 95.25 U |
| HEXACHLOROETHANE | UG/KG | 110 - 110 | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U | 110.0 U |
| METHANAMINE, N-METHYL-N-NITROSO | UG/KG | 84 - 88 | 86.0 U | 86.0 U | 85.75 U | 86.0 U | 86.0 U | 85.67 U | 86.67 U | 86.33 U | 86.0 U | 86.25 U |
| N-NITROSODI-N-PROPYLAMINE | UG/KG | 97 - 100 | 99.0 U | 99.0 U | 98.25 U | 99.0 U | 99.0 U | 98.67 U | 99.67 U | 99.0 U | 99.0 U | 99.0 U |
| N-NITROSODIPHENYLAMINE | UG/KG | 85 - 89 | 87.0 U | 87.0 U | 86.5 U | 87.0 U | 87.0 U | 86.67 U | 87.67 U | 87.33 U | 87.0 U | 87.25 U |
| NITROBENZENE | UG/KG | 98 - 100 | 100.0 U | 100.0 U | 99.0 U | 99.67 U | 99.5 U | 99.67 U | 100.0 U | 99.33 U | 99.33 U | 99.5 U |
| P-CHLOROANILINE | UG/KG | 83 - 87 | 85.0 U | 85.0 U | 84.5 U | 85.0 U | 85.0 U | 84.67 U | 85.67 U | 85.33 U | 85.0 U | 85.25 U |
| P-NITROANILINE | UG/KG | 75 - 79 | 77.0 U | 77.0 U | 76.5 U | 77.0 U | 77.0 U | 76.67 U | 77.67 U | 77.33 U | 77.0 U | 77.25 U |
| PENTACHLOROPHENOL | UG/KG | 140 - 140 | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U | 140.0 U |
| PHENOL | UG/KG | 87 - 91 | 89.0 U | 89.0 U | 88.5 U | 89.0 U | 129.0 | 88.67 U | 89.67 U | 89.33 U | 89.0 U | 89.25 U |
| PYRIDINE | UG/KG | 210 - 210 | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U | 210.0 U |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NA = not analyzed.

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

| | | | | | Brewerton | | | | Craighill | | | |
|--------------------|-------|------------------|-------------|----------|-----------|-----------|-----------|-----------|-----------|---------|---------|------------|
| | | | Inside Site | Outside | Eastern | | Craighill | Craighill | Upper | Cutoff | Swan | |
| | | REACH ID: | 104 | Site 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | IPLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| 4,4'-DDD | UG/KG | 0.49 - 1.2 | 0.63 | 0.5 U | 0.78 | 0.5 U | 0.51 U | 1.27 | 0.5 U | 0.5 U | 0.5 U | 0.5 U |
| 4,4'-DDE | UG/KG | 0.93 - 2.3 | 1:27 | 0.96 U | 1.0 | 0.95 U | 0.96 U | 1.99 | 0.96 U | 0.96 U | 0.96 U | 0.96 U |
| 4,4'-DDT | UG/KG | 0.64 - 1.6 | 0.66 U | 0.66 U | 0.66 U | 0.66 U | 0.66 U | 0.98 U | 0.66 U | 0.66 U | 0.66 U | 0.66 U |
| ALDRIN | UG/KG | 0.23 - 0.58 | 0.33 | 0.24 U | 0.31 | 0.24 U | 0.24 U | 0.35 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U |
| ALPHA-BHC | UG/KG | 0.29 - 0.71 | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.43 U | 0.29 U | 0.3 U | 0.3 U | 0.29 U |
| BETA-BHC | UG/KG | 0.21 - 0.53 | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.32 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U |
| CHLORDANE | UG/KG | 6.3 - 16 | 6.4 U | 6.4 U | 6.38 U | 6.37 U | 6.4 U | 9.63 U | 6.43 U | 6.43 U | 6.43 U | 6.42 U |
| CHLOROBENSIDE | UG/KG | 0.65 - 1.6 | 3.3 U | 3.3 U | 0.66 U | 0.66 U | 0.67 U | 0.98 U | 0.67 U | 0.66 U | 0.67 U | 0.67 U |
| DACTHAL | UG/KG | 0.65 - 1.6 | 10.0 U | 10.0 U | 0.66 U | 0.66 U | 0.67 U | 0.98 U | 0.67 U | 0.66 U | 0.67 U | 0.67 U |
| DELTA-BHC | UG/KG | 0.18 - 0.44 | 0.18 U | 0.18 U | 0.18 U | 0.18 U | 0.18 U | 0.27 U | 0.18 U | 0.18 U | 0.18 U | 0.18 U |
| DIELDRIN | UG/KG | 0.43 - 1.1 | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.45 U | 0.66 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U |
| ENDOSULFAN I | UG/KG | 0.25 - 0.63 | 0.28 | 0.26 U | 0.41 | 0.26 U | 0.26 U | 0.38 U | 0.26 U | 0.26 U | 0.26 U | 0.26 U |
| ENDOSULFAN II | UG/KG | 0.42 - 1 | 0.43 U | 0.43 U | 0.43 U | 0.43 U | 0.44 U | 0.62 U | 0.43 U | 0.43 U | 0.43 U | 0.43 U |
| ENDOSULFAN SULFATE | UG/KG | 0.57 - 1.4 | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.85 U | 0.58 U | 0.59 U | 0.58 U | 0.58 U |
| ENDRIN | UG/KG | 0.41 - 1 | 0.42 U | 0.5 | 0.42 U | 0.42 U | 0.42 U | 0.61 U | 0.42 U | 0.42 U | 0.42 U | 0.42 U |
| ENDRIN ALDEHYDE | UG/KG | 0.68 - 1.7 | 0.81 | 0.91 | 0.71 | 0.7 U | 0.71 U | 1.04 U | 0.71 U | 0.71 U | 0.71 U | 0.71 U |
| GAMMA-BHC | UG/KG | 0.23 - 0.58 | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.35 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U |
| HEPTACHLOR | UG/KG | 0.29 - 0.73 | 0.3 U | 0.3 U | 0.31 | 0.3 U | 0.3 U | 0.44 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U |
| HEPTACHLOR EPOXIDE | UG/KG | 0.21 - 0.51 | 1.08 | 0.41 | 1.03 | 0.21 U | 0.21 U | 0.31 U | 0.24 | 0.21 U | 0.22 | 0.44 |
| METHOXYCHLOR | UG/KG | 3.1 - 7.8 | 3.2 U | 3.65 | 3.2 U | 3.2 U | 3.2 U | 4.73 U | 3.2 U | 3.17 U | 3.2 U | 3.2 U |
| MIREX | UG/KG | 0.65 - 1.6 | 3.3 U | 3.3 U | 0.66 U | 0.66 U | 0.67 U | 0.98 U | 0.67 U | 0.66 U | 0.67 U | 0.67 U |
| TOXAPHENE | UG/KG | 62 - 150 | 63.0 U | 63.0 U | 62.75 U | 62.67 U | 63.0 U | 92.33 U | 63.33 Ū | 63.33 U | 63.33 U | 63.25 U |

TABLE 3-9 MEAN CONCENTRATIONS OF CHLORINATED PESTICIDES IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

U = not detected in any sample within reach; value represents mean detection limit (DL).

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.





TABLE 3-10MEAN CONCENTRATIONS OF ORGANOPHOSPHORUS PESTICIDES (UG/KG) IN SEDIMENTFROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | | | | Brewerton | | | | Craighill | | | |
|------------------|-------|------------------|-------------|--------------|-----------|-----------|-----------|-----------|-----------|---------|---------|------------|
| | | | Inside Site | Outside Site | Eastern | | Craighill | Craighill | Upper | Cutoff | Swan | |
| | | REACH ID: | 104 | 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| | SA | MPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | PLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| AZINPHOS METHYL | UG/KG | 33 - 34 | 33.0 U | 33.0 U | 33.0 U | 33.0 U | 33.5 U | 33.33 U | 33.67 U | 33.67 U | 33.67 U | 33.75 U |
| DEMETON | UG/KG | 33 - 34 | 33.0 U | 33.0 U | 33.0 U | 33.0 U | 33.5 U | 33.33 U | 33.67 U | 33.67 U | 33.67 U | 33.75 U |
| ETHYL PARATHION | UG/KG | 33 - 34 | 33.0 U | 33.0 U | 33.0 U | 33.0 U | 33.5 U | 33.33 U | 33.67 U | 33.67 U | 33.67 U | 33.75 U |
| MALATHION | UG/KG | 33 - 34 | 33.0 U | 33.0 U | 33.0 U | 33.0 U | 33.5 U | 33.33 U | 33.67 U | 33.67 U | 33.67 U | 33.75 U |
| METHYL PARATHION | UG/KG | 33 - 34 | 33.0 U | 33.0 U | 33.0 U | 33.0 U | 33.5 U | 33.33 U | 33.67 U | 33.67 U | 33.67 U | 33.75 U |

TABLE 3-11 MEAN CONCENTRATIONS OF PCB AROCLORS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | | | | Brewerton | | | | Craighill | | | |
|--------------|-------|------------------|-------------|----------|-----------|-----------|-----------|-----------|----------------|--------|--------|------------|
| | | | Inside Site | Outside | Eastern | | Craighill | Craighill | Upper | Cutoff | Swan | |
| | | REACH ID: | 104 | Site 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | PLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| AROCLOR 1016 | UG/KG | 1.1 - 1.2 | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.1 U | 1.13 U |
| AROCLOR 1221 | UG/KG | 9.8 - 11 | 10.0 U | 10.0 U | 9.93 U | 9.93 U | 9.95 U | 9.97 U | 10.0 U | 9.93 U | 9.93 U | 10.25 U |
| AROCLOR 1232 | UG/KG | 2.9 - 3.3 | 3.0 U | 3.0 U | 2.98 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 2.97 U | 3.0 U | 3.08 U |
| AROCLOR 1242 | UG/KG | 0.98 - 1.1 | 1.0 U | 1.0 U | 0.99 U | 0.99 U | 1.0 U | 1.0 U | 1. Jack 1. 7.1 | 0.99 U | 0.99 U | 1.02 U |
| AROCLOR 1248 | UG/KG | 1.8 - 2 | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 2.53 | 1.8 U | 1.85 U |
| AROCLOR 1254 | UG/KG | 2.5 - 2.9 | 4.7 | 3.33 | 3.52 | 2.6 U | 2.6 U | 7.93 | 12.33 | 5.1 | 2.6 U | 2.67 U |
| AROCLOR 1260 | UG/KG | 1.1 - 1.2 | 4.03 | 2.4 | 6.88 | 1.1 U | 1.1 U | 3.57 | 7.6 | 2.63 | 1.33 | 1.43 |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.





TABLE 3-12 MEAN CONCENTRATIONS OF PCB CONGENERS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | S. | REACH ID: | Inside Site 104 | Outside Site 104 | Brewerton Eastern Extension | Craighill | Craighill Angle | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester |
|------------------------|-------|----------------|--------------------|---------------------|-----------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|------------|
| | SAM | DIFSIZE (N) | 1 | 1 | 1 | Orab | Grad | Grab | Grab | Grab | Grab | Grab |
| | SAM | The Sine (11). | | | 1 | 0 | 0 | 2 | 3 | 1 | 0 | 0 |
| ANALYTE | UNIT | DL (RANGE) | | | 12.12.22 | | | | | | | |
| BZ# 8* | UG/KG | 0.95 - 0.98 | 0.97 U | 0.97 U | 0.95 U | NA | NA | 0.96 U | 0.98 U | 0.98 U | NA | NA |
| BZ# 18* | UG/KG | 0.71 - 0.73 | 0.72 U | 0.72 U | 0.71 U | NA | NA | 0.71 U | 0.73 U | 0.73 U | NA | NA |
| BZ# 28* | UG/KG | 0.82 - 0.85 | 0.84 U | 0.84 U | 0.82 U | NA | NA | 0.83 U | 0.85 U | 0.85 U | NA | NA |
| BZ# 44* | UG/KG | 0.71 - 0.73 | 0.72 U | 0.72 U | 0.71 U | NA | NA | 0.71 U | 0.73 U | 0.73 U | NA | NA |
| BZ# 49 | UG/KG | 0.95 - 0.98 | 0.97 U | 0.97 U | 0.95 U | NA | NA | 0.96 U | 0.98 U | 0.98 U | NA | NA |
| BZ# 52* | UG/KG | 1.3 - 1.3 | 1.3 U | 1.3 U | 1.3 U | NA | NA | 1.3 U | 1.3 U | 1.3 U | NA | NA |
| BZ# 66* | UG/KG | 0.87 - 0.9 | 0.89 U | 0.89 U | 0.87 U | NA | NA | 0.89 U | 0.9 U | 0.9 U | NA | NA |
| BZ# 77* | UG/KG | 1.8 - 1.8 | 1.8 U | 1.8 U | 1.8 U | NA | NA | 1.8 U | 1.8 U | 1.8 U | NA | NA |
| BZ# 87 | UG/KG | 0.8 - 0.83 | 0.82 U | 0.82 U | 0.8 U | NA | NA | 0.81 U | 0.83 U | 0.83 U | NA | NA |
| BZ# 101* | UG/KG | 0.81 - 0.84 | 0.83 U | 0.83 U | 0.81 U | NA | NA | 0.88 | 0.84 U | 0.84 U | NA | NA |
| BZ# 105* | UG/KG | 0.84 - 0.87 | 0.86 U | 0.86 U | 0.84 U | NA | NA | 0.85 U | 0.87 U | 0.87 U | NA | NA |
| BZ# 118* | UG/KG | 0.98 - 1 | 1.0 U | 1.0 U | 0.98 U | NA | NA | 1.0 U | 1.0 U | 1.0 U | NA | NA |
| BZ# 126* | UG/KG | 1.3 - 1.3 | 1.3 U | 1.3 U | 1.3 U | NA | NA | 1.3 U | 1.3 U | 1.3 U | NA | NA |
| BZ# 128* | UG/KG | 0.98 - 1 | 1.0 U | 1.0 U | 0.98 U | NA | NA | 1.0 U | 1.0 U | 1.0 U | NA | NA |
| BZ# 138* | UG/KG | 0.98 - 1 | 1.0 U | 1.0 U | 1.6 | NA | NA | 1.0 U | 1.0 U | 1.0 U | NA | NA |
| BZ# 153* | UG/KG | 0.98 - 4.8 | 0.99 U | 0.99 U | 2.7 | NA | NA | 0.98 U | 1.0 U | 1.0 U | NA | NA |
| BZ# 156 | UG/KG | 1.2 - 1.2 | 1.2 U | 1.2 U | 1.2 U | NA | NA | 1.2 U | 1.2 U | 1.2 U | NA | NA |
| BZ# 169* | UG/KG | 1.6 - 1.6 | 1.6 U | 1.6 U | 1.6 U | NA | NA | 1.6 U | 1.6 U | 1.6 U | NA | NA |
| BZ# 170* | UG/KG | 0.96 - 0.99 | 0.98 U | 0.98 U | 1.0 | NA | NA | 0.97 U | 0.99 U | 0.99 U | NA | NA |
| BZ# 180* | UG/KG | 1.1 - 1.1 | 1.1 U | 1.1 U | 2.1 | NA | NA | 1.1 U | 1.1 U | 1.1 U | NA | NA |
| BZ# 183 | UG/KG | 0.62 - 0.64 | 0.63 U | 0.63 U | 0.62 U | NA | NA | 0.63 U | 0.64 U | 0.64 U | NA | NA |
| BZ# 184 | UG/KG | 0.77 - 0.8 | 0.79 U | 0.79 U | 0.77 U | NA | NA | 0.79 U | 0.8 U | 0.8 U | NA | NA |
| BZ# 187* | UG/KG | 0.76 - 0.79 | 0.78 U | 0.78 U | 1.4 | NA | NA | 0.78 U | 0.79 U | 0.79 U | NA | NA |
| BZ# 195 | UG/KG | 1.2 - 1.2 | 1.2 U | 1.2 U | 1.2 U | NA | NA | 1.2 U | 1.2 U | 1.2 U | NA | NA |
| BZ# 206 | UG/KG | 1.2 - 1.2 | 1.2 U | 1.2 U | 1.2 U | NA | NA | 1.2 U | 1.27 | 1.2 U | NA | NA |
| BZ# 209 | UG/KG | 0.98 - 1 | NA | NA | 0.98 U | NA | NA | 1.25 | 1.47 | 1.0 U | NA | NA |
| TOTAL PCB (ND=0)** | UG/KG | - | 0.0 | 0.0 | 17.6 | NA | NA | 0.9 | 0.0 | 0.0 | NA | NA |
| TOTAL PCB (ND=1/2DL)** | UG/KG | | 18.7 | 18.7 | 31.3 | NA | NA | 19.13 | 18.7 | 18.8 | NA | NA |
| TOTAL PCB (ND=DL) | UG/KG | - | 37.4 | 37.4 | 44.9 | NA | NA | 37.3 | 37.5 | 37.6 | NA | NA |

U = not detected in any sample within reach; value represents mean detection limit (DL).

* = PCB congeners used for total PCB summation, as per Table 9-3 of the ITM (USEPA/USACE 1998). Total multipled by a factor of 2 as per NOAA 1993.

* *= Note that the mean of total PCBs for individual samples is not equivalent to the sum of mean individual PCBs for ND=0 and ND=1/2DL.

NA = not analyzed.

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

TABLE 3-13MEAN CONCENTRATIONS OF PAHs (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR
APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | | | | Brewerton | | | | Craighill | | | |
|------------------------|-------|------------------|----------|----------|-----------|-----------|-----------|-----------|-----------|---------|--------|------------|
| | | | Inside | Outside | Eastern | | Craighill | Craighill | Upper | Cutoff | Swan | |
| | | REACH ID: | Site 104 | Site 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| | S. | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | IPLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | 3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| ACENAPHTHENE | UG/KG | 16 - 16 | 16.0 U | 16.0 U | 73.5 | 16.0 U | 20.5 | 16.0 U | 16.0 U | 16.0 U | 16.0 U | 65.5 |
| ACENAPHTHYLENE | UG/KG | 35 - 45 | 40.0 | 35.75 U | 37.75 U | 36.0 U | 36.5 U | 35.67 U | 36.0 U | 35.67 U | 36.0 U | 36.0 U |
| ANTHRACENE | UG/KG | 0.86 - 0.9 | 27.82 | 2.91 | 20.23 | 1.52 | 5.75 | 2.53 | 2.53 | | 2.86 | 14.0 |
| BENZ[A]ANTHRACENE | UG/KG | 0.79 - 0.83 | 36.95 | 7.6 | 32.67 | 1.44 | 12.75 | 5.9 | 8.4 | 8.33 | 5.43 | 24.9 |
| BENZO[A]PYRENE | UG/KG | 0.76 - 2.8 | 42.2 | 11.88 | 47.0 | 1.45 U | 16.0 | 5.97 | 9.33 | 9.8 | 5.5 | 34.52 |
| BENZO[B]FLUORANTHENE | UG/KG | 1.7 - 1.7 | 73.33 | 15.43 | 104.0 | 9.43 | 29.5 | 13.83 | 13.33 | 12.33 | 9.7 | 105.0 |
| BENZO[G,H,I]PERYLENE | UG/KG | 1.9 - 1.9 | 27.65 | 9.4 | 36.2 | 1.9 U | 11.95 | 3.4 | 5.83 | 6.03 | 3.33 | 31.5 |
| BENZO[K]FLUORANTHENE | UG/KG | 0.81 - 0.85 | 19.0 | 5.23 | 19.55 | 0.92 | 8.5 | 3.0 | 4:37 | 4.6 | 2.43 | 16.18 |
| CHRYSENE | UG/KG | 1.1 - 1.1 | 65.67 | 13.38 | 25.82 | 1.7 | 10.75 | 5.3 | 7.0 | 6.87 | 5.97 | 34.53 |
| DIBENZ[A,H]ANTHRACENE | UG/KG | 1.9 - 1.9 | 5.13 | 2.6 | 3.58 | 1.9 U | 1.9 U | 1.9 U | 1.9 U | 1.9 U | 1.9 U | 2.78 |
| FLUORANTHENE | UG/KG | 2.8 - 3 | 95.9 | 15.82 | 79.75 | 4.1 | 33.0 | 11.33 | 15.33 | 15.0 | 11.3 | 66.45 |
| FLUORENE | UG/KG | 3.5 - 3.7 | 115.55 | 10.15 | 99.25 | 3.6 U | 21.35 | 8.37 | 6.73 | 10.17 | 12.53 | 84.4 |
| INDENO[1,2,3-CD]PYRENE | UG/KG | 1.7 - 1.7 | 20.6 | 7.45 | 22.35 | 2.13 | 9.15 | 4.07 | 5.67 | 6.77 | 1.7 U | 22.13 |
| NAPHTHALENE | UG/KG | 17 - 17 | 79.0 | 18.75 | 93.0 | 17.0 U | 37.0 | 19.33 | 22.67 | 20.0 | 19.0 | 65.25 |
| PHENANTHRENE | UG/KG | 0.68 - 0.82 | 87.13 | 10.3 | 57.25 | 2.24 | 23.0 | 8.47 | 9.77 | 8.97 | 7.53 | 48.9 |
| PYRENE | UG/KG | 0.83 - 0.87 | 102.5 | 16.1 | 72.5 | 1.93 | 21.5 | 6.27 | 9.43 | 9.13 | 9.53 | 63.17 |
| TOTAL PAH (ND=0)* | UG/KG | | 807 | 129 | 782 | 18.3 | 251 | 83.8 | 112 | 109 | 82 | 665 |
| TOTAL PAH (ND=1/2DL)* | UG/KG | - | 830.56 | 164.11 | 803.05 | 60.79 | 274.98 | 117.55 | 143.16 | 141.93 | 116.20 | 690.03 |
| TOTAL PAH (ND=DL) | UG/KG | - | 854.42 | 198.75 | 824.40 | 103.25 | 299.10 | 151.33 | 174.30 | 174.97 | 150.73 | . 715.20 |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

* = Note that the mean of total PAHs for individual samples is not equivalent to the sum of mean individual PAHs for ND=0 and ND=1/2DL.





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TABLE 3-14MEAN CONCENTRATIONS OF METALS (MG/KG) IN SEDIMENT FROM BALTIMOREHARBOR APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

| | | REACH ID: | Inside Site 104 | Outside Site 104 | Brewerton Eastern Extension | Craighill | Craighill Angle | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester |
|-----------|-------|---------------|--------------------|---------------------|-----------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|------------|
| | SA | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | PLE SIZE (N): | 4 | 4 | 4 | 3 | 2 | 3 | 3 | 3 | -3 | 4 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| ALUMINUM | MG/KG | 5.5 - 127 | 12245.0 | 13425.0 | 26250.0 | 9146.67 | 10725.0 | 12583.33 | 17033.33 | 18766.67 | 18400.0 | 20000.0 |
| ANTIMONY | MG/KG | 0.09 - 0.11 | 3.96 | 0.96 | 1.26 | 0.62 | 1.02 | 0.85 | 1.0 | 1.02 | | 0.79 |
| ARSENIC | MG/KG | 0.2 - 0.22 | 42.38 | 13.38 | 14.03 | 6.67 | 13.55 | 12.87 | 11.63 | 12.67 | 10.63 | 11.85 |
| BERYLLIUM | MG/KG | 0.02 - 0.02 | 1.28 | 1.33 | 1.77 | 0.72 | 1.25 | 1.09 | 1.4 | 1.6 | 1.43 | 1.88 |
| CADMIUM | MG/KG | 0.06 - 0.67 | 15.13 | 8.73 | 0. 62 U | 0.06 U | 0.3 U | 0.08 U | 0.3 U | 0.3 U | 0.61 U | 0.46 U |
| CHROMIUM | MG/KG | 0.39 - 0.45 | 68.3 | 37.52 | 41.92 | 18.1 | 37.25 | 34.13 | 45.77 | 51.73 | 33.83 | 33.05 |
| COPPER | MG/KG | 0.2 - 0.22 | 135.07 | 38:22 | 45.27 | 20.17 | 37.75 | 34.87 | 38.5 | 45.17 | 37.23 | 44.95 |
| IRON | MG/KG | 5.1 - 520 | 58175.0 | 34625.0 | 56425.0 | 27980.0 | 34600.0 | 33533.33 | 33566.67 | 39166.67 | 36166.67 | 42100.0 |
| LEAD | MG/KG | 0.09 - 0.11 | 246.82 | 42.83 | 56.4 | 16.9 | 47.05 | 40.0 | 47.2 | 57.33 | 39.6 | 40.2 |
| MANGANESE | MG/KG | 0.78 - 8 | 894.0 | 2037.5 | 1095.0 | 548.67 | 1050.0 | 1014.0 | 3096.67 | 3756.67 | 1090.0 | 1006.0 |
| MERCURY | MG/KG | 0.04 - 0.05 | 其同于1114 | 0.23 | 0.23 | 0.05 U | 0.17 | 0.2 | 0.24 | 0.26 | 0.15 | 0.18 |
| NICKEL | MG/KG | 0.49 - 0.56 | 36.52 | 40.63 | 56.43 | 19.97 | 36:75 | 31.93 | 42.23 | 50.07 | 39.57 | 53.73 |
| SELENIUM | MG/KG | 0.2 - 0.22 | 4.95 | 1.75 | 1.83 | 0.2 U | 1.4 | 0.8 | 0.92 | 1.01 | 0.87 | 0.69 |
| SILVER | MG/KG | 0.09 - 0.11 | NA | NA | 1.05 | 0.1 U | 0.34 | 0.1 U | 0.19 | 0.27 | 0.74 | 0.48 |
| THALLIUM | MG/KG | 0.08 - 0.55 | 0.16 | 0.15 U | 0.14 | 0.15 | 0.11 | 0.17 | 0.11 | 0.15 | 0.13 | 0.23 |
| ZINC | MG/KG | 1.2 - 1.3 | 640.25 | 228.25 | 279.0 | 99.73 | 233.0 | 183.0 | 235.0 | 260.33 | 198.33 | 206.43 |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NA = not analyzed.

NOTE: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

TABLE 3-15CONCENTRATIONS OF BUTYLTINS* (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR
APPROACH CHANNELS (1998) AND PLACEMENT SITE 104 (1997)

NA

NA

NA

| | | REACH ID: | Inside Site 104 | Outside Site 104 | Brewerston Eastern Extension | Craighill | Craighill Angle | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester |
|--------------|-------|----------------|--------------------|---------------------|------------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|------------|
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab | Grab |
| | SAM | IPLE SIZE (N): | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | _ 1 | 1 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | |
| MONOBUTYLTIN | UG/KG | 2.05 - 2.05 | NA | NA | NA | 2.05 U | 2.05 U | 2.05 U | 2.05 U | 2.05 U | 2.05 U | 2.05 U |
| DIBUTYLTIN | UG/KG | 3.11 - 3.11 | NA | NA | NA | 3.63 | 3.4 | 3.11 U | 5.03 | 3.11 U | 3.11 U | 3.11 U |

7.07

1.88 U

5.05

7.65

3.12

3.68

94.27

TRIBUTYLTIN U = not detected.

* = Butyltin analyzed only at one location within each channel reach.

UG/KG

1.88 - 1.88

NA = not analyzed.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 3-16 MEAN CONCENTRATIONS OF ANALYTES EXCEEDING TELs (1997-1998)

| ANALYTETEL UNITInside VALUEOutside Site 104Eastern Site 104Craighill ExtensionCraighill AngleUpper EntranceCutoff AngleSwan PointARSENICMG/KG7.2442.3813.3814.0313.5512.8711.6312.6710.63CADMIUMMG/KG0.67615.138.73< | Tolchester 11.85 44.95 40.20 0.18 53.73 206.43 |
|--|--|
| ANALYTE UNIT VALUE Site 104 Site 104 Extension Craighill Angle Entrance Range Angle Point ARSENIC MG/KG 7.24 42.38 13.38 14.03 13.55 12.87 11.63 12.67 10.63 CADMIUM MG/KG 0.676 15.13 8.73 | Tolchester 11.85 44.95 40.20 0.18 53.73 206.43 |
| ARSENIC MG/KG 7.24 42.38 13.38 14.03 13.55 12.87 11.63 12.67 10.63 CADMIUM MG/KG 0.676 15.13 8.73 10.63 | 11.85 44.95 40.20 0.18 53.73 206.43 |
| CADMIUM MG/KG 0.676 15.13 8.73 Image: Chromitian constraints Image: Chromitian constrain | 44.95 40.20 0.18 53.73 206.43 |
| CHROMIUM MG/KG 52.3 68.30 Image: Complex constraints | 44.95 40.20 0.18 53.73 206.43 |
| COPPER MG/KG 18.7 135.08 38.23 45.28 20.17 37.75 34.87 38.50 45.17 37.23 | 44.95 40.20 0.18 53.73 206.43 |
| | 40.20 0.18 53.73 206.43 |
| LEAD MG/KG 30.24 246.83 42.83 56.40 47.05 40.0 47.20 57.33 39.60 | 0.18 53.73 206.43 |
| MERCURY MG/KG 0.13 1.14 0.23 0.23 0.18 0.20 0.24 0.26 0.15 | 53.73 206.43 |
| NICKEL MG/KG 15.9 36.53 40.63 56.43 19.97 36.75 31.93 42.23 50.07 39.57 | 206.43 |
| SILVER MG/KG 0.73 1.05 0.74 | 206.43 |
| ZINC MG/KG 124 640.25 228.25 279.0 233.0 183.0 235.0 260.33 198.33 | (1.1 |
| ACENAPHTHENE UG/KG 6.71 16.0 U 16.0 U 73.5 16.0 U 20.5 16.0 U 16.0 U 16.0 U 16.0 U | 65.5 |
| ACENAPHTHYLENE UG/KG 5.87 40.0 35.75 U 37.75 U 36.0U 36.5U 35.67 U 36.67U 36.67 U 36.07 U 36.0 | 36.0 U |
| ANTHRACENE UG/KG 46.85 | |
| BENZ[A]ANTHRACENE UG/KG 74.83 | |
| BENZO[A]PYRENE UG/KG 88.81 | |
| CHRYSENE UG/KG 107.77 | |
| DIBENZ[A,H]ANTHRACENE UG/KG 6.22 | |
| FLUORANTHENE UG/KG 112.82 | |
| FLUORENE UG/KG 21.17 115.55 99.25 21.35 | 84.40 |
| NAPHTHALENE UG/KG 34.57 79.0 93.0 37.0 | 65.25 |
| PHENANTHRENE UG/KG 86.68 87.13 | |
| PYRENE UG/KG 152.66 | |
| TOTAL PAH (ND=0) UG/KG 1684.06 | |
| TOTAL PAH (ND=1/2DL) UG/KG 1684.06 | |
| TOTAL PAH (ND=DL) UG/KG 1684.06 | |
| 4,4'-DDD UG/KG 1.22 1.27 | |
| 4,4'-DDE UG/KG 2.07 | |
| 4,4'-DDT UG/KG 1.19 | |
| CHLORDANE UG/KG 2.26 6.40 U 6.40 U 6.40 U 6.40 U 6.40 U 9.60 U 6.40 U 6.40 U 6.40 U | 6.40 U |
| DIELDRIN UG/KG 0.715 | |
| GAMMA-BHC UG/KG 0.32 0.35 | |
| TOTAL PCB (ND=0) UG/KG 21.55 | |
| TOTAL PCB (ND=1/2DL) UG/KG 21.55 31.27 | |
| TOTAL PCB (ND=DL) UG/KG 21.55 37.36 U 37.36 U 44.94 37.33 37.49 37.56 U | |
| 2-METHYLNAPHTHALENE UG/KG 20.21 77.25 29.0 U 38.5 29.0 U 29.0 U 28.67 U 29.0 U | 75.5 |
| BIS(2-ETHYLHEXYL) PHTHALATE UG/KG 182.16 203.0 212.67 | 235.50 |

U = not detected; mean detection limit exceeds TEL value.

| | | | | | Brewerton | | | | Craighill | | | |
|-----------------------|-------|---------|----------|----------|-----------|-----------|-----------|-----------|-----------|--------|---------------------------------------|------------|
| | | PEL | Inside | Outside | Eastern | | Craighill | Craighill | Upper | Cutoff | Swan | |
| ANALYTE | UNIT | VALUE | Site 104 | Site 104 | Extension | Craighill | Angle | Entrance | Range | Angle | Point | Tolchester |
| ARSENIC | MG/KG | 41.6 | 42.38 | | | | | | | | | |
| CADMIUM | MG/KG | 4.21 | 15.13 | 8.73 | | | | | | | | |
| CHROMIUM | MG/KG | 160.4 | | | | | | | | | | |
| COPPER | MG/KG | 108.2 | 135.08 | | | | | | | | | |
| LEAD | MG/KG | 112.18 | 246.83 | | | | | | | | | |
| MERCURY | MG/KG | 0.696 | 1.14 | | | | | | | | | |
| NICKEL | MG/KG | 42.8 | | | 56.43 | | | | | 50.07 | | 53.73 |
| SILVER | MG/KG | 1.7 | | | | | | | | | | |
| ZINC | MG/KG | 271 | 640.25 | | 279.0 | | | | | | | |
| ACENAPHTHENE | UG/KG | 88.9 | | | | | | | | | | |
| ACENAPHTHYLENE | UG/KG | 127.87 | | | | | | | | | | |
| ANTHRACENE | UG/KG | 245 | | | | | | | | | | |
| BENZ[A]ANTHRACENE | UG/KG | 692.53 | | | | | | | | | | |
| BENZO[A]PYRENE | UG/KG | 763.22 | | | | | | | | | | |
| CHRYSENE | UG/KG | 845.98 | | | | | | | | | | |
| DIBENZ[A,H]ANTHRACENE | UG/KG | 134.61 | | | | | T T | | | | | |
| FLUORANTHENE | UG/KG | 1493.54 | | | | | T | | | | | |
| FLUORENE | UG/KG | 144.35 | | | | | | | | | | |
| NAPHTHALENE | UG/KG | 390.64 | | | | | | | | | | |
| PHENANTHRENE | UG/KG | 543.53 | | | | | | | | | | |
| PYRENE | UG/KG | 1397.6 | | | | | | | | | | |
| TOTAL PAH (ND=0) | UG/KG | 16770.4 | | | | | | | | | | |
| TOTAL PAH (ND=1/2DL) | UG/KG | 16770.4 | | | | | | | | | | |
| TOTAL PAH (ND=DL) | UG/KG | 16770.4 | | | | | | | | | | |
| 4,4'-DDD | UG/KG | 7.81 | | | | | | | | | · · · · · · · · · · · · · · · · · · · | |
| 4,4'-DDE | UG/KG | 374.17 | | | | | | | | | | |
| 4,4'-DDT | UG/KG | 4.77 | | | | | | | | | | |
| CHLORDANE | UG/KG | 4.79 | 6.40 U | 6.40 U | 6.40 U | 6.40 U | 6.40 U | 9.63 U | 6.40 U | 6.40 U | 6.40 U | 6.40 U |
| DIELDRIN | UG/KG | 4.3 | | | | | | | | | | |
| GAMMA-BHC | UG/KG | 0.99 | | | | | | | | | | |
| TOTAL PCB (ND≈0) | UG/KG | 188.79 | | | | | | | | | | |
| TOTAL PCB (ND=1/2DL) | UG/KG | 188.79 | | | | | | | | | | |
| TOTAL PCB (ND=DL) | UG/KG | 188.79 | | | | | | | | | | |
| 2-METHYLNAPHTHALENE | UG/KG | 201.28 | | | | | | | | | | |
| BIS(2-ETHY) | UG/KG | 2646.51 | | | | | | | | | | |

TABLE 3-17 MEAN CONCENTRATIONS OF ANALYTES EXCEEDING PELs (1997-1998)

U = not detected, action limit exceeds PEL value.
TABLE 3-18 FREQUENCY OF DETECTION ^(a) BY ANALYTICAL FRACTION FOR SEDIMENTS FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE SITE 104, AND OUTSIDE SITE 104 (1997 -1998)

| | DETECT | ION FREQ | UENCY | PERCENT DETECT | | | |
|-----------------------------|-------------------------------------|--------------------|---------------------|-------------------------------------|--------------------|---------------------|--|
| ANALYTICAL FRACTION | Approach Channels ^(b) | Inside Site 104 | Outside Site 104 | Approach Channels ^(b) | Inside Site 104 | Outside Site 104 | |
| VOCs | 20/850 | 0/136 | 0/136 | 2.4% | 0% | 0% | |
| SVOCs | 28/1275 | 2/208 | 2/208 | 2.1% | 1% | 1% | |
| Chlorinated Pesticides | 20/550 | 7/88 | 4/88 | 3.6% | 8% | 4.5% | |
| Organophosphorus Pesticides | 0/145 | 0/20 | 0/20 | 0% | 0% | 0% | |
| PCB Aroclors | 27/174 | 3/28 | 2/28 | 15.4% | 11% | 7% | |
| PCB Congeners | 11/182 | 0/25 | 0/25 | 6% | 0% | 0% | |
| PAHs | 282/400 | 55/64 | 43/64 | 71% | 86% | 67.2% | |
| Metals | 337/400 | 57/60 | 56/60 | 84% | 95% | 93% | |
| Butyltins | 9/24 | NT | NT | 37.5% | NT | NT | |
| TOTAL | 734/4000 | 124/629 | 107/629 | 18.4% | 19.7% | 17.0% | |

(a) = total number of detected analytes / total number of analytical tests.

(b) = combined total for all approach channels.

NT = not tested

TABLE 3-19NUMBER OF MEAN CONCENTRATIONS IN TARGET ANALYTE FRACTIONS IN
SEDIMENTS THAT EXCEED TELs (1997-1998)

| ANALYTE | Inside Site 104 | Outside Site 104 | Brewerton Eastern Extension | Craighill | Craighill Angle | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester |
|--------------------------------|--------------------|---------------------|-----------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|------------|
| METALS | 8 | 7 | 7 | 2 | 6 | 6 | 6 | 6 | 7 | 6 |
| PAHs | 5 | 2 | 4 | 2 | 4 | 2 | 2 | 2 | 2 | 4 |
| PESTICIDES | 1 | 1 | 1 | 1 | 1 | 3 | 1 | 1 | 1 | 1 |
| TOTAL PCBs (ND=DL) | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SEMIVOLATILE ORGANIC COMPOUNDS | 0 | 0 | 1 | 2 | 1 | 2 | 1 | 1 | 1 | 2 |
| TOTAL # OF TEL EXCEEDANCES | 14 | 10 | 14 | 7 | 12 | 13 | 10 | 10 | 11 | 13 |



TABLE 3-20 NUMBER OF MEAN CONCENTRATIONS IN TARGET ANALYTE FRACTIONS IN
SEDIMENTS THAT EXCEED PELs (1997-1998)

| ANALYTE | Inside Site 104 | Outside Site 104 | Brewerton Eastern Extension | Craighill | Craighill Angle | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester |
|--------------------------------|--------------------|---------------------|-----------------------------------|-----------|--------------------|-----------------------|-----------------------------|-----------------|---------------|------------|
| METALS | 6 | 1 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| PAHs | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| PESTICIDES | 1 | l | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| PCBs, TOTAL | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SEMIVOLATILE ORGANIC COMPOUNDS | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL # OF PEL EXCEEDANCES | 7 | 2 | 3 | 1 | 1 | 1 | 1 | 2 | 1 | 2 |

TABLE 3-21 CONCENTRATIONS OF INORGANIC NON-METALS (MG/L) IN ELUTRIATES* FROMBALTIMORE HARBOR APPROACH CHANNELS (1998)

| | | | Brewerton Eastern | Craighill | Craighill Entrance / | Craighill Upper Range / | | |
|-------------------------------|------|--------|----------------------|-----------|-------------------------|----------------------------|------------|------------|
| | REA | CH ID: | Extension | Angle | Craighill | Cutoff Angle | Swan Point | Tolchester |
| ANALYTE | UNIT | DL | | | | | | |
| CHROMIUM (HEXAVALENT) | MG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| CYANIDE | MG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| NITROGEN, AMMONIA | MG/L | 0.1 | 6.3 | 8.8 | 1.7 | 5.8 | A | 3.3 |
| NITROGEN, NITRATE AND NITRITE | MG/L | 0.05 | 0.13 | 0.07 | 0.05U | 0.09 | 0.07 | 0.14 |
| NITROGEN, TOTAL KJELDAHL | MG/L | 0.25 | 0.26 | 7.3 | 1.9 | 2.2 | 0.25U | 1.7 |
| OXYGEN DEMAND, BIOCHEMICAL | MG/L | 1 | A 4 | 4.6 | 5.3 | 6.8 | 4.8 | 4.1 |
| OXYGEN DEMAND, CHEMICAL | MG/L | 10 | 99.5 | 124 | 113 | 384 | 102 | 137 |
| PHOSPHORUS, TOTAL | MG/L | 0.05 | 0.09 | 0.38 | 0.21 | 0.24 | 0.31 | 0.05U |
| SULFIDE, TOTAL | MG/L | 1 | 2U | 2U | 2U | 2U | 1U | 1U |
| TOTAL ORGANIC CARBON | MG/L | 1 | 1.8 | 1U | 1U | 1U | 1.2 | 1.6 |

U = not detected.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.



TABLE 3-22 CONCENTRATIONS OF VOLATILE ORGANIC COMPOUNDS (UG/L) IN ELUTRIATES* FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

| | REA | CH ID: | Brewerton Eastern Extension | Craighill | Craighill Entrance / Craighill | Craighill Upper Range / Cutoff Angle | Swan Point | Talchester |
|---------------------------|------|--------|-----------------------------------|-----------|--------------------------------------|--|-------------|----------------------|
| ANALVTE | UNIT | DI | BATCHOION | THE | craight | Cuton Angle | Swan I Unit | Torenester |
| 1 1 1-TRICHLOROFTHANE | UG/L | 2 | 2111 | 211 | 211 | 211 | | 211 |
| 1 1 2 2 TETPACHLOROETHANE | UG/L | 0.0 | 0.011 | 0.011 | 0.011 | 0.011 | 0.011 | 20 |
| 1 1 2-TRICHLOROFTHANE | UG/L | 0.5 | 0.50 | 0.50 | 0.90 | 0.50 | 0.90 | 0.90 |
| LI-DICHLOROFTHANE | UG/L | 0.5 | 0.50 | 0.50 | 0.50 | 0.50 | 0.50 | 0.30 |
| 1 1 DICHLOROETHVI ENE | UG/L | 1 | 111 | 0.50 | 0.80 | 0.00 | 0.80 | 0.80 |
| 1.2 DICHLOROPENZENE | UG/L | 2 | 211 | 211 | 211 | 211 | 10 | 10 |
| 1.2-DICHLOROETHANE | UG/L | 0.4 | 0.411 | 0.411 | 0.411 | 0.411 | 0.411 | 20 |
| 1.2-DICHLOROPPOPANE | UG/L | 0.4 | 0.40 | 0.40 | 0.40 | 0.40 | 0.40 | 0.40 |
| 1 4-DICHLOROBENZENE | UG/L | 3 | 311 | 311 | 311 | 311 | 0.50 | 0.30 |
| 2-CHLOROETHYL VINYL ETHER | UG/L | 2 | 211 | 211 | 211 | 211 | - 30 | 211 |
| ACROLEIN | UG/L | 18 | 18U | 1811 | 1811 | 1811 | 1811 | 1911 |
| ACRYLONITRILE | UG/L | 12 | 12U | 1211 | 120 | 1211 | 1211 | 1211 |
| BENZENE | UG/L | 0.5 | 0.5U | 0.50 | 0.5U | 0.511 | 0.511 | 0.511 |
| BROMODICHLOROMETHANE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.50 |
| BROMOMETHANE | UG/L | 2 | 2U | 2U | 2U | 2U | 2U | 211 |
| CARBON DISULFIDE | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| CARBON TETRACHLORIDE | UG/L | 1 | iU | 1U | 10 | 10 | 10 | 10 |
| CFC-11 | UG/L | 2 | 2U | 2U | 2U | 2U | 2U | 2U |
| CFC-12 | UG/L | 1 | 1U | 1U | 1U | 1U | 1U | 10 |
| CHLOROBENZENE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U |
| CHLORODIBROMOMETHANE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| CHLOROETHANE | UG/L | 2 | 2U | 2U | 2U | 2U | 2U | 2U |
| CHLOROFORM | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| CHLOROMETHANE | UG/L | 1 | 1U | 1U | 1U | 1U | IU | 10 |
| CIS-1,3-DICHLOROPROPENE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U |
| DICHLOROMETHANE | UG/L | 0.7 | | 0.7U | 6 | 0.7U | 8 | 10 - rolling - 1 - 6 |
| ETHYLBENZENE | UG/L | 0.5 | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U |
| M-DICHLOROBENZENE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| METHYLBENZENE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U |
| TETRACHLOROETHENE | UG/L | 0.6 | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U |
| TRANS-1,2-DICHLOROETHENE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| TRANS-1,3-DICHLOROPROPENE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U |
| TRIBOMOMETHANE | UG/L | 0.4 | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U | 0.4U |
| TRICHLOROETHYLENE | UG/L | 0.5 | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U |
| VINYL CHLORIDE | UG/L | 0.9 | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U |

U = not detected.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 3-23CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/L) INELUTRIATES* FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

| | ſ | | Brewerton | | Craighill | Craighill | | |
|------------------------------|------|--------|-----------|-----------|------------|---------------------|------------|------------|
| | | | Eastern | Craighill | Entrance / | Upper Range / | | |
| | REA | CH ID: | Extension | Angle | Craighill | Cutoff Angle | Swan Point | Tolchester |
| ANALYTE | UNIT | DL | | | | | | |
| 1,2,4-TRICHLOROBENZENE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 1,2-DIPHENYLHYDRAZINE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| 1-METHYLNAPHTHALENE | UG/L | 5 | | 5U | 5U | 5U | 5U | 10U |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 2,4,5-TRICHLOROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2,4,6-TRICHLOROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2,4-DICHLOROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2,4-DIMETHYLPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2,4-DINITROPHENOL | UG/L | 6 | 6U | 6U | 6U | 6U | 6U | 6U |
| 2,4-DINITROTOLUENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 2,6-DINITROTOLUENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 2-CHLORONAPHTHALENE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2-CHLOROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2-METHYL-4,6-DINITROPHENOL | UG/L | 6 | 6U | 6U | 6U | 6U | 6U | 6U |
| 2-METHYLNAPHTHALENE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 2-METHYLPHENOL | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 2-NITROANILINE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| 2-NITROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| 3,3'-DICHLOROBENZIDINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 3,4-METHYLPHENOL | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 3-NITROANILINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 4-BROMOPHENYL PHENYL ETHER | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| 4-CHLORO-3-METHYLPHENOL | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 4-CHLOROPHENYL PHENYL ETHER | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| 4-NITROPHENOL | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| BENZIDINE | UG/L | 18 | 18U | 18U | 18U | 18U | 18U | 18U |
| BENZYL ALCOHOL | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | ·4U |
| BENZYL BUTYL PHTHALATE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| BIS(2-CHLOROETHOXY)METHANE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| BIS(2-CHLOROETHYL) ETHER | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |



TABLE 3-23 (CONTINUED)

.

| | | | Brewerton | G | Craighill | Craighill | | |
|---------------------------------|------|--------|------------|-----------|------------|---------------|------------|------------|
| | | | Eastern | Craighill | Entrance / | Upper Range / | | |
| | REA | CH ID: | Extension | Angle | Craighill | Cutoff Angle | Swan Point | Tolchester |
| ANALYTE | UNIT | DL | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 7 | 7U | 7U | 7U | 7U | 7U | 7U |
| CARBAZOLE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| CYCLOHEXANONE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| DI-N-BUTYL PHTHALATE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| DI-N-OCTYL PHTHALATE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| DIBENZOFURAN | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| DIETHYL PHTHALATE | UG/L | 3 | | 3U | 3U | 3U | 3U | 3U |
| DIMETHYL PHTHALATE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3 U |
| HEXACHLORO-1,3-BUTADIENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| HEXACHLOROBENZENE | UG/L | 6 | 6U | 6U | 6U | 6U | 6U | 6U |
| HEXACHLOROCYCLOPENTADIENE | UG/L | 2 | 2 U | 2U | 2U | 2U | 2U | 2U |
| HEXACHLOROETHANE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| METHANAMINE, N-METHYL-N-NITROSO | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| N-NITROSODI-N-PROPYLAMINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| N-NITROSODIPHENYLAMINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| NITROBENZENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| P-CHLOROANILINE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| P-NITROANILINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| PENTACHLOROPHENOL | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| PHENOL | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| PYRIDINE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |

U = not detected.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

| | ſ | | Brewerton | | Craighlll | Cralghill | | |
|--------------------|------|--------|-----------|-----------|------------|---------------|------------|------------|
| | | | Eastern | Cralghlll | Entrance / | Upper Range / | | |
| | REA | CH ID: | Extension | Angle | Craighill | Cutoff Angle | Swan Point | Tolchester |
| ANALYTE | UNIT | DL | | | | | | |
| 4,4'-DDD | UG/L | 0.002 | 0.002U | 0.002U | 0.002U | 0.002U | 0.002U | 0.002U |
| 4,4'-DDE | UG/L | 0.004 | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U |
| 4,4'-DDT | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U |
| ALDRIN | UG/L | 0.009 | 0.00899U | 0.00899U | 0.00899U | 0.00899U | 0.00899U | 0.00899U |
| ALPHA-BHC | UG/L | 0.001 | 0.001U | 0.0034J | 0.0042J | 0.0061J | 0.0051J | 0.0094J |
| BETA-BHC | UG/L | 0.001 | 0.0037B | 0.001UJ | 0.0047B | 0.0025B | 0.0049J | 0.0046B |
| CHLORDANE | UG/L | 0.04 | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U |
| CHLORBENSIDE | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02 | 0.03 |
| DACTHAL | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| DELTA-BHC | UG/L | 0.001 | 0.001U | 0.0011J | 0.001U | 0.001U | 0.001U | 0.001U |
| DIELDRIN | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U |
| ENDOSULFAN I | UG/L | 0.001 | 0.0063J | 0.001UJ | 0.001UJ | 0.001UJ | 0.0077J | 0.0067J |
| ENDOSULFAN II | UG/L | 0.004 | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U |
| ENDOSULFAN SULFATE | UG/L | 0.004 | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U | 0.004U |
| ENDRIN | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U |
| ENDRIN ALDEHYDE | UG/L | 0.006 | 0.006U | 0.006U | 0.006U | 0.006U | 0.006U | 0.006U |
| GAMMA-BHC | UG/L | 0.001 | 0.001U | 0.001U | 0.001U | 0.0027B | 0.001U | 0.01B |
| HEPTACHLOR | UG/L | 0.006 | 0.00869B | 0.01B | 0.01B | 0.02B | 0.02B | 0.02B |
| HEPTACHLOR EPOXIDE | UG/L | 0.001 | 0.0071B | 0.0067B | 0.0027B | 0.0076B | 0.0042B | 0.0031B |
| METHOXYCHLOR | UG/L | 0.009 | 0.00899U | 0.00899U | 0.00899U | 0.00899U | 0.00899U | 0.00899U |
| MIREX | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| TOXAPHENE | UG/L | 0.47 | 0.47U | 0.47U | 0.47U | 0.47U | 0.47U | 0.47U |

TABLE 3-24 CONCENTRATIONS OF CHLORINATED PESTICIDES (UG/L) INELUTRIATES* FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

U = not detected. B = found in blank. J = estimated value.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.



TABLE 3-25CONCENTRATIONS OF ORGANOPHOSPHORUS PESTICIDES (UG/L) INELUTRIATES* FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

| | REA | CH ID: | Brewerton Eastern Extension | Craighill Angle | Craighill Entrance / Craighill | Craighill Upper Range / Cutoff Angle | Swan Point | Tolchester |
|------------------|------|--------|-----------------------------------|--------------------|--------------------------------------|--|------------|------------|
| ANALYTE | UNIT | DL | | | | | | |
| AZINPHOS METHYL | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| DEMETON | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| ETHYL PARATHION | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| MALATHION | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| METHYL PARATHION | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |

 $\mathbf{U} =$ not detected.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 3-26 CONCENTRATIONS OF PCB AROCLORS (UG/L) IN ELUTRIATES* FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

| | REACH ID: | | Brewerton Eastern Extension | Craighill Angle | Craighill Entrance / Craighill | Craighill Upper Range / Cutoff Angle | Swan Point | Tolchester |
|--------------|-----------|------|-----------------------------------|--------------------|--------------------------------------|--|------------|------------|
| ANALYTE | UNIT | DL | | | | | | |
| AROCLOR 1016 | UG/L | 0.12 | 0.12U | 0.12U | 0.12U | 0.12U | 0.12U | 0.12U |
| AROCLOR 1221 | UG/L | 0.12 | 0.12U | 0.12U | 0.12U | 0.12U | 0.12U | 0.12U |
| AROCLOR 1232 | UG/L | 0.15 | 0.15U | 0.15U | 0.15U | 0.15U | 0.15U | 0.15U |
| AROCLOR 1242 | UG/L | 0.11 | 0.11U | 0.11U | 0.11U | 0.11U | 0.11U | 0.11U |
| AROCLOR 1248 | UG/L | 0.11 | 0.11U | 0.11U | 0.11U | 0.11U | 0.11U | 0.11U |
| AROCLOR 1254 | UG/L | 0.14 | 0.14U | 0.14U | 0.14U | 0.14U | 0.14U | 0.14U |
| AROCLOR 1260 | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |

U = not detected. **B** = found in blank. **J** = estimated value.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 3-27CONCENTRATIONS OF PAHs (UG/L) IN ELUTRIATES* FROM BALTIMOREHARBOR APPROACH CHANNELS (1998)

| | | ſ | Brewerton | | Craighill | Craighill | | · · · · · · · · · · · · |
|------------------------|------|--------|------------|-----------|------------|---------------|------------|-------------------------|
| | | | Eastern | Craighill | Entrance / | Upper Range / | | |
| | REA | CH ID: | Extension | Angle | Craighill | Cutoff Angle | Swan Point | Tolchester |
| ANALYTE | UNIT | DL | | | | | | |
| ACENAPHTHENE | UG/L | 3 | 3 U | 3U | 3U | 3U | 3U | 3U |
| ACENAPHTHYLENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| ANTHRACENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| BENZ[A]ANTHRACENE | UG/L | 3 | 3 U | 3U | 3U | 3U | 3U | 3U |
| BENZO[A]PYRENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| BENZO[B]FLUORANTHENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| BENZO[G,H,I]PERYLENE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| BENZO[K]FLUORANTHENE | UG/L | 3 | 3U | 3U | 3 U | 3U | 3U | 3U |
| CHRYSENE | UG/L | 3 | <u> </u> | <u>3U</u> | 3U | 3U | 3U | 3U |
| DIBENZ[A,H]ANTHRACENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| FLUORANTHENE | UG/L | 5 | 5U | 5U | 5Ú | 5U | 5U | 5U |
| FLUORENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| INDENO[1,2,3-CD]PYRENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4U |
| NAPHTHALENE | UG/L | 3 | 3U | 3U | 3U | 3U | 3U | 3U |
| PHENANTHRENE | UG/L | 5 | 5U | 5U | 5U | 5U | 5U | 5U |
| PYRENE | UG/L | 4 | 4U | 4U | 4U | 4U | 4U | 4Ū |
| TOTAL PAH (ND=0)** | UG/L | - | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL PAH (ND=1/2DL)** | UG/L | - | 31 | 31 | 31 | 31 | 31 | 31 |
| TOTAL PAH (ND=DL) | UG/L | - | 62 | 62 | 62 | 62 | 62 | 62 |

U = not detected.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

* = Results from one elutriate sample per reach.

** = Note that the mean of total PAHs for individual samples is not equivalent to the sum of mean individual PAHs for ND=0 and ND=1/2DL.

| | REA | CH ID: | Brewerton Eastern Extension | Craighill Angle | Craighill Entrance / Craighill | Craighill Upper Range / Cutoff Angle | Swan Point | Tolchester |
|-----------|------|--------|-----------------------------------|--------------------|--------------------------------------|--|------------|------------|
| ANALYTE | UNIT | DL | | | | | | |
| ALUMINUM | UG/L | 56 | 145J | 153J | 176J | 168J | 183J | 162J |
| ANTIMONY | UG/L | 1 | 4.2B | 4B | 2.2B | 3.2B | 3.9B | 4B |
| ARSENIC | UG/L | 2 | 13.1 | 21.7 | 7.4J | 17.8 | 23 | 9.8J |
| BERYLLIUM | UG/L | 0.2 | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U |
| CADMIUM | UG/L | 0.6 | 0.6UJ | 0.6UJ | 0.6UJ | 0.6UJ | 0.6UJ | 0.6UJ |
| CHROMIUM | UG/L | 0.7 | 1.9B | 45B | 0.7UE | 0.7UE | 1.8B | 1.7B |
| COPPER | UG/L | 2 | 2U | 2U | 2U | 20 | 2U | 2U |
| IRON | UG/L | 52 | 52U | 1000B | 52U | 684B | 52U | 52U |
| LEAD | UG/L | 1 | 1U | 1U | 10 | 1U | 1U | 1U |
| MANGANESE | UG/L | 8 | 8260J | 7180E | 746J | 6980J | 11200J | 6850J |
| MERCURY | UG/L | 0.1 | 0.1U | 0.82J | 0.97B | 0.12B | 0.1U | 0.1U |
| NICKEL | UG/L | 2 | 2U | 20.2B | 2U. | I 2UJ | 2U | 2UJ |
| SELENIUM | UG/L | 2 | 5.1B | 3.9B | 20 | 2.7B | 7.2B | 3B |
| SILVER | UG/L | 1 | 1.6B | 1U | 10 | 2.1B | 1.6B | 2.1B |
| THALLIUM | UG/L | 1 | 10UJ | 1UJ | 1U. | I IUJ | 20UJ | 10UJ |
| ZINC | UG/L | 12 | 12U | 12U | 120 | 12U | 12U | 12U |

TABLE 3-28CONCENTRATIONS OF METALS (UG/L) IN ELUTRIATES* FROM
BALTIMORE HARBOR APPROACH CHANNELS (1998)

U = not detected. **B** = value <RL but >IDL/MDL. **E** = amount detected is > Method Calibration Limit. **J** = estimated value.

DL = detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 3-29 FREQUENCY OF DETECTION ^{(a)(b)} BY ANALYTICAL FRACTION FOR ELUTRIATES FROM BALTIMORE HARBOR APPROACH CHANNELS (1998)

| ANALYTICAL FRACTION | FREQUENCY | PERCENT DETECT |
|-----------------------------|-----------|-------------------|
| VOCs | 4/210 | 2% |
| SVOCs | 0/324 | 0% |
| Chlorinated Pesticides | 30/132 | 22.7% |
| Organophosphorus Pesticides | 0/30 | 0% |
| PCB Aroclors | 0/42 | 0% |
| PCB Congeners | NT | NT |
| PAHs | 0/90 | 0% |
| Metals | 43/96 | 45% |
| Butyltins | NT | NT |
| TOTAL | 77/924 | 8.3% |

(a) = total number of detected analytes / total number of analytical tests.

(b) = combined total for all approach channels.

NT = not tested

TABLE 3-30ACONCENTRATIONS OF DETECTED ANALYTES IN ELUTRIATES EXCEEDINGUSEPA ACUTE SALTWATER CRITERIA - 1998CHANNEL DATA

| | | EPA | Brewerton | | Craighill | Craighill Upper | | |
|--------------------|------|-----------|-----------|-----------------|------------|-----------------|------------|------------|
| | | SALTWATER | Eastern | | Entrance / | Range / Cutoff | | |
| ANALYTE | UNIT | ACUTE | Extension | Craighill Angle | Craighill | Angle | Swan Point | Tolchester |
| Non-Metals | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | 43 | | | | | | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | а | | | | | | |
| Pesticides | | | | | | | | |
| ALPHA-BHC | UG/L | a | | | | | | |
| BETA-BHC | UG/L | а | | | | | | |
| ENDOSULFAN I | UG/L | 0.034 | | | | | | |
| GAMMA-BHC | UG/L | 0.16 | | | <u>.</u> . | | | |
| HEPTACHLOR | UG/L | 0.053 | | | | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | | | | | | |
| Metals | | | | - | | | | |
| ANTIMONY | UG/L | а | | | | | | |
| ARSENIC | UG/L | 69 | | | | | | |
| CHROMIUM | UG/L | 1100 | | | | | | |
| MANGANESE | UG/L | а | | | | | | - |
| MERCURY | UG/L | 1.8 | | | | | | |
| NICKEL | UG/L | 74 | | | | | | |
| SELENIUM | UG/L | 290 | | | | | | |
| SILVER | UG/L | 1.9 | | | | 2.1 | | 2.1 |

a = no USEPA acute saltwater criterion.

Acute criteria based on 1-hr. average exposure concentrations.



TABLE 3-30BMIXING FACTORS REQUIRED TO COMPLY WITH USEPA ACUTE
SALTWATER CRITERIA - 1998 CHANNEL DATA

| | | EPA | Brewerton | | Craighill | Craighill | | |
|--------------------|------|-----------|-----------|-----------|------------|---------------------|-------|------------|
| | | SALTWATER | Eastern | Craighill | Entrance / | Upper Range / | Swan | |
| ANALYTE | UNIT | ACUTE | Extension | Angle | Craighill | Cutoff Angle | Point | Tolchester |
| Non-Metals | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | 43 | | | | | | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | а | | | | | | |
| Pesticides | | | | | | | | |
| ALPHA-BHC | UG/L | а | | | | | | |
| BETA-BHC | UG/L | а | | | | | | |
| ENDOSULFAN I | UG/L | 0.034 | | | | | | |
| GAMMA-BHC | UG/L | 0.16 | | | | | | |
| HEPTACHLOR | UG/L | 0.053 | | | | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | | | | | | |
| Metals | | | | | | | | |
| ANTIMONY | UG/L | а | | | | | | |
| ARSENIC | UG/L | 69 | | | | | | |
| CHROMIUM | UG/L | 1100 | | | | | | |
| MANGANESE | UG/L | а | | | | | | |
| MERCURY | UG/L | 1.8 | | | | | | |
| NICKEL | UG/L | 74 | | | | | | |
| SELENIUM | UG/L | 290 | | | | | | |
| SILVER | UG/L | 1.9 | | | | 1 | | 1 |

a = no USEPA acute saltwater criterion.

Acute criteria based on 1-hr. average exposure concentrations.

TABLE 3-31ACONCENTRATIONS OF DETECTED ANALYTES IN ELUTRIATE EXCEEDINGUSEPA CHRONIC SALTWATER CRITERIA - 1998CHANNEL DATA

| | | EPA | Brewerton | | Craighill | Craighill Upper | | |
|--------------------|------|-----------|-----------|-----------|------------|-----------------|------------|------------|
| | | SALTWATER | Eastern | Craighill | Entrance / | Range / Cutoff | | |
| ANALYTE | UNIT | CHRONIC | Extension | Angle | Craighill | Angle | Swan Point | Tolchester |
| Non-Metals | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | 6.4 | | 8.8 | | | 7 | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | a | | | | | | |
| Pesticides | | | | | | | | |
| ALPHA-BHC | UG/L | a | | | | | | |
| BETA-BHC | UG/L | a | | | | | | |
| ENDOSULFAN I | UG/L | 0.0087 | | | | | | |
| GAMMA-BHC | UG/L | a | | | | | | |
| HEPTACHLOR | UG/L | 0.0036 | 0.00869 B | 0.01 B | 0.01 B | 0.02 B | 0.02 B | 0.02 B |
| HEPTACHLOR EPOXIDE | UG/L | 0.0036 | 0.0071 B | 0.0067 B | | 0.0076 B | 0.0042 B | |
| Metals | | | | | | | | |
| ANTIMONY | UG/L | a | | | | | | |
| ARSENIC | UG/L | 36 | | | | | | |
| CHROMIUM | UG/L | 50 | | | | | | |
| MANGANESE | UG/L | a | | | | | | |
| MERCURY | UG/L | 0.94 | | | 0.97 | | | |
| NICKEL | UG/L | 8.2 | | 20.2 | | | | |
| SELENIUM | UG/L | 71 | | | | | | |
| SILVER | UG/L | a | | | | | | |

a = no USEPA chronic saltwater criterion.

B= detected in laboratory blank

Chronic criteria based on 4-day average exposure concentrations.



TABLE 3-31BMIXING FACTORS REQUIRED TO COMPLY WITH USEPACHRONIC SALTWATER CRITERIA - 1998 CHANNEL DATA

| | | EPA | Brewerton | | Craighill | Craighill | | |
|--------------------|------|-----------|-----------|-----------|------------|---------------|-------|---------------------------------------|
| | | SALTWATER | Eastern | Craighill | Entrance / | Upper Range / | Swan | |
| ANALYTE | UNIT | CHRONIC | Extension | Angle | Craighill | Cutoff Angle | Point | Tolchester |
| Non-Metals | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | 6.4 | | 1 | | | 1 | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | a | | | | | | |
| Pesticides | | | | | | | | · · · · · · · · · · · · · · · · · · · |
| ALPHA-BHC | UG/L | a | | | | | | |
| BETA-BHC | UG/L | а | | | | | | |
| ENDOSULFAN I | UG/L | 0.0087 | | | | | | |
| GAMMA-BHC | UG/L | а | | | | | | |
| HEPTACHLOR | UG/L | 0.0036 | 3 B | 3 B | 3 B | 6 B | 6 B | 6 B |
| HEPTACHLOR EPOXIDE | UG/L | 0.0036 | 2 B | 2 B | | 2 B | 1 B | |
| Metals | | | | | | | | |
| ANTIMONY | UG/L | а | | | | | | |
| ARSENIC | UG/L | 36 | | | | | | |
| CHROMIUM | UG/L | 50 | | | | | | |
| MANGANESE | UG/L | а | | | | | | |
| MERCURY | UG/L | 0.94 | | | 1 | | | |
| NICKEL | UG/L | 8.2 | | 2 | | | | |
| SELENIUM | UG/L | 71 | | | | | | |
| SILVER | UG/L | a | | | | | | ſ |

a = no USEPA chronic saltwater criterion.

B = constituent dected in laboratory blank.

Chronic criteria based on 4-day average exposure concentrations.

TABLE 3-32A MEAN CONCENTRATIONS OF DETECTED ANALYTES IN ELUTRIATE EXCEEDING USEPA HUMAN HEALTH CRITERIA -1998 CHANNEL DATA

| | | EPA | Brewerton | | Craighill | Craighill Upper | | |
|--------------------|------|--------|-----------|-----------|------------|-----------------|----------|------------|
| | | HUMAN | Eastern | Craighill | Entrance / | Range / Cutoff | Swan | |
| ANALYTE | UNIT | HEALTH | Extension | Angle | Craighill | Angle | Point | Tolchester |
| Non-Metal | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | а | | | | | | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | 16000 | | | | | | |
| Pesticides | | | | | | | | |
| ALPHA-BHC | UG/L | | | | | | | |
| BETA-BHC | UG/L | | | | | | | |
| ENDOSULFAN I | UG/L | | | | | | | |
| GAMMA-BHC | UG/L | | | | | | | |
| HEPTACHLOR | UG/L | 0.0021 | 0.00869 B | 0.01 B | 0.01B | 0.02 B | 0.02 B | 0.02 B |
| HEPTACHLOR EPOXIDE | UG/L | 0.0011 | 0.0071 B | 0.0067 B | 0.0027 B | 0.0076 B | 0.0042 B | 0.0031 B |
| Metals | | | | | | | | |
| ANTIMONY | UG/L | 4300 | | | | | | |
| ARSENIC | UG/L | 1.4 | 13.1 | 21.7 | 7.4 | 17.8 | 23 | 9.8 |
| CHROMIUM | UG/L | а | | | | | | |
| MANGANESE | UG/L | 100 | 8260 | 7180 | 746 | 6980 | 11200 | 6850 |
| MERCURY | UG/L | 0.051 | | 0.82 | 0.97 | 0.12 | | |
| NICKEL | UG/L | 4600 | | 20.2 | | | | |
| SELENIUM | UG/L | а | | | | | | |
| SILVER | UG/L | а | | | | | | |

a = no USEPA human health criterion.

B = detected in laboratory blank

Human health criteria based on daily lifetime (70-year) average consumption of aquatic organisms. Comparisons to criteria for detected concentrations only.



TABLE 3-32BMIXING FACTORS REQUIRED TO COMPLY WITH USEPA HUMAN HEALTH
CRITERIA - 1998 CHANNEL DATA

| | | | Brewerton | | Craighill | Craighill Upper | | |
|--------------------|------|-----------|-----------|-----------|------------|-----------------|------------|------------|
| | | EPA HUMAN | Eastern | Craighill | Entrance / | Range / Cutoff | | |
| ANALYTE | UNIT | HEALTH | Extension | Angle | Craighill | Angle | Swan Point | Tolchester |
| Non-Metal | | | | | | | | <u> </u> |
| NITROGEN, AMMONIA | MG/L | а | | | | | | |
| VOCs | | | | | | | | |
| DICHLOROMETHANE | UG/L | 16000 | | | | | | |
| Pesticides | | | | | | | | |
| ALPHA-BHC | UG/L | 0.13 | | | | | | |
| BETA-BHC | UG/L | 0.46 | | | | | | |
| ENDOSULFAN I | UG/L | 240 | | | | | | |
| GAMMA-BHC | UG/L | 0.63 | | | | | | |
| HEPTACHLOR | UG/L | 0.0021 | 4 B | 5 B | 5 B | 10 B | 10 B | 10 B |
| HEPTACHLOR EPOXIDE | UG/L | 0.0011 | 6 B | 6 B | 2 B | 7 B | 4 B | 3 B |
| Metals | | | | | | | | |
| ANTIMONY | UG/L | 4300 | | | | | | |
| ARSENIC | UG/L | 1.4 | 9 | 16 | 5 | 13 | 16 | 7 |
| CHROMIUM | UG/L | а | | | | | | |
| MANGANESE | UG/L | 100 | 83 | 72 | 7 | 70 | 112 | 69 |
| MERCURY | UG/L | 0.051 | | 16 | 19 | 2 | | |
| NICKEL | UG/L | 4600 | | | | | | |
| SELENIUM | UG/L | а | | | | | | |
| SILVER | UG/L | а | | | | | | |

a = no USEPA human health criterion.

B= constituent detected in laboratory blank.

Human health criteria based on daily lifetime (70-year) average consumption of aquatic organisms.

TABLE 3-33 NUMBER OF CONSTITUENTS EXCEEDING APPLICABLE WQC*- 1998 ELUTRIATE DATA

| WATER QUALITY CRITERIA | Brewerton Eastern Extension | Craighill Angle | Craighill Entrance / Craighill | Craighill Upper Range / Cutoff Angle | Swan Point | Tolchester |
|--|-----------------------------------|--------------------|--------------------------------------|--|------------|------------|
| EPA ACUTE EXCEEDANCES | 0 | 0 | 0 | 1 | 0 | 1 |
| EPA CHRONIC EXCEEDANCES ^(a) | 0 | 2 | 1 | 2 | 1 | 0 |
| EPA HUMAN HEALTH EXCEEDANCES ^(a) | 2 | 3 | 3 | 5 | 2 | 2 |
| TOTAL # OF WQC EXCEEDANCES* | 2 | 5 | 4 | 8 | 3 | 3 |
| MAXIMUM MIXING FACTOR (MF) REQUIRED TO MEET ALL APPLICABLE WQC | 83 | 72 | 19 | 70 | 112 | 69 |
| CONSTITUENT REQUIRING MAXIMUM MF | Manganese | Manganese | Mercury | Manganese | Manganese | Manganese |

(a) constituents detected in laboratory blanks not included in exceedance count

* Exceedances based on 100% elutriate without consideration of mixing (see Tables 3-30B, 3-31B, and 3-32B for mixing factors)

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CHAPTER 5 -ANALYTICAL TESTING METHODS

5. ANALYTICAL TESTING OF BULK SEDIMENT, WATER, AND TISSUE

Analytical testing of sediment, water, elutriates, and tissue was conducted by STL-Baltimore (formerly EA Laboratories) located in Sparks, Maryland. Additional analytical support was provided by EBA Engineering (grain size and Atterberg Limits); EA's Ecotoxicology Laboratory (moisture content in tissue); STL-Burlington, Vermont (butyltins); and Paradiagm Analytical Laboratory (PAL) (dioxin and furan congeners).

Sediments, receiving water, and elutriates were tested for the following chemical fractions:

- volatile organic compounds (VOCs),
- semivolatile organic compounds (SVOCs),
- chlorinated and organophosphorus pesticides,
- polychlorinated biphenyl (PCB) congeners,
- polynuclear aromatic hydrocarbons (PAHs),
- metals,
- butyltins,
- dioxin and furan congeners, and
- cyanide.

In addition, sediments, receiving water, and elutriates were tested for the following nutrient and general chemistry parameter:

- ammonia
- nitrate+nitrite
- total Kjeldahl nitrogen (TKN)
- total phosphorus (TP),
- total organic carbon (TOC),
- total sulfide
- chemical oxygen demand (COD), and
- biological oxygen demand (BOD).

The following physical and general chemistry analyses were conducted for bulk sediments:

- grain size,
- Atterberg Limits,
- specific gravity,
- moisture content, and
- simultaneously extracted metals (SEM) / acid volatile sulfides (AVS).

In the first round of sampling (September 1999), bulk sediments collected from each station were tested for all of the above referenced fractions with the exception of VOCs and dioxin and



furan congeners. In the second round of sampling (December 1999), VOCs were tested in sediment collected from each station; and dioxin and furan congeners, TOC, moisture content, and grain size determinations were conducted for reach composites. A sample-by-sample breakdown of the analytical testing components for sediment is provided in Table 5-1A.

Elutriates were tested from each reach in both rounds of sampling. Elutriates were tested for all chemical parameters in both rounds of sampling with the exception of dioxin and furan congeners, which were tested only in the second round. Receiving waters from both Inside Site 104 and Outside Site 104 were also collected and tested for chemical constituents. A sample-by-sample breakdown of analytical testing for receiving water and elutriate samples is provided in Table 5-1B.

Target fractions for tissue analysis were determined in conjunction with USEPA Region III and USACE–Baltimore District following review of bulk sediment data and are discussed in detail in Chapter 9.

Target chemical analytes, target detection limits (TDLs), analytical methods, elutriate preparation procedures, and sample holding times were derived from the following guidance documents:

- USEPA/USACE, 1998 (EPA-823-B-98-004). Evaluation of Dredged Material Proposed for Discharge in Waters of the U.S.–Testing Manual (Inland Testing Manual–ITM).
- USEPA/USACE, 1995 (EPA-823-B-95-001). QA/QC Guidance for Sampling and Analysis of Sediments, Water, and Tissues for Dredged Material Evaluations.
- USEPA/USACE, 1991 (EPA-503/8-91/001). Evaluation of Dredged Material Proposed for Ocean Disposal, Testing Manual (The Green Book).

The analytical program for this project is described in detail in the Analytical Quality Assurance Project Plan (QAPP) (EA and STL 2000) (Appendix B). The QAPP was reviewed and approved by USACE–Baltimore District prior to initiation of the analytical testing program. A sample-bysample breakdown of analytical testing is provided in Tables 5-1A and 5-1B. Required sample containers, preservation techniques, and holding time requirements are provided in Chapter 4 (Tables 4-8 and 4-9 for aqueous and sediment samples, respectively) and Chapter 9 (Table 9-7 for tissues). Key components of the testing program are outlined in the following sections.

5.1 ANALYTICAL METHODS

Inorganic and organic compounds for this project were determined using the methods listed in Table 5-2. To meet program-specific regulatory requirements for chemicals of concern, methods are followed as stated with exceptions noted below:

5.1.1 PCB Congeners

PCBs for this project were analyzed and quantified as individual congeners by SW846 Method 8082. Tables 5-3 (sediment), 5-4 (aqueous/elutriate), and 5-5 (tissue) provide a list of the 26 congeners that were tested for in the various matrices. These 26 congeners include all of the "summation" and "highest priority" congeners, plus several of the "secondary priority" congeners, specified in Table 9-3 of the ITM (USEPA/USACE 1998).

5.1.2 Semivolatile Organics and PAHs

In order to achieve the TDLs referenced in QA/QC Guidance for Sampling and Analysis of Sediments, Water, and Tissues for Dredged Material Evaluations- Chemical Evaluations [EPA 823-B-95-001, April 1995] for sediment samples, PAHs in sediments were determined utilizing the SW846 Methods 3540C/8310 (HPLC).

5.1.3 Metals

Metals were determined utilizing Inductively Coupled Plasma (ICP) according to the methodology specified, including the use of TRACE ICP, with the following exceptions:

- For thallium, samples were analyzed by Graphite Furnace Atomic Absorption (GFAA) method (SW846 7841)
- For mercury, samples were analyzed by Cold Vapor Atomic Absorption (CVAA) method (SW846 7470A (aqueous) or 7471A (soil)). Preparation of samples for mercury analyses was modified to use autoclave digestion procedures in place of water bath according to STL-M-7470/1.

5.1.4 Dioxin and Furan Congeners

Dioxin and furan congeners were determined using method SW8290 for sediment, water, and elutriate matrices. USEPA method 1613 was used for dioxin and furan congener determination in the tissue samples. USEPA method 1613 requires more internal standards than SW8290 and is the recommended method in the ITM for dioxin and furan analysis in tissues. In general, tissue matrices are difficult to analyze due to lipid interferences and other complexities associated with the tissue media.

5.2 DETECTION LIMITS

The detection limit is a statistical concept that corresponds to the minimum concentration of an analyte above which the net analyte signal can be distinguished with a specified probability from the signal due to the noise inherent in the analytical system. The method detection limit (MDL) was developed by the USEPA, and is defined as "the minimum concentration of a substance that

can be measured and reported with 99% confidence that the analyte concentration is greater than zero" (40 CFR 136, Appendix B). Detection limits applicable to this project are listed in Tables 5-3, 5-4, and 5-5 for sediment samples, aqueous/elutriate samples, and tissue samples, respectively. These tables include the TDLs referenced in the ITM, and the laboratory MDLs for each parameter by matrix.

The sample quantitation limit (SQL), as specified by USACE, is the analyte MDL adjusted for any method modifications which will allow the laboratory to effectively achieve the TDLs, dilutions, and percent moisture. Analytical results for this project are reported to the SQL.

Chemical concentrations in sediments are reported on a dry weight basis. In order to achieve the lowest possible reporting limit (RL) for sediment samples, which characteristically have moisture contents in excess of 20%, the sample weight taken for analysis was adjusted for the percent moisture in the sample (up to 50% moisture), prior to analysis.

5.3 LABORATORY QUALITY CONTROL SAMPLES

Quality control samples specified in the ITM were analyzed at the frequency stated below for each matrix. Standard Reference Materials (SRMs) were obtained from National Institute of Standards and Technology (NIST) or a comparable source, if available. Acceptance criteria for laboratory quality control samples are listed in Appendix A of the Analytical QAPP (EA and STL 2000) (Appendix B).

| QC Sample | Frequency |
|-------------------------------------|--|
| Standard Reference Material | 1 per analytical batch of 1-20 samples, where available |
| Method Blanks | 1 per analytical batch of 1-20 samples |
| Laboratory Control Sample | 1 per analytical batch of 1-20 samples |
| Surrogates | Spiked into all field and QC samples (Organic Analyses) |
| Sample Duplicates | 1 per analytical batch of 1-20 samples (Inorganic Analyses) |
| Matrix Spike/Matrix Spike Duplicate | 1 per analytical batch of 1-20 samples |

5.3.1 Standard Reference Material

Standard Reference Materials (SRM) represent performance-based QA/QC. A standard reference material is a soil/solution with a certified concentration that is analyzed as a sample and is used to monitor analytical accuracy. SRMs (if available) were analyzed for the following matrix/fractions:

- Sediment: Pesticides, PCB Congeners, Metals, PAHs, and Dioxin/Furan Congeners
- Water: Pesticides, PCB Congeners, and PAHs
- Tissue: Metals and Dioxin/Furan Congeners

Control criteria applied only to those analytes having SRM true values ≥ 10 times the MDL established for the method.

5.3.2 Method Blanks

The method (reagent) blank (MB) is used to monitor laboratory contamination. This is usually a sample of laboratory reagent water processed through the same analytical procedure as the sample (i.e., digested, extracted, distilled). Method blanks were analyzed at a frequency of one every analytical preparation batch of twenty (20) or fewer samples.

5.3.3 Laboratory Control Sample

The Laboratory Control Sample (LCS) is a fortified method blank consisting of reagent water or solid fortified with the analytes of interest for single-analyte methods and selected analytes for multi-analyte methods according to the appropriate analytical method. LCS were prepared and analyzed with each analytical batch, and analyte recoveries were used to monitor analytical accuracy and precision.

5.3.4 Matrix Spike (MS) / Matrix Spike Duplicate (MSD)

A fortified sample (matrix spike) is an aliquot of a field sample that is fortified with the analyte(s) of interest and analyzed to monitor matrix effects associated with a particular sample. Samples to be spiked were chosen at random. The final spiked concentration of each analyte in the sample was at least ten times the calculated MDL. A matrix spike and a duplicate-fortified sample (matrix spike duplicate) was performed for every batch of twenty (20) or fewer samples.

5.3.5 Laboratory Duplicates

A laboratory duplicate is a second aliquot of a field sample that is analyzed to monitor analytical precision associated with that particular sample. Laboratory duplicates were performed for only analytes for which MS/MSD analyses were not appropriate (i.e., nutrients and physical chemistry). One sample was analyzed in duplicate for every batch of twenty (20) or fewer samples.





5.3.6 Surrogates

Surrogates are organic compounds that are similar to, but not the same as, analytes of interest in chemical composition, extraction, and chromatography, but are not normally found in environmental samples. These compounds were spiked into all blank, standards, samples, and spiked samples prior to analysis for organic parameters. Generally, surrogates are not used for inorganic analyses. Percent recoveries were calculated for each surrogate. Surrogates were spiked into samples according to the requirements of the reference analytical method (Section 7 of the Analytical QAPP, EA and STL 2000). Surrogate spike recoveries were evaluated against the limits provided in Appendix A of Analytical QAPP, (EA and STL 2000), and were used to assess method performance and sample measurement bias. If sample dilution caused the surrogate concentration to fall below the quantitation limit, surrogate recoveries were not calculated.

5.3.7 Trip Blanks

Trip blanks (also called transport blanks) were analyzed to evaluate the effect of ambient site conditions and sample transport on sample integrity and to ensure proper container preparation and handling techniques. Trip blanks are samples that originate as analyte-free water placed in organic vials (preserved with HCl) in the laboratory and analyzed for VOCs. One trip blank was analyzed per sampling day or sample processing day. All volatiles samples were stored in the same cooler as the trip blank. Analytical results for trip blanks are provided in Attachment III.

5.4 ANALYTICAL DATA VALIDATION

Data validation was conducted for dioxin and furan congeners in sediment, water, elutriate, and tissue matrices. Metals, pesticides, PCB congeners, and sulfide fractions were validated in the receiving site water and elutriate samples (for both 1998 and 1999 data). Organic chemical fractions (SVOCs, pesticides, PCB congeners, and PAHs) were validated in the tissue samples.

Validation of the analytical data was conducted by Environmental Data Services, Inc. (EDS) located in Concord, New Hampshire. The data validation protocols were derived from the following USEPA guidelines allowing for the quality control requirements specific to the methods used for this project:

- USEPA, Region III. 1999. *Standard operating procedure for dioxin/furan data validation*. Region III Central Regional Laboratory, QA Branch. Annapolis, MD.
- USEPA, Region III. 1995. Innovative approaches to data validation. Region III Central Regional Laboratory, QA Branch. Annapolis, MD.
- USEPA, Region III. 1994. Region III modifications to national functional guidelines for organic data review: multi-media, multi-concentration (OLM01.0-OLM01.9). Region III Central Regional Laboratory, QA Branch. Annapolis, MD

• USEPA, Region III. 1993. *Region III modifications to the laboratory data validation functional for evaluating inorganics analysis*. Region III Central Regional Laboratory, QA Branch. Annapolis, MD.

Analytical data were reported at a Level IV data validation which requires checking 10% of the raw data (i.e., calculations, concentrations of analytes, detection limits, % RSD, % D, % R values, etc.).

The data validation reports (DVRs) for each Sample Delivery Group (SDG) are provided in Attachment IV. Each report contains the following information:

- An overview and summary of the DVR that includes findings by analyses type and a report content statement;
- Copies of US EPA Form I's and/or equivalents;
- A report for each parameter group for the SDG including an introduction, full sample IDs, and technical review comments for each required performance criterion with the actions taken; and
- Data limitations including data usability statements.

| Someting Baseh | Receiving Water / | | | Chlorinated & Organophosphorus | PCB | | | | Dioxins & | | Nitrate | | Total | | | | | |
|-----------------------------|------------------------|----------|----------|-----------------------------------|---------------------|--------------|----------|-----------|-----------------------|----------|-----------|----------------------|----------|------------|----------|----------|-------------|----------|
| Sampling Keach | Elutriate Sample ID | VUCs | SVOCs | Pesticides | Congeners | PAHs | Metals | Butyltins | Furans ^(a) | Ammonia | + Nitrite | Cyanide | Sulfide | TKN | TOC | ТР | COD | BOD |
| Brewerton Eastern Extension | BE-EL | <u> </u> | X | X | X | X | X | X | X | X | X | Х | X | Х | Х | X | X | X |
| C&D Approach Channels | CD-EL | X | X | X | Х | X | X | X | X | X | X | x | X | X | x | X | X | X |
| | CD-VC-EL | <u> </u> | X | <u>X</u> | X | X | X | X | X | X | X | X | X | X | X | X | X | X |
| Craighill | CR-EL | X | X | X | X | X | X | X | X | X | X | x | x | x | X | x | | x |
| Craighill Angle-East | CRA-E-VC-EL | X | X | X | X | Х | X | X | X | x | x | x | x | x | X | x | | |
| Craighill Angle-West | CRA-W-VC-EL | X | X | X | X | X | X | X | x | X | x | $-\frac{\Lambda}{Y}$ | X I | X | X Y | - X | | - X |
| Craighill Entrance | CRE-VC-EL | X | X | X | x | X | X | X | X | x | Y | | - A | - A Y | v A | | | |
| Craighill Upper Range | CRU-EL | X | X | X | x | X | x | X | x | x | Y Y | × × | × I | × A | | X X | | |
| Cutoff Angle | CUT-EL | X | X | X | x | X | x | x | X | Y | x X | | | | | | | |
| Inside Placement Site 104 | KI-EL | Х | X | X | x | <u> </u> | x | x | Y | - A Y | v l | | | | | | | |
| | KI-3WAT ^(b) | | x | x | Y · | v | v | v | | | | <u> </u> | | - <u>^</u> | <u>^</u> | <u> </u> | | <u> </u> |
| | KI-7-FI (b) | | v | v | X | | <u>A</u> | | A | A | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u>X</u> | | <u> </u> |
| | KI-7WAT | v | | <u>x</u> | | - <u>X</u> - | X | <u> </u> | <u> </u> | <u>X</u> | <u>X</u> | X | <u> </u> | <u> </u> | <u>X</u> | X | <u> </u> | <u> </u> |
| | KI-7-WATED | X | X | X | $-\hat{\mathbf{x}}$ | - A V | X V | A V | X | <u> </u> | <u>X</u> | <u>X</u> | <u>X</u> | X | <u> </u> | X | <u>X</u> | <u> </u> |
| Outside Placement Site 104 | KI JAWAT(C) | | | A | <u>A</u> | <u>^</u> | | | <u> </u> | <u> </u> | X | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | <u> </u> | <u> </u> |
| Outside Placement Sile 104 | | X | <u> </u> | <u>X</u> | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | X | X | X | X | X | <u> </u> | X | Х | X | X |
| | KI-OUT-EL | X | <u>X</u> | <u>X</u> | X | X | _X | X | X | X | Х | x | x | x | x | x | x | x |
| Swan Point | SWP-VC-EL | X | X | X | X | X | X | X | X | X | X | X | X | X | X | X | X | X |
| Tolchester Channel - North | TLC-N-VC-EL | X | X | X | X | X | X | X | X | X | XI | X | X | X | X | x | | X |
| Tolchester Channel - South | TLC-S-VC-EL | X | X | X | X | X | X | X | X | X | X | | X | x | X | X | | |
| Tolchester Straightening | TLS-VC-EL | X | X | X | X | X | X | X | X | X | X | X | x | X | x | X | X | |

TABLE 5-1B SAMPLE-BY-SAMPLE SUMMARY OF ANALYTICAL TESTING FOR RECEIVING WATER AND ELUTRIATES

(a) = fractions tested for December sampling only.(b) = tested in September only.

(c) = tested in December only.

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TABLE 5-1A SAMPLE-BY-SAMPLE SUMMARY OF ANALYTICAL TESTING FOR SEDIMENT

.

| Sample Red Sample | | | | | Chlorinated & Organophosphorus | РСВ | | | | Dioxins & | <u> </u> | Nitrate | | Total | 1 | | | SEM (| | | Carrie | Attanham | Specific | Moistur |
|---|-----------------------------|---------------------------|---------------------|----------------------|-----------------------------------|-----------|--------------|----------|------------------|-------------------------|----------------|----------------------|----------------------|----------------------|-----------------------|------------------------|---------------------------------------|---------------------|----------------------|---------------------|----------------------|---------------------------|---------------------|---------------------|
| Beyeder Each Anne Parie | Sampling Reach | Sediment Sample ID | VOCs ^(a) | SVOCs | Pesticides | Congeners | PAHs | Metals | Butyltins | Furans ^(a,b) | Ammonia | + Nitrite | Cvanide | Sulfide | TKN | TOC | тр | AVS | COD | BOD | Grain | Limits | Gravity | Conten |
| Billion x </td <td>Brewerton Eastern Extension</td> <td>BEISED</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>Х</td> <td>X</td> <td>X</td> <td></td> <td>x</td> <td>X</td> <td>X</td> <td>V</td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td>Size</td> <td></td> <td>V</td> <td>Conten</td> | Brewerton Eastern Extension | BEISED | X | X | X | X | Х | X | X | | x | X | X | V | | | | | | | Size | | V | Conten |
| Bisson Bisson Discontrol X <td></td> <td>BE2SED</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>Х</td> <td>X</td> <td>X</td> <td></td> <td>X</td> <td>X</td> <td>x</td> <td>X</td> <td></td> <td><u> </u></td> <td><u>+</u></td> <td>├</td> <td>$-\frac{\Lambda}{v}$</td> <td>· ^</td> <td>÷ ÷</td> <td><u>↓ Ŷ</u></td> <td>+</td> <td><u>+</u></td> | | BE2SED | X | X | X | X | Х | X | X | | X | X | x | X | | <u> </u> | <u>+</u> | ├ | $-\frac{\Lambda}{v}$ | · ^ | ÷ ÷ | <u>↓ Ŷ</u> | + | <u>+</u> |
| BBSODD X <td></td> <td>BE3SED</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td></td> <td>X</td> <td>X</td> <td>x</td> <td>$\frac{\Lambda}{X}$</td> <td><u>x</u></td> <td>÷</td> <td>$\frac{\hat{\mathbf{x}}}{\mathbf{x}}$</td> <td>├─Ŷ</td> <td></td> <td>$\frac{\Lambda}{Y}$</td> <td>- A</td> <td></td> <td>$\frac{\Lambda}{Y}$</td> <td>+</td> | | BE3SED | X | X | X | X | X | X | X | | X | X | x | $\frac{\Lambda}{X}$ | <u>x</u> | ÷ | $\frac{\hat{\mathbf{x}}}{\mathbf{x}}$ | ├ ─ Ŷ | | $\frac{\Lambda}{Y}$ | - A | | $\frac{\Lambda}{Y}$ | + |
| Biology X </td <td></td> <td>BE3SEDFD</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td></td> <td>X</td> <td>X</td> <td>x</td> <td></td> <td>$\frac{\Lambda}{X}$</td> <td>Y</td> <td>X</td> <td>- Â</td> <td>Ŷ</td> <td>Y X</td> <td>$\frac{2}{Y}$</td> <td>X Y</td> <td><u> </u></td> <td>$\frac{\Lambda}{Y}$</td> | | BE3SEDFD | X | X | X | X | X | X | X | | X | X | x | | $\frac{\Lambda}{X}$ | Y | X | - Â | Ŷ | Y X | $\frac{2}{Y}$ | X Y | <u> </u> | $\frac{\Lambda}{Y}$ |
| BECOMP ² Low Low <thlow< th=""> Low <thlow< th=""> <thlow<< td=""><td></td><td>BE4SED</td><td><u> </u></td><td><u> </u></td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td></td><td>X</td><td>X</td><td>X</td><td>x</td><td>X</td><td>X X</td><td>X</td><td>-<u>^</u></td><td>Y Y</td><td>Y X</td><td>X X</td><td>Î X</td><td>X</td><td></td></thlow<<></thlow<></thlow<> | | BE4SED | <u> </u> | <u> </u> | X | X | X | X | X | | X | X | X | x | X | X X | X | - <u>^</u> | Y Y | Y X | X X | Î X | X | |
| CAD Agreent Chaices Cad Agreent Chaices X | | BE-COMP ^(a) | | | | | | | | x | | | | <u> </u> | | | <u> </u> | <u>^</u> | <u> </u> | <u> </u> | | <u> </u> | | |
| Coulding X X X X </td <td>C&D Approach Channels</td> <td>CD003SED</td> <td>X</td> <td>X</td> <td>X</td> <td>x</td> <td>Х</td> <td>x</td> <td>x</td> <td></td> <td>X</td> <td>Y</td> <td>V</td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td></td> <td><u> </u></td> <td></td> <td><u> </u></td> | C&D Approach Channels | CD003SED | X | X | X | x | Х | x | x | | X | Y | V | | | | | | | | | <u> </u> | | <u> </u> |
| Boolessing X | 1 | CD004SED | X | X | X | х | X | X | x | | <u>x</u> | <u> </u> | | $-\frac{\lambda}{v}$ | | X | | <u>X</u> | <u> </u> | <u>X</u> | <u> </u> | | | |
| EDOSSID X </td <td></td> <td>CD005SED</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>Х</td> <td>X</td> <td>X</td> <td></td> <td>x</td> <td><u>x</u></td> <td>X X</td> <td>- Ŷ</td> <td>$-\hat{\mathbf{v}}$</td> <td>-</td> <td>- Â-</td> <td>$-\frac{1}{v}$</td> <td>× ×</td> <td></td> <td></td> <td></td> <td></td> <td>-</td> | | CD005SED | X | X | X | X | Х | X | X | | x | <u>x</u> | X X | - Ŷ | $-\hat{\mathbf{v}}$ | - | - Â- | $-\frac{1}{v}$ | × × | | | | | - |
| DD.01 V(SBD) X <t< td=""><td>1</td><td>CD006SED</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td></td><td>X</td><td>X</td><td>-<u>^</u></td><td>$\frac{\Lambda}{X}$</td><td>- A Y</td><td>Ŷ</td><td>- Â</td><td></td><td>×</td><td>- Â</td><td>- Â</td><td>× · · ·</td><td>Ŷ</td><td><u>+</u></td></t<> | 1 | CD006SED | X | X | X | X | X | X | X | | X | X | - <u>^</u> | $\frac{\Lambda}{X}$ | - A Y | Ŷ | - Â | | × | - Â | - Â | × · · · | Ŷ | <u>+</u> |
| DDD/VSED X< | | CD-001VCSED | X | X | X | X | X | X | X | | X | X | X | X | X | Y Y | X | · A Y | - Â | A Y | ├ ─ र | + | - Ŷ | $\frac{1}{v}$ |
| Craghil L X </td <td></td> <td>CD-002VCSED</td> <td><u> </u></td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>Х</td> <td>X</td> <td></td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td><u> </u></td> <td></td> <td>$-\hat{\mathbf{x}}$</td> <td>X X</td> <td><u>^</u></td> <td>- Â</td> <td>t</td> <td>x x</td> <td>Ŷ</td> | | CD-002VCSED | <u> </u> | X | X | X | X | Х | X | | X | X | X | X | X | <u> </u> | | $-\hat{\mathbf{x}}$ | X X | <u>^</u> | - Â | t | x x | Ŷ |
| Craghil QLSED X <th< td=""><td></td><td>CD-COMP^(a)</td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td>x</td><td></td><td></td><td></td><td></td><td><u>^</u></td><td><u>^</u></td><td></td><td><u> </u></td><td>~^</td><td><u>^</u></td><td><u></u></td><td></td><td><u> </u></td><td></td></th<> | | CD-COMP ^(a) | | | | | | | | x | | | | | <u>^</u> | <u>^</u> | | <u> </u> | ~^ | <u>^</u> | <u></u> | | <u> </u> | |
| Image: First product of the second | Craighill | CRISED | X | X | x | x | x | x | Y | | V I | v | v | | | <u> </u> | | | | | <u>X</u> | | | |
| Craghil Angle Fast X | | CR2SED | X | Х | X | <u> </u> | X | x | X | ······ | Ŷ | - <u>^</u> | | X | <u> </u> | X | <u>X</u> | <u>X</u> | <u>X</u> | <u>X</u> | <u>X</u> | X | <u>X</u> | <u>X</u> |
| CongMill Auge-Est CRACOMP ⁴⁰ N X< | | CR3SED | X | Х | X | X | X | X | x | | X Y | $-\hat{\mathbf{x}}$ | ^ | | X | X | <u> </u> | <u>X</u> | <u>X</u> | X | <u> </u> | <u> </u> | <u>X</u> | X |
| Craighill Angle-Base CRA-E-DOT(VSED X | | CR-COMP ^(a) | | | | | | | | v | | | ^ | | <u> </u> | <u> </u> | X | <u> </u> | X | <u> </u> | <u>X</u> | X | X | <u> </u> |
| CRAFE GODYCSED X | Craighill Angle-East | CRA-E-001VCSED | X | x | Y | | v | v | | | | | | | | <u> </u> | | | | | <u>X</u> | | | <u> </u> |
| CRA-E-00YCSED X < | 5 | CRA-E-002VCSED | X | x | <u> </u> | | | | | | <u>X</u> | <u>X</u> | <u>X</u> | X | X | <u>X</u> | X | <u> </u> | X | X | X | X | X | X |
| CRA6-COMP ¹⁰ CR CR CR CA A A A A A X | | CRA-E-003VCSED | X | x | <u> </u> | x | - <u>^</u> - | - A Y | - ÷ | | | <u> </u> | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | X | X | <u> </u> | <u> </u> | X | X | X | <u>X</u> |
| Chill Angle-Weat Chi A, WAQUY (SED) X | | CRA-E-COMP ^(a) | | | <u>A</u> | | | ^ | <u>^</u> | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u>X</u> | X | <u>X</u> | X | X | X |
| Characterization Characterization< | ghill Angle-West | CRA-W-001VCSED | v | | | | | | | <u>X</u> | | | | | | <u> </u> | | | | | Х | | | X |
| TAL M-GOURDED X < | | CRA-W-001VCSED | - Â | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | | X | X | X | Х | X | X | X | X | X | X | X | Х | Х | Х |
| CRAW-COMP ²⁰ N < | | CRA-W-002VCSED | X | $-\frac{\Lambda}{Y}$ | × × | | | <u> </u> | <u> </u> | | <u> </u> | <u> </u> | X | X | X | X | X | X | X | X | X | X | X | Х |
| Craghill Extrace CRE 001VSED X </td <td>-</td> <td>CRA-W-COMP^(a)</td> <td><u> </u></td> <td><u> </u></td> <td>A</td> <td></td> <td></td> <td><u>X</u></td> <td><u> </u></td> <td></td> <td><u> </u></td> <td><u> </u></td> <td><u> </u></td> <td><u> </u></td> <td><u> </u></td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>Х</td> <td>Х</td> | - | CRA-W-COMP ^(a) | <u> </u> | <u> </u> | A | | | <u>X</u> | <u> </u> | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | X | X | X | X | X | X | Х | Х |
| Chapman landary Chap of A A A X | Cenighill Entrance | CRE MIVCSED | | | | | | | | <u> </u> | | | | | | X | | | | | X | | | Х |
| CRE.000VCSED X <t< td=""><td></td><td>CRE-001VCSED</td><td>A V</td><td></td><td>X</td><td><u> </u></td><td><u>X</u></td><td>X</td><td>X</td><td></td><td>X</td><td><u>X</u></td><td>X</td><td>X</td><td>Х</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>Х</td><td>X</td><td>X</td><td>Х</td></t<> | | CRE-001VCSED | A V | | X | <u> </u> | <u>X</u> | X | X | | X | <u>X</u> | X | X | Х | X | X | X | X | X | Х | X | X | Х |
| CRE-DOMVCSED X <t< td=""><td></td><td>CRE-002VCSED</td><td>- A V</td><td></td><td><u> </u></td><td><u> </u></td><td><u>X</u></td><td>X</td><td>X</td><td></td><td>X</td><td>X</td><td>X</td><td>X</td><td>Х</td><td>X</td><td>X</td><td>X</td><td>X</td><td>Х</td><td>Х</td><td>x</td><td>Х</td><td>Х</td></t<> | | CRE-002VCSED | - A V | | <u> </u> | <u> </u> | <u>X</u> | X | X | | X | X | X | X | Х | X | X | X | X | Х | Х | x | Х | Х |
| CRU-COMP ⁰ x x <th< td=""><td></td><td>CRE-004VCSED</td><td></td><td></td><td><u> </u></td><td>X</td><td><u> </u></td><td><u> </u></td><td><u> </u></td><td></td><td><u> </u></td><td><u> </u></td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>Х</td><td>Х</td><td>X</td></th<> | | CRE-004VCSED | | | <u> </u> | X | <u> </u> | <u> </u> | <u> </u> | | <u> </u> | <u> </u> | X | X | X | X | X | X | X | X | X | Х | Х | X |
| Craighill Upper Range CRUISED X< | | CRE-COMP ^(a) | <u>^</u> | | ^ | <u> </u> | <u> </u> | <u> </u> | <u> </u> | | <u> </u> | <u>X</u> | <u> </u> | X | <u> </u> | Х | X | X | X | X | Х | X | X | X |
| Christel X< | | CRECOMP | | | | | | | | <u>X</u> | | | | | | x | | | | | Х | | | x |
| LK028DP A X </td <td>Jraigmii Opper Kange</td> <td>CRUISED</td> <td><u>X</u></td> <td><u> </u></td> <td><u> </u></td> <td><u>x</u></td> <td>X</td> <td>X</td> <td>X</td> <td></td> <td>X</td> <td>X</td> <td>Х</td> <td>X</td> | Jraigmii Opper Kange | CRUISED | <u>X</u> | <u> </u> | <u> </u> | <u>x</u> | X | X | X | | X | X | Х | X | X | x | X | X | X | X | X | X | x | X |
| Lecoustion A X | | CRU2SED | | <u> </u> | <u>X</u> | X | <u> </u> | X | X | | X | X | X | Х | X | X | X | X | Х | X | X | x | X | X |
| CRU-OMP ⁴⁰ A X <th< td=""><td></td><td>CRUZSEDFU</td><td></td><td><u>X</u></td><td><u> </u></td><td>X</td><td><u> </u></td><td><u> </u></td><td>X</td><td></td><td>X</td><td>X</td><td>Х</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td><td>X</td></th<> | | CRUZSEDFU | | <u>X</u> | <u> </u> | X | <u> </u> | <u> </u> | X | | X | X | Х | X | X | X | X | X | X | X | X | X | X | X |
| Cutoff Angle CUT (SED X | | CRUSSED | ^ | + | <u> </u> | X | <u> </u> | <u> </u> | <u> </u> | | _ X | X | Х | X | X | Х | X | X | X | X | X | X | x | X |
| Cutoff Ange Cutoff Ange Cutoff Ange X <t< td=""><td></td><td>CKU-COMP**</td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td>X</td><td></td><td></td><td></td><td></td><td></td><td>х</td><td></td><td></td><td></td><td></td><td>x</td><td></td><td></td><td>X</td></t<> | | CKU-COMP** | | | | | | | | X | | | | | | х | | | | | x | | | X |
| CUT2SED X </td <td>Cutoff Angle</td> <td>CUTISED</td> <td>X</td> <td>X</td> <td>X</td> <td>x</td> <td>X</td> <td>Х</td> <td>X</td> <td>T</td> <td>X</td> <td>X</td> <td>X</td> <td>x</td> <td>X I</td> <td>x</td> <td>x 1</td> <td>X</td> <td>X</td> <td>x</td> <td>Y I</td> <td>- Y</td> <td>v l</td> <td></td> | Cutoff Angle | CUTISED | X | X | X | x | X | Х | X | T | X | X | X | x | X I | x | x 1 | X | X | x | Y I | - Y | v l | |
| CUT3SED X </td <td></td> <td>CUT2SED</td> <td>X</td> <td><u> </u></td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td></td> <td>x</td> <td>X</td> <td>x</td> <td>X</td> <td>x</td> <td>X</td> <td>$\frac{\pi}{x}$</td> <td>$-\frac{x}{x}$</td> <td>x</td> <td>$\frac{x}{x}$</td> <td>$\frac{\Lambda}{X}$</td> <td>v</td> <td></td> <td>- <u>^</u></td> | | CUT2SED | X | <u> </u> | X | X | X | X | X | | x | X | x | X | x | X | $\frac{\pi}{x}$ | $-\frac{x}{x}$ | x | $\frac{x}{x}$ | $\frac{\Lambda}{X}$ | v | | - <u>^</u> |
| CUT-COMP ^{and} X < | | CUT3SED | <u> </u> | <u> </u> | <u>x</u> | X | X | X | X | · · · · · · | Х | X | x | X | x | X | X | X | x | X | - x | $-\frac{\Lambda}{X}$ | - Ŷ | - <u>x</u> |
| KL-7N-SED X | | CUT-COMP ^(a) | | | | | | | | x | | | | | | x | | | | | - | <u> </u> | | ~ |
| KL-7S-SED X K1-7wersed X </td <td>nside Placement Site 104</td> <td>KI-7N-SED</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>X</td> <td>x</td> <td>x</td> <td></td> <td>X</td> <td>Y I</td> <td></td> <td></td> <td></td> <td>$-\hat{\mathbf{v}}$</td> <td></td> <td></td> <td></td> <td></td> <td><u> </u></td> <td></td> <td></td> <td><u> </u></td> | nside Placement Site 104 | KI-7N-SED | X | X | X | X | X | x | x | | X | Y I | | | | $-\hat{\mathbf{v}}$ | | | | | <u> </u> | | | <u> </u> |
| KI-7E-SED X | | KI-7S-SED | Х | X | X | x | x | x | x | | X | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{v}}$ | × × | + | + | | | ~~~ + | <u> </u> | <u> </u> | - <u>×</u> | <u> </u> | <u> </u> |
| KI-7W-SED X K1-7.REFSED X< | | KI-7E-SED | X | X | X | X | X | x | - x + | | x | $-\frac{\Lambda}{X}$ | $-\frac{\Lambda}{Y}$ | $-\hat{\mathbf{x}}$ | $-\frac{\Lambda}{Y}$ | $-\hat{\mathbf{x}}$ | + | | - | ~~ + | | | <u>X</u> | <u>X</u> |
| KI-3-SED X < | | KI-7W-SED | X | X | X | x | x | x | <u>x</u> | | x + | $-\frac{x}{x}$ | X | $-\frac{\Lambda}{Y}$ | Y Y | - Ŷ | | + - + | $-\frac{\Lambda}{V}$ | - + | | - <u>×</u> | | <u>X</u> |
| KI-5-SED X < | | KI-3-SED | X | X | X | x | x | X | - <u>x</u> | | x | $-\frac{\alpha}{x}$ | $\frac{\Lambda}{X}$ | $-\frac{\Lambda}{X}$ | $-\frac{\Lambda}{X}$ | $\frac{1}{\mathbf{x}}$ | | | $-\hat{\mathbf{v}}$ | - | | | | <u>X</u> |
| KI-7-REFSED X | l | KI-5-SED | <u>x</u> | X | <u> </u> | X | x | x | x | | X | X | $\frac{1}{x}$ | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{x}} +$ | $-\frac{\alpha}{x}$ | | + | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{v}}$ | _ { | | |
| KI-7-REFSEDFD X | | KI-7-REFSED | X | X | x | x | x | x | X | | - <u>x</u> -+ | | $\frac{1}{x}$ | <u>-</u> | $-\frac{\alpha}{x}$ + | $-\frac{\alpha}{x}$ | | + + | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{x}}$ | | | | |
| KI-S-1-SED X | l | KI-7-REFSEDFD | X | X | X | x | x | x | X | | X | x | | $-\frac{x}{x}$ | $-\frac{2}{x}$ | $-\hat{\mathbf{x}}$ | + | | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{x}}$ | - \$ | | | |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | | KI-S-I-SED | <u> </u> | X | X | X | X | x | - <u>x</u> | | x | x + | $\frac{1}{x}$ | $\frac{1}{x}$ | $\frac{1}{x}$ | x | - | - | $\frac{1}{x}$ | | + | | - <u>~</u> | ÷ |
| $ \mathbf{K} = COMP^{(a)}$ | | KI-S-2-SED | X | X | <u>x</u> | X | x | x | X | | - <u>x</u> | X | x + | $-\frac{1}{x}$ | $\frac{1}{x}$ | $\frac{1}{x}$ | + | + + | $-\frac{\alpha}{x}$ | $-\hat{\mathbf{x}}$ | $-\frac{\Lambda}{V}$ | <u>-</u> | | |
| | | KI-COMP ^(a) | | | | | | | | x | | +- | | | + | + | <u>^-</u> | - <u>^</u> + | <u> </u> | | <u></u> | | - ^ | <u> </u> |

TABLE 5-1A (CONTINUED)

| | | 1 | 1 | Chlorinated & | | | | | T | T | | Τ | T | | T | T | T | | | | | T | |
|----------------------------|----------------------------|---------------------|----------------------|------------------|------------|----------------------|------------------------------|----------------------|-------------------------|----------------|----------------------|---------------------|---------------------|--|---------------------|---------------------|----------|--|-------------|----------|-------------|----------|---------------|
| | | | | Organophosphorus | РСВ | | | ł | Dioxins & | | Nitrate | | Total | | | | SEM / | | | Grain | Atterborg | Specific | Moisture |
| mpung Keach | Sediment Sample ID | VOCs'" | SVOCs | Pesticides | Congeners | PAHs | Metals | Butyltins | Furans ^(a,b) | Ammonia | + Nitrite | Cvanide | Sulfide | TKN | TOC | тр | AVS | 00 | ROD | Sizo | Limito | Crowitz | Contont |
| Outside Placement Site 104 | KI-11-SED | <u> </u> | X | X | X | X | X | Х | | X | X | X | Y | V | | - V | AVS | | | Size | | Gravity | Content |
| | KI-14-SED | <u>X</u> | <u> </u> | X | X | X | X | Х | | X | <u> </u> | | | ~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~~ | | <u>↓</u> | <u> </u> | | - <u> </u> | X | <u> </u> | X | X |
| | KI-15-SED | <u> </u> | <u> </u> | X | X | X | X | Х | | x | x | X | - Ŷ | | | | <u>├</u> | | ~ <u>~</u> | <u>λ</u> | <u> </u> | <u>X</u> | |
| | KI-16-SED | <u> </u> | X | X | X | X | X | Х | | x | X | <u> </u> | Y Y | $-\frac{2}{v}$ | $\frac{\lambda}{v}$ | ÷ | | - <u>^</u> | - Â | | <u> </u> | <u> </u> | X |
| 1 | KI-3-SED | <u> </u> | X | X | X | Х | X | X | | X | x | X | Y Y | $-\hat{v}$ | | - A | | | | <u>^</u> | <u> </u> | | |
| | KI-5-SED | <u> </u> | X | X | X | Χ | X | Х | | x | X | <u>x</u> | X X | | | - A V | <u>├</u> | - ÷ | - <u>^</u> | <u> </u> | <u> </u> | | |
| | KI-OUT-COMP ^(a) | | | | | | | ÷., | Y | | | <u>^</u> | | | <u>^</u> | | <u> </u> | | | <u> </u> | <u> </u> | X | <u>↓ ×</u> |
| Swan Point | SWP-001VCSED | X | X | X | X | X | x | x | <u> </u> | | v | | <u></u> | | <u> </u> | | | | | <u> </u> | | | |
| | SWP-002VCSED | X | X | X | x | x | X | <u> </u> | | <u>+ </u> | <u>A</u> | <u>×</u> | <u> </u> | <u>X</u> | <u> </u> | <u>X</u> | <u>X</u> | X | <u>X</u> | X | <u> </u> | X | <u> </u> |
| | SWP-003VCSED | X | Х | X | x | X | X | - <u>x</u> | | ⊢-≎I | $-\frac{\lambda}{v}$ | <u>X</u> | <u>X</u> | <u> </u> | <u>X</u> | <u>X</u> | <u> </u> | <u>X</u> | X | <u>X</u> | X | X | <u> </u> |
| | SWP-004VCSED | X | Х | X | <u>x</u> | X | $-\frac{x}{x}$ | - X | | ├ ─ | | <u> </u> | <u>X</u> | <u>X</u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | X |
| | SWP-005VCSED | X | х | X | - <u>x</u> | X | X X | | | | <u> </u> | <u> </u> | <u> </u> | <u>X</u> | X | X | <u>X</u> | X | <u>X</u> | <u>X</u> | <u>X</u> | X | X |
| | SWP-006VCSED | X | X | x | - <u>x</u> | X | $\frac{\Lambda}{\mathbf{x}}$ | - Ŷ | | | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u>X</u> | <u> </u> | X | <u>X</u> | X | Х | X | X |
| | SWP-COMP ^(a) | | | | | | <u> </u> | | | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | <u> </u> | <u> </u> | <u> </u> | X | X | X | X |
| Tolchester Channel - North | TLC-N-005VCSED | x | X | Y | | | | | <u> </u> | | | | | | X | | | | | <u>x</u> | | | x |
| | TLC-N-006VCSED | X | x | | | $-\frac{\Lambda}{V}$ | `` | <u> </u> | | X | <u>X</u> | <u> </u> | <u> </u> | X | X | X | X | X | X | X | X | X | X |
| | TLC-N-007VCSED | X | x | <u>^</u> Y | | | | | | X | <u>X</u> | <u> </u> | <u> </u> | X | X | X | X | X | Х | X | X | X | X |
| | TLC-N-008VCSED | X | $\frac{x}{x}$ | <u> </u> | | | | $-\frac{\lambda}{v}$ | | X | <u> </u> | <u> </u> | <u> </u> | X | X | X | X | x | X | X | X | X | X |
| | TLC-N-009VCSED | X | x | X | | $-\hat{\mathbf{v}}$ | $-\hat{\mathbf{v}}$ | | | <u>X</u> | <u> </u> | <u> </u> | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | X | x | X | X | X | X | X |
| | TLC-N-010VCSED | X | <u>x</u> | | <u>^</u> | $-\hat{\mathbf{v}}$ | | | | $-\frac{x}{y}$ | <u> </u> | <u>X</u> | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | X | X | X | X | X | X | X |
| | TLC-N-COMP ^(a) | | | <u>A</u> | <u> </u> | | | - ^ - + | | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | <u> </u> | X | X | X | X | X | X | X |
| Tolchester Channel - South | TICS COLVESED | | | | | | | | <u> </u> | | | | | | x | | | | | x | х | x | X |
| rolenester Chaimer - South | TLC S 002VCSED | $-\hat{\mathbf{v}}$ | | X | <u> </u> | X | <u> </u> | <u> </u> | | x | X | X | X | X | X | X | X | X | X | X | x | x | X |
| | TLC S 003VCSED | | <u>^</u> | <u> </u> | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | | X | X | X | X | X | x | X | X | x | X | x | x | x | X |
| | TLC-S-003VCSED | | $-\frac{\lambda}{v}$ | X | <u> </u> | <u>X</u> | <u> </u> | <u> </u> | | x | X | X | X | X | X | X | X | X | X | X | x | x | $\frac{1}{x}$ |
| | TLC-S-004VCSED | | | X | <u> </u> | <u> </u> | <u> </u> | X | | X | X | X | X | x | X | X | X | x | X | X | x | X | X |
| | TLC-S-COMP** | | | | | | | | Х | | | | | | x | | | | | Y | | | |
| Tolchester Straightening | TLS-00IVCSED | <u>X</u> | X | X | X | Х | X | X | | X I | X | x | x | X I | - X | | v | ······································ | | | | | <u> </u> |
| | TLS-002VCSED | <u> </u> | <u> </u> | X | x | X | x | x | | - <u>x</u> | $-\frac{1}{x}$ | $-\hat{\mathbf{x}}$ | $-\hat{\mathbf{x}}$ | - <u>x</u> | $-\hat{\mathbf{v}}$ | $-\hat{\mathbf{v}}$ | | + - + | | | | | `` |
| | TLS-COMP ^(a) | | | | | | | | x | | | | | | $-\frac{2}{u}$ | | | <u> </u> | | | X | <u> </u> | <u> </u> |
| | | | | | | | | ا | <u> </u> | | | | | | X | | | | | X | | | X |

(a) = December sampling only.(b) = Composite samples only.

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TABLE 5-2 BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104: ANALYTICAL METHODS

| Parameter | Method | Method # | Matrix | Reference |
|---|---|----------|------------|-----------|
| SAMPLE PREPARATION | | | | |
| Semivolatile (PAHs, Pest, PCBs, SVOA) | Continuous Extraction | 3502C | W, E | EPA, 1997 |
| Semivolatile (PAHs, Pest, PCBs, SVOA) | Soxhlet Extraction | 3540C | S, T | EPA, 1998 |
| Semivolatile (Dioxins/Furans) | Pressurized Fluid Extraction | 3545 | S, T | EPA, 1999 |
| Volatiles | Purge and Trap (40°C) | 5030A | S | EPA, 2000 |
| Volatiles | Purge and Trap | 5030B | W, E | EPA, 2001 |
| Soluble Salts Extractions | Aqueous Extraction | 10-2 | S | ASA, 1992 |
| Total Metals Digestion (ICP) | Nitric Acid - Hydrochloric Acid | 3010A | W, E | EPA, 1997 |
| Total Metals Digestion (GFAA) | Nitric Acid | 3020A | W, E | EPA, 1997 |
| Total Metals Digestion | Nitric Acid - Hydrogen Peroxide | 3050B | S, T | EPA, 1997 |
| ORGANIC - EXTRACTION CLEANUP | Liquid-liquid Partitioning | 3665A | W, E, S, T | EPA, 1997 |
| Sulfuric Acid Cleanup | Treatment with Cu, Hg, or TBA-sulfite | 3660A/B | S | EPA, 1997 |
| Sulfuric Cleanup | Molecular Size Exclusion Chromatography | 3640A | Т | EPA, 1997 |
| Gel Permeation Chromatography (GPC) Cleanup | Adsorption Column Chromatography | 3620B | W, E, S, T | EPA, 1997 |
| Florisil Cleanup | - | | | |
| ORGANICS | | | | |
| Semivolatile Organic Compounds | Gas Chromatography/Mass Spectrometry | 8270C | W, E, S, T | EPA, 1997 |
| Volatile Organic Compounds | Gas Chromatography/Mass Spectrometry | 8260B | ' W, E, S | EPA, 1997 |
| Polynuclear Aromatic Hydrocarbons (PAH) | HPLC - UV, fluorescence | 8310 | W, E, S, T | EPA, 1997 |
| Halogenated Hydrocarbon Pesticides | Gas Chromatography - ECD | 8081A | W, E, S, T | EPA, 1997 |
| Organophosphorus Pesticides | Gas Chromatography - NPD/FPD | 8141A | W, E, S, T | EPA, 1997 |
| PCB Aroclors | Gas Chromatography - ECD | 8082 | W, E, S, T | EPA, 1997 |
| PCB Congeners | Gas Chromatography - ECD | 8082 | W, E, S, T | EPA, 1997 |
| Dioxins/Furans | HRGC/HRMS | 8290 | W, E, S | EPA, 1997 |
| Dioxins/Furans | HRGC/HRMS | 1613 | Т | EPA, 1990 |
| Organotins | Gas Chromatography - FPD | STL-SOP | W, E, S, T | |
| METALS | | | | |
| Aluminum | Atomic Emission - 1CP | 6010B | W, E, S, T | EPA, 1997 |
| Antimony | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Arsenic | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Beryllium | Atomic Emission - Trace 1CP | 6010B | W, E, S, T | EPA, 1997 |
| Cadmium | Atomic Emission - Trace 1CP | 6010B | W, E, S, T | EPA, 1997 |
| Chromium | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |

| Parameter | Method | Method # | Matrix | Reference |
|-----------------------------|---|------------|------------|-------------|
| Copper | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Iron | Atomic Emission - ICP | 6010B | W, E, S, T | EPA, 1997 |
| Lead | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Manganese | Atomic Emission - ICP | 6010B | W, E, S, T | EPA, 1997 |
| Mercury | Atomic Absorption - Cold Vapor | 7470A | W, E | EPA, 1997 |
| Mercury | Atomic Absorption - Cold Vapor | 7471A | S, T | EPA, 1997 |
| Nickel | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Selenium | Atomic Emission - Trace ICP | 6010B | W, E, S | EPA, 1997 |
| Silver | Atomic Emission - Trace ICP | 6010B | W, E, S, T | EPA, 1997 |
| Thallium | Atomic Absorption - Furnace | 7841 | W, E, S, T | EPA, 1997 |
| Zinc | Atomic Emission - ICP | 6010B | W, E, S, T | EPA, 1997 |
| INORGANIC NONMETALS | | | | |
| Cyanide, Total | Colorimetric - Automated UV | 9012A | W, E, S | EPA, 1997 |
| Sulfide, Total | Titrimetric | 9034 | W, E | EPA, 1997 |
| Sulfide, Total | Distillation/Titrimetric | 9030B/9034 | S | EPA, 1997 |
| Total Organic Carbon | Oxidation - Infrared | 9060 | W, E, S | EPA, 1997 |
| Biochemical Oxygen Demand | BOD (% day, 20°C) | 405.1 | W, E, S | EPA, 1979 |
| Chemical Oxygen Demand | Colorimetric - Manual | 410.4 | W, E, S | EPA, 1979 |
| Nitrogen, Ammonia | Colorimetric - Automated Phenate | 350.1 | W, E, S | EPA, 1979 |
| Nitrogen, Total Kjeldhal | Colorimetric - Autoanalyzer II | 351.2 | W, E, S | EPA, 1979 |
| Nitrogen, Nitrate + Nitrite | Colorimetric - Cadmium Reduction | 353.2 | W, E, S | EPA, 1979 |
| AVS/SEM | | | S | EPA, 1991 |
| Total Phosphorus | Colorimetric | 365.3 | W, E, S | EPA, 1979 |
| PHYSICAL PARAMETERS | | | | |
| Grain Size | Sieve Analysis | D422 | S | ASTM, 1995 |
| Atterberg Limits | Physical Measurement | D4318 | S | ASTM, 1995 |
| Moisture Content | Gravimetric | D4959 | S | ASTM, 1995 |
| Moisture Content | Gravimetric | EA-SOP | Т | |
| Specific Gravity | Hydrometer | | S | Plumb, 1981 |
| Percent Lipids | Extraction/Gravimetric | STL-SOP | Т | |

TABLE 5-2 (CONTINUED)





TABLE 5-2 (CONTINUED)

Matrix codes:

- W Water
- E Elutriate
- S Sediment
- T Tissue

References:

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- EPA, 1979 United States Environmental Protection Agency. 1979. Methods for Chemical Analysis of Water and Wastes. EPA-600/4-79-020. U.S. EPA, Cincinnati, Ohio.
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- EPA, 1997 United States Environmental Protection Agency. 1997. Test Methods for Evaluating Solid Waste. Physical/Chemical Methods. EPA SW-846, 3rd edition, including Final Update III. U.S. EPA, Washington, D.C.
- Plumb, 1981 Plumb, R.H. Jr. 1981. Procedures for handling and Chemical Analysis of Sediment and Water Samples. Technical Report EPA/CE-81-1. U.S. Army Engineer Waterways Experiment Station, CE, Vicksburg, Miss.

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|--|-----------|----------------------------------|-----------------------------------|
| Organochlorine Pesticides - GC/ECD - (SW846 354 | OC/8081A) | | |
| Aldrin | ug/kg | 0.52 | 10 |
| α-BHC | ug/kg | 0.38 | - |
| ß-BHC | ug/kg | 0.49 | - |
| A-BHC | ug/kg | 0.49 | _ |
|) PUC (Lindene) | ug/kg | 0.45 | 10 |
| Chlorbergide | ug/kg | 0.45 | 10 |
| Chlordena (Technical) | ug/kg | 3.3 | 2 |
| Chiordane (Technical) | ug/kg | 1.0 | 10 |
| | ug/kg | 10 (** | 2 |
| 4,4°-DDD | ug/kg | 0.42 | 10 |
| 4,4°-DDE | ug/kg | 0.40 | 10 |
| 4,4'-DD1 | ug/kg | 0.66 | 10 |
| Dieldrin | ug/kg | 0.43 | 10 |
| Endosulfan 1 | ug/kg | 0.72 | 10 |
| Endosulfan II | ug/kg | 0.36 | 10 |
| Endosulfan sulfate | ug/kg | 0.84 | 10 |
| Endrin | ug/kg | 1.5 | 5 |
| Endrin aldehyde | ug/kg | 0.94 | 5 |
| Heptachlor | ug/kg | 0.60 | 10 |
| Heptachlor epoxide | ug/kg | 0.81 | 10 |
| Mirex | ug/kg | 3.3 ^(c) | |
| Methoxychlor | ug/kg | 2.6 | 10 |
| Toxaphene | ug/kg | 14 | 50 |
| PCB Aroclors – GC/ECD – (SW846 3540C/8082) | | | |
| Aroclor 1016 | ug/kg | 5.4 | |
| Aroclor 1221 | ug/kg | 6.8 | |
| Aroclor 1232 | ug/kg | 12 | |
| Aroclor 1242 | ug/kg | 8.8 | |
| Aroclor 1248 | ug/kg | 2.6 | |
| Aroclor 1254 | ug/kg | 7.8 | |
| Aroclor 1260 | ug/kg | 4.9 | |
| PCB Congeners - GC/ECD - (SW846 3540C/8082) | | | |
| 2,4'-Dichlorobiphenyl (BZ # 8) | ug/kg | 0.10 | 1 |
| 2,2',5-Trichlorobiphenyl (BZ # 18) | ug/kg | 0.10 | 1 |
| 2,4,4'-Trichlorobiphenyl (BZ # 28) | ug/kg | 0.037 | 1 |
| 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44) | ug/kg | 0.11 | 1 |
| 2.2'.4.5'-Tetrachlorobiphenvl (BZ # 49) | ug/kg | 0.17 | 1 |
| 2.2'.5.5'-Tetrachlorobiphenvl (BZ # 52) | ug/kg | 0.10 | 1 |
| 2.3' 4.4'-Tetrachlorobinhenvl (BZ # 66) | ug/kg | 0.056 | 1 |
| 3.3'.4.4'-Tetrachlorobinhenvi (BZ # 77) | ug/kg | 0.082 | 1 |
| 2.2', $3.4.5'$ -Pentachlorobiphenyl (BZ # 87) | ug/kg | 0.042 | 1 |
| 2.2' 4.5.5'-Pentachlorobiphenyl (BZ # 101) | ug/kg | 0.058 | 1 |
| 2 3 3' 4 4'-Pentachlorobiphenyl (BZ # 101) | ug/kg | 0.18 | 1 |
| 2 3' 4 4' 5-Pentachlorobinhenvl (BZ # 118) | ug/kg | 0.069 | 1 |
| 2,3,4,4,5 - Ciliacinorobiphenyi (D2 # 116) 3,3',4,4',5 Deptahlorobiphenyi (D2 # 126) | ug/kg | 0.009 | 1 |
| 2 2' 3 3' 4 4' Herechlorobishenyl (DZ # 120) | ug/kg | 0.049 | 1 |
| 2,2,3,3,7,7,7,7,17,7,10,0000000000000000000 | ug/kg | 0.040 | 1 |
| 2,2,3, 4 ,4,3-mcxaciliorodipitenyi (BZ # 130) 2,2',4,4',5,5' Heveeblorobinbanui (B7 # 152) | ug/kg | 0.043 | 1 |
| 2,2,4,4,3,3+nexachiorodipnenyi (BZ # 133) | ug/kg | 1.03 | 1 |
| 2,3,3,4,4,3-Hexacniorobipnenyi (BZ # 156) | ug/kg | 0.080 | 1 |
| 3,3,4,4,5,5 -Hexachlorobiphenyl (BZ # 169) | ug/kg | 0.095 | l |
| 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) | ug/kg | 0.071 | I |
| 2,2,3,4,4',5,5'-Heptachlorobiphenyl (BZ # 180) | ug/kg | 0.087 | 1 |
| 2,2',3,4,4',5',6-Heptachlorobiphenyl (BZ # 183) | ug/kg | 0.051 | 1 |
| 2,2',3,4,4',6,6'-Heptachlorobiphenyl (BZ # 184) | ug/kg | 0.056 | 1 |
| 2,2',3,4',5,5',6-Heptachlorobiphenyl (BZ # 187) | ug/kg | 0.060 | 1 |

TABLE 5-3 METHOD DETECTION LIMITS (MDLs) FOR SEDIMENT SAMPLES

⁽a) Method Detection Limit (MDL) STL-Baltimore for standard solid matrix determined according to the procedure in 40 CFR 136 Appendix B.

⁽b) Target Detection Limit (USEPA/USACE 1995).

⁽c) For these compounds, no laboratory MDL has been determined by STL-Baltimore. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits).

TABLE 5-3 (CONTINUED)

| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|--|--------|----------------------------------|-----------------------------------|
| 2,2,3,3,4,4,5,5,6,-Decachlorobiphenyl (BZ # 209) ug/kg 0.13 1 2,2,3,3,4,4,5,5,6,-Decachlorobiphenyl (BZ # 209) ug/kg 21 - Demeton ug/kg 21 - Etyl parathon (Parathion) ug/kg 33 (°) 6 Outhon (Azinphos methyl) ug/kg 22 - Malathon ug/kg 20 - Acrolein ug/kg 20 - Acrolein ug/kg 0.6 10 Benzene ug/kg 0.6 - Bromodichloromethane ug/kg 0.6 - Bromodichloromethane ug/kg 0.6 - Bromodichloromethane ug/kg 0.6 - Bromodichloromethane ug/kg 0.6 - Carbon transhoride ug/kg 0.6 - Carbon transhoride ug/kg 0.6 - Carbon transhoride ug/kg 0.6 - Choroberane ug/kg 0.6 - Choroberane ug/kg 0.6 - Choroberane | 2,2',3,3',4,4',5,6-Octachlorobiphenyl (BZ # 195) | ug/kg | 0.087 | 1 |
| 2,2'3,3',4',5',5',6'-Decachlorobiphenyl (BZ # 209) ug/kg 0.16 1 Organophosphorus Pesiticides GC/NPD - (SW 3540C/8141A) - - Ethyl parathion (Parathion) ug/kg 23 6 Guhion (Azimphos methyl) ug/kg 22 - Mathion ug/kg 17 6 Volatile organies - GC/NS (SW845 5030A/ 8260B) - - Acrolein ug/kg 8 - Benzene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.6 - Bromodichloromethane ug/kg 0.6 - Bromomethane ug/kg 0.5 - Bromomethane ug/kg 0.5 - Carbon disulfide ug/kg 1 - Carbon disulfide ug/kg 0.6 10 Choroethane ug/kg 0.6 - Carbon disulfide ug/kg 0.6 - Diobroromethane ug/kg 0.6 - Diobroromethane ug/kg 0.6 - 1.2-Dichorobetnzene <t< td=""><td>2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl (BZ # 206)</td><td>ug/kg</td><td>0.13</td><td>1</td></t<> | 2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl (BZ # 206) | ug/kg | 0.13 | 1 |
| Organophosphorus Pesticides GC/NPD - (SW 3540C/8141A) | 2,2',3,3',4,4',5,5',6,6'-Decachlorobiphenyl (BZ # 209) | ug/kg | 0.16 | 1 |
| Demetron ug/kg 21 Ethyl parathion (Parathion) ug/kg 23 60 Guthion (Azinphos methyl) ug/kg 16 5 Methyl parathion ug/kg 17 6 Volatile organics - GC/MS (SW846 5030.4/ 8260B) - - Acroloninile ug/kg 8 - Benzene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.6 - Carbon on disulfide ug/kg 0.4 - Carbon oterachloride ug/kg 0.5 - Chlorobenzene ug/kg 0.6 - Carbon oterachloride ug/kg 0.6 - Carbon oterachloride ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromodichloromethane ug/kg 0.6 - Lobaltobenzene ug/kg 0.6 - Lobalobenzene ug/k | Organophosphorus Pesticides GC/NPD - (SW 3540C/8 | 8141A) | | |
| Ethyl parathion (Parathion) ug/kg 23 Malathion ug/kg 16 5 Malathion ug/kg 16 5 Mathion ug/kg 17 6 Volatile organics - GC/MS (SW846 5030 / 8260B) - - Acrolein ug/kg 8 - Benzene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.7 - Bromodichloromethane ug/kg 0.6 - Bromomothane ug/kg 0.6 - Bromomothane ug/kg 0.5 - Carbon disulfide ug/kg 0.5 - Chorobenzene ug/kg 0.6 10 Chorobenzene ug/kg 0.6 10 Chorobenzene ug/kg 0.6 10 Chorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 - | Demeton | ug/kg | 21 | |
| Guiltion (Azinphos methyl) ug/kg 22 Malahin ug/kg 16 5 Methyl parathion ug/kg 17 6 Volatile organics - CC/MS (SW846 5030A/ 8260B) Acrolein ug/kg 20 - Acrolein ug/kg 0.6 10 Benzene ug/kg 0.7 - Benzene ug/kg 0.7 - Benzene ug/kg 0.6 - Bromodichloromethane ug/kg 0.5 - Bromodichloromethane ug/kg 0.5 - Carbon disulfide ug/kg 0.5 - Carbon disulfide ug/kg 0.5 - Chlorobenzene ug/kg 0.5 - Chlorobenzene ug/kg 0.5 - Chlorobenzene ug/kg 0.6 10 Chlorotentane ug/kg 1 - - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.5 - L1-Diblorobenzene ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.5 - L1-Diblorobenzene ug/kg 0.5 - L1-Diblorobenzene ug/kg 0.5 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.5 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.5 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.6 - Dibloromethane ug/kg 0.6 - Dibromothoromethane ug/kg 0.5 - Dibromothoromethane ug/kg 0.5 - Dibromothoromethane ug/kg 0.5 - Dibromothoromethane ug/kg 0.6 - Dibromot | Ethyl parathion (Parathion) | ug/kg | 33 (0) | 6 |
| Malathon ug/kg 16 5 Molthyl parathion ug/kg 17 6 Volatile organics - GC/MS (SW846 50304/ 8260B) - - Acrolein ug/kg 8 - Benzene ug/kg 0.6 10 Bromoform ug/kg 0.6 10 Bromosthane ug/kg 0.9 - Bromosthane ug/kg 0.4 - Carbon disulfide ug/kg 0.4 - Carbon tetrachloride ug/kg 0.9 - Chloroethyn ug/kg 0.6 10 Chloroethane ug/kg 0.6 10 Chloroethane ug/kg 0.6 10 Chloroethane ug/kg 0.6 20 1,2-Dichlorobenzene ug/kg 0.6 20 1,3-Dichlorobenzene ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - | Guthion (Azinphos methyl) | ug/kg | 22 | |
| Methyl paratholon ug/kg 17 6 Volatile organics - GC/MS (SW846 5030A/ 8260B) - - Acrolein ug/kg 8 - Benzene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.7 - Bromodichloromethane ug/kg 0.6 - Carbon tertachloride ug/kg 0.5 - Chlorobenzene ug/kg 0.6 10 Chlorobenzene ug/kg 0.6 10 Chlorotoftane ug/kg 0.6 1 Chlorotoftane ug/kg 0.6 1 Dibromochloromethane ug/kg 0.6 - Lobithorobenzene ug/kg 0.6 - L>Dibromochloromethane ug/kg 0.6 - L>Dibrohorobenzene ug/kg 0.6 - L>Dibrohorobenzene ug/kg 0.6 - L>Dibrohorobenzene ug/kg 0.6 - L>Dibrohorobenzene ug/kg | Malathion | ug.kg | 16 | 5 |
| Volatile organics - GC/MS (SW846 50304/ 8260B) ug/kg 20 - Acroloin ug/kg 8 - Benzme ug/kg 0.6 10 Bromodichloromethane ug/kg 0.7 - Bromodichloromethane ug/kg 0.6 - Carbon tetrachloride ug/kg 0.6 - Carbon tetrachloride ug/kg 0.5 - Chlorobenzene ug/kg 0.6 10 Chlorobenzene ug/kg 0.6 10 Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - 1.2-Dichlorobenzene ug/kg 0.6 - 1.3-Dichlorobenzene ug/kg 0.6 - 1.3-Dichloroponpa | Metnyl parathion | ug/Kg | 17 | 6 |
| Acroinini ug/kg 20 Acroinini ug/kg 8 Benzene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.7 - Bromomethane ug/kg 0.9 - Carbon disulfide ug/kg 0.4 - Carbon terachloride ug/kg 0.9 - Chlorobenzene ug/kg 0.9 - Chloromethane ug/kg 0.9 - Chloromethane ug/kg 0.6 10 Chloromethane ug/kg 0.6 10 Chloromethane ug/kg 0.6 10 Chloromethane ug/kg 0.6 20 Liboromethane ug/kg 1 - 1.2-Dichlorobenzene ug/kg 1 - 1.3-Dichlorobenzene ug/kg 0.6 20 Liboromethane ug/kg 0.6 - Li-Dichlorobenzene ug/kg 0.6 - Li-Dichlorobenzene ug/kg 0.6 - Li-Dichlorobenzene ug/kg 0.5 - Li-Dichlorobenzene ug/kg 0.5 - Li-Dichloropenae ug/kg 0.5 - Li-Dichloropenae ug/kg 0.6 - Li-Dichloropenae ug/kg 0.6 - Li-Dichloropenae ug/kg 0.5 - Li-Dichloropenae ug/kg 0.6 - Li-Dichloropenae ug/kg 0.5 - Li-Dichloropenae ug/kg 0.6 - Li-Dichloropenae ug/kg 0.5 - | Volatile organics - GC/MS (SW846 5030A/ 8260B) | | • | |
| Acryonithie ug/kg 8 Berizene ug/kg 0.6 10 Bromodichloromethane ug/kg 0.7 Bromomethane ug/kg 0.7 Bromomethane ug/kg 0.9 - Bromomethane ug/kg 0.6 - Carbon teracholoride ug/kg 0.5 - Chlorobenzene ug/kg 0.9 - Chlorobenzene ug/kg 0.9 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Chlorobenzene ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - Dibromochlorobenzene ug/kg 0.6 - Dichlorobenzene ug/kg 0.6 - L]Dichlorobenzene ug/kg 0.6 - L]Dichloroptane ug/kg 0.6 - L]Dichloroptane ug/kg 0.6 - L]Dichloroptane ug/kg 0.6 - LTans-L].3-Dichloroptane ug/kg 0.6 - LTans-L].3-Dichloroptane ug/kg 0.6 - L].2-Dichloroptane ug/kg 0.6 - L].1-Dichloroptane ug/kg 0.6 - L].2-Dichloroptane ug/kg 0.5 - L].2-Dichloroptane ug/kg 0.6 - Tetrachloropthane ug/kg 0.5 - L].2-Dichloroptane ug/kg 0.6 - Tetrachloropthane ug/kg 0.5 - L].2-Tichloroptane ug/kg 0.6 - Tetrachloropthane ug/kg 0.5 - L].2-Tichloroptane ug/kg 0.5 - L].2-Tichloroptane ug/kg 0.6 - Tetrachloropthane ug/kg 0.5 - L].2-Dichloroptane ug/kg 0.5 - Tetrachloropthane ug/kg 0.5 - L].2-Tichloroptane ug/kg 0.5 - L].2-Tichloroptane ug/kg 0.5 - L].2-Tichloropthane ug/kg 0.5 - Semivolatile crganics GC/MS - (SW846 3540C/8270C) Semivolatile crganics GC/MS - (SW846 354 | Acrolein | ug/kg | 20 | - |
| boundary in the second | Activioninie Denzene | ug/kg | 8 | - |
| Bromotoring ug/kg 0.7 Bromorem ug/kg 0.7 Bromorem ug/kg 0.9 Carbon disulfide ug/kg 0.4 - Carbon disulfide ug/kg 0.5 - Chlorobenzene ug/kg 0.9 - Chlorotentane ug/kg 1 - 2-Chlorotentane ug/kg 0.6 - Chlorotentane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 - Dichlorotentane ug/kg 0.6 - Dichlorotentane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - Dichlorotentane ug/kg 0.6 - Dichlorotentane ug/kg 0.6 - Dichlorotenzene ug/kg 0.6 - Dichlorotentane ug/kg 0.6 - Li-Dichlorotentane ug/kg 0.7 - Ethylbenzene ug/kg 0.7 - Ethylbenzene ug/kg 0.6 - Tetrachlorotentane ug/kg 0.6 - Tetrachlorotentane ug/kg 0.6 - Tetrachlorotentane ug/kg 0.6 - Tetrachlorotentane ug/kg 0.5 - Li,1,2-Trichlorotentane ug/kg 0.5 - Li,1,2-Trichlorotentane ug/kg 0.5 - Semivotentene ug/kg 0.5 - Semivotenene ug/kg 0.5 - Semivotenene ug/kg 0.5 - S | Benzene | ug/kg | 0.0 | 10 |
| Domostini ug/kg 0.5 - Carbon etrachoride ug/kg 0.6 - Carbon etrachoride ug/kg 0.5 - Chlorobenzene ug/kg 0.9 - Chlorobethane ug/kg 0.9 - 2-Chlorobthyl vinyl ether ug/kg 0.6 10 Chlorobethane ug/kg 0.6 - Dibromochloromethane ug/kg 1 - 1,2-Dichlorobenzene ug/kg 0.6 - 1,3-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,1-Dichlorobenzene ug/kg 0.6 - 1,1-Dichlorobenzene ug/kg 0.6 - 1,1-Dichlorobenzene ug/kg 0.6 - 1,1-Dichlorobenzene ug/kg 0.6 - 1,1-Dichloropropane ug/kg 0.6 - 1,1-Dichloropropene | Bromoform | ug/kg | 0.7 | - |
| Carbon disulfide ug/kg 0.4 - Carbon disulfide ug/kg 0.5 - Chloroentane ug/kg 0.9 - Chloroethane ug/kg 1 - 2-Chloroethane ug/kg 0.6 10 Chloroethane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloroptopane ug/kg 0.6 - 1,1-Dichloroptopene ug/kg 0.6 - 1,2-Dichloroptopene ug/kg 0.6 - 1,1,2,2-Tetracholoroethane | Bromomethane | ug/kg | 0.9 | - |
| Carbon terrachloride Ug/kg 0.7 - Chlorobenzene ug/kg 0.9 - Chlorobenzene ug/kg 1 - 2-Chloroethyl vinyl ether ug/kg 0.6 10 Chloroothane ug/kg 0.6 10 Chloroothane ug/kg 0.6 - Dibromochloromethane ug/kg 1 20 1,3-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - 1,3-Dichlorobenzene ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropene ug/kg 0.6 - 1,2-Dichloropropene ug/kg 0.6 10 Uphenzene | Carbon disulfide | ug/kg | 0.0 | - |
| University Ug/kg 0.9 - Chloroethane ug/kg 1 - 2-Choroethane ug/kg 2 - Chloroform ug/kg 0.6 10 Chloroform ug/kg 0.6 - Dibromochloromethane ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloroethane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - trans-1,3-Dichloropropane ug/kg 0.6 - trans-1,2-Dichloroethane | Carbon tetrachloride | ug/kg | 0.5 | - |
| Chloroethane ug/kg 1 - 2-Chloroethyl vinyl ether ug/kg 2 - Chloroform ug/kg 0.6 10 Chloroform ug/kg 0.6 - Dibromochloromethane ug/kg 1 - 1,2-Dichlorobenzene ug/kg 1 20 1,3-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - Liborodifluoromethane ug/kg 0.6 - Lj-Dichlorobenzene ug/kg 0.6 - Lj-Dichloroethane ug/kg 0.6 - Lj-Dichloroethane ug/kg 0.6 - Lj-Dichloroethane ug/kg 0.6 - Lj-Dichloropropane ug/kg 0.6 - Lj-Dichloropropane ug/kg 0.6 - Lj-Dichloropropane ug/kg 0.6 - Lj-Dichloropropene ug/kg 0.6 - Lj-Dichloropropene ug/kg 0.6 - Lj-Dichloropropene ug/kg | Chlorobenzene | ug/kg | 0.9 | - |
| 2-Chloroethyl vinyl etherug/kg2-Chlorooformug/kg0.610Chloroofthaneug/kg0.6-Dibromochloromethaneug/kg1201,2-Dichlorobenzeneug/kg0.6201,4-Dichlorobenzeneug/kg0.6201,4-Dichlorobenzeneug/kg0.6-1,3-Dichlorobenzeneug/kg0.6-1,2-Dichlorotheneug/kg0.5-1,1-Dichlorothaneug/kg0.5-1,2-Dichlorotethaneug/kg0.6-1,1-Dichlorotethaneug/kg0.6-1,2-Dichlorotethaneug/kg0.6-1,2-Dichloropropaneug/kg0.6-1,2-Dichloropropaneug/kg0.6-1,2-Dichloropropaneug/kg0.6-1,2-Dichloropropaneug/kg0.6-1,1,2,2-Tetxhloroethaneug/kg0.510Methylen chlorideug/kg0.5-1,1,2,2-Tetxhloroethaneug/kg0.6-1,1,1,2-Trichloroethaneug/kg0.6-1,1,1,2-Trichloroethaneug/kg0.5-1,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg | Chloroethane | ug/kg | 1 | - |
| Chloroform ug/kg 0.6 10 Chloromethane ug/kg 0.6 - 1,2-Dichlorobenzene ug/kg 1 - 1,3-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - 1,4-Dichlorobenzene ug/kg 0.6 - Dichlorodifluoromethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloroethane ug/kg 0.6 - 1,2-Dichloropthane ug/kg 0.6 - 1,2-Dichloropthane ug/kg 0.6 - 1,2-Dichloropthane ug/kg 0.6 - 1,2-Dichloroptopene ug/kg 0.6 - 1,2-Dichloroptopene ug/kg 0.6 - 1,2-Dichloroptopene ug/kg 0.6 - 1,2-Dichloroptopene ug/kg 0.6 - 1,1,2-Tetrachloroethane ug/kg 0.6 - 1,1,2,2-Tetrachloroethane <td>2-Chloroethyl vinyl ether</td> <td>ug/kg</td> <td>2</td> <td>-</td> | 2-Chloroethyl vinyl ether | ug/kg | 2 | - |
| Chloromethane ug/kg 0.6 - Dibromochloromethane ug/kg 1 - 1,2-Dichlorobenzene ug/kg 0.6 20 1,3-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.6 - Dichlorodifluoromethane ug/kg 0.6 - Dichlorodifluoromethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloroethane ug/kg 0.6 - 1,2-Dichloroptene ug/kg 0.6 - 1,2-Dichloroptene ug/kg 0.6 - 1,2-Dichloroptopane ug/kg 0.6 - cis-1,3-Dichloroptopene ug/kg 0.6 - trans-1,3-Dichloroptopene ug/kg 0.6 - trans-1,3-Dichloroptene ug/kg 0.6 - trans-1,3-Dichloroptene ug/kg 0.6 - trans-1,3-Dichloroptene ug/kg 0.6 - 1,1,2,2-Tetrachloroethane ug/kg 0.6 - | Chloroform | ug/kg | 0.6 | 10 |
| Dibromochloromethane ug/kg 1 1,2-Dichlorobenzene ug/kg 1 20 1,3-Dichlorobenzene ug/kg 0.6 20 1,4-Dichlorobenzene ug/kg 0.8 20 trans-1,2-Dichloroethene ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.5 - 1,1-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,1-Dichloropthane ug/kg 0.6 - 1,1-Dichloropthane ug/kg 0.6 - 1,2-Dichloropthane ug/kg 0.6 - trans-1,3-Dichloropthene ug/kg 0.7 10 Methylene chloride ug/kg 0.6 - Tetrachloroethane ug/kg 0.6 - Toluene ug/kg 0.6 - Toluene ug/kg 0.7 10 T,1,2-Titchloroethane ug/kg 0.7 10 T,1,1-Titchloroethane ug/kg 0.7 10 T,1,1-Titchloroethane ug/kg 0.7 10 T,1,1-Titchloroethane ug/kg 0.7 10 T,1,2-Titchloroethane ug/kg 0.7 10 T,1,2-Titchloroethane ug/kg 0.7 10 T,1,1-Titchloroethane ug/kg 0.7 10 T,1,2-Titchloroethane ug/kg 0.7 5 - Semivolatile organics GC/MS - (SW846 3540C/8270C) Benzoic acid ug/kg 1600 100 Benzyl alcohol ug/kg 58 50 Bis(2-chloroethyl) ether ug/kg 63 - Horophylphenyl ether ug/kg 61 - Horophylphenyl ether ug/kg 55 50 | Chloromethane | ug/kg | 0.6 | - |
| 1,2-Dichlorobenzeneug/kg1201,3-Dichlorobenzeneug/kg0.6201,4-Dichlorobenzeneug/kg0.820trans-1,2-Dichloroetheneug/kg0.6-Dichlorodifluoromethaneug/kg0.6-1,1-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloropthaneug/kg0.6-1,2-Dichloropthaneug/kg0.6-1,2-Dichloropthaneug/kg0.6-1,2-Dichloroptopeneug/kg0.6-trans-1,3-Dichloroptopeneug/kg0.6-trans-1,3-Dichloroptopeneug/kg0.6-trans-1,3-Dichloroptopeneug/kg0.6-trans-1,3-Dichloropthaneug/kg0.6-trans-1,1,2-Zietrachloroethaneug/kg0.6-trans-1,1,1,1-Trichloroethaneug/kg0.6-trichloroethaneug/kg0.5trichloroethaneug/kg0.5trichloroethaneug/kg0.6trichloroethaneug/kg0.6trichloroethaneug/kg0.6trichloroethaneug/kg0.6trichloroethaneug/kg0.6trichloroethaneug/kg0.6< | Dibromochloromethane | ug/kg | 1 | - |
| 1,3-Dichlorobenzeneug/kg0.6201,4-Dichlorobenzeneug/kg0.8201,4-Dichlorobenzeneug/kg0.6-Dichlorodifluoromethaneug/kg0.5-1,1-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloroethaneug/kg0.6-1,2-Dichloroptpaneug/kg0.6-cis-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.510Methylene chlorideug/kg0.6-1,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.6-1,1,1,1-Trichloroethaneug/kg0.6-1,1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.5-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzyl alcoholug/kg5850Bis(2-chloroethy)) etherug/kg63-Bis(2-chloroethy))nethaneug/kg63- <t< td=""><td>l,2-Dichlorobenzene</td><td>ug/kg</td><td>1</td><td>20</td></t<> | l,2-Dichlorobenzene | ug/kg | 1 | 20 |
| 1,4-Dichlorooenzene ug/kg 0.8 20 trans-1,2-Dichloroethene ug/kg 0.6 - Dichlorodifluoromethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropene ug/kg 0.6 - trans-1,3-Dichloropropene ug/kg 0.6 - trans-1,3-Dichloropropene ug/kg 0.6 - trans-1,3-Dichloropropene ug/kg 0.9 - thylbenzene ug/kg 0.5 10 Methylene chloride ug/kg 0.6 - 1,1,2,2-Tetrachloroethane ug/kg 0.6 10 Toluene ug/kg 0.6 10 Toluoroethane ug/kg 0.6 - 1,1,1,2-Trichloroethane ug/kg 0.6 - 1,1,1,2-Trichloroethane ug/kg 0.6 - Tr | 1,3-Dichlorobenzene | ug/kg | 0.6 | 20 |
| Hans-1,2-Dichlorotententeug/kg0.6-1,1-Dichlorotentaneug/kg0.5-1,1-Dichlorotentaneug/kg0.6-1,1-Dichlorotentaneug/kg0.6-1,1-Dichlorotentaneug/kg0.6-1,2-Dichloropropaneug/kg0.6-cis-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.9-ttans-1,3-Dichloropropeneug/kg0.9-ttans-1,3-Dichloropropeneug/kg0.9-ttans-1,3-Dichloropropeneug/kg0.6-ttans-1,3-Dichloropropeneug/kg0.6-ttans-1,3-Dichloropropeneug/kg0.6-ttans-1,3-Dichloropropeneug/kg0.6-ttans-1,3-Dichloropropeneug/kg0.6-ttans-1,3-Dichloropthaneug/kg0.6-1,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.6-1,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-trichlorofluoromethaneug/kg0.5-trichloroetheneug/kg0.6-trichloroetheneug/kg0.6-trichloroethaneug/kg0.5-trichloroethaneug/kg0.5-trichloroethaneug/kg0.5-trichloroethaneug/kg0.5-trichlor | 1,4-Dichlorobenzene | ug/kg | 0.8 | 20 |
| Diction D | Dichlorodifluoromethene | ug/kg | 0.6 | - |
| 1,2-Dichloroethane ug/kg 0.4 - 1,2-Dichloroethane ug/kg 0.6 - 1,1-Dichloroethane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - 1,2-Dichloropropane ug/kg 0.6 - trans-1,3-Dichloropropene ug/kg 0.9 - trans-1,3-Dichloropropene ug/kg 0.5 10 Methylene chloride ug/kg 0.6 - trans-1,2-Tetrachloroethane ug/kg 0.6 - Tetrachloroethane ug/kg 0.6 - 1,1,2,2-Tetrachloroethane ug/kg 0.6 - 1,1,2,1-Trichloroethane ug/kg 0.6 - 1,1,2-Trichloroethane ug/kg 0.5 - 1,1,2-Trichloroethane ug/kg 0.6 - Trichlorofluoromethane ug/kg 0.6 - Trichlorofluoromethane ug/kg 0.5 - Vinyl chloride ug/kg 0.5 - Semivolatile organics GC/MS - (SW846 3540C/8270C) - - | 1 I-Dichloroethane | ug/kg | 0.5 | • |
| 1,1-Dichloroethene ug/kg 0.5 - 1,2-Dichloropropane ug/kg 0.6 - cis-1,3-Dichloropropene ug/kg 0.6 - trans-1,3-Dichloropropene ug/kg 0.9 - Ethylbenzene ug/kg 0.9 - Methylene chloride ug/kg 0.6 - 1,1,2,2-Tetrachloroethane ug/kg 0.6 - Tetrachloroethane ug/kg 0.6 - Toluene ug/kg 0.6 10 1,1,1-Trichloroethane ug/kg 0.6 - Trichloroethane ug/kg 0.6 - 1,1,2-Trichloroethane ug/kg 0.6 - Trichlorofluoromethane ug/kg 0.6 - Trichlorofluoromethane ug/kg 0.6 - Vinyl chloride ug/kg 0.5 - Semivolatile organics GC/MS - (SW846 3540C/8270C) - - Benzoic acid ug/kg 58 50 Bis(2-chloroethyl) ether ug/kg 63 - Bis(2-chloroethyl) | 1.2-Dichloroethane | ug/κg | 0.4 | - |
| 1,2-Dichloropropane ug/kg 0.6 - cis-1,3-Dichloropropene ug/kg 0.9 - Ethylbenzene ug/kg 0.9 - Methylene chloride ug/kg 0.9 - 1,1,2,2-Tetrachloroethane ug/kg 0.6 - Tetrachloroethane ug/kg 0.6 - Tetrachloroethane ug/kg 0.6 10 Toluene ug/kg 0.7 10 1,1,2-Trichloroethane ug/kg 0.6 - Trichloroethane ug/kg 0.5 - 1,1,2-Trichloroethane ug/kg 0.6 - Trichloroethane ug/kg 0.6 - Trichloroethane ug/kg 0.5 - 1,1,2-Trichloroethane ug/kg 0.6 - Trichlorofluoromethane ug/kg 0.6 - Vinyl chloride ug/kg 0.5 - Semivolatile organics GC/MS - (SW846 3540C/8270C) - - Benzoic acid ug/kg 58 50 Bis(2-chloroethyl) ether < | 1.1-Dichloroethene | ug/kg | 0.5 | • |
| cis-1,3-Dichloropropeneug/kg0.6-trans-1,3-Dichloropropeneug/kg0.9-Ethylbenzeneug/kg0.510Methylene chlorideug/kg0.9-1,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.610Tolueneug/kg0.7101,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichlorofluoromethaneug/kg0.6-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg5650 | 1,2-Dichloropropane | ug/kg | 0.6 | - |
| trans-1,3-Dichloropropeneug/kg0.9-Ethylbenzeneug/kg0.510Methylene chlorideug/kg0.510I,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.610Tolueneug/kg0.7101,1,1-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-Trichloroethaneug/kg0.6-Trichloroetheneug/kg0.5-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg5850Benzoic acidug/kg63-Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Burylbenzylphthalateug/kg5650 | cis-1,3-Dichloropropene | ug/kg | 0.6 | - |
| Ethylbenzeneug/kg0.510Methylene chlorideug/kg0.9-1,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.610Tolueneug/kg0.7101,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.5-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg5850Bis(2-chloroethoxy)methaneug/kg63-Bis(2-chloroethoxy)methaneug/kg63-Bis(2-chloroethoxy)methaneug/kg53504-Bromophenyl phenyl etherug/kg5650 | trans-1,3-Dichloropropene | ug/kg | 0.9 | - |
| Methylene chlorideug/kg 0.9 -1,1,2,2-Tetrachloroethaneug/kg 0.6 -Tetrachloroethaneug/kg 0.6 10Tolueneug/kg 0.7 101,1,1-Trichloroethaneug/kg 0.5 -1,1,2-Trichloroethaneug/kg 0.6 -1,1,2-Trichloroethaneug/kg 0.6 -1,1,2-Trichloroethaneug/kg 0.6 -1,1,2-Trichloroethaneug/kg 0.6 -1,1,2-Trichloroethaneug/kg 0.6 -1,2-Trichloroethaneug/kg 0.6 -Trichloroethaneug/kg 0.6 -Vinyl chlorideug/kg 0.4 -Vinyl chlorideug/kg 0.5 -Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg 1600 100 Benzyl alcoholug/kg 58 50 Bis(2-chloroethxy) etherug/kg 63 -Bis(2-chloroethxy)methaneug/kg 53 50 Bis(2-chloroethxy) phthalateug/kg 56 50 | Ethylbenzene | ug/kg | 0.5 | 10 |
| 1,1,2,2-Tetrachloroethaneug/kg0.6-Tetrachloroethaneug/kg0.610Tolueneug/kg0.7101,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-1,2-Trichloroethaneug/kg0.6-Vinyl chlorideug/kg5850Benzoic acidug/kg5350Bis(2-chloroethyl) etherug/kg5350Bis(2-chloroethoxy)methaneug/kg5350Bis(2-ethylhexyl) phthalateug/kg61-Butylbenzylphthalateug/kg5650 | Methylene chloride | ug/kg | 0.9 | - |
| Tetrachloroetheneug/kg0.610Tolueneug/kg0.7101,1,1-Trichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-Trichloroethaneug/kg0.6-Trichloroethaneug/kg0.6-Trichloroethaneug/kg0.710Trichloroetheneug/kg0.4-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethxy)methaneug/kg69-Bis(2-chloroethxyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | 1,1,2,2-Tetrachloroethane | ug/kg | 0.6 | - |
| 101uene ug/kg 0.7 10 1,1,1-Trichloroethane ug/kg 0.5 - 1,1,2-Trichloroethane ug/kg 0.6 - Trichloroethane ug/kg 0.7 10 Trichloroethane ug/kg 0.6 - Trichloroethene ug/kg 0.4 - Vinyl chloride ug/kg 0.5 - Semivolatile organics GC/MS - (SW846 3540C/8270C) Benzoic acid ug/kg 1600 Benzyl alcohol ug/kg 58 50 Bis(2-chloroethyl) ether ug/kg 63 - Bis(2-chloroethoxy)methane ug/kg 53 50 Bis(2-chloroethoxy)methane ug/kg 53 50 Bis(2-chloroethoxy)methane ug/kg 53 50 Bis(2-chloroethoxy)methane ug/kg 61 - Butylbenzylphthalate ug/kg 56 50 | Tetrachloroethene | ug/kg | 0.6 | 10 |
| 1,1,1-1fichloroethaneug/kg0.5-1,1,2-Trichloroethaneug/kg0.6-Trichloroethaneug/kg0.710Trichlorofluoromethaneug/kg0.4-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Butylbenzylphthalateug/kg5650 | Toluene | ug/kg | 0.7 | 10 |
| 1, 1, 2-1 Inchioreethaneug/kg0.0-Trichloroetheneug/kg0.710Trichlorofluoromethaneug/kg0.4-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Butylbenzylphthalateug/kg5650 | 1,1,1-1 Inchloroethane | ug/kg | 0.5 | - |
| Inclusionug/kg0.710Trichlorofluoromethaneug/kg0.4-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600Benzoic acidug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethoxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Butylbenzylphthalateug/kg5650 | Trichloroethene | ug/kg | 0.0 | - |
| Intention deformationug/kg0.4-Vinyl chlorideug/kg0.5-Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethyl) methaneug/kg69-Bis(2-chloroethyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Bis(2-ethylhexyl) phthalateug/kg5350Butylbenzylphthalateug/kg5650 | Trichlorofluoromethane | ug/kg | 0.7 | 10 |
| Semivolatile organics GC/MS - (SW846 3540C/8270C)Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethoxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Vinyl chloride | ug/kg | 0.5 | - |
| Benzoic acidug/kg1600100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethoxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Semivolatile organics GC/MS - (SW846 3540C/2770C) | | | |
| Benzyl alcoholug/kg100Benzyl alcoholug/kg5850Bis(2-chloroethyl) etherug/kg63-Bis(2-chloroethoxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Benzoic acid | uø/kø | 1600 | 100 |
| Bis(2-chloroethyl) etherug/kg5350Bis(2-chloroethoxy)methaneug/kg69-Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Benzyl alcohol | ug/kg | 58 | 50 |
| Bis(2-chloroethoxy)methaneug/kg69Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Bis(2-chloroethyl) ether | ug/kg | 63 | - |
| Bis(2-ethylhexyl) phthalateug/kg53504-Bromophenyl phenyl etherug/kg61-Butylbenzylphthalateug/kg5650 | Bis(2-chloroethoxy)methane | ug/kg | 69 | - |
| 4-Bromophenyl phenyl etherug/kg61Butylbenzylphthalateug/kg5650 | Bis(2-ethylhexyl) phthalate | ug/kg | 53 | 50 |
| Butylbenzylphthalate ug/kg 56 50 | 4-Bromophenyl phenyl ether | ug/kg | 61 | • |
| | Butylbenzylphthalate | ug/kg | 56 | 50 |

(a) Method Detection Limit (MDL) STL-Baltimore for standard solid matrix determined according to the procedure in 40 CFR 136 Annendix B

Appendix B.(b) Target Detection Limit (USEPA/USACE 1995).

(c) For these compounds, no laboratory MDL has been determined by STL-Baltimore. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits).
| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|--|----------------------------|----------------------------------|-----------------------------------|
| 4-Chloro-3-methylphenol | ug/kg | 70 | - |
| 2-Chloronaphthalene | ug/kg | 74 | - |
| 2-Chlorophenol | ug/kg | 64 | - |
| 4-Chlorophenyl phenyl ether | ug/kg | 71 | - |
| Dibenzofuran | ug/kg | 73 | 50 |
| Di-n-butyl phthalate | ug/kg | 46 | 50 |
| 1.2-Dichlorobenzene | ug/kg | 54 | 20 |
| 1.3-Dichlorobenzene | ug/kg | 72 | 20 |
| 1.4-Dichlorobenzene | ug/kg | 73 | 20 |
| 3.3'-Dichlorobenzidine | ug/kg | 47 | - |
| 2.4-Dichlorophenol | ug/kg | 68 | - |
| Diethyl phthalate | ug/kg | 47 | 50 |
| 4.6-Dinitro-2-Methylphenol | ug/kg | 64 | - |
| 2.4-Dimethylphenol | ug/kg | 130 | 20 |
| Dimethyl phthalate | ug/kg | 55 | 50 |
| 2.4-Dinitrophenol | ug/kg | 630 | - |
| 2.4-Dinitrotoluene | ug/kg | 51 | - |
| 2.6-Dinitrotoluene | ug/kg | 58 | - |
| 1 2-Dinhenvlhydrazine | ug/kg | 55 | - |
| Di-n-octyl phthalate | ug/kg | 64 | 50 |
| Hexachlorobenzene | ng/kg | 59 | 10 |
| Hexachlorobutadiene | ng/kg | 67 | 20 |
| Heyachloroethane | ng/kg | 130 | 100 |
| Hexachlorocyclonentadiene | ug/kg | 61 | |
| Isophorone | ng/kg | 81 | |
| 2-Methylphenol | ng/kg | 77 | 50 |
| 4-Methylphenol | ug/kg | 160 | 100 |
| Nitrohenzene | ng/kg | 72 | - |
| 2-Nitrophenol | ng/kg | 55 | - |
| A-Nitrophenol | ug/kg | 52 | - |
| N-Nitrosodinbenvlamine | ug/kg | 57 | 20 |
| N-Nitrosodimethylamine | ng/kg | 57 | - |
| N-Nitroso-di-n-propylamine | ng/kg | 86 | - |
| 2 2'-Oxybis(1-chloropropage) | ug/kg | 83 | - |
| Pentachlorophenol | ug/kg | 310 | 100 |
| Phenol | ng/kg | 66 | 100 |
| 1 2 4-Trichlorobenzene | ng/kg | 75 | 10 |
| 2,4,6-Trichlorophenol | ug/kg | 77 | - |
| Polynuclear Aromatic Hydrocarbons (PAH | s) - HPLC (SW846 3540C/83) | 10) | |
| Acenaphthene | ug/kg | 9.6 | 20 |
| Acenaphthylene | ug/kg | 20 | 20 |
| Anthracene | ug/kg | 0.73 | 20 |
| Benzo[a]anthracene | ug/kg | 0.85 | 20 |
| Benzo[b]fluoranthene | ug/kg | 1.8 | 20 |
| Benzo[k]fluoranthene | ug/kg | 2.1 | 20 |
| Benzo[a]pyrene | ug/kg | 1.1 | 20 |
| Benzo[ghi]perylene | ug/kg | 2.1 | 20 |
| Chrysene | ug/kg | 1.3 | 20 |

TABLE 5-3 (CONTINUED)



Method Detection Limit (MDL) STL-Baltimore for standard solid matrix determined according to the procedure in 40 CFR 136 (a)

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Appendix B. Target Detection Limit (USEPA/USACE 1995). For these compounds, no laboratory MDL has been determined by STL-Baltimore. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits). (c)

⁽b)

TABLE 5-3 (CONTINUED)

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|-------|----------------------------------|-----------------------------------|
| Dibenzofa,h]anthracene | ug/kg | 3.4 | 20 |
| Fluoranthene | ug/kg | 2.6 | 20 |
| Fluorene | ug/kg | 1.3 | 20 |
| Indeno[1.2.3-cd]pyrene | ug/kg | 1.1 | 20 |
| 1-Methylnaphthalene | ug/kg | 12 | 20 |
| 2-Methylnaphthalene | ug/kg | 7.8 | 20 |
| Nanhthalene | ug/kg | 7.4 | 20 |
| Phenanthrene | ug/kg | 1.3 | 20 |
| Pyrene | ug/kg | 0.63 | 20 |
| Organotins by GC/FPD (STL-Burlington SOP) | | | |
| Monobutyltins | ug/kg | 1.0 (c) | 10 |
| Dibutyltins | ug/kg | 1.3 (c) | 10 |
| Tributyltins | ug/kg | 1.5 (c) | 10 |
| Dioxins/Furans-HRGC/HRMS (SW846 3545/8290) | | | |
| 2.3.7.8-TCDF | ng/kg | 0.36 | 1 |
| 2.3.7.8-TCDD | ng/kg | 0.29 | 1 |
| 1.2.3.7.8-PeCDF | ng/kg | 0.52 | 2.5 |
| 2.3.4.7.8-PeCDF | ng/kg | 0.78 | 2.5 |
| 1.2.3.7.8-PeCDD | ng/kg | 0.87 | 2.5 |
| 1.2.3.4.7.8-HxCDF | ng/kg | 0.90 | 5 |
| 1 2 3 6 7 8-HxCDF | ng/kg | 1.19 | 5 |
| 2.3.4.6.7.8-HxCDF | ng/kg | 1.07 | 5 |
| 123478-HxCDD | ng/kg | 1.26 | 5 |
| 1 2 3 6 7 8-HxCDD | ng/kg | 0.93 | 5 |
| 1.2.3.7.8 9-HxCDD | ng/kg | 1.64 | 5 |
| 1 2 3 7 8 9-HxCDF | ng/kg | 0.70 | 5 |
| 1 2 3 4 6 7 8-HpCDF | ng/kg | 1.37 | 5 |
| 1 2 3 4 6 7 8-HnCDD | ng/kg | 1.29 | 5 |
| 1 2 3 4 7 8 9-HnCDF | ng/kg | 1.38 | 5 |
| OCDD | ng/kg | 10.48 | 10 |
| OCDF | ng/kg | 2.16 | 10 |
| Inorgania Nonmetals (Conoral Organics | | | |
| Cuonido (SW846 0012 A) | malka | 0.065 | 2.0 |
| Nitragen ammonio (EBA 350.1) | mg/kg | 1 4 | 0.1 |
| Nitrogen, nitrate \pm nitrite (EPA 353.2) | mg/kg | 0.50 (c) | |
| Nitrogen, Intrate - Intrate (EFA 353.2) | malka | 43.2 | |
| Sulfda (SW846 0020P/0024) | malka | 30.7 | 0.1 |
| Suinde (SW846 90500/9054) | mg/kg | 547 | 1000 |
| POD(EBA 405.1) | mg/kg | 60 (c) | |
| COD (EPA 400.1) | mg/kg | 1000 (c) | - |
| Total Phosphorus (EPA 365.3M) | mg/kg | 2.1 | |
| Matala Cald Varian (SW046 7471 A) | | | |
| Mercury | mg/kg | 0.027 | 0.2 |
| | | | |
| Metals - Furnace (SW846 3050B/7841) | malka | 0.15 | 0.2 |
| i naniuni | mg/Kg | 0.15 | 0.2 |





Method Detection Limit (MDL) for standard solid matrix determined according to the procedure in 40 CFR 136 Appendix B.

Target Detection Limit .

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|--------------------------------------|-------|----------------------------------|-----------------------------------|
| | | | |
| Metals - ICP (SW846 3050B/6010B) | | | 50 |
| Aluminum | mg/kg | 3.7 | 50 |
| lron | mg/kg | 3.1 | 50 |
| Manganese | mg/kg | 0.78 | 5.0 |
| Zinc | mg/kg | 0.79 | 15 |
| Metals-TRACE ICP (SW846 3050B/6010B) | | | |
| Antimony | mg/kg | 0.22 | 2.5 |
| Arsenic | mg/kg | 0.093 | 5.0 |
| Bervilium | mg/kg | 0.0080 | 2.5 |
| Cadmium | mg/kg | 0. 022 | 0.3 |
| Chromium | mg/kg | 0.091 | 5.0 |
| Copper | mg/kg | 0.17 | 5.0 |
| Lead | mg/kg | 0.093 | 5.0 |
| Nickel | mg/kg | 0.25 | 5.0 |
| Selenium | mg/kg | 0.13 | 1.0 |
| Silver | mg/kg | 0.28 | 0.2 |

TABLE 5-3 (CONTINUED)



Method Detection Limit (MDL) for standard solid matrix determined according to the procedure in 40 CFR 136 Appendix B.

Target Detection Limit.

⁽a) (b) (c) For these compounds, no laboratory MDL has been determined. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits).

TABLE 5-4 METHOD DETECTION LIMITS FOR SITE WATER ANDELUTRIATE SAMPLES

| Laboratory MDL ^(a) | Recommended TDL ^(b) |
|----------------------------------|-----------------------------------|
| | |
| 0.023 | 0.04 |
| 0.010 | |
| 0.011 | |
| 0.012 | |
| 0.0081 | 0.1 |
| $0.10^{(c)}$ | 0.002 |
| 0.10 | 0.14 |
| 0.30 (c) | 0.01 |
| 0.018 | 0.01 |
| 0.010 | 0.1 |
| 0.024 | 0.1 |
| 0.020 | 0.1 |
| 0.010 | 0.02 |
| 0.019 | 0.1 |
| 0.024 | 0.1 |
| 0.029 | 0.1 |
| 0.033 | 0.1 |
| 0.032 | 0.1 |
| 0.023 | 0.1 |
| 0.019 | 0.1 |
| 0.085 | 0.5 |
| 0.10 | <u>.</u> |
| 0.49 | 0.5 |
| | |
| 0.33 | |
| 0.32 | |
| 0.29 | |
| 0.30 | |
| 0.094 | |
| 0.44 | |
| 0.41 | |
| | |
| 0.0030 | 0.01 |
| 0.0064 | 0.01 |
| 0.0065 | 0.01 |
| 0.0055 | 0.01 |
| 0.0030 | 0.01 |
| 0.0022 | 0.01 |
| 0.00045 | 0.01 |
| 0.0025 | 0.01 |
| 0.0012 | 0.01 |
| 0.0026 | 0.01 |
| 0.0034 | 0.01 |
| 0.0018 | 0.01 |
| 0.0022 | 0.01 |
| 0.0013 | 0.01 |
| 0.0013 | 0.01 |
| 0.0030 | 0.01 |
| 0.0012 | 0.01 |
| 0.0022 | 0.01 |
| 0.0014 | 0.01 |
| 0.0015 | 0.01 |
| 0.0017 | 0.01 |
| 0.0017 | 0.01 |
| | 0.0015 0.0017 0.00099 |

(a) Method Detection Limit (MDL) STL-Baltimore for standard water matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).

TABLE 5-4 (CONTINUED)

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|--|--------------|----------------------------------|-----------------------------------|
| 2,2',3,4',5,5',6-Heptachlorobiphenyl (BZ # 187) | ug/L | 0.0053 | 0.01 |
| 2,2',3,3',4,4',5,6-Octachlorobiphenyl (BZ # 195) | ug/L | 0.0017 | 0.01 |
| 2,2',3,3',4,4',5,5',6-Nonachlorobiphenyl (BZ # 206) | ug/L | 0.0024 | 0.01 |
| 2,2',3,3',4,4',5,5',6,6'-Decachlorobiphenyl (BZ # 209) | ug/L | 0.0026 | 0.01 |
| Organophosphorus Pesticides GC/NPD/FPD - (SW 3520C/8 | 3141A) | • • | |
| Demeton | ug/L | 2.0 | |
| Ethyl parathion (Parathion) | ug/L | 1 | 0.8 |
| Guthion (Azinphos methyl) | ug/L | 0.58 | |
| Malathion | ug/L | 0.22 | 0.8 |
| Methyl parathion | ug/L | 0.24 | 0.8 |
| Volatile organics GC/MS - 5 mL purge (SW846 5030B/826 | 0B) | | |
| Acrolein | ug/L | 6 | |
| Acrylonitrile | ug/L | 0 | |
| Benzene | ug/L | 0.0 | 5 |
| Bromodichloromethane | ug/L | 0.0 | 5 |
| Bromoform | ug/L | 1 | |
| Bromomethane | ug/L | 1 0 7 | |
| Carbon disulfide | ug/L | 0.7 | |
| Carbon tetrachloride | ug/L | 0.8 | |
| Chlorobenzene | ug/L | 0.8 | |
| Chloroethane | ug/L | 0.0 2 | |
| 2-Chioroethyi vinyi ether | ug/L | 0.6 | 5 |
| Chloromethene | ug/L ng/I | 0.5 | |
| Dihamashlasamathana | ug/L ug/I | 0.5 | |
| trans 1 2 Dishlarasthana | ug/L | 1 | |
| Dichlorodifluoromethane | ug/L | 2 | |
| 1 L-Dichloroethane | ug/L | 0.6 | |
| 1,7-Dichloroethane | ug/L | 0.8 | |
| 1 1-Dichloroethene | ug/L | 1 | |
| 1.2-Dichloropropane | ug/L | 0.7 | |
| cis-1.3-Dichloropropene | ug/L | 0.7 | |
| trans-1.3-Dichloropropene | ug/L | 0.7 | |
| Ethylbenzene | ug/L | 0.7 | 5 |
| Methylene chloride | ug/L | 1 | |
| 1,1,2,2-Tetrachloroethane | ug/L | 1 | |
| Tetrachloroethene | ug/L | 1 | 5 |
| Toluene | ug/L | 0.7 | 5 |
| 1,1,1-Trichloroethane | ug/L | 1 | |
| 1,1,2-Trichloroethane | ug/L | 0.7 | |
| Trichloroethene | ug/L | 1 | 5 |
| Trichlorofluoromethane Vinvl chloride | ug/L ug/L | 0.9 | |
| | -6-2 | | |
| Semivolatile organics GC/MIS - (SW840 3520C/82/0C) Bonzoia asid | μσ/Ι | 34 | 50 |
| Denzole acid Renzyl alcohol | ug/L | 2 | 50 |
| Benzyi alconor Bis(2-chloroethyl) ether | ug/I. | - 2 | |
| Bis(2-chloroethovy)methane | | 2 | |
| Bis(2-ethylberyl) phthalate | <u>-е</u> | 2 | 10 |
| 4-Bromonhenvl nhenvl ether | ug/L | 3 | |
| Rutvlhenzvlnhthalate | ug/L | 2 | 10 |
| 4-Chloro-3-methylphenol | ug/L | 1 | |
| 2-Chloronaphthalene | ug/L | 2 | |
| 2-Chlorophenol | ug/L | 1 | |

(a) Method Detection Limit (MDL) STL-Baltimore for standard water matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).

TABLE 5-4 (CONTINUED)

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|-----------------------|----------------------------------|-----------------------------------|
| 4-Chlorophenyl phenyl ether | ug/L | 2 | |
| Dibenzofuran | ug/L | 2 | 10 |
| Di-n-butyl phthalate | ug/L | 4 | 10 |
| 1,2-Dichlorobenzene | ug/L | 2 | 10 |
| 1,3-Dichlorobenzene | ug/L | 2 | 10 |
| 1,4-Dichlorobenzene | ug/L | 2 | 10 |
| 3,3'-Dichlorobenzidine | ug/L | 7 | |
| 2,4-Dichlorophenol | ug/L | 2 | |
| Diethyl phthalate | ug/L | 3 | 10 |
| 4.6-Dinitro-2-Methylphenol | ug/L | 5 | |
| 2,4-Dimethylphenol | ug/L | 4 | 10 |
| Dimethyl phthalate | ug/L | 3 | 10 |
| 2.4-Dinitrophenol | ug/L | 23 | |
| 2.4-Dinitrotoluene | ug/L | 2 | |
| 2.6-Dinitrotoluene | ug/L | 2 | |
| 1.2-Diphenvlhvdrazine | ug/L | 3 | |
| Di-n-octyl phthalate | ug/L | 3 | 10 |
| Hexachlorobenzene | ug/L | 3 | 10 |
| Hexachlorobutadiene | ug/L | 2 | 50 |
| Hexachloroethane | ug/L | 2 | 50 |
| Hexachlorocyclopentadiene | ug/L | 4 | •• |
| Isophorone | ug/L | 2 | |
| 2-Methylphenol | ug/L | 2 | |
| 4-Methylphenol | ug/L | 2 | |
| Nitrobenzene | ug/L | 3 | |
| 2-Nitrophenol | ug/L | 3 | 10 |
| 4-Nitrophenol | ug/L | 4 | 10 |
| N-Nitrosodinbenylamine | ug/I | 4 | 50 |
| N-Nitrosodimethylamine | ug/L | 3 | |
| N-Nitroso-di-n-propylamine | ug/L | 4 | |
| 2 2'-Oxybis(1-chloropropage) | ug/L | 1 | |
| Pentachlorophenol | ug/L | 2 | 50 |
| Phenol | ug/L | 2 | 10 |
| 1 2 4-Trichloroben zene | ug/L | 2 | 10 |
| 2 4 6-Trichlorophenol | ug/L | 2 | |
| | | - | |
| 'olynuclear Aromatic Hydrocarbons (PAHs) - HP | LC (SW846 3520C/8310) | 0.30 | 10 |
| Acenaphinene | ug/L | 0.39 | 10 |
| Acenaphthylene | ug/L | 0.38 | 10 |
| Anthracene | ug/L | 0.034 | 10 |
| Benzolajanthracene | ug/L | 0.031 | 10 |
| Benzo[b]fluoranthene | ug/L | 0.034 | 10 |
| Benzo[k]fluoranthene | ug/L | 0.053 | 10 |
| Benzo[a]pyrene | ug/L | 0.047 | 10 |
| Benzo[ghi]perylene | ug/L | 0.066 | 10 |
| Chrysene | ug/L | 0.024 | 10 |
| Dibenzo[a,h]anthracene | ug/L | 0.065 | 10 |
| Fluoranthene | ug/L | 0.047 | 10 |
| Fluorene | ug/L | 0.064 | 10 |
| Indeno[1,2,3-cd]pyrene | ug/L | 0.035 | 10 |
| I-Methylnaphthalene | ug/L | 0.31 | 10 |
| 2-Methylnaphthalene | ug/L | 0.21 | 10 |
| Naphthalene | ug/L | 0.32 | 10 |
| Phenanthrene | ug/L | 0.034 | 10 |
| | · 0 = | | |

Dioxins/Furans - HRGC/HRMS (SW846 3520C/8290)

⁽a) Method Detection Limit (MDL) STL-Baltimore for standard water matrix determined according to the procedure in 40 CFR 136 Appendix B.

⁽b) Target Detection Limit (USEPA/USACE 1995).

⁽c) For these compounds, no laboratory MDL has been determined by STL-Baltimore. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits).

TABLE 5-4 (CONTINUED)

| 2,3,7,8-TCDF rg/L 0.0023 0.01 2,3,7,8-TCDD rg/L 0.0038 0.01 1,3,3,7,8-TCDD rg/L 0.0038 0.0125 2,3,4,7,8-FCDF rg/L 0.0026 0.025 1,2,3,7,8-HCDF rg/L 0.0064 0.025 1,2,3,4,7,8-HCDF rg/L 0.0064 0.025 1,2,3,4,7,8-HCDF rg/L 0.0086 0.05 1,2,3,4,7,8-HCDD rg/L 0.0086 0.05 1,2,3,4,7,8-HCDD rg/L 0.0086 0.05 1,2,3,4,7,8-HCDD rg/L 0.0083 0.05 1,2,3,4,7,8-HCDD rg/L 0.018 0.05 1,2,3,4,6,7,8-HCDD rg/L 0.0108 0.05 1,2,3,4,6,7,8-HCDF rg/L 0.0128 0.05 1,2,3,4,6,7,8-HCDF rg/L 0.0128 0.05 1,2,3,4,6,7,8-HCDF rg/L 0.018 0.05 1,2,3,4,6,7,8-HCDF rg/L 0.018 0.05 1,2,3,4,6,7,8-HCDF rg/L 0.031 0.01 <th>Parameter</th> <th>Units</th> <th>Laboratory MDL ^(a)</th> <th>Recommended TDL^(b)</th> | Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|---|----------------------|----------------------------------|-----------------------------------|
| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | 2 3 7 8-TCDF | ng/L | 0.0023 | 0.01 |
| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | 2 3 7 8-TCDD | ng/L | 0.0038 | 0.01 |
| 2.3,47,5,9,000 ng/L 0.0122 0.025 1,2,3,47,5,8+4CDP ng/L 0.0064 0.025 1,2,3,47,5,8+4CDF ng/L 0.0043 0.035 1,2,3,47,5,8+4CDF ng/L 0.0085 0.055 1,2,3,47,5,8+4CDF ng/L 0.0085 0.055 1,2,3,47,5,8+4CDF ng/L 0.0083 0.055 1,2,3,47,5,8+4CDD ng/L 0.0083 0.055 1,2,3,4,7,8-94CDF ng/L 0.0135 0.055 1,2,3,4,7,8-94CDF ng/L 0.0104 0.055 1,2,3,4,7,8-94CDF ng/L 0.0102 0.055 1,2,3,4,7,8-94CDF ng/L 0.0124 0.055 0,2,3,4,7,8,9-41pCDF ng/L 0.0124 0.055 0,2,3,4,7,8,9-41pCDF ng/L 0.0184 0.01 0,02D ng/L 0.0383 0.1 0,02F ng/L 0.031 (c) 0.01 Diburytins ug/L 0.031 (c) 0.01 Triburytins ug/L 0.038 (c) 0.01 Triburytins ug/L 0.035 0.1 | 1 2 3 7 8-PeCDF | ng/L | 0.0080 | 0.025 |
| 1.3.3.7.5-PCDD ng/L 0.0064 0.025 1.2.3.7.5-PCDD ng/L 0.0095 0.05 1.2.3.4.7.8-HxCDF ng/L 0.0043 0.05 1.2.3.4.7.8-HxCDF ng/L 0.0083 0.05 1.2.3.4.7.8-HxCDD ng/L 0.0096 0.05 1.2.3.4.7.8-HxCDD ng/L 0.0093 0.05 1.2.3.4.7.8-HxCDD ng/L 0.0102 0.05 1.2.3.4.7.8-HxCDF ng/L 0.0124 0.05 CDD ng/L 0.0341 0.1 DCDF ng/L 0.0341 0.1 CDD ng/L 0.035 0.1 Dbutylins ug/L 0.035 0.01 Tributylins ug/L 0.035 0.1 Nimogen, anmonia (EPA 350.1)< | 2 3 4 7.8-PeCDF | ng/L | 0.0122 | 0.025 |
| 1.2.3.4.7,8+HCDF mg/L 0.0043 0.05 1.2.3.4.7,8+HCDF mg/L 0.0043 0.05 1.2.3.4.7,8+HCDF mg/L 0.0085 0.05 1.2.3.4.7,8+HCDF mg/L 0.0083 0.05 1.2.3.4.7,8+HCDF mg/L 0.0083 0.05 1.2.3.7,8.9+HCDF mg/L 0.0135 0.05 1.2.3.4.7,8.9-HCDF mg/L 0.0102 0.05 1.2.3.4.7,8.9-HCDF mg/L 0.0104 0.05 1.2.3.4.7,8.9-HCDF mg/L 0.0124 0.05 0.2.3.4.6.7,8-HCDF mg/L 0.0124 0.05 0.7.3.4.7,8.9-HCDF mg/L 0.0124 0.05 0.7.2.3.4.7,8.9-HCDF mg/L 0.0383 0.1 0.7.2.3.4.7,8.9-HCDF mg/L 0.0341 0.1 0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0.0. | 1 2 3 7 8-PeCDD | ng/L | 0.0064 | 0.025 |
| 1.23.6.7.8-HxCDF mg/L 0.0043 0.05 2.3.46,7.8-HxCDF ng/L 0.0085 0.05 1.2.3.47,8-HxCDD ng/L 0.0096 0.05 1.2.3.4,7,8-HxCDD ng/L 0.0093 0.05 1.2.3.4,7,8-HxCDD ng/L 0.0093 0.05 1.2.3.4,6,7.8-HxCDF ng/L 0.012 0.05 1.2.3.4,6,7.8-HxCDF ng/L 0.012 0.05 1.2.3.4,6,7.8-HxCDD ng/L 0.012 0.05 1.2.3.4,6,7.8-HxCDF ng/L 0.012 0.05 1.2.3.4,6,7.8-HxCDF ng/L 0.0124 0.05 1.2.3.4,6,7.8-HyCDF ng/L 0.0124 0.05 0.2.3.4,6,7.8-HyCDF ng/L 0.0341 0.1 0.CDF ng/L 0.031 (c) 0.01 DCganotins by GC/FPO (STL-Burlington SOP) W 0.038 (c) 0.01 Tributyltins ug/L 0.038 (c) 0.01 Dibutyltins ug/L 0.038 (c) 0.01 Nirrogen, nitract + nitric (EPA 351.2) mg/L 0.020 Nitridgen, nitract + nitric (EPA 351.2) <td>1,2,3,4,7,8-HxCDF</td> <td>ng/L</td> <td>0.0095</td> <td>0.05</td> | 1,2,3,4,7,8-HxCDF | ng/L | 0.0095 | 0.05 |
| 2.3.4.6.7.8.+inCDP ng/L 0.0085 0.05 1.2.3.6.7.8.+inCDD ng/L 0.0086 0.05 1.2.3.6.7.8.+inCDD ng/L 0.0083 0.05 1.2.3.7.8.9.+inCDF ng/L 0.0083 0.05 1.2.3.7.8.9.+inCDF ng/L 0.0135 0.05 1.2.3.7.8.9.+inCDF ng/L 0.0102 0.05 1.2.3.4.6.7.8.+inCDF ng/L 0.0124 0.05 1.2.3.4.7.8.+inCDF ng/L 0.0381 0.1 Organetins by CC/FPD (STL-Barlington SOP) ng/L 0.031 (c) 0.01 Monoburytins ug/L 0.031 (c) 0.01 0.01 Diburytins ug/L 0.031 (c) 0.01 0.01 Inorgen, ammonia (EPA 350.1) mg/L 0.032 0.03 Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.020 - Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.020 - Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.033 0.1 OC (SW846 900) mg/L <t< td=""><td>1.2.3.6.7.8-HxCDF</td><td>ng/L</td><td>0.0043</td><td>0.05</td></t<> | 1.2.3.6.7.8-HxCDF | ng/L | 0.0043 | 0.05 |
| 1.2.3.4.7.8.HICDD ng/L 0.0096 0.05 1.2.3.7.8.HICDD ng/L 0.0083 0.05 1.2.3.7.8.HICDD ng/L 0.0083 0.05 1.2.3.7.8.HICDD ng/L 0.0105 0.05 1.2.3.4.7.8.HICDF ng/L 0.0108 0.05 1.2.3.4.7.8.HICDF ng/L 0.0108 0.05 1.2.3.4.7.8.HICDF ng/L 0.0108 0.05 1.2.3.4.7.8.HICDF ng/L 0.0124 0.05 0.CDF ng/L 0.0341 0.1 OCDF ng/L 0.0316 (c) 0.01 Dibutyltins ug/L 0.031 (c) 0.01 Dibutyltins ug/L 0.038 (c) 0.01 Dibutyltins ug/L 0.038 (c) 0.01 Tributyltins ug/L 0.044 (c) 0.01 Dibutyltins ug/L 0.0050 5 Nitrogen, annnonia (EPA 35.1) mg/L 0.028 0.03 Nitrogen, initat + nitrific (EPA 35.2) mg/L 0.020 Nitrogen, initat + nitrific (EPA 35.1) mg/L 0.028 0 | 2.3.4.6.7.8-HxCDF | ng/L | 0.0085 | 0.05 |
| 123.67.7.8-HrCDD ng/L 0.0083 0.05 12.3.7.8.9-HrCDD ng/L 0.0093 0.05 12.3.7.8.9-HrCDF ng/L 0.0102 0.05 12.3.4.67.8-HrCDF ng/L 0.0102 0.05 12.3.4.7.8.+HrCDF ng/L 0.0102 0.05 12.3.4.7.8.+HrCDF ng/L 0.0124 0.05 0.CDD ng/L 0.0383 0.1 Organotics by GC/FPD (STL-Burlington SOP) mg/L 0.031 (c) 0.01 Monoburyltins ug/L 0.031 (c) 0.01 10 Diburyltins ug/L 0.038 (c) 0.01 10 Inorganic nonmetal/general organics ug/L 0.044 (c) 0.01 Inorganic nonmetal/general organics ug/L 0.020 - Cyanide (SW846 9012A) mg/L 0.020 - Nitrogen, nitrate + nitrite (EPA 35.2) mg/L 0.020 - Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.020 - Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.03 0.1 ODD (EPA 405.1) mg/L 0.03 <td>1.2.3.4.7.8-HxCDD</td> <td>ng/L</td> <td>0.0096</td> <td>0.05</td> | 1.2.3.4.7.8-HxCDD | ng/L | 0.0096 | 0.05 |
| 12.3.7.89-HxCDD mg/L 0.0093 0.05 12.3.7.89-HxCDF mg/L 0.0135 0.05 12.3.3,4.67,8-HpCDF mg/L 0.0102 0.05 12.3.4,67,8-HpCDF mg/L 0.0124 0.05 12.3.4,67,8-HpCDF mg/L 0.0124 0.05 OCDD mg/L 0.0341 0.1 OCDF mg/L 0.0383 0.1 Organotics by GC/FP0 (STL-Burlington SOP) | 1.2.3.6.7.8-HxCDD | ng/L | 0.0083 | 0.05 |
| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | 1.2.3.7.8.9-HxCDD | ng/L | 0.0093 | 0.05 |
| 1,2,3,4,6,7,8-HpCDF ng/L 0,0102 0.05 1,2,3,4,7,8-HpCDD ng/L 0,0108 0.05 1,2,3,4,7,8-HpCDF ng/L 0,0124 0.05 OCDD ng/L 0,0341 0.1 OCDF ng/L 0,0383 0.1 Organotins by GC/FPD (STL-Burlington SOP) | 1.2.3.7.8.9-HxCDF | ng/L | 0.0135 | 0.05 |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 1.2.3.4.6.7.8-HpCDF | ng/L | 0.0102 | 0.05 |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 1.2.3.4.6.7.8-HpCDD | ng/L | 0.0108 | 0.05 |
| OCDD ng/L 0.0341 0.1 OCDF ng/L 0.0383 0.1 Organotias by GC/FPD (STL-Burlington SOP) Monoburyltins ug/L 0.038 (c) 0.01 Diburyltins ug/L 0.038 (c) 0.01 Tributyltins ug/L 0.044 (c) 0.01 Inorganic nonmetals/general organics Cyanide (SW846 9012A) mg/L 0.028 0.03 Nitrogen, nitrate + nitrite (EPA 353.2) mg/L 0.020 Nitrogen, nitrate + nitrite (EPA 351.2) mg/L 0.15 0.1 TOC (SW846 9060) mg/L 0.37 - Sulfide (SW846 9034) mg/L 0.37 - ToC (SW846 9060) mg/L 0.039 0.2 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) - Mercury ug/L 0.039 0.2 Metals - ICP (SW846 3010A/6010B) - - Aluminum ug/L 2.3 1.0 | 1.2.3.4.7.8.9-HpCDF | ng/L | 0.0124 | 0.05 |
| OCDF ng/L 0.0383 0.1 Organotins by GC/FPD (STL-Burlington SOP) | OCDD | ng/L | 0.0341 | 0.1 |
| Organotins by GC/FPD (STL-Burlington SOP) Monobutyltins ug/L 0.031 (c) 0.01 Dibutyltins ug/L 0.038 (c) 0.01 Inorganic nonmetals/general organics U 0.044 (c) 0.01 Cyanide (SW846 9012A) mg/L 0.0050 5 Nitrogen, ammonia (EPA 350.1) mg/L 0.028 0.03 Nitrogen, ammonia (EPA 351.2) mg/L 0.020 Nitrogen, total Kjeldahl (EPA 353.2) mg/L 0.035 0.1 TOC (SW846 9034) mg/L 0.37 - - Sulfide (SW846 9060) mg/L 0.37 - - TOC (SW846 9060) mg/L 0.037 - - DOD (EPA 405.1) mg/L 0.014 - Total Phosphorus (EPA 365.3) mg/L 0.014 - Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 2.4 1.0 Metals - Furnace (SW846 3010A/6010B) | OCDF | ng/L | 0.0383 | 0.1 |
| Monobutyltins ug/L 0.031 (c) 0.01 Dibutyltins ug/L 0.038 (c) 0.01 Tributyltins ug/L 0.038 (c) 0.01 Inorganic nonmetals/general organics Cyanide (SW 846 9012A) mg/L 0.028 0.03 Nitrogen, ammonia (EPA 350.1) mg/L 0.028 0.03 Nitrogen, intrate + nitrite (EPA 353.2) mg/L 0.19 Nitrogen, notal Kjeldahl (EPA 351.2) mg/L 0.35 0.1 TOC (SW 846 9060) mg/L 0.080 1000 BOD (EPA 405.1) mg/L 0.37 - - - COD (EPA 405.1) mg/L 0.37 - - - - Mercury ug/L 0.37 - - - - - Mercury ug/L 0.37 - - - - - Metals - Autoclave Digestion - Cold Vapor (SW 846 7470A) mg/L 0.039 0.2 - Metals - Furnace (SW 846 3010A/6010B) - - - | Organotins by GC/FPD (STL-Burlington SOP) | | | |
| Dibutyltins Tributyltins ug/L $0.038 (c)$ 0.01 Inorganic nometals/general organics ug/L $0.044 (c)$ 0.01 Inorganic nometals/general organics mg/L 0.0050 5 Cyanide (SW846 9012A) mg/L 0.028 0.03 Nitrogen, ammonia (EPA 350.1) mg/L 0.020 $-$ Nitrogen, total Kjeldahl (EPA 351.2) mg/L 0.020 $-$ Sulfde (SW846 9060) mg/L 0.035 0.1 TOC (SW846 9060) mg/L 0.37 $-$ COD (EPA 410.4) mg/L 0.37 $-$ Total Phosphorus (EPA 365.3) mg/L 0.0014 $-$ Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3010A/7841) ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 2.4 1.0 Manganese ug/L 3.0 3.0 3.0 Zinc ug/L 0.74 1.0 | Monobutyltins | ug/L | 0.031 (c) | 0.01 |
| Tributyltins ug/L 0.044 (c) 0.01 Inorganic nonmetals/general organics | Dibutyltins | ug/L | 0.038 (c) | 0.01 |
| Ioorganic nonmetals/general organics mg/L 0.0050 5 Cyanide (SW846 9012A) mg/L 0.0028 0.03 Nitrogen, annunonia (EPA 350.1) mg/L 0.028 0.03 Nitrogen, nitrate + intrine (EPA 351.2) mg/L 0.19 Nitrogen, nitrate + intrine (EPA 351.2) mg/L 0.35 0.1 TOC (SW846 9060) mg/L 0.37 Sulfide (SW846 9060) mg/L 0.37 COD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 0.0014 Total Phosphorus (EPA 365.3) mg/L 0.039 0.2 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW866 3020A/7841) mg/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 6.6 1.0 Zinc ug/L 3.0 3.0 Manganese ug/L 1.7 1.0 | Tributyltins | ug/L | 0.044 (c) | 0.01 |
| Cyanide (SW846 9012A) mg/L 0.0050 5 Nitrogen, ammonia (EPA 350.1) mg/L 0.028 0.03 Nitrogen, nitrate + intrite (EPA 353.2) mg/L 0.020 Nitrogen, nitrate + intrite (EPA 351.2) mg/L 0.19 Sulfide (SW846 9034) mg/L 0.35 0.1 TOC (SW846 9060) mg/L 0.030 1000 BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 0.37 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) Metals - Furnace (SW846 3010A/6010B) Aluminum ug/L 57.6 40 Iron ug/L 57.6 40 Metals - ICP (SW846 3010A/6010B) < | Inorganic nonmetals/general organics | | | |
| Nitrogen, ammonia (EPA 350.1) mg/L 0.028 0.03 Nitrogen, nitrate + nitrite (EPA 353.2) mg/L 0.020 Nitrogen, total Kjeldahl (EPA 351.2) mg/L 0.19 Sulfide (SW846 9034) mg/L 0.35 0.1 TOC (SW846 9060) mg/L 0.37 COD (EPA 405.1) mg/L 0.37 COD (EPA 405.1) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) mg/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 57.6 40 Iron ug/L 6.6 1.0 Manganese ug/L 6.6 1.0 Zinc ug/L 1.7 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.7 1.0 Antimony ug/L 0.63 0.2 3.0 Antimony ug/L 0.74 1.0 0.083 0.2 <td< td=""><td>Cyanide (SW846 9012A)</td><td>mg/L</td><td>0.0050</td><td>5</td></td<> | Cyanide (SW846 9012A) | mg/L | 0.0050 | 5 |
| Nitrogen, nitrate + nitrite (EPA 353.2) mg/L 0.020 Nitrogen, total Kjeldahl (EPA 351.2) mg/L 0.19 Sulfide (SW846 9034) mg/L 0.35 0.1 TOC (SW846 9060) mg/L 0.080 1000 BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 0.37 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) Antimony ug/L 3.0 3.0 Arsenic ug/L 1.7 1.0 Beryllium ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 | Nitrogen, ammonia (EPA 350.1) | mg/L | 0.028 | 0.03 |
| Nirrogen, total Kjeldahl (EPA 351.2) mg/L 0.19 Sulfide (SW846 9034) mg/L 0.35 0.11 TOC (SW846 9060) mg/L 0.35 0.11 BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 4.4 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 57.6 40 Iron ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.7 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 0.24 1.0 Cadmium ug/L 0.24 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 | Nitrogen, nitrate + nitrite (EPA 353.2) | mg/L | 0.020 | |
| Sulfide (SW846 9034) mg/L 0.35 0.1 TOC (SW846 9060) mg/L 0.080 1000 BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 0.0014 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) mg/L 2.4 1.0 Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 42.8 10 Aluminum ug/L 57.6 40 Iron ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.7 1.0 Arsenic ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Copper ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 | Nitrogen, total Kjeldahl (EPA 351.2) | mg/L | 0.19 | |
| TOC (SW846 9060) mg/L 0.080 1000 BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 4.4 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 2.4 1.0 Aluminum ug/L 57.6 40 Iron ug/L 6.6 1.0 Zinc ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 0.083 0.2 Cadmium ug/L 0.083 0.2 1.0 Motosci ug/L 1.0 0.083 0.2 Metals-Trace ICP (SW846 3010A/6010B) ug/L 0.74 1.0 Cadmium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Copper ug/L 1.4 <t< td=""><td>Sulfide (SW846 9034)</td><td>mg/L</td><td>0.35</td><td>0.1</td></t<> | Sulfide (SW846 9034) | mg/L | 0.35 | 0.1 |
| BOD (EPA 405.1) mg/L 0.37 COD (EPA 410.4) mg/L 4.4 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) mg/L 2.4 1.0 Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 42.8 10 Aluminum ug/L 6.6 1.0 Iron ug/L 3.0 3.0 Manganese ug/L 1.7 1.0 Zinc ug/L 1.7 1.0 Metals-Trace ICP (SW846 3010A/6010B) | TOC (SW846 9060) | mg/L | 0.080 | 1000 |
| COD (EPA 410.4) mg/L 4.4 Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) mg/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) ug/L 2.4 1.0 Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.7 1.0 Antimony ug/L 0.083 0.2 Antimony ug/L 0.083 0.2 Antimony ug/L 0.74 1.0 Cadmium ug/L 0.74 1.0 Copper ug/L 0.74 1.0 Nickel ug/L 1.1 1.0 | BOD (EPA 405.1) | mg/L | 0.37 | *- |
| Total Phosphorus (EPA 365.3) mg/L 0.0014 Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) Mercury ug/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) Aluminum ug/L 57.6 40 Iron ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 Nickel ug/L 1.9 1.0 | COD (EPA 410.4) | mg/L | 4.4 | |
| Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) Mercury ug/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) Aluminum ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.0 3.0 Antimony ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 Nickel ug/L 1.9 1.0 | Total Phosphorus (EPA 365.3) | mg/L | 0.0014 | *= |
| Mercury ug/L 0.039 0.2 Metals - Furnace (SW846 3020A/7841) Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) Aluminum ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) Antimony ug/L 3.0 3.0 Arsenic ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 1.9 1.0 Nickel ug/L 1.9 1.0 Nickel ug/L 1.1 1.0 | Metals - Autoclave Digestion - Cold Vapor (SW846 7470A) | | | |
| Metals - Furnace (SW846 3020A/7841) Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Beryllium ug/L 0.0833 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 0.74 1.0 Nickel ug/L 2.4 1.0 | Mercury | ug/L | 0.039 | 0.2 |
| Thallium ug/L 2.4 1.0 Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 1.7 1.0 Antimony ug/L 1.7 1.0 Arsenic ug/L 0.0833 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Nickel ug/L 2.4 1.0 | Metals - Furnace (SW846 3020A/7841) | | | |
| Metals - ICP (SW846 3010A/6010B) ug/L 57.6 40 lron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 10 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 3.0 3.0 Arsenic ug/L 0.083 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Nickel ug/L 1.9 1.0 Nickel ug/L 1.0 1.0 | Thallium | ug/L | 2.4 | 1.0 |
| Aluminum ug/L 57.6 40 Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Ug/L 1.9 1.0 1.0 | Metals - ICP (SW846 3010A/6010B) | a | 67 6 | 40 |
| Iron ug/L 42.8 10 Manganese ug/L 6.6 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 1.9 1.0 Nickel ug/L 1.9 1.0 Lead ug/L 1.1 1.0 | | ug/L | 5/.0 42.9 | 40 |
| Manganese ug/L 0.0 1.0 Zinc ug/L 2.3 1.0 Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.74 1.0 Chromium ug/L 1.9 1.0 Nickel ug/L 1.9 1.0 Lead ug/L 1.1 1.0 | Iron | ug/L | 42.8 | 10 |
| Metals-Trace ICP (SW846 3010A/6010B) ug/L 3.0 3.0 Antimony ug/L 1.7 1.0 Arsenic ug/L 0.083 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 1.9 1.0 Lead ug/L 1.1 1.0 | Manganese Zinc | ug/L ug/L | 2.3 | 1.0 |
| Antimony ug/L 3.0 3.0 Arsenic ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 | Metals-Trace ICP (SW846 3010 & /6010B) | | | |
| Arsenic ug/L 1.7 1.0 Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 | Antimony | ng/L | 3.0 | 3.0 |
| Beryllium ug/L 0.083 0.2 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Lead ug/L 1.1 1.0 | A reenic | ug/L | 17 | 1.0 |
| Cadmium ug/L 0.00 0.12 Cadmium ug/L 0.24 1.0 Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Lead ug/L 1.1 1.0 | Bervllium | ug/L | 0.083 | 0.2 |
| Chromium ug/L 0.74 1.0 Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Lead ug/L 1.1 1.0 | Cadmium | ug/L | 0.24 | 1.0 |
| Copper ug/L 1.9 1.0 Nickel ug/L 2.4 1.0 Lead ug/L 1.1 1.0 | Chromium | - <i>g ~</i> ug/L | 0.74 | 1.0 |
| Nickel ug/L 2.4 1.0 Lead ug/L 1.1 1.0 | Copper | ug/L | 1.9 | 1.0 |
| Lead ug/L 1.1 1.0 | Nickel | ug/L | 2.4 | 1.0 |
| | Lead | ug/L | 1.1 | 1.0 |
| Selenium ug/L 1.8 2.0 | Selenium | ug/L | 1.8 | 2.0 |
| Silver ug/L 3.2 1.0 | Silver | ug/L | 3.2 | 1.0 |

(a) Method Detection Limit (MDL) STL-Baltimore for standard water matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).

TABLE 5-5 METHOD DETECTION LIMITS FOR TISSUE SAMPLES

| Organochlorine Pesticides - GC/ECD - (SW846 3540C/8081A)Aldrin yg/kg α -BHC yg/kg β -BHC yg/kg λ -BHC (Lindane) yg/kg λ -Det yg/kg λ -A'-DDD yg/kg λ -A'-DDT yg/kg Endosulfan 1 yg/kg Endosulfan 1 yg/kg Endosulfan 1 yg/kg Endrin aldehyde yg/kg Heptachlor epoxide yg/kg Mirex yg/kg Methoxychlor yg/kg PCB Arcolors - GC/ECD - (SW846 3540C/8082) Aroclor 1016 yg/kg Aroclor 1221 yg/kg Aroclor 1242 yg/kg Aroclor 1242 yg/kg Aroclor 1254 yg/kg Aroclor 1260 yg/kg 2,2',5'.5'-Tetrachlorobiphenyl (BZ # 18) yg/kg 2,2',5'.5'-Tetrachlorobiphenyl (BZ # 20) yg/kg 2,2',5'.5'-Tetrachlorobiphenyl (BZ # 105) yg/kg 2,2',3,4'-Tetrachlorobiphenyl (BZ # 105) yg/kg 2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 105) yg/kg 2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 101) yg/kg 2,2',3,4,4'-S-Hentachlorobiphenyl (BZ # 126) yg/kg 2,2',3,4,4'-S-Hentachlorobiphenyl (BZ # 105) | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|----------------------------------|-----------------------------------|
| Aldrin ug/kg α -BHC ug/kg β -BHC ug/kg β -BHC ug/kg β -BHC ug/kg α -DDD ug/kg α -DDD ug/kg α -DDD ug/kg α -DDD ug/kg α -DDT ug/kg Endosulfan 1 ug/kg α -DDT ug/kg α -Dot ug/kg α -Dot ug/kg α -Color 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1242 ug/kg Aroclor 1242 ug/kg Aroclor 1242 ug/kg Aroclor 1244 ug/kg α -Dot 1254 ug/kg α -Aroclor 1260 ug/kg | | |
| a-BHCug/kg β -BHCug/kg β -BHCug/kg λ -BHC (Lindane)ug/kg λ -BHC (Lindane)ug/kg λ -BHC (Lindane)ug/kg λ -BHC (Lindane)ug/kg λ -DDrug/kg λ -DDDug/kg λ -DDDug/kg λ -DDTug/kg λ -DDTug/kgDieldrinug/kgEndosulfan Iug/kgEndosulfan IIug/kgEndosulfan IIug/kgEndosulfan sulfateug/kgEndrin idehydeug/kgHeptachlorug/kgWirexug/kgVethoxychlorug/kgFocapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1232ug/kgAroclor 1243ug/kgAroclor 1254ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)PCB Congeners - GC/ECD - (SW846 3540C/8082)Aroclor 1248ug/kgAroclor 1254ug/kgAroclor 1254ug/kg2,2,5,5-Tertachlorobiphenyl (BZ # 18)ug/kg2,2,3,5'-Tetrachlorobiphenyl (BZ # 28)ug/kg2,2,3,5'-Tetrachlorobiphenyl (BZ # 40)ug/kg2,2,3,4,4'-Pentachlorobiphenyl (BZ # 101)ug/kg2,3,4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3,4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3,4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3,4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg </td <td>0.53</td> <td>10</td> | 0.53 | 10 |
| 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg 3-BHC ug/kg Chlordane (Technical) ug/kg 3-Chorbenzide ug/kg 3-Arcolor 10D ug/kg 4-DDD ug/kg 3-DDD ug/kg 3-DD ug/kg 3 | 0.72 | - |
| | 0.77 | - |
| PDIC Generation of the set of t | 0.69 | _ |
| PCB Congeners - GC/ECD - (SW846 3540C/8082) $PCB Congeners - GC/ECD - (SW846 3540C/8082)$ $PCB Congeners - GC/ECD + (SW846 3540C/8082)$ $PCB Congeners - GC/E$ | 0.82 | 10 |
| Definition of the set | 2 2 (c) | 2 |
| Chordane (rechnical)up/kgActhalug/kg1,4'-DDDug/kg1,4'-DDTug/kgEndosulfan Iug/kgEndosulfan Iug/kgEndosulfan IIug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgEndosulfan Sulfateug/kgHeptachlor epoxideug/kgVierxug/kgVierxug/kgVierxug/kgVacolor 1016ug/kgAroclor 1221ug/kgAroclor 1232ug/kgAroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)PCB Congeners - GC/ECD - (SW846 3540C/8082)P2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,4'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 128)ug/kg <t< td=""><td>3.3</td><td>10</td></t<> | 3.3 | 10 |
| DactnalUp Kg44-DDDug/kg44-DDTug/kgbieldrinug/kgEndosulfan Iug/kgEndosulfan Iug/kgEndosulfan sulfateug/kgEndosulfan sulfateug/kgEndosulfan sulfateug/kgEndrinug/kgEndrinug/kgEndrinug/kgEndrinug/kgSendrin aldehydeug/kgHeptachlorug/kgHeptachlor epoxideug/kgWirexug/kgVirexug/kgVethoxychlorug/kgFoxapheneug/kgAroclor 1016ug/kgAroclor 1221ug/kgAroclor 1232ug/kgAroclor 1243ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,2',5.Trichlorobiphenyl (BZ # 8)ug/kg2,2',5.Trichlorobiphenyl (BZ # 8)ug/kg2,2',5.Trichlorobiphenyl (BZ # 18)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,4'-Tetrachlorobiphenyl (BZ # 87)ug/kg2,2',3,4'-Tetrachlorobiphenyl (BZ # 101)ug/kg2,2',3,4'-Tetrachlorobiphenyl (BZ # 105)ug/kg2,2',3,4'-Fentachlorobiphenyl (BZ # 105)ug/kg2,2',3,4'-Fentachlorobiphenyl (BZ # 105)ug/kg2,2',3,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,4'-S-Pen | 10 (c) | 10 |
| A_4 -DDDug/kg A_4 -DDEug/kg A_4 -DDTug/kgDieldrinug/kgEndosulfan Iug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgEndosulfan 11ug/kgHeptachlorug/kgHeptachlor epoxideug/kgWethoxychlorug/kgOxapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1222ug/kgAroclor 1248ug/kgAroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,2',5,5'-Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 101)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',4,4'-Tetrachlorobiphenyl (BZ # 101)ug/kg2,2',4,4'-Tetrachlorobiphenyl (BZ # 118)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',4,4'-S-Pentachlorobiphenyl (BZ # 128) | 15 | 10 |
| A^{+} -DDug/kgDieldrinug/kgDieldrinug/kgEndosulfan IIug/kgEndosulfan IIug/kgEndosulfan sulfateug/kgEndrinug/kgEndrinug/kgEndrinug/kgHeptachlorug/kgHeptachlor epoxideug/kgMirexug/kgMethoxychlorug/kgFOX apheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1221ug/kgAroclor 1221ug/kgAroclor 1221ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1260ug/kg2,4-Dichlorobiphenyl (BZ # 8)ug/kg2,2,5.5.*Tetrachlorobiphenyl (BZ # 44)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 49)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 77)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 77)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 10)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 118)ug/kg2,2,3.5.*Tetrachlorobiphenyl (BZ # 128)ug/kg2,2,3.3,4,4.*S-Pentachlorobiphenyl (BZ # 118)ug/kg2,3,4,4.*S-Pentachlorobiphenyl (BZ # 118)ug/kg2,3,4,4.*S-Hetachlorobiphenyl (BZ # 128)ug/kg2,2,3,3,4,4.*S-Hetachlorobiphenyl (BZ # 118)ug/kg< | 1.5 | 10 |
| A_{4} -DD1ug/kgDieldrinug/kgEndosulfan Iug/kgEndosulfan IIug/kgEndosulfan sulfateug/kgEndrinug/kgEndrinug/kgEndrinug/kgHeptachlorug/kgHeptachlor epoxideug/kgMirexug/kgMethoxychlorug/kgFoxapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1016ug/kgAroclor 1221ug/kgAroclor 1222ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1260ug/kg2,2',5.5'.Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',5,5'.Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5'.Tetrachlorobiphenyl (BZ # 45)ug/kg2,2',4,5'.Pentachlorobiphenyl (BZ # 101)ug/kg2,2',4,5,5'.Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,3',4,4'.Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 165)ug/kg2,2',3,3',4,4'.S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'.S-Heptachlorobiphenyl (BZ # 165)ug/kg2,2',3,3',4,4'.S-Heptachlorobiphenyl (BZ # 166)ug/kg2,2',3,3',4,4'.S-Heptachlorobiphenyl (BZ # | 1.3 | 10 |
| Deldrinug/kgEndosulfan Iug/kgEndosulfan IIug/kgEndosulfan sulfateug/kgEndosulfan sulfateug/kgEndrin aldehydeug/kgHeptachlorug/kgHeptachlorug/kgHeptachlorug/kgMirexug/kgMethoxychlorug/kgFoxapheneug/kgAroclor 1016ug/kgAroclor 1221ug/kgAroclor 1242ug/kgAroclor 1244ug/kgAroclor 1254ug/kgAroclor 1260ug/kgAroclor 1260ug/kgAroclor 1260ug/kgAroclor 1274ug/kgAroclor 1284ug/kgAroclor 1260ug/kg2,2',5-Tritholorobiphenyl (BZ # 8)ug/kg2,2',5-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5-Tetrachlorobiphenyl (BZ # 45)ug/kg2,2',5,5-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',5,5-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',5,5-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',5,5-Tetrachlorobiphenyl (BZ # 101)ug/kg2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 177)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg </td <td>1.2</td> <td>10</td> | 1.2 | 10 |
| Endosulfan Iug/kgEndosulfan IIug/kgEndosulfan sulfateug/kgEndrinug/kgEndrinug/kgHeptachlorug/kgHeptachlor epoxideug/kgVirexug/kgMirexug/kgMethoxychlorug/kgToxapheneug/kgVCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1222ug/kgAroclor 1232ug/kgAroclor 1243ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 87)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',4,5,5'-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,4,4'-5-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'-5-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Hexachlorobipheny | 0.77 | 10 |
| Endosulfan II ug/kg Endosulfan sulfate ug/kg Endrin aldehyde ug/kg Heptachlor epoxide ug/kg Heptachlor epoxide ug/kg Mirex ug/kg Mirex ug/kg Methoxychlor ug/kg Oxaphene ug/kg PCB Aroclors - GC/ECD - (SW846 3540C/8082) Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1222 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1248 ug/kg Aroclor 1254 ug/kg Aroclor 1254 ug/kg Aroclor 1254 ug/kg Aroclor 1250 ug/kg 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 28) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 45) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 45) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 45) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Fentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4'-Fentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 10) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 105) ug/kg | 0.61 | 10 |
| Endosulfan sulfateug/kgEndrinug/kgEndrin aldehydeug/kgHeptachlorug/kgHeptachlorug/kgHeptachlorug/kgMirexug/kgMirexug/kgMethoxychlorug/kgFoxapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1260ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 44)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,4,4'-Fertachlorobiphenyl (BZ # 105)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,4,4'-S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-Fertachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4',5 | 0.86 | 10 |
| Endrinug/kgEndrin aldehydeug/kgHeptachlorug/kgHeptachlor epoxideug/kgMirexug/kgMirexug/kgMirexug/kgMirexug/kgMirexug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1222ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 18)ug/kg2,2',5-Trichlorobiphenyl (BZ # 28)ug/kg2,2',5-Trichlorobiphenyl (BZ # 44)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 18)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 10)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 10)ug/kg2,2',3,4'-Tetrachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 118)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 118)ug/kg2,2',4,4',5'-Hexachlorobiphenyl (BZ # 128)ug/kg2,2',4,4',5'-Hexachlorobiphenyl (BZ # 133)ug/kg2,3',4,4',5'-Hexac | 0.95 | 10 |
| Endrin aldehyde ug/kg Heptachlor ug/kg Heptachlor epoxide ug/kg Mirex ug/kg Methoxychlor ug/kg Foxaphene ug/kg PCB Aroclors - GC/ECD - (SW846 3540C/8082) Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1248 ug/kg Aroclor 1254 ug/kg Aroclor 1254 ug/kg Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 44) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 87) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 87) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 87) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 87) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 118) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 126) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 138) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 138) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 138) ug/kg 2,3',4,4'-S-Hexachlorobiphenyl (BZ # 169) ug/kg 2,3',4,4'-S-Hexachlorobiphenyl (BZ # 169) ug/kg 2,3',4,4'-S-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 170) ug/kg | 0.98 | 10 |
| Heptachlorug/kgHeptachlor epoxideug/kgMirexug/kgMirexug/kgFoxapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1260ug/kgAroclor 1260ug/kgAroclor 127ug/kgAroclor 1283ug/kgAroclor 1264ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,2',5-Trichlorobiphenyl (BZ # 44)ug/kg2,2',5-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5.Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5.Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 87)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,3',4,4'S-Pentachlorobiphenyl (BZ # 105)ug/kg2,3',4,4'S-Pentachlorobiphenyl (BZ # 105)ug/kg2,3',4,4'S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'S-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4'S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',4,4'S-Pentachlorobiphenyl (BZ # 126)ug/kg2,3',4,4'S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',4,4'S-Y-Hexachlorobiphenyl (BZ # 126)ug/kg2,3',4,4'S-Hexachlorobiphenyl (BZ # 126)ug/kg2,3',4,4'S | 1.2 | 10 |
| Heptachlor epoxideug/kgWirexug/kgMirexug/kgFoxapheneug/kgPCB Aroclors - GC/ECD - (SW846 3540C/8082)Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1222ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1254ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 8)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 87)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 87)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg <td>0.68</td> <td>10</td> | 0.68 | 10 |
| Mirex ug/kg Methoxychlor ug/kg Foxaphene ug/kg PCB Aroclors - GC/ECD - (SW846 3540C/8082) aroclor 1016 Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1248 ug/kg Aroclor 1254 ug/kg Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) ug/kg 2,4'-Dichlorobiphenyl (BZ # 18) ug/kg 2,2',5,5'-Trichlorobiphenyl (BZ # 18) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',4,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 18) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',4,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg | 0.73 | 10 |
| Methoxychlor ug/kg Foxaphene ug/kg PCB Aroclors - GC/ECD - (SW846 3540C/8082) Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1248 ug/kg Aroclor 1254 ug/kg Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) ug/kg 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 8) ug/kg 2,2,',5-Trichlorobiphenyl (BZ # 8) ug/kg 2,2,',4,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',3,4,4'-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,4,4',5-Hexachlorobiphenyl (BZ # 153) ug/kg <t< td=""><td>3.3 ^(c)</td><td></td></t<> | 3.3 ^(c) | |
| Foxaphene ug/kg PCB Aroclors - GC/ECD - (SW846 3540C/8082) ug/kg Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1244 ug/kg Aroclor 1254 ug/kg Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) ug/kg 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,2',3,3',4,4'-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 153) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 153) ug/kg 2,2',3,3',4,4',5-Hexachl | 3.0 | 10 |
| PCB Aroclors - GC/ECD - (SW846 3540C/8082) Aroclor 1016 ug/kg Aroclor 1221 ug/kg Aroclor 1232 ug/kg Aroclor 1242 ug/kg Aroclor 1248 ug/kg Aroclor 1254 ug/kg Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) ug/kg 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',4,5'S'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 118) ug/kg 2,3',4,4'S-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',4,4',5,5'-Pentachlorobiphenyl (BZ # 128) ug/kg 2,3',4,4',5,5'-Pentachlorobiphenyl (BZ # 128) ug/kg 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',4,4',5,5'-Hexachl | 13 | 50 |
| Aroclor 1016ug/kgAroclor 1221ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1243ug/kgAroclor 1244ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,3',4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3',4,4'-S-Pentachlorobiphenyl (BZ # 118)ug/kg2,2',3,3',4,4'-S-Hexachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170)ug/kg | | |
| Aroclor 1221ug/kgAroclor 1232ug/kgAroclor 1242ug/kgAroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082)2,4'-Dichlorobiphenyl (BZ # 8)2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,2,3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 87)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170)ug/kg | 15 | |
| Aroclor 1232ug/kgAroclor 1242ug/kgAroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kgPCB Congeners - GC/ECD - (SW846 3540C/8082) $2,4$ - Dichlorobiphenyl (BZ # 8)ug/kg $2,4$ - Dichlorobiphenyl (BZ # 18)ug/kg $2,4$ - Trichlorobiphenyl (BZ # 18)ug/kg $2,4,4$ - Trichlorobiphenyl (BZ # 44)ug/kg $2,2,3,5$ - Tetrachlorobiphenyl (BZ # 49)ug/kg $2,2,4,5$ - Tetrachlorobiphenyl (BZ # 49)ug/kg $2,2,3,5$ - Tetrachlorobiphenyl (BZ # 77)ug/kg $2,2,3,5,5$ - Pentachlorobiphenyl (BZ # 77)ug/kg $2,2,3,5,5$ - Pentachlorobiphenyl (BZ # 101)ug/kg $2,2,3,5,5$ - Pentachlorobiphenyl (BZ # 105)ug/kg $2,2,3,5,5,5$ - Pentachlorobiphenyl (BZ # 105)ug/kg $2,2,3,5,5,5,5,5,5,5,5,5,5,5,5,5,5,5,5,5,$ | 3.8 | |
| Aroclor 1242ug/kgAroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,2',5-Trichlorobiphenyl (BZ # 28)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,3',4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3',4,4',5-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg | 10 | |
| Aroclor 1248ug/kgAroclor 1254ug/kgAroclor 1260ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,2',5-Trichlorobiphenyl (BZ # 28)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',3,3',4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,2',3,3',4,4'-5-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 153)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170)ug/kg | 4.9 | |
| Aroclor 1254ug/kgAroclor 1260ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,4,4'-Trichlorobiphenyl (BZ # 28)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,2',3,4,4'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,3',4,4'-S-Pentachlorobiphenyl (BZ # 118)ug/kg2,3',4,4',5-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 153)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Heptachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5'-Heptachlorobiphenyl (BZ # 170)ug/kg | 8.7 | |
| Aroclor 1260 ug/kg PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg 2,4,4'-Trichlorobiphenyl (BZ # 28) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 77) ug/kg 2,2',4,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 2,2',3,3',4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 156) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 7.2 | |
| PCB Congeners - GC/ECD - (SW846 3540C/8082) 2,4'-Dichlorobiphenyl (BZ # 8) ug/kg 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg 2,4,4'-Trichlorobiphenyl (BZ # 28) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,5'-Pentachlorobiphenyl (BZ # 77) ug/kg 2,2',4,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',4,5'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 128) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 14 | |
| 2,4'-Dichlorobiphenyl (BZ # 8)ug/kg2,2',5-Trichlorobiphenyl (BZ # 18)ug/kg2,4,4'-Trichlorobiphenyl (BZ # 28)ug/kg2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg2,2',4,5'-Tetrachlorobiphenyl (BZ # 49)ug/kg2,2',5,5'-Tetrachlorobiphenyl (BZ # 52)ug/kg2,3',4,4'-Tetrachlorobiphenyl (BZ # 77)ug/kg2,2',3,5'-Pentachlorobiphenyl (BZ # 77)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 87)ug/kg2,2',4,5'-Pentachlorobiphenyl (BZ # 101)ug/kg2,3',4,4'-Pentachlorobiphenyl (BZ # 105)ug/kg2,3',4,4'-S-Pentachlorobiphenyl (BZ # 126)ug/kg2,2',3,3',4,4'-S-Pentahlorobiphenyl (BZ # 128)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138)ug/kg2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 153)ug/kg2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 169)ug/kg2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170)ug/kg | | |
| 2,2',5-Trichlorobiphenyl (BZ # 18) ug/kg $2,4,4'$ -Trichlorobiphenyl (BZ # 28) ug/kg $2,2',3,5'$ -Tetrachlorobiphenyl (BZ # 44) ug/kg $2,2',4,5'$ -Tetrachlorobiphenyl (BZ # 49) ug/kg $2,2',5,5'$ -Tetrachlorobiphenyl (BZ # 52) ug/kg $2,3',4,4'$ -Tetrachlorobiphenyl (BZ # 52) ug/kg $2,3',4,4'$ -Tetrachlorobiphenyl (BZ # 77) ug/kg $2,2',3,4,5'$ -Pentachlorobiphenyl (BZ # 77) ug/kg $2,2',3,4,5'$ -Pentachlorobiphenyl (BZ # 101) ug/kg $2,3',4,4'$ -Pentachlorobiphenyl (BZ # 105) ug/kg $2,3',4,4'$ -S-Pentachlorobiphenyl (BZ # 118) ug/kg $3,3',4,4',5$ -Pentachlorobiphenyl (BZ # 126) ug/kg $2,2',3,3',4,4'$ -Fetxachlorobiphenyl (BZ # 138) ug/kg $2,2',3,4,4',5'$ -Hexachlorobiphenyl (BZ # 153) ug/kg $2,3',4,4',5$ -Hexachlorobiphenyl (BZ # 156) ug/kg $2,3',4,4',5'$ -Hexachlorobiphenyl (BZ # 169) ug/kg $2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170)ug/kg$ | 0.24 | 2 |
| 2,4,4'-Trichlorobiphenyl (BZ # 28) ug/kg 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44) ug/kg 2,2',4,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg 2,2',5,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 52) ug/kg 2,3',4,4'-Tetrachlorobiphenyl (BZ # 66) ug/kg 2,2',3,4,5'-Pentachlorobiphenyl (BZ # 77) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4'-S-Pentachlorobiphenyl (BZ # 118) ug/kg 2,2',3,3',4,4'-S-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.12 | 2 |
| 2,2',3,5'-Tetrachlorobiphenyl (BZ # 44)ug/kg $2,2',4,5'$ -Tetrachlorobiphenyl (BZ # 49)ug/kg $2,2',5,5'$ -Tetrachlorobiphenyl (BZ # 52)ug/kg $2,3',4,4'$ -Tetrachlorobiphenyl (BZ # 66)ug/kg $3,3',4,4'$ -Tetrachlorobiphenyl (BZ # 77)ug/kg $2,2',3,4,5'$ -Pentachlorobiphenyl (BZ # 77)ug/kg $2,2',4,5,5'$ -Pentachlorobiphenyl (BZ # 101)ug/kg $2,2',4,5,5'$ -Pentachlorobiphenyl (BZ # 105)ug/kg $2,3',4,4'$ -Pentachlorobiphenyl (BZ # 105)ug/kg $2,3',4,4',5$ -Pentachlorobiphenyl (BZ # 126)ug/kg $2,2',3,3',4,4',5$ -Pentahlorobiphenyl (BZ # 128)ug/kg $2,2',3,4,4',5'$ -Hexachlorobiphenyl (BZ # 138)ug/kg $2,2',4,4',5,5'$ -Hexachlorobiphenyl (BZ # 153)ug/kg $2,3',4,4',5$ -Hexachlorobiphenyl (BZ # 169)ug/kg $2,2',3,3',4,4',5$ -Heptachlorobiphenyl (BZ # 169)ug/kg $2,2',3,3',4,4',5$ -Heptachlorobiphenyl (BZ # 170)ug/kg | 0.12 | 2 |
| 2,2',4,5'-Tetrachlorobiphenyl (BZ # 49) ug/kg $2,2',5,5'$ -Tetrachlorobiphenyl (BZ # 52) ug/kg $2,3',4,4'$ -Tetrachlorobiphenyl (BZ # 66) ug/kg $3,3',4,4'$ -Tetrachlorobiphenyl (BZ # 77) ug/kg $2,2',3,4,5'$ -Pentachlorobiphenyl (BZ # 87) ug/kg $2,2',4,5,5'$ -Pentachlorobiphenyl (BZ # 101) ug/kg $2,3,3',4,4'$ -Pentachlorobiphenyl (BZ # 105) ug/kg $2,3',4,4',5$ -Pentachlorobiphenyl (BZ # 105) ug/kg $2,3',4,4',5$ -Pentachlorobiphenyl (BZ # 126) ug/kg $2,2',3,3',4,4',5$ -Pentachlorobiphenyl (BZ # 128) ug/kg $2,2',3,4,4',5'$ -Hexachlorobiphenyl (BZ # 138) ug/kg $2,2',4,4',5,5'$ -Hexachlorobiphenyl (BZ # 153) ug/kg $2,3',4,4',5$ -Hexachlorobiphenyl (BZ # 169) ug/kg $2,2',3,3',4,4',5$ -Heptachlorobiphenyl (BZ # 170) ug/kg | 0.11 | 2 |
| 2,2',5,5'-Tetrachlorobiphenyl (BZ # 52) ug/kg $2,3',4,4'$ -Tetrachlorobiphenyl (BZ # 66) ug/kg $3,3',4,4'$ -Tetrachlorobiphenyl (BZ # 77) ug/kg $2,2',3,4,5'$ -Pentachlorobiphenyl (BZ # 87) ug/kg $2,2',4,5,5'$ -Pentachlorobiphenyl (BZ # 101) ug/kg $2,3,3',4,4'$ -Pentachlorobiphenyl (BZ # 105) ug/kg $2,3',4,4'$ -Pentachlorobiphenyl (BZ # 118) ug/kg $3,3',4,4',5$ -Pentachlorobiphenyl (BZ # 126) ug/kg $2,2',3,3',4,4',5$ -Pentachlorobiphenyl (BZ # 128) ug/kg $2,2',3,4,4',5'$ -Hexachlorobiphenyl (BZ # 138) ug/kg $2,2',4,4',5,5'$ -Hexachlorobiphenyl (BZ # 153) ug/kg $2,3',4,4',5$ -Hexachlorobiphenyl (BZ # 156) ug/kg $2,3',4,4',5,5'$ -Hexachlorobiphenyl (BZ # 169) ug/kg $2,2',3,3',4,4',5$ -Heptachlorobiphenyl (BZ # 170) ug/kg | 0.26 | 2 |
| 2,3',4,4'-Tetrachlorobiphenyl (BZ # 66) ug/kg 3,3',4,4'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,4,5'-Pentachlorobiphenyl (BZ # 87) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4'-Pentachlorobiphenyl (BZ # 118) ug/kg 2,2',3,3',4,4'-S-Pentahlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Fentachlorobiphenyl (BZ # 128) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.14 | 2 |
| 3,3',4,4'-Tetrachlorobiphenyl (BZ # 77) ug/kg 2,2',3,4,5'-Pentachlorobiphenyl (BZ # 87) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 2,2',3,3',4,4'-Fentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4',5-Pentahlorobiphenyl (BZ # 128) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.33 | 2 |
| 2,2',3,4,5'-Pentachlorobiphenyl (BZ # 87) ug/kg 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 3,3',4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.21 | 2 |
| 2,2',4,5,5'-Pentachlorobiphenyl (BZ # 101) ug/kg 2,3,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 3,3',4,4',5-Pentachlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,3',4,4'-Fentachlorobiphenyl (BZ # 128) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.13 | 2 |
| 2,3,3',4,4'-Pentachlorobiphenyl (BZ # 105) ug/kg 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 3,3',4,4',5-Pentahlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,3',4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.18 | 2 |
| 2,3',4,4',5-Pentachlorobiphenyl (BZ # 118) ug/kg 3,3',4,4',5-Pentahlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',4,4',5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.16 | 2 |
| 3,3',4,4',5-Pentahlorobiphenyl (BZ # 126) ug/kg 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 2,3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.21 | 2 |
| 2,2',3,3',4,4'-Hexachlorobiphenyl (BZ # 128) ug/kg 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.20 | 2 |
| 2,2',3,4,4',5'-Hexachlorobiphenyl (BZ # 138) ug/kg 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 153) ug/kg 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Hexachlorobiphenyl (BZ # 170) ug/kg | 0.14 | 2 |
| 2,2',4,4',5,5'-Hexachlorobiphenyl (BZ # 153) 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.20 | 2 |
| 2,3,3',4,4',5-Hexachlorobiphenyl (BZ # 156) ug/kg 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.12 | 2 |
| 3,3',4,4',5,5'-Hexachlorobiphenyl (BZ # 169) ug/kg 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.088 | 2 |
| 2,2',3,3',4,4',5-Heptachlorobiphenyl (BZ # 170) ug/kg | 0.11 | 2 |
| | 0.075 | 2 |
| 2 2' 3 4 4' 5 5'-Hentachlorohinhenvl (BZ # 180) ug/kg | 0.085 | - 2 |
| 2 2' 3 4 4' 5' 6 Hentachlarahinhenvi (B7 # 183) ug/kg | 0.076 | 2 |
| $(D_2 + 103)$ ug/s 2 2 4 4 6 6' Hentachlarahinhenvi (D7 + 184) ug/ba | 0.070 | 2 |

(a) Method Detection Limit (MDL) STL-Baltimore for standard tissue matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).

TABLE 5-5 (CONTINUED)

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) | _ |
|--|---------|----------------------------------|-----------------------------------|---|
| 2 2' 3 4' 5 5' 6-Hentachlorobinhenvl (BZ # 187) | ug/kg | 0.088 | 2 | |
| 2 2' 3 3' 4 4' 5 6-Octachlorobinhenyl (BZ # 195) | ug/kg | 0.13 | 2 | |
| 2,2',3,3',4,4',5,5' 6-Nonachlorobinhenvl (BZ # 206) | ug/kg | 0.11 | 2 | |
| 2,2',3,3',4,4',5,5',6,6'-Decachlorobiphenyl (BZ # 209) | ug/kg | 0.15 | 2 | |
| Organophosphorus Pesticides GC/NPD - (SW 3540C/ | '8141A) | | | |
| Dementon | ug/kg | 33 ^(c) | | |
| Ethyl parathion (Parathion) | ug/kg | 33 ^(c) | 6 | |
| Guthion (Azinphos methyl) | ug/kg | 33 ^(c) | | |
| Malathion | ug/kg | 33 ^(c) | 5 | |
| Methyl parathion | ug/kg | 33 ^(c) | 6 | |
| Semivolatile organics GC/MS - (SW846 3540C/8270C) |) | | | |
| Benzoic Acid | ug/kg | 250 | 100 | |
| Benzyl alcohol | ug/kg | 100 | 100 | |
| Bis(2-chloroethyl) ether | ug/kg | · 110 | - | |
| Bis(2-chloroethoxy)methane | ug/kg | 130 | - | |
| Bis(2-ethylhexyl) phthalate | ug/kg | 180 | 20 | |
| 4-Bromophenyl phenyl ether | ug/kg | 120 | - | |
| Butylbenzylphthalate | ug/kg | 150 | 20 | |
| 4-Chloro-3-methylphenol | ug/kg | 77 | - | |
| 2-Chloronaphthalene | ug/kg | 100 | - | |
| 2-Chlorophenol | ug/kg | 91 | - | |
| 4-Chlorophenyl phenyl ether | ug/kg | 94 | - | |
| Dibenzofuran | ug/kg | 100 | 20 | |
| Di-n-butyl nhthalate | ug/kg | 110 | 20 | |
| 1.2-Dichlorobenzene | ug/kg | 140 | 20 | |
| 1.3 Dichlorobenzene | ug/kg | 140 | 20 | |
| 1.4-Dichlorobenzene | ug/kg | 140 | 20 | |
| 2.2! Dichlerchenzidine | ug/kg | 280 | 20 | |
| 2.4 Dishlaranhanal | ug/kg | 100 | - | |
| Disthul antholoto | ug/kg | 110 | 20 | |
| 4.6 Dinitro 2 Mathulahanal | ug/kg | 100 | - | |
| 2.4. Dimethylphenol | ug/kg | 84 | . 20 | |
| 2,4-Dimethylphenol | ug/kg | 64 | 20 | |
| 2.4 Disites has l | ug/kg | 100 | 20 | |
| 2,4-Dinitrophenol | ug/kg | 190 | - | |
| 2,4-Dinitrotoluene | ug/kg | 57 | - | |
| 2,6-Dinitrotoluene | ug/kg | 07 | - | |
| 1,2-Diphenylhydrazine | ug/kg | 09 | - | |
| Di-n-octyl phthalate | ug/kg | 370 | 20 | |
| Hexachlorobenzene | ug/kg | 89 | 20 | |
| Hexachlorobutadiene | ug/kg | 120 | 40 | |
| Hexachloroethane | ug/kg | 110 | 40 | |
| Hexachlorocyclopentadiene | ug/kg | 59 | | |
| lsophorone | ug/kg | 120 | | |
| 2-Methylphenol | ug/kg | 87 | 20 | |
| 4-Methylphenol | ug/kg | 6/ | 20 | |
| Nitrobenzene | ug/kg | 110 | - | |
| 2-Nitrophenol | ug/kg | 120 | - | |
| 4-Nitrophenol | ug/kg | 140 | - | |
| N-Nitrosodiphenylamine | ug/kg | 110 | 20 | |
| N-Nitrosodimethylamine | ug/kg | 120 | - | |
| N-Nitroso-di-n-propylamine | ug/kg | 120 | - | |
| 2,2'-Oxybis(1-chloropropane) | ug/kg | 150 | - | |
| Pentachlorophenol | ug/kg | 210 | 100 | |
| Phenol | ug/kg | 98 | 20 | |
| 1,2,4-Trichlorobenzene | ug/kg | 120 | 20 | |

(a) Method Detection Limit (MDL) STL-Baltimore for standard tissue matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).

TABLE 5-5 (CONTINUED)

| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) |
|---|--------------------|----------------------------------|-----------------------------------|
| 2,4,6-Trichlorophenol | ug/kg | 83 | - |
| Polynuclear Aromatic Hydrocarbons (PAHs) - HP | LC (SW846 3540C/83 | (10) | |
| Acenaphthene | ug/kg | 2.8 | 20 |
| Acenaphthylene | ug/kg | 21 | 20 |
| Anthracene | ug/kg | 0.54 | 20 |
| Benzo[a]anthracene | ug/kg | 0.76 | 20 |
| Benzo[b]fluoranthene | ug/kg | 0.78 | 20 |
| Benzo[k]fluoranthene | ug/kg | 0.44 | 20 |
| Benzo[a]pyrene | ug/kg | 0.41 | 20 |
| Benzo[ghi]nervlene | ug/kg | 0.41 | 20 |
| Chrysene | ug/kg | 0.22 | 20 |
| Dibenzo[a h]anthracene | ug/kg | 0.24 | 20 |
| Fluoranthene | ug/kg | 0.78 | 20 |
| Fluorene | ug/kg | 1.0 | 20 |
| Indeno[1 2 3-cd]pyrene | ug/kg | 0.80 | 20 |
| 1-Methylnanhthalene | ug/kg | 48 | 20 |
| 2-Methylnaphthalene | ug/kg | 4.0 | 20 |
| Nonhthalene | ug/kg | 3.6 | 20 |
| Phenanthrene | ug/kg | 5.0 | 20 |
| Pyrane | ug/kg | 0.44 | 20 |
| ryrene | ug/kg | 0.44 | 20 |
| Organotins - GC/FPD (STL Burlington SOP) | | | |
| Monobutyltins | ug/kg | 1.0 ^(c) | 10 |
| Dibutyltins | ug/kg | 1.3 ^(c) | 10 |
| Tributyltins | ug/kg | 1.5 ^(c) | 10 |
| Dioxins/Furans – HRGC/HRMS (EPA 1613) | | | |
| 2.3.7.8-TCDF | ng/kg | 0.73 | 1 |
| 2 3 7 8-TCDD | ng/kg | 0.19 | 1 |
| 1 2 3 7 8-PeCDF | ng/kg | 0.12 | 25 |
| 2 3 4 7 8-PeCDF | ng/kg | 0.52 | 2.5 |
| 1 2 3 7 8-PeCDD | ng/kg | 1.20 | 2.5 |
| 1 2 3 4 7 8-HxCDF | ng/kg | 0.50 | 5 |
| 1 2 3 6 7 8-HxCDF | ng/kg | 0.35 | 5 |
| 2 3 4 6 7 8-HxCDF | ng/kg | 0.39 | 5 |
| 1 2 3 4 7 8-HxCDD | ng/kg | 0.32 | 5 |
| 1 2 3 6 7 8-HxCDD | ng/kg | 0.52 | 5 |
| 1 2 3 7 8 9-HyCDD | ng/kg | 0.74 | 5 |
| 1 2 3 7 8 9-HxCDF | ng/kg | 0.50 | 5 |
| 1 2 3 4 6 7 8-HpCDF | ng/kg | 1.63 | 5 |
| 1 2 3 4 6 7 8-HnCDD | ng/kg | 2.06 | 5 |
| 1 2 3 4 7 8 9-HpCDF | ng/kg | 0.66 | 5 |
| OCDD | ng/kg | 18.17 | 10 |
| OCDF | ng/kg | 2.20 | 10 |
| | | | |
| Metals - Cold Vapor (SW846 7471A) | malia | 0.14 | 0.01 |
| Mercury | mg/kg | 0.14 | 0.01 |
| Metals - Furnace (SW846 3050B/7000 series) | | | |
| Thallium | mg/kg | 0.30 | 0.1 |
| Metals - ICP (SW846 3050B/6010B) | | | |
| Aluminum | malka | 11.0 | 1.0 |
| Iron | ma/ka | 16.0 | 10 |
| Manganese | malka | 20 | 0.5 |
| Zinc | mg/kg | 9.1 | 2.0 |
| | | | |

(a) Method Detection Limit (MDL) STL-Baltimore for standard tissue matrix determined according to the procedure in 40 CFR 136 Appendix B.

(b) Target Detection Limit (USEPA/USACE 1995).



| Parameter | Units | Laboratory MDL ^(a) | Recommended TDL ^(b) | |
|--------------------------------------|-------|----------------------------------|-----------------------------------|--|
| | | | | |
| Metals-TRACE ICP (SW846 3050B/6010B) | | | | |
| Antimony | mg/kg | 0.90 | 0.1 | |
| Arsenic | mg/kg | 0.42 | 0.1 | |
| Beryllium | mg/kg | 0.30 | 0.1 | |
| Cadmium | mg/kg | 0. 045 | 0.1 | |
| Chromium | mg/kg | 0.27 | 0.1 | |
| Copper | mg/kg | 0.42 | 0.1 | |
| Lead | mg/kg | 0.23 | 0.1 | |
| Nickel | mg/kg | 0.22 | 0.1 | |
| Silver | mg/kg | 0.27 | 0.1 | |
| Selenium | mg/kg | 0.16 | 0.2 | |

TABLE 5-5 (CONTINUED)



⁽a) Method Detection Limit (MDL) STL-Baltimore for standard tissue matrix determined according to the procedure in 40 CFR 136 Appendix B.

⁽b) Target Detection Limit (USEPA/USACE 1995).

⁽c) For these compounds, no laboratory MDL has been determined by STL-Baltimore. A Reporting Limit is used based upon the low calibration standard concentration (Organotins are lab reporting limits).

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CHAPTER 6 -BULK SEDIMENT

6. BULK SEDIMENT CHEMISTRY

This chapter presents a Tier II sediment chemistry evaluation for the approach channels, Inside Placement Site 104, Outside Site 104, and the Ocean Reference area. The chapter presents data for sediments that were specifically collected in 1999 for Tier II and Tier III evaluations. The 1999 field program is described in Chapter 4. The following topics are presented and discussed: (1) constituents tested and detected within the sediments; (2) comparisons of detected constituents to sediment quality guidelines; (3) comparisons of detected constituents to background and reference concentrations; (4) comparisons of detected constituents to available Tier I information presented in Chapter 3; and (5) results of Theoretical Bioaccumulation Potential (TBP) calculations.

6.1.1 SAMPLE RECEIPT AND HOMOGENIZATION PROCEDURES

The sediment cores and grab samples were transported from the field staging areas (either Fort McHenry, Baltimore or Sandy Point State Park, Annapolis) to EA's laboratory facility in Sparks, Maryland at the end of each workday. The cores and grab samples were chilled with ice during transport. The sediment cores and the grab samples designated for ecotoxicological testing were stored in a walk-in refrigeration unit cooled to 4°C until all sampling was complete. Grab samples designated for chemical analysis were hand-delivered to the analytical laboratory on the day of collection. Upon completion of all sampling activities, the cores were sorted, visually inspected, and the labeled core sleeves were compared against the chain-of-custody record prior to processing. Grab samples that were received at the analytical laboratory were compared against the chain-of-custody upon receipt.

Core processing was conducted at EA's warehouse facility on 30 September to 01 October 1999 and 17 December 1999. Sediment samples for each station were processed for analytical testing and channel composites were created for ecotoxicological testing.

Sediments from each station were extracted from the core sleeves and were composited and homogenized in pre-cleaned, 5-gallon stainless steel bowls. Multiple cores for each station were extracted and homogenized with decontaminated stainless-steel mixing spoons until the sediment was thoroughly mixed and was of uniform consistency. When compositing and homogenization of sediment from each station was complete, subsamples were removed for bulk chemistry analysis. The sub-samples were placed into pre-cleaned glass jars using stainless steel spoons, and were labeled for analytical testing. The remainder of the sediment was placed in a 55-gallon decontaminated fiberglass holding container to create the channel composite for ecotoxicological testing. When not actively being processed, the core and grab samples were stored in a secured walk-in cooler, in the dark at 4°C. A second chain-of-custody form was completed for the homogenized core sediment that was designated for chemical analysis, and the sample homogenization time was recorded as the initiation of the sample holding time.

The bulk sediment samples were hand-delivered to STL-Baltimore, where the samples were compared against the chain-of-custody form. The samples were logged into the analytical

laboratory and were assigned a unique accession number. Samples were stored in walk-in refrigeration units (cooled to 4°C) following receipt and prior to analysis. Copies of the bulk sediment chain-of-custody forms are provided in Attachment II. The compositing scheme for the ecotoxicological testing is discussed in Chapters 4 and 8.

6.2 ANALYTICAL METHODS AND DETECTION LIMITS

Bulk sediments were analyzed for target analytes identified in the approved Analytical Chemistry QAPP (EA and STL 2000) and outlined in Chapter 5. Project-specific analytical methods and detection limits for sediment samples are provided in Tables 5-2 and 5-3, respectively. A sample-by-sample breakdown of bulk sediment analyses is provided in Table 5-1A.

6.3 DATA ANALYSIS

For sediments, data were evaluated based on mean concentrations of constituents detected within each channel. The mean concentrations best represent the concentrations that would be expected when the material is dredged, mixed together, and placed in large volumes.

6.3.1 Mean Calculations

Mean concentrations of detected analytes were calculated for each sampling reach (channels, Inside Site 104, Outside Site 104, and Ocean Reference). The mean concentration for each reach is most representative of the concentration expected in the field during dredging and placement and is most representative of the expected concentration in the reach composite samples that were utilized in the elutriate, toxicity, and bioaccumulation testing. The detection limit (DL) was substituted for non-detected (ND) analytes in the calculations of the mean for each reach. Using the detection limit gives the highest possible mean value. If an analyte was not detected in any sample within a sampling reach, the mean detection limit is reported and qualified with a "U."

The mean analyte concentrations for Inside Placement Site 104 were calculated with data from eight analyzed samples [the five targeted stations (Figure 4-11), one station located south of KI-7 (KI-7REF) (Figure 4-12) where visual signs of contamination were present, and two field duplicate samples (KI-7FD and KI-7REFFD)]. The mean calculations for Craighill Upper Range and Brewerton Eastern Extension each included results from one field duplicate sample in addition to the results from the targeted locations. Analyte concentrations for the Ocean Reference sediment represent the actual detected concentrations in one composited sample. Dioxin and furan results for each channel, reference, or placement site reach represent the actual detected concentrations in one composited sample.

For individual samples, PCB concentrations were determined by summing the 18 summation congeners (as specified in Table 9-3 of the ITM). The total summed concentration was then multiplied by a factor of 2 following the NOAA (1993) approach for total PCB determinations. Total PAHs were determined by summing the concentrations of 16 PAHs in each sample. In the

summation calculations for both total PCBs and total PAHs, three total values are presented in the data tables:

- Non-detects = zero (ND=0);
- Non-detects = $\frac{1}{2}$ of the detection limit (ND= $\frac{1}{2}$ DL); and
- Non-detects = the detection limit (ND=DL).

The substitution of the detection limit (ND=DL) provides the most conservative approach to calculating and evaluating the data. However, in cases where few PCB congeners or PAHs are detected, the detection limit drives the total value and overestimates the actual expected concentration.

Mean total PCB and total PAH concentrations were determined by averaging the total that was calculated for each sample. In the PCB and PAH tables (Table 6-9 and 6-10, respectively), note that the average of the sums does not equal the sum of the average concentrations for each congener or analyte for calculations with ND=0 and ND=1/2.

6.3.2 Comparisons to Sediment Quality Guidelines (SQGs)

Mean concentrations of detected analytes in sediment samples were compared against Sediment Quality Guidelines (SQG) for marine sediments. Threshold Effects Levels (TEL) and Probable Effects Levels (PEL) (Buchman 1999; MacDonald 1994; MacDonald et al. 1996) are among the most commonly used of the methods that attempt to provide sediment contaminant concentration values that differentiate sediments of little concern from those predicted to have adverse biological effects.

TELs represent the contaminant concentration below which adverse biological effects rarely occur. PELs represent the contaminant concentration above which adverse biological effects frequently occur. Values that fall between the TEL and PEL represent the concentrations at which adverse biological effects occasionally occur. TEL and PEL screening values are provided in Table 3-1.

O'Connor et al. (1998) and O'Connor and Paul (1999) quantitatively evaluated the reliability of sediment toxicity predictions based on Effect Range Low (ERL) / Effect Range Median (ERM) values, which are derived by a process very similar to the TEL/PEL process and have similar values. Both papers attempt to validate the values using large independent datasets that contain both sediment chemistry and sediment toxicity data for each sample. Both papers define a toxic sediment as one that produces less than 80 percent survival of amphipods after a 10-day exposure to whole sediment, the same criterion of whole sediment toxicity used in the ITM. Both papers reach the same conclusions. O'Connor and Paul (1999) found that within a 2,475-sample dataset, 2,087 (84%) of the samples were not toxic. A total of 730 samples did not exceed any ERL (comparable to TEL), and 697 (95.5%) of these were not toxic. This indicates that not exceeding an ERL is a reliable predictor of non-toxicity, and the same should be true for

the closely related TEL. Of the 453 samples that exceeded at least on ERM, 186 (41%) actually produced toxicity. Therefore, exceeding an ERM (much less an ERL) is less than 50% accurate as a predictor of sediment toxicity, and the same would be expected to be true for the closely related PEL (and TEL). This independent evaluation indicates that:

- not exceeding a TEL should reliably predict the absence of whole sediment toxicity,
- exceeding a PEL (much less a TEL) does not reliably indicate toxicity, and
- many, perhaps even most, sediments that exceed one or more PELs are not toxic.

Because TELs/PELs have been widely used despite their recently demonstrated low reliability in predicting toxicity, the mean concentrations of contaminants in the sediments proposed for placement at Site 104 were compared to the TEL and PEL values for all analytes for which TEL/PEL values have been developed. Comparison of sediment chemistry to SQGs is not a part of the tiered testing evaluations in the ITM (USEPA/USACE 1998) or the Green Book (USEPA/USACE 1991). For dredged material evaluations, SQGs are used as a tool to assist with identification of COPCs and to provide additional weight of evidence in the evaluation (USACE-WES 1998). Comparisons to TEL/PEL values were used only for these purposes in this evaluation of the sediments proposed for placement at Site 104.

6.4 BULK SEDIMENT RESULTS

Results of the bulk sediment chemistry analyses are presented in the following sub-sections. Sample weights were adjusted for percent moisture (up to 50%) prior to analysis to achieve the lowest possible detection limits. Because sediments contain a large proportion of moisture, each analyte has a sample-specific detection limit. The detection limit range is provided in the analytical summary tables (Tables 6-3 through 6-13). Analytical results are reported on a dry weight basis. Definitions of organic and inorganic data qualifiers are provided in Tables 3-4 and 3-5, respectively. Qualifiers for dioxin and furan analysis are described in Table 6-1. Analytical narratives that include an evaluation of laboratory quality assurance/quality control results are provided in Attachment II. STL-Baltimore will retain and archive the results of these analyses for 7 years from the date of issuance of the final results.

Mean analyte concentrations are provided by analytical fraction in Tables 6-2 through 6-13. Results of TEL and PEL screening comparisons for mean concentrations are provided in Tables 6-14 and 6-15, respectively. Frequency of detection by analytical fraction for each channel is provided in Table 6-16. Frequency of detection by analytical fraction for Inside Site 104, Outside Site 104, and the Ocean Reference is provided in Table 6-17. Data for individual stations and summary statistics that include minimum and maximum concentrations for each reach are provided in Appendices D and E, respectively.

6.4.1 Physical Analyses

Results of physical analyses are provided in Table 6-2. Grain size determinations indicated that the channel sediments were primarily comprised of silt and clay (Figure 6-1). The Craighill



Channel exhibited the highest proportion of sand (67%), followed by Craighill Upper Range (32%). Inside Site 104, Outside Site 104, and the Ocean Reference area contained higher proportions of sand than the channels, with the exception of the Craighill Channel and Craighill Upper Range (Figure 6-1). Percent moisture in the channel sediments ranged from 40.8% to 72%. The ocean sediment had the lowest moisture content (21.5%).

6.4.2 Inorganic Non-metals

Results of the inorganic non-metal analyses are provided in Table 6-3. Mean concentrations of TOC in the channels ranged from 2.7 to 13.4 percent (Figure 6-2). These concentrations are similar to those reported for the channels in the 1998 study (3.4 to 14.1 percent) (see Chapter 3). Surficial sediments from the channels that have been most recently dredged (Craighill and Craighill Upper Range) contained the lowest mean concentrations of TOC. The highest mean TOC in the sediment was from the Tolchester Straightening, where the deepest cores were collected (approximately 10 ft in depth). Mean TOC concentrations in the Inside and Outside Site 104 reference areas were 7.2% and 11.4%, respectively, and fell within the range of mean TOC reported for the channels. The Ocean Reference sediment contained the lowest percentage of TOC (0.5%).

Mean sulfide concentrations in the channels ranged from non-detect (C&D Approach Channel cores) to 1,536 mg/kg (Swan Point). Mean sulfide concentrations were generally higher in sediments from the eastern side of the upper Bay (Inside and Outside Site 104, Swan Point, and Tolchester Straightening). These areas are more prone to low dissolved oxygen conditions and sulfide formation. The high sulfide concentrations reported for Inside and Outside Site 104 are consistent with results reported for surficial sediments in the 1997-1998 Site 104 and channel studies (see Chapter 3). Sulfides were not detected in the Ocean Reference grab sample or C&D Canal Approach core sample sediments.

Mean total phosphorus (TP) ranged from 198 mg/kg (Craighill Upper Range) to 632 mg/kg (Craighill Angle West). TP detected in the reference and placement areas fell within the range of concentrations reported for the channels, with the Ocean Reference sediment at the lower end.

Mean concentrations of ammonia-nitrogen (total ammonia) ranged from 12 mg/kg (Craighill Upper Range) to 223 mg/kg (Craighill Entrance). Mean ammonia-nitrogen concentrations in Inside and Outside Site 104 fell within the range for the channels. Mean ammonia-nitrogen concentrations Outside Site 104 (182.4 mg/kg) were nearly 5 times higher than Inside Site 104 (36.7 mg/kg). The 1997 Site 104 studies revealed a similar pattern, with ammonia-nitrogen concentrations Outside Site 104 (137.7 mg/kg) approximately 2.6 times higher than Inside Site 104 (51.2 mg/kg) (see Chapter 3). Ammonia-nitrogen was detected just above the detection limit in the Ocean Reference sediment (2 mg/kg).

Mean TKN (ammonia + inorganic nitrogen) concentrations ranged from 225.2 mg/kg (Craighill) to 3,714 mg/kg (Brewerton Eastern Extension). Mean TKN concentrations for both Inside and Outside Site 104 fell within the range reported for the channel sediments, and mean TKN

concentration Outside Site 104 was more than 2 times higher than Inside Site 104. TKN concentration was lowest in the Ocean Reference sediment (180 mg/kg).

6.4.3 Volatile Organic Compounds (VOCs)

Results of the volatile organic compound analyses are presented in Table 6-4. VOCs were detected in 23 of 1,540 cases (1.5%) in the channel sediments. VOCs were detected in 8 of 175 cases (4.6%) Inside Site 104 and in 5 of 140 cases (3.6%) Outside Site 104 (see Table 6-17). Only 2 of the 35 tested VOCs, carbon disulfide and dichloromethane (methylene chloride), were detected, and they were detected within both the channel and Inside/Outside Site 104 samples. The highest mean concentration of carbon disulfide was detected Inside Site 104 (7.98 μ g/kg) and the highest mean concentration of dichloromethane was detected for Brewerton Eastern Extension (15.75 ug/kg). Although neither compound was detected in the laboratory method blanks, both analytes are common laboratory contaminants. Overall, VOCs were not detected in 4 of the 13 channel reaches (C&D Canal Approach, Craighill Angle West, Craighill Entrance, and Tolchester North). No VOCs were detected in the Ocean Reference sediment. No TEL or PEL values exist for VOCs. A few scattered occurrences of carbon disulfide and dichloromethane were also detected in channel sediments tested in 1998 (EA 2000c) (see Chapter 3).

6.4.4 Semivolatile Organic Compounds (SVOCs)

Results of the semivolatile organic compound analyses are presented in Table 6-5. Seven of the 47 tested SVOCs were detected in the channel sediments. Overall, SVOCs were detected in 164 of 2,350 cases (7%) in the channel sediments. SVOCs were detected in 23 of 376 cases (6.1%) Inside Site 104 and in 18 of 188 cases (9.6%) Outside Site 104. At least one SVOC was detected within every tested channel reach. No SVOCs were detected in the Ocean Reference sediment. Six of seven SVOCs detected in the channel sediments were also detected Inside and Outside Site 104 [1-methylnaphthalene, 2-methylnaphalene, 3,4-methylphenol, bis(2-ethylhexyl)phthalate, di-n-butyl phthalate, and phenol]. Four SVOCs were detected Outside Site 104, but not within the channel reaches [2,2'-oxybis(1-chloropropane), 4-nitrophenol, di-noctyl phthalate, and diethyl phthalate]. Dibenzofuran was detected in the Tolchester Straightening area and Outside Site 104. Benzyl butyl phthalate was the only SVOC detected in the channels (Brewerton Eastern Extension only) that was not detected Inside or Outside Site 104.

2-Methylnaphthalene and bis(2-ethylhexyl)phthalate are the only SVOCs that have TEL/PEL values (Tables 6-14 and 6-15). Mean concentrations of 2-methylnaphthalene exceeded the TEL in 11 of the 13 channel reaches and in both Inside and Outside Site 104. Mean concentrations of 2-methylnaphthalene exceeded the PEL only for the Tolchester Straightening area and Inside Site 104 sediments. Although bis-(2-ethylhexyl) phthalate was detected in 5 of the 13 channel reaches and in both Inside Site 104 sediments, none of the concentrations exceeded the TEL or PEL value.

Six of the seven SVOCs detected in the channel sediments in the 1999 testing program were also detected in the channel sediments in the 1998 testing program [1-methylnaphthalene, 2-methylnaphalene, 3,4-methylphenol, benzyl butyl phthalate, bis(2-ethylhexyl)phthalate, and phenol] (see Chapter 3).

6.4.5 Chlorinated Pesticides

Results of the chlorinated pesticides analyses are presented in Table 6-6. Approximately 98% of the channel pesticide analyses yielded no detectable concentrations of chlorinated pesticides. Chlorinated pesticides were detected in 16 of 1,100 cases (1.5%) in the channel sediments. Chlorinated pesticides were detected in 5 of the 13 channels reaches [C&D Canal Approach (cores), Craighill Angle West, Tolchester North, Tolchester South, and Tolchester Straightening]. Only 3 of the 22 tested chlorinated pesticides were detected in channel sediments (heptachlor epoxide, 4,4'-DDD and 4,4'-DDE). These three pesticides were also detected in the channel sediments tested in the 1998 program (EA 2000d) (see Chapter 3).

Chlorinated pesticides were detected in 8 of 176 cases (4.5%) Inside Site 104 and in 3 of 88 cases (3.4%) Outside Site 104. Five of the 22 tested chlorinated pesticides were detected Inside Site 104 (4,4'-DDE, aldrin, endosulfan II, endrin aldehyde, and heptachlor epoxide). Heptachlor epoxide was the only chlorinated pesticide detected Outside Site 104. None of the 22 targeted chlorinated pesticides was detected in the Ocean Reference sediment.

TEL and PEL values exist for 6 of the 22 tested chlorinated pesticides (4,4'-DDD, 4,4'-DDE, 4,4'-DDT, chlordane, dieldrin, and gamma-BHC (Tables 6-14 and 6-15, respectively). None of the mean detected concentrations of DDD or DDE exceeded PEL or TEL values. The detection limits for several chlorinated pesticides exceeded the TEL value: gamma-BHC, dieldrin, chlordane, and 4,4-DDT. However, none of the mean detection limits exceeded the PEL values for these pesticides. In cases where the detection limit exceeded the TEL value, it is not possible to determine whether these pesticides were present at concentrations between the TEL and PEL.

6.4.6 Organophosphorus Pesticides

Results of the organophosphorus pesticide analyses are provided in Table 6-7. Organophosphorus pesticides were not detected in any of the channel sediments (0 of 250 cases), Inside Site 104 (0 of 40 cases), Outside Site 104 (0 of 20 cases), or in the Ocean Reference (0 of 5 cases). There are no TEL/PEL values for organophosphorus pesticides.

6.4.7 PCB Aroclors and Congeners

Results of the PCB aroclor analyses are provided in Table 6-8. PCB aroclors (mixtures of congeners) were not detected in any of the channel sediments (0 of 350 cases), Outside Site 104 (0 of 28 cases), or in the Ocean Reference (0 of 7 cases). Two of the seven tested PCB aroclors (Aroclor 1248 and Aroclor 1254) were detected Inside Site 104 (2 of 56 cases or 3.6%). There are no TEL/PEL values for PCB aroclors. Aroclors 1254 and 1260 were also detected Inside

Site 104, Outside Site 104, and in several channel reaches in the 1997 and 1998 testing programs (EA 1998a and 2000c) (see Chapter 3).

Results of the PCB congener analyses are provided in Table 6-9. Twenty-five of the 26 tested congeners were detected in the channel sediments. PCB congeners were detected in 654 of 1,350 cases (48.4%) in the channel sediments. PCB congeners were detected in 110 of 208 cases (53%) Inside Site 104 and in 33 of 104 cases (32%) Outside Site 104. One PCB congener (BZ#8) was detected in the Ocean Reference sediment (1 of 26 cases or 3.8%). Importantly, the majority of detected congeners were detected at concentrations that were below the recommended USEPA/USACE (1995) Target Detection Limit (TDL).

Mean total PCB concentrations for the channels and reference areas are depicted in Figure 6-3. Mean total PCB concentrations in the channels (ND=DL) ranged from 3.45 μ g/kg (Craighill Channel) to 22.6 μ g/kg [C&D Canal Approach (cores)]. The mean total PCB concentration Inside Site 104 was approximately 3.3 times higher than Outside Site 104. The total PCB values for both Inside and Outside Site 104 fell within the range for the channels. Only the mean total PCB concentration for the C&D Approach (cores) exceeded the TEL (21.55 μ g/kg) and that by 1 μ g/kg (Table 6-14).

The detection limits for PCB congeners in the 1997 and 1998 sampling programs (EA 1998a and 2000c) were 2 to 3 times higher than the detection limits in the 1999 sampling program. Therefore, total PCBs (ND=DL) for the 1997-1998 testing exceeded the TEL, even when no congeners were detected (see Chapter 3). Importantly, the TDL for each individual PCB congener is 1 μ g/kg (USEPA/USACE 1995). If total PCBs are calculated using the TDL, the total exceeds the TEL value.

6.4.8 Polynuclear Aromatic Hydrocarbons

Results of the PAH analyses are provided in Table 6-10. Each of the tested PAHs was detected in at least one of the channel reaches. PAHs were detected in 696 of 800 cases (87%) in the channel sediments. PAHs were detected in 116 of 128 cases (91%) Inside Site 104 and in 55 of 64 cases (86%) Outside Site 104. None of the sixteen tested PAHs was detected in the Ocean Reference sediment. Acenaphthylene and dibenz[a,h]anthracene were the only two PAHs that were not detected in all of the channel reaches. All of the tested PAHs were detected Inside Site 104, and dibenz[a,h]anthracene was the only PAH that was not detected Outside Site 104.

Mean total PAH concentrations (ND=DL) are depicted in Figure 6-4. The mean total PAH concentration in Tolchester Straightening was nearly 3 times higher than the next highest mean channel concentration (Tolchester North). Overall, the mean concentrations of total PAHs Inside and Outside Site 104 were nearly equivalent and were higher than each of the channels with the exception the Tolchester Straightening sediments. Mean total PAH concentrations reported for Inside Site 104, Outside Site 104, and Tolchester Straightening exceeded the TEL value of $1,684.06 \mu g/kg$ (Table 6-14), but concentrations were well below the PEL value of

16,770.4 μ g/kg (Table 6-15). None of the mean total PAHs concentrations reported in the 1997 and 1998 testing programs exceeded the TEL value (EA 1998a and 2000c) (see Chapter 3).

Overall, of the channel reaches, Tolchester Straightening had the greatest number of PAH analytes with mean concentrations exceeding TELs (12 analytes) and PELs (4 analytes). Mean concentrations of four PAHs from the Tolchester Straightening area exceeded TELs only in that channel and in no other channel reach (anthracene, benz(a)anthracene, benzo(a)pyrene, and pyrene).

Inside and Outside Site 104 had mean concentrations for 9 and 10 analytes, respectively, that exceeded TEL values. Both the Inside and Outside Site 104 reference areas had only one PAH analyte with a mean concentration that exceeded a PEL (acenaphthene).

6.4.9 Metals

Results of the metals analyses are provided in Table 6-11. Metals were detected in 731 of 800 cases (91%) in the channel sediments. Metals were detected in 118 of 128 cases (92%) Inside Site 104 and 58 of 64 cases (91%) Outside Site 104. Ten of the 16 tested metals (62.5%) were detected in the Ocean Reference sediment. Six metals (arsenic, copper, lead, mercury, nickel, and zinc) exceeded TEL values in 11 of the 13 channel reaches and in both Inside and Outside Site 104 (Table 6-14). Craighill Channel and Craighill Upper Range were the only channels that did not have any TEL or PEL exceedances for metals. Cadmium exceeded the TEL in three channel reaches [C&D Canal Approach (cores), Craighill Angle West, and Tolchester Channel North], and chromium exceeded the TEL in one channel reach (Cutoff Angle). Nickel and zinc were the only metals in the channel sediment that exceeded PEL values (7 and 2 channel reaches, respectively) (Table 6-15). Nickel and zinc were also the only two metals detected in the channel sediment in the 1998 testing program that exceeded the PELs (see Chapter 3).

In addition to arsenic, copper, lead, mercury, and zinc, mean silver concentrations also exceeded the TEL in both Inside and Outside Site 104 (Table 6-14). Mean cadmium and chromium concentrations also exceeded the TEL value Inside Site 104. Mean concentrations of lead, silver, and zinc concentrations exceeded the PEL value Inside Site 104. None of the metals detected Outside Site 104 exceeded PEL values (Table 6-15). None of the metals exceeded TELs at the Ocean Reference site.

The mean simultaneously extracted metals /acid volatile sulfide (SEM/AVS) ratio was less than 1 for all channel reaches, with the exception of Brewerton Eastern Extension and the C&D Canal Approach Channel (surficial sediments). Ratios that are less than 1 indicate that metals are bound to organic material and are not bioavailable. Ratios that are greater than 1 indicate that metals could be bioavailable to aquatic organisms.



6.4.10 Butyltins

Results of the butyltin analyses are provided in Table 6-12. Butyltins were detected in 4 of 200 cases (2%) in the channel sediments. Low concentrations of dibutyltin were detected in one sample each from Craighill Angle East (11 μ g/kg) and Tolchester Straightening (4.9 μ g/kg), and a low concentration of tributyltin was detected in Craighill Upper Range (4 μ g/kg). Tributyltin was detected below the lowest method calibration limit in the Ocean Reference sediment (3 μ g/kg). Butyltins were not detected either Inside Site 104 (0 of 32 cases) or Outside Site 104 (0 of 16 cases).

6.4.11 Dioxin and Furan Congeners

Results of the dioxin and furan analyses and associated Toxicity Equivalency Factors (TEFs) and Toxicity Equivalency Quotients (TEQs) are provided in Table 6-13. The TEFs represent the toxicity of each congener relative to 2,3,7,8 TCDD. All seven of the dioxin congeners and ten of the tested furan congeners were detected within channel sediments. In the 13 channel reaches, dioxin and furan congeners were detected in 106 of 221 cases (48%). All 17 of the dioxin and furan congeners were detected Inside Site 104, 4 of 17 (24%) were detected Outside Site 104, and 5 of 17 (29%) were detected in the Ocean Reference sediment.

The highest and most frequent concentrations were detected for the congeners with the lowest TEF potency (1,2,3,4,6,7,8-HpCDD, OCDD, 1,2,3,4,6,7,8-HPCDF, and OCDF). TEQs (ND=1/2 DL) ranged from 0.50 ng/kg to 6.53 ng/kg. Craighill Entrance and Craighill Channel tended to have the highest detected congener concentrations, and the highest TEQ value (ND=1/2 DL) was reported for Craighill Channel (6.53 ng/kg) (Figure 6-5). Seven of the 13 channel reaches had a TEQ value that was less than 2 ng/kg (parts per quadrillion), and 2 of the 7 concentrations were below 1 ng/kg [C&D Canal Approach (cores) and Craighill Upper Range].

TEQs for Inside Site 104, Outside Site 104, and the Ocean Reference site were 1.5 ng/kg, 0.84 ng/kg, and 0.16 ng/kg, respectively. There are no TEL/PEL values for dioxin.

6.5 THEORETICAL BIOACCUMULATION POTENTIAL (TBP)

Results of the bulk sediment analyses were used to calculate Theoretical Bioaccumulation Potential (TBP). TBP is a screening tool that provides a partial basis for selecting appropriate tissue analyses for quantification of bioaccumulation (Chapter 9). The TBP represents the approximate equilibrium tissue concentration that would be expected if the sediment or dredged material were the only source of contaminants. TBP estimates the potential concentration of a neutral organic substance that would accumulate in an organism from continuous exposure to the contaminated sediment (USACE–WES 1999). TBP is only determined for nonpolar organic compounds (pesticides, PAHs, PCBs, and dioxin and furan congeners) and is not calculated for metals, organic acids or salts, organotins, or methyl mercury. According to the ITM, TBP is an environmentally conservative value (USEPA/USACE 1998) and a conservative predictor of bioaccumulation (USACE–WES 1999); that is, calculated TBP values tend to be higher than the actual bioaccumulation values measured in tissues of organisms exposed to the same sediment. Although a substance may have the potential to cause an adverse effect, the actual likelihood of an adverse effect is a function of : (1) physical and chemical properties of the constituent, (2) actual concentration in the tissue, and (3) the period of exposure (USACE–WES 1999).

The TBP calculation requires the concentration of the contaminant found in the sediment, the percent TOC in each sediment sample, and the organisms' percent lipid content. TBP was calculated using the methods described in the ITM (USEPA/ USACE 1998). The equation for determining the TBP is as follows:

$$TBP = BSAF (C_s / \% TOC) \%L$$

Where TBP is expressed in the same concentration units as the C_s and

C_s = Concentration found in the sediment (expressed in any unit); these data are provided in Tables 6-6 (pesticides), 6-7 (PCB congeners), 6-8 (PAHs), and 6-11 (dioxin and furan congeners);

BSAF = Biota Sediment Accumulation Factor = 4 (Ankley et al. 1992);

- % TOC = Total organic carbon in the sediment (expressed as a decimal fraction); these data are provided in Table 6-3;
- % L = Lipid content of the organism (expressed as a decimal fraction of whole body weight) (USEPA/USACE 1998).

For screening purposes, TBP conservatively identifies contaminants in dredged material that may cause unacceptable bioaccumulation in aquatic organisms. TBP calculations have known limitations and uncertainty associated with estimating PAH bioaccumulation (MacFarland and Clarke 1999). Analyte-specific BSAF values may be generated and used within the calculations (Citation). A BSAF value of 4 (as recommended by the ITM) for all analytes has been referred to as "unreasonably conservative" for predicting bioaccumulation (citation). Although a BSAF value of 1 has been justified for calculating dioxin TBP (USEPA 1993; Pruell et al. 1993), a BSAF value of 4 was used for all analytical fractions in this project to identify a worst-case bioaccumulation potential.

TBP values were calculated using the mean sediment contaminant concentrations from each sampling reach. If a compound was not detected in the sediment within the channel reach, TBP was not calculated. Lipid concentrations for soft-bodied invertebrates vary depending upon the test organisms, but can range up to as high as 1-2 percent of total body wet weight (USEPA/USACE 1998). A 2 percent lipid value was used for the TBP calculations, assuming a

worst-case bioaccumulation potential.

For each analytical fraction (chlorinated pesticides, organophosphorus pesticides, PCB aroclors, PCB congeners, PAHs, and dioxin/furan congeners), three sets of numerical comparisons were conducted: approach channels vs. Inside Site 104, approach channels vs. Outside Site 104, and approach channels vs. the Ocean Reference.

TBPs of chemical constituents that exceed the placement site/reference site TBPs will be identified as COPCs to further evaluate in Tier III bioaccumulation studies.

6.5.1 Pesticides and PCB Aroclors

TBP results and comparisons for pesticides are provided in Table 6-18A (Inside Site 104), 6-18B (Outside Site 104), and 6-18C (Ocean Reference). In the 13 channel reaches, the TBP for pesticides (both chlorinated and organophosphorus) and PCB aroclors exceeded the Inside Site 104 TBP in only 4 of 286 cases (1.4%), the Outside Site 104 TBP in 8 of 286 cases (2.8%), and the Ocean Reference TBP in 9 of 286 cases (3.1%). 4,4'-DDD, 4,4'-DDE, and heptachlor epoxide were the only three pesticides that exceeded the placement site/reference site TBP values (Table 6-24). There were no TBP exceedances for either organophosphorus pesticides or PCB aroclors (see Tables 6-18A, 6-18B, and 6-18C). These results indicate only three pesticides (4,4'-DDD, 4,4'-DDE, and heptachlor epoxide) have *the potential to* bioaccumulate more in tissue exposed to channel sediments than to tissues exposed to the placement site/reference site sediments.

6.5.2 PCB Congeners

TBP results and comparisons for PCB congeners and total PCBs are provided in Tables 6-19A (Inside Site 104), 6-19B (Outside Site 104), and 6-19C (Ocean Reference). In the 13 channel reaches, the TBP for PCB congeners numerically exceeded the Inside Site 104 TBP in 46 of 338 cases (13.6%), the Outside Site 104 TBP in 210 of 338 cases (62%), and the Ocean Reference TBP in 253 of 338 cases (75%) (Table 6-24).

The C&D Approach (core) was the only channel reach where the TBP for total PCBs (ND=0) numerically exceeded the TBP for total PCBs Inside Site 104. The TBP for total PCBs (ND=DL and ND=1/2DL) in each of the 13 reaches numerically exceeded the TBP for total PCBs Outside Site 104. The TBP for total PCBs (ND=0) in 6 of the 13 reaches numerically exceeded the TBP for total PCBs in the Ocean Reference sediment. These results indicate that total PCBs in several channel reaches have *the potential* to bioaccumulate more in tissue exposed to channel sediments than to tissues exposed to placement site/reference site sediments.

6.5.3 Polynuclear Aromatic Hydrocarbons

TBP results and comparisons for PAHs are provided in Tables 6-20A (Inside Site 104), 6-20B (Outside Site 104), and 6-20C (Ocean Reference). In the 13 channel reaches, the TBP for PAHs

numerically exceeded the Inside Site 104 TBP in 11 of 208 cases (5.3%), the Outside Site 104 TBP in 28 of 208 cases (13.5%), and the Ocean Reference TBP in 194 of 208 cases (93%) (Table 6-24).

TBP values for 7 of the 16 tested PAHs numerically exceeded the TBP for Inside Site 104. Tolchester Straightening had the highest number of PAHs that exceeded Inside Site 104 TBP values (6 PAHs) and had the highest TBP values. Tolchester Straightening was the only channel reach where the TBP for Total PAHs exceeded TBP value for Inside Site 104.

TBP values for 10 of the 16 targeted PAHs exceeded the TBP for Outside Site 104. Dibenz[a,h]anthracene was the PAH with the most Outside Site 104 exceedances (7 of the 13 channel reaches). Tolchester Straightening had the highest number of PAHs that exceeded Outside Site 104 TBP values (10 PAHs), followed by Tolchester Channel North (7 PAH exceedances). Tolchester Straightening was the only channel reach where the TBP for Total PAHs exceeded the TBP value for Outside Site 104.

TBP values for all 16 of the tested PAHs exceeded the TBP for the Ocean Reference area. Fourteen of the 16 tested PAHs exceeded the Ocean Reference TBP for all 13 channel reaches. The TBP value for total PAHs (ND=0) in the every channel exceeded the Ocean Reference TBP value.

These results suggest that only PAHs have the potential to bioaccumulate more in tissue exposed to sediments from the Tolchester Straightening than from tissue exposed to sediments Inside and Outside Site 104. In addition, comparisons to the Ocean Reference data indicate that PAHs have the *potential* to bioaccumulate more in tissue exposed to all channels than to tissues exposed to the Ocean Reference sediment.

6.5.4 Dioxin and Furan Congeners

TBP results and comparisons for dioxin and furan congeners are provided in Tables 6-21A (Inside Site 104), 6-21B (Outside Site 104), and 6-21C (Ocean Reference). In the 13 channel reaches, the TBP for dioxin and furan congeners exceeded the Inside Site 104 TBP in 67 of 221 cases (30%), the Outside Site 104 TBP in 95 of 221 cases (43%), and the Ocean Reference TBP in 74 of 221 cases (34%). Overall, the Craighill Channel had the highest TBP values and the most TBP exceedances for all of the reference areas. Comparisons of TBP for TEQs indicated that dioxins and furans in several channel reaches have *the potential* to bioaccumulate more in tissue exposed to channel sediments than tissue exposed to placement site/reference area sediments.



6.6 DISCUSSION AND TIER II TBP EVALUATION

6.6.1 Frequency of Detection

Overall, of the three reference areas, the Ocean Reference sediment had the fewest detected organic constituents (Table 6-17). Metals, many of which naturally occur in sediments, were the constituents that were most frequently detected in the approach channels, Inside Site 104, Outside Site 104, and in the Ocean Reference area. PAHs were the most frequently detected organic constituents in the sediments from the approach channels, Inside Site 104, and Outside Site 104. Overall, few chlorinated VOCs, SVOCs, chlorinated pesticides, and PCB aroclors were detected in the channels and placement site/reference areas. No organophosphorus pesticides were detected in any of the sediments tested.

6.6.2 Comparisons to TELs and PELs

The number of TEL and PEL exceedances for mean concentrations in each channel reach are summarized in Tables 6-22 (TEL) and 6-23 (PEL). Inside Site 104 and Outside Site 104 had the highest number of TEL exceedances for mean concentrations, 21 and 22 respectively. Of the 13 channels, Tolchester Straightening had the highest number of TEL exceedances (20), followed by the Brewerton Eastern Extension, C&D Approach (surficial), Cutoff Angle, and Tolchester North (15). Inside Site 104 and Tolchester Straightening had the highest number of PEL exceedances (5 and 6 exceedances, respectively).

Although the number of TEL or PEL exceedances within a given channel reach can not predict toxicity, none of the tested sediments may be ruled out as non-toxic without additional evaluation (O'Connor et al. 1998).

6.6.3 Chemical Concentrations

Sediments serve as a sink and a source for natural materials, as well as organic contaminants which bind to fine particulates that may be deposited and buried within sediments. Disturbance by dredging and placement can re-mobilize contaminants and particulates from the sediment into the water column. Areas proposed for dredging in urbanized watersheds can contain measurable quantities of contaminants. Contaminants originate from both point-sources (e.g., industrial and municipal effluents) and non-point sources (e.g., stormwater runoff, agricultural runoff, and atmospheric deposition). The sediments and sediment quality of the Upper Bay is primarily influenced by non-point sources within the Chesapeake Bay watershed.

According to the 1999 State of the Chesapeake Bay Report (USEPA-Chesapeake Bay Program 1999), there is no evidence of system-wide toxic problems within the sediments of the Chesapeake Bay. Currently, there are three Chesapeake Bay Regions of Concern: Baltimore Harbor/Patapsco River, the Anacostia River, and the Elizabeth River. Sediments in these systems are severely contaminated with anthropogenic constituents and are targeted for sediment clean-up and remediation. The mainstem navigation channels proposed for dredging and

placement at Site 104 are not within the immediate vicinity of these areas and are not considered areas of concern. Overall, in the channels proposed for dredging and in sediments from Outside Site 104, there are only scattered hits of contaminants, as would be expected from sediments outside areas of concern. Sediment quality Inside Site 104 has been influenced by past placement activities and some of the chemical concentrations reported in the sediments are likely to have been influenced by these past placement activities.

The major types of contaminants that potentially occur in sediments include bulk organics (hydrocarbons that include oil and grease), halogenated hydrocarbons (persistent organics that degrade slowly), polycyclic aromatic hydrocarbons (PAHs, organics that include petroleum products and petroleum by-products), metals and nutrients.

6.6.3.1 Metals

Metals were consistently detected in the approach channels, Inside Site 104, Outside Site 104, and in the Ocean Reference sediment, and metals tended to have the most TEL exceedances. Although not statistically compared, mean concentrations of metals in the channel sediments were generally less than or comparable to the concentrations reported for both Outside and Inside Site 104.

The majority of metals detected in the sediments are naturally occurring within the environment (e.g., arsenic, cadmium, copper, lead, manganese, nickel, and zinc), and small quantities of some of these metals are essential nutrients for aquatic organisms (USEPA-CBP 1995). Metals tend to be naturally elevated in the Upper Bay region, and Eskin et al. (1996) noted that, Bay-wide, the highest concentrations and greatest variability of trace metals occur in the Upper Bay region from Pooles Island to the Bay Bridge.

Arsenic may be naturally released to the environment through volcanic eruption or by the weathering of arsenic-containing rocks. Anthropogenic sources of arsenic include fossil fuel burning and manufacturing of pesticides, wood preservatives, and fertilizers. Elevated arsenic concentrations occur throughout the Upper Bay region (Eskin et al., 1996). The concentration of arsenic reported in the Upper Bay approach channels are within the range of values reported in previous Upper Bay studies for sediments with similar grain-size characteristics (Eskin et al., 1996).

Copper may be naturally released through the weathering of rocks or release of copper sulfide. Man-made sources of copper include wood preservatives, anti-fouling paint, copper pipesand fungicides (MacDonald 1993). Mercury is released to aquatic environments from naturally occurring mercury in rocks and from anthropogenic sources such as paper mills and chemical facilities (USEPA 1999c). Incineration and fossil fuel combustion release mercury into the atmosphere and it is redeposited on land and surface waters, then adsorbed by soils and sediments. Lead primarily originates from industrial uses, including paints, batteries, leaded fuels, and metal manufacturing. Nickel and zinc are trace metals that are found in soils and sediments, but can also originate from industrial manufacturing of metals and metal alloys. Previous studies have indicated that nickel and zinc occur at naturally elevated levels in sediments of the Upper Chesapeake Bay (Eskin et al. 1996). The primary man-made source of nickel is combustion of fossil fuels, and refining and electroplating processes. Zinc is detected at high concentrations in urban stormwater, and stormwater runoff is considered to be a major source of zinc to the Upper Bay (Eskin et al. 1996)

Generally, metals accumulate in organism tissues, but most, with the exception of mercury, do not biomagnify in the food chain (Suedel et al. 1994). The bioavailability of divalent metals to aquatic organisms is influenced by the ratio of SEM/AVS. In low oxygenated environments, metals may precipitate with sulfides, making them unavailable for uptake by aquatic organisms. Brewerton Eastern Extension and the C&D Canal Approach Channel (surficial sediments) were the only reaches where the SEM/AVS ratio was greater than 1, indicating that some metals (particularly cadmium, copper, lead, nickel, and zinc) may be bioavailable in these reaches, but are not in any other reaches.

6.6.3.2 PAHs

PAHs are found throughout the environment (U.S. Department of Health and Human Services 1995; Menzie et al. 1992) and are widespread throughout the Chesapeake Bay sediments (Eskin et al., 1996; USEPA-CBP 1995). PAHs originate from both natural and anthropogenic sources. Forest fires and volcanic eruptions are the primary natural sources of PAHs while fuel combustion processes are the primary anthropogenic source. The majority of PAHs are distributed to aquatic environments via atmospheric deposition. PAHs are divided into two categories: high molecular weight (HMW) and low molecular weight (LMW) PAHs. The HMW PAHs originate from the combustion of fossil fuels and include fluoranthene, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene, benzo(g,h,i)perylene, dibenzo(a,h)anthracene, ideno(1,2,3,-cd)pyrene, and pyrene. The LMW PAHs originate from both natural sources and petroleum products and include acenaphthene, naphthalene, acenaphthylene, anthracene, fluorene, 1-methylnaphthalene, and 2-methylnaphthalene.

In the 13 channel reaches, the highest concentrations of PAHs were reported for the Tolchester Straightening. The sediments from this location represent the deepest material proposed for dredging (10-ft cores). All of the PAHs that exceeded PELs in the Tolchester Straightening were LMW PAHs. These concentrations, however, fall below the mean total PAH concentration of 4,766 ppb that Eskin et al. (1996) found in this region of the Bay. The elevated concentration of PAHs in this region of the Bay may be related to high TOC values, as PAHs have a high affinity for particulates. There are no significant point sources for PAHs in the near vicinity of the Tolchester Straightening. Eskin et al. (1994) found that PAH concentrations in Bay sediments peak in the upper Bay from Turkey Point to the Patapsco River.

The high PAHs Inside Site 104 (particularly throughout the southern end of the site) may have originated from historical dredged material placement activities that occurred at the site from 1924 through 1975. Sediments placed from 1924 to 1975 were placed prior to implementation of testing programs. These sediments may have originated from historic maintenance dredging of

the approach channels and from within Baltimore Harbor. Sediments from Inside Site 104 contained elevated concentrations of LMW PAHs compared to the majority of the channel sediments. Observations of sediment collected from the vicinity of KI-7 indicated that a petroleum-like odor and oily sheen was present approximately 4-5 inches below the sediment surface.

Although there were signs of visual contamination Inside Site 104, the mean concentrations of PAHs in sediments were similar Inside Site 104 and Outside Site 104. Inside Site 104 tended to have higher concentrations of LMW PAHs, and Outside Site 104 tended to have higher concentrations of HMW PAHs. Although not statistically compared, concentrations of PAHs in the channels proposed for maintenance dredging are comparable to or lower than the mean concentrations for both Outside and Inside Site 104, with the exception of Tolchester Straightening.

6.6.3.3 PCBs

PCBs are man-made chemicals that were historically used in electrical transformers, are widespread in the mainstem Upper Bay, are persistent in the environment (USEPA 1999a), and are known to bioaccumulate in aquatic organisms (USEPA 1999a). The mean concentrations of PCBs throughout the channels proposed for dredging were substantially lower than the TEL values. Because they are present in such low concentrations, PCBs in the channel sediments are not expected to cause adverse effects to aquatic organisms. Studies by Eskin et al. (1994) also detected low concentrations of PCB congeners in the mainstem sediments.

6.6.3.4 Pesticides

Pesticides, such as DDT and DDE, are persistent within the environment and have the potential to bioaccumulate in aquatic organisms and biomagnify in the food chain (Suedel et al. 1994). Only a few pesticides were detected in the channel sediments, and the majority of detected pesticides were present in the eastern channel reaches (C&D Approach, Tolchester North, Tolchester South, and Tolchester Straightening) and Inside Site 104. Concentrations of pesticides in the eastern channel reaches may originate from agricultural applications of pesticides or atmospheric deposition.

6.6.3.5 Dioxin and Furan Congeners

Dioxin and furan congeners are found throughout the environment (USEPA 1999b), and small quantities may be detected in any type of environmental sample (USACE–WES 1992ab). 2,3,7,8 TCDD is the most toxic dioxin congener and is the most frequently studied congener in published literature. Both natural and man-made processes may produce dioxins. Forest fires are a natural source of dioxin to the environment. The majority of polychlorinated dioxin and furan congeners, however, are the product of incomplete combustion in the presence of chlorine or the product of industrial chlorination processes (Miller, Norris, and Hawkes 1973). The most

common anthropogenic sources of dioxins include incinerators and pulp and paper mills (USEPA 1999b).

Dioxins bind tightly to particulates and are not water-soluble (USEPA 1993c); therefore, dioxin impacts are more likely to be associated with sediments than with the water column. Toxicity Equivalency Quotients (TEQs) represent a weighted summation of all dioxin and furan congeners based on the toxicity of each congener relative to 2,3,7,8, TCDD (the most toxic congener). OCDD (octochlorodibenzo-p-dioxin), the least toxic congener, was frequently detected and was the congener detected in the highest concentrations in all of the channel sediments. Sediment quality studies by Eskin et al. (1996) detected OCDD in 13 of 16 mainstem stations with concentrations ranging from 100 to 2670 ng/kg. The concentrations of OCDD reported in the channel sediments are similar to those reported by Eskin et al. (1996), ranging from 85.1 to 1040 ng/kg. 2,3,7,8-TCDD was detected in only one channel sample (Tolchester North).

There are no known point sources of dioxin to the Upper Bay. Atmospheric deposition is the most likely source of this contaminant to the Upper Bay region. The distribution of dioxins and furans does not appear to follow any consistent pattern. This suggests a more ubiquitous source and likely represents general background values throughout the sampling area.

6.6.4 Comparisons to 1998 Channel and 1997 Site 104 Sediment Data

Comparisons of the 1997 Site 104 data (see Chapter 3), the 1998 channel data (see Chapter 3), and the 1999 channel data reveal similar results for the majority of tested and detected constituents. Metal TEL/PEL comparisons for the 1999 data yielded consistently similar results to the 1998 channel data comparisons (Brewerton Eastern Extension, Craighill, Craighill Angle, Craighill Entrance, Craighill Upper Range, Cutoff Angle, Swan Point, and Tolchester). PAH TEL/PEL comparisons for the 1999 data also yielded similar results to the 1998 data comparisons. In the 1999 samples, however, a greater number of PAHs were detected above TEL values.

Comparisons of chlorinated and organophosphorus pesticides indicated that a similar combination of pesticides was detected Inside Site 104 both sampling years. Overall, fewer pesticides were detected in Brewerton Eastern Extension and Outside Site 104 in the 1999 sampling compared to the 1998 sampling. PCB aroclors were detected in 6 of the 8 tested channels in 1998, but were not detected in 1999. In addition, comparisons of VOCs and SVOCs results in 1998 and 1999 indicated that a similar suite of analytes was detected in both fractions in the channel sediments both sampling years.

Overall, comparison of the two sets of sediment chemistry data indicates that the analytical methods are consistently detecting a similar or equivalent subset of analytes in the channel reaches (for those channels that were sampled both years). This consistency yields a level of confidence that the sampling is adequately characterizing the areas proposed for dredging.

6.6.5 Theoretical Bioaccumulation Potential

Results of the TBP calculations and reference site comparisons indicated that several pesticides and the majority of the PCB congeners, PAHs, and dioxin/furan congeners have the potential to bioaccumulate to higher levels in aquatic organisms exposed to channel sediments than in organisms exposed to one or more of the placement site/reference sediments. Although a substance may have the potential to bioaccumulate and cause an adverse effect, the actual likelihood of an adverse effect is a function of: (1) physical and chemical properties of the constituent, (2) actual concentration in the tissue, and (3) the period of exposure (USACE–WES 1999).

Results of the TBP comparisons identified organic COPCs which require further evaluation in the Tier III bioaccumulation testing.

6.6.6 Tier II Sediment Conclusions

According to the ITM (USEPA/USACE 1998), after consideration of the Tier II TBP data in a manner comparable to that which would be used to make a decision in higher tiers, one of the following two conclusions is reached:

- 1. The dredged material is predicted to not result in unacceptable adverse effects due to bioaccumulation of the measured non-polar organic compounds, or
- 2. The available information is not sufficient to predict whether the dredged material will result in unacceptable adverse effects due to bioaccumulation of the measured non-polar organic compounds, and further evaluation of bioaccumulation in Tier III is necessary to furnish information to make determinations under the guidelines.

Results of the TBP evaluation indicate that bioaccumulation testing in Tier III is warranted to determine the actual bioaccumulation of COPCs in tissues exposed to the channel sediments and placement site/reference sediments. Methodology and results of the bioaccumulation studies are provided in Chapter 9.

Results of the TEL/PEL comparisons indicated that none of the sediments could be ruled out as non-toxic. Sediments from each channel and placement/reference area were directly tested for acute water column and sediment toxicity in Tier III. Methodology and results of the toxicity testing are provided in Chapter 8.
















TABLE 6-1 DIOXIN AND FURAN DATA QUALIFIERS

.

| A or J | Amount detected is less than the Method Calibration Limit. |
|--------|--|
| B | Detected in the blank. |
| Ε | Amount detected is over the Method Calibration Limit. |
| DPE | Denotes the presence of possible polychlorinated diphenyleters. |
| EDL | "Estimated Detection Limit" |
| EMPC | "Estimated Maximum Possible Concentration" |
| ppt | Parts-per-trillion (pg/g; ng/L) |
| Q | Indicated the presence of quantitative interferences. They generally result in an underestimation of the affected total homologue groups. |
| V | Recovery is lower than 40%. The data has been validated based upon a favorable signal-to-noise and detection limit. |

TABLE 6-2MEAN VALUES FOR PHYSICAL CHARACTERISTICS IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE
PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | Ocean | Outside | Inside | Brewerton Eastern | C&D | C&D | | Craighill | Cralghill | Cralphill | Craighill Upper | Cutoff | Swan | Tolchester | Tolchester | Tolebester |
|------------------|-----------|-----------|----------|----------|----------------------|----------|----------|-----------|------------|------------|-----------|--------------------|--------|-------|------------|------------|---------------|
| RE | EACH ID: | Reference | Site 104 | Site 104 | Extension | Approach | Approach | Cralghiil | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| SAMPL | .E TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| SAMPLE | SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | 4 | 2 |
| % M(| DISTURE | 21.5 | 71.2 | 60.1 | 71.7 | 62.4 | 56.5 | 42.1 | 69.1 | 72.0 | 69.6 | 40.8 | 68.4 | 68.3 | 62.1 | 64.3 | 55.0 |
| ANALYTE | UNIT | | | | | | | | | | | | | | | | |
| CLAY | % | 11.4 | 36.58 | 34.3 | 44.42 | 27.92 | 48.23 | 22.94 | 55.5 | 54.43 | 44.34 | 17.34 | 30.75 | 48.06 | 45.54 | 37.84 | 50.77 |
| GRAVEL | % | 0 | 0 | 4.04 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5.95 | 0 | 0 | 0 | 0 |
| SAND | % | 26.6 | 26.1 | 19.47 | 3.87 | 5.96 | 1.07 | 66.74 | 1.67 | 7.3 | 10.1 | 31.7 | 13.98 | 7.4 | 1.17 | 2.06 | <u></u> |
| SILT | % | 62 | 37.32 | 42.19 | 51.72 | 66.12 | 50.7 | 10.32 | 42.83 | 38.27 | 45.56 | 50.96 | 49.33 | 44.54 | 53.29 | 60.1 | 48.13 |
| SPECIFIC GRAVITY | T/4C | 1.7 | 1.25 | 1.35 | 1.22 | 1.33 | 1.4 | 1.57 | 1.2 | 1.23 | 1.2 | 1.73 | 1.2 | 1.22 | 1.33 | 1 29 | 40:15 |







TABLE 6-3MEAN CONCENTRATIONS OF INORGANIC NON-METALS (MG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDEPLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | S STATI | REACH ID: AMPLE TYPE: IONS SIZE (N): | Ocean Reference Grab | Outside Site 104 Grab 4 | Inside Site 104 Grab 8 | Brewerton Eastern Extension Grab 5 | C&D Approach Grab 4 | C&D Approach Core 2 | CraighIII Grab 3 | Craighili Angle East Core 3 | Craighill Angle West Core 3 | Craighill Entrance Core 4 | Craighill Upper Range Grab 4 | Cutoff Angle Grab 3 | Swan Point Core 6 | Tolchester North Core 6 | Tolchester South Core 4 | Tolchester Straightening Core 2 |
|-------------------------------|------------|--|----------------------------|----------------------------------|---------------------------------|--|------------------------------|------------------------------|------------------------|--------------------------------------|--------------------------------------|------------------------------------|--|------------------------------|----------------------------|----------------------------------|----------------------------------|--|
| ANALYTE | UNIT | DL (RANGE) | EN | | | | | | | | | | _ | | | | | |
| CYANIDE | MG/KG | 0.04 - 0.22 | 0.06 U | EE:26.1.44 | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.11 U | 6.14 | 0.06 U | 0.06 U | 0.2 | 0.09 | 0.09 U | 0.33 |
| NITROGEN, AMMONIA | MG/KG | 1-3.4 | 2.0 B | 2182.35 | 36.7 | 104.56 | 19.48 | Ar. 4 35.5 | 验表题 14.1 | 113.97 | 131.23 | 223.0 | 200 30 12.0 | Later 63.6 | 146.17 | 87.02 | 160.0 | 67.0 |
| NITROGEN, NITRATE AND NITRITE | MG/KG | 0.02 - 0.08 | Asiate 0.52 | 0.82 | A \$ 0.32 | \$- ExeCx1.47 | JAN 1.1.01 | 28:061.6 | 0.99 | 0.25 | 0.37 | State 0.54 | Sep. 40.71 | 2.46 | 0.05 U | 0.86 | | 0.03 U |
| NITROGEN, TOTAL KJELDAHL | MG/KG | 4.2 - 91.5 | Mar 180.0 | 2990.0 | 1326.38 | 3714.0 | AN 561.0 | 1825.0 | 225.17 | 734.67 | 2863.33 | 1093.75 | 759.25 | 1816.67 | 1031.0 | 1121.17 | 1138.75 | 339.0 |
| OXYGEN DEMAND, BIOCHEMICAL | MG/KG | 0.37 - 3175 | ALS: 118.0 | 2015.25 | 2169.5 | 846.66 | ses 1004.5 | 1049.0 | 1025.67 | 1162.67 | 1150.0 | 1270.0 | an 774.75 | 1470.0 | 1640.18 | 1134.33 | 1625.0 | 1540.0 |
| OXYGEN DEMAND, CHEMICAL | MG/KG | 243 - 8900 | 6150.0 | 69625.0 | 123287.5 | 141940.0 | -173000.0 | # 41100.0 | 25841.0 | 219800.0 | 305333.33 | 142350.0 | 24402.5 | 127666.67 | 162933.33 | 211233.33 | 144450.0 | 213000.0 |
| PHOSPHORUS, TOTAL | MG/KG | 1.8 - 5.2 | 218.0 | 478.25 | 345.75 | 501.6 | 425.75 | Mar 435.5 | 258.7 | 537.67 | 1616 632.9 | 588.0 | 198.0 | 514.0 | 571.0 | 472.5 | 523.5 | 421.0 |
| SULFIDE, TOTAL | MG/KG | 30.6 - 98.4 | 30.7 U | Se 1735.0 | 1146.5 | 93.56 | 331.07 | 30.7 U | 643.0 | 591.67 | ÷ 1409.33 | 317.9 | 195.63 | Aut 668.97 | 1535.67 | 695.9 | 559.8 | A |
| TOTAL ORGANIC CARBON | MG/KG | 474 - 7200 | 5310.0 | 114480.0 | 72222.2 | 105333.33 | 94630.0 | \$4400.0 | 29997.5 | 102425.0 | 95475.0 | 74320.0 | 27530.0 | 79400.0 | 86827.14 | * 86371.43 | 89660.0 | 133666.67 |

U = not detected in any sample within reach; value represents mean detection limit (DL). B = value <RL but >IDL/MDL.

| | | ſ | | | | Brewerton | | | | | | | Cralabili | | | | | |
|---------------------------|--------|---------------|-----------|----------|----------|-----------|----------|----------|-----------|------------|------------|-----------|-----------|----------|----------|------------|------------|---------------|
| | | | Ocean | Outside | Inside | Factorn | C&D | C&D | | Cralabill | Cralebill | Craighill | Linner | Cutoff | Swan | Tolebester | Tolebester | Talabaataa |
| | | REACHID: | Reference | Site 104 | Site 104 | Extension | Annroach | Annroach | Craighill | Angle Fast | Angle West | Entrance | Range | Angle | Point | North | South | Strolohtoning |
| | S | MPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | SAM | PLE SIZE (N): | 1 | 4 | 5. | 4* | 4 | 2 | 3 | 3 | 3 | 4 | 30 | 3 | 6 | 6 | Core | 2 |
| ANALYTE | LINHT | DI (DANCE) | | | | | | | | | | | | | | | | |
| | UG/KG | 07.2 | 0.60.11 | 1.7511 | 1 19 11 | 2.00.11 | 1.0011 | 1.0011 | 0.90 11 | 1 50 11 | 2 00 11 | 1 50 11 | 0.022.11 | 2 00 11 | 1 60 11 | 11711 | 2 00 11 | 1.00.11 |
| 1 1 2 2 TETRACULOROETHANE | LIC/KG | 0.7-2 | 0.00 0 | 2 25 11 | 1.10 0 | 2.00 U | 1.000 | 1.50 11 | 0.90 0 | 1.90 0 | 2.00 U | 2.00 11 | 1 667 11 | 2.00 U | 1.50 U | 1.170 | 2.00 U | 1.00 U |
| | UG/KG | 0.8 - 3 | 0.80 U | 2.250 | 1.00 U | 2.00 U | 1.75 0 | 1.50 U | 0.93 U | 1.03 U | 2.00 U | 2.00 U | 1.0070 | 2.00 U | 2.00 U | 1.6/ U | 2.00 U | 1.00 U |
| 1 1 DICHI OPOETHANE | UG/KG | 0.8-3 | 0.50 [] | 1 50 11 | 1.000 | 1.00 [] | 1.0011 | 0.9511 | 0.73 0 | 1.00 [1] | 1.6711 | 2.00 U | 0.00.11 | 2.00 U | 2.00 0 | 1.67 U | 2.00 U | 1.00 U |
| | UG/KG | 0.7-2 | 0.50 11 | 1.500 | 1.12 0 | 2.00 11 | 1.0011 | 1.0011 | 0.770 | 1.50 11 | 2.00 11 | 1.00 0 | 0.90 0 | 2.00 11 | 1.170 | 0.970 | 1.00 U | 0.80 U |
| 1.2 DICITLOROPENZENE | UG/KG | 1.5 | 1.0011 | 3 25 11 | 3.00 11 | 3 50 11 | 2 7511 | 2 50 11 | 1.6711 | 3 17 11 | 4.00 11 | 1.50 0 | 2 22 11 | 2.00 0 | 2 17 11 | 2.22.11 | 2.00 U | 1.00 U |
| 1.2-DICHLOROFTHANE | UG/KG | 08-3 | 0.8011 | 2 25 11 | 1.80 11 | 2 00 11 | 1.7511 | 1.50 11 | 0.93 11 | 1 83 11 | 2.00 U | 2.00.11 | 1.6711 | 2.0011 | 3.170 | 2.33 0 | 3.23 U | 2.00 U |
| | UG/KG | 0.8 - 3 | 0 80 11 | 2.25 0 | 1.80 11 | 2.00 U | 1.7511 | 1.50 1 | 0.93 11 | 1.83 [] | 2.00 U | 2.00 0 | 1.6711 | 2.00 U | 2.00 0 | 1.07 U | 2.00 U | 1.00 U |
| 1.4.DICIII OROBENZENE | UG/KG | 1 - 4 | 1.0011 | 3.0011 | 2 20 11 | 2 50 11 | 2.0011 | 2 00 11 | 16711 | 2 50 11 | 3.00 [] | 2.00 0 | 1.6711 | 2.00 0 | 2.00 0 | 2.0011 | 2.00 U | 1.00 U |
| 2-CHLOROFTHYL VINYL FTHER | UG/KG | 3 - 10 | 2.0011 | 7.2511 | 5.80 U | 6.50 U | 5.2511 | 5.00 11 | 4.0011 | 61711 | 7 67 11 | 6.00 11 | 4.6711 | 6 67 11 | 6 50 11 | 5 17 11 | 6 75 11 | 2.00 U |
| ACROLEIN | UG/KG | 27-95 | 25.00 U | 70.75 U | 56.80 U | 66.25 U | 52.50 U | 49.00 U | 37.67 U | 61.3311 | 77.00 11 | 59.00 U | 46 33 11 | 66 67 11 | 63.00.11 | 51 1711 | 68.00.11 | 41.00 11 |
| ACRYLONITRILE | UG/KG | 11 - 38 | 10.00 U | 28.25 U | 22.80 U | 26.75 U | 20.75 U | 20.00 11 | 15.00 U | 24.6711 | 31.00 11 | 23 75 11 | 18 33 11 | 27.0011 | 25 33 11 | 20.50 11 | 27 50 11 | 41.00 U |
| BENZENE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.67 U | 2.00 11 | 2 00 11 | 11711 | 2 00 11 | 1.0011 |
| BROMODICHLOROMETHANE | UG/KG | 0.9 - 3 | 0.90 U | 2.50 U | 2.00 U | 2.00 U | 2.00 U | 2.00 U | 1.30 U | 2.00 U | 2.17 U | 2.00 U | L.17 U | 2.00 U | 2.17 U | 2.0011 | 2.00 0 | 1.0011 |
| BROMOMETHANE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 2 00 11 | 1671 | 2.00 U | 1.00 11 |
| CARBON DISULFIDE | UG/KG | 0.5 - 9 | 0.50 U | 5.50 | 7.98 | 3.75 | 2.25 | 0.95 U | 3.43 | 1.00 U | 1.17 U | 1.00 U | 5.90 | 5.33 | 7.33 | 0.9711 | 1.75 | 3.90 |
| CARBON TETRACHLORIDE | UG/KG | 0.7 - 2 | 0.60 U | 1.75 U | 1.18 U | 2.00 U | 1.00 U | 1.00 U | 0.90 U | 1.50 U | 2.00 U | 1.50 U | 0.93 U | 2.00 U | 1.50 U | L.17U | 2.00 U | 1.0011 |
| CFC-11 | UG/KG | 0.5 - 2 | 0.50 U | 1.50 U | 1.12 U | 1.00 U | 1.00 U | 0.95 U | 0.77 U | 1.00 U | 1.17 U | 1.00 U | 0.90 U | 1.00 U | 1.17 U | 0.97 U | 1.00 U | 0.80 U |
| CFC-12 | UG/KG | 0.7 - 2 | 0.60 U | 1.75 U | 1.18 U | 2.00 U | 1.00 U | 1.00 U | 0.90 U | 1.50 U | 2.00 U | 1.50 U | 0.93 U | 2.00 U | 1.50 U | 1.17 U | 2.00 U | 1.00 U |
| CHLOROBENZENE | UG/KG | 1-4 | 1.00 U | 3.25 U | 2.60 U | 3.00 U | 2.00 U | 2.00 U | 1.17 U | 2.50 U | 3.67 U | 2.75 U | 2.00 U | 3.00 U | 3.00 U | 2.17 U | 3.00 U | 2.00 U |
| CHLORODIBROMOMETHANE | UG/KG | 1-5 | 1.00 U | 3.25 U | 3.00 U | 3.50 U | 2.75 U | 2.50 U | 1.17 U | 3.17 U | 4.00 U | 3.00 U | 2.33 U | 3.33 U | 3.17 U | 2.33 U | 3.25 U | 2.00 U |
| CHLOROETHANE | UG/KG | 1-5 | 1.00 U | 3.25 U | 3.00 U | 3.50 U | 2.75 U | 2.50 U | 1.17 U | 3.17 U | 4.00 U | 3.00 U | 2.33 U | 3.33 U | 3.17 U | 2.33 U | 3.25 U | 2.00 U |
| CHLOROFORM | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17 Ú | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 1.00 U |
| CHLOROMETHANE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 1.00 U |
| CIS-1,3-DICHLOROPROPENE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17U | 2.00 U | 2.00 U | 1.I7U | 2.00 U | 1.00 U |
| DICHLOROMETHANE | UG/KG | 1-6 | 1.00 U | 12.25 | 12.20 | 15.75 | 8.00 | 2.00 U | 8.00 | 4.83 | 3.67 U | 2.75 U | 2.00 U | 3.00 U | 3.00 U | 2.17 U | 3.00 U | 2.00 U |
| ETHYLBENZENE | UG/KG | 0.7 - 2 | 0.60 U | 1.75 U | 1.18 U | 2.00 U | 1.00 U | 1.00 U | 0.90 U | 1.50 U | 2.00 U | 1.50 U | 0.93 U | 2.00 U | 1.50 U | 1.17 U | 2.00 U | 1.00 U |
| M-DICHLOROBENZENE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 1.00 U |
| METHYLBENZENE | UG/KG | 0.9 - 3 | 0.90 U | 2.50 U | 2.00 U | 2.00 U | 2.00 U | 2.00 U | 1.30 U | 2.00 U | 2.17 U | 2.00 U | 1.17 U | 2.00 U | 2.17 U | 2.00 U | 2.00 U | 1.00 U |
| TETRACHLOROETHENE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 1.00 U |
| TRANS-1,2-DICHLOROETHENE | UG/KG | 0.8 - 3 | 0.80 U | 2.25 U | 1.80 U | 2.00 U | 1.75 U | 1.50 U | 0.93 U | 1.83 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 2.00 U | 1.17 U | 2.00 U | 1.00 U |
| TRANS-1,3-DICHLOROPROPENE | UG/KG | 1-4 | 1.00 U | 3.25 U | 2.60 U | 3.00 U | 2.00 U | 2.00 U | 1.17U | 2.50 U | 3.67 U | 2.75 U | 2.00 U | 3.00 U | 3.00 U | 2.17 U | 3.00 U | 2.00 U |
| TRIBOMOMETHANE | UG/KG | 1-4 | 1.00 U | 3.25 U | 2.60 U | 3.00 U | 2.00 U | 2.00 U | 1.17 U | 2.50 U | 3.67 U | 2.75 U | 2.00 U | 3.00 U | 3.00 U | 2.17 U | 3.00 U | 2.00 U |
| TRICHLOROETHYLENE | UG/KG | 0.9 - 3 | 0.90 U | 2.50 U | 2.00 U | 2.00 U | 2.00 U | 2.00 U | 1.30 U | 2.00 U | 2.17U | 2.00 U | 1.17U | 2.00 U | 2.17 U | 2.00 U | 2.00 U | 1.00 U |
| VINYL CHLORIDE | UG/KG | 0.7 - 2 | 0.60 U | 1.75 U | 1.18 U | 2.00 U | 1.00 U | 1.00 U | 0.90 U | 1.50 U | 2.00 U | 1.50 U | 0.93 U | 2.00 U | 1.50 U | 1.17 U | 2.00 U | 1.00 U |

TABLE 6-4 MEAN CONCENTRATIONS OF VOLATILE ORGANIC COMPOUNDS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

U = not detected in any sample within reach; value represents mean detection limit (DL).

* = volatiles analyzed for 5 sample Inside Site 104, 4 samples Brewerton Eastern Extension, and 3 samples Craighill Upper Range.







TABLE 6-5 MEAN CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | | | | | Brewerton | | | | | | | Cralphill | | | | | |
|-------------------------------------|--------|--------------|-----------|-----------|------------|-----------|--|-----------|------------|------------|------------|-------------|-----------|------------|-------------|------------|------------|---------------|
| | | | Ocean | Outside | Incide | Fastern | C&D | C&D | | Craighill | Cralabili | Craiabill | linner | Cutoff | | Toleberter | Talabartas | Talabastas |
| | | BEACH ID. | Deference | Site 104 | Site 104 | Extension | Approach | Annroach | Crainbill | Angle Fact | Angle West | Entrance | Deper | Angle | Sugar Balat | Name | Couth | 1 oicnester |
| | | MOLE TYPE. | Cash | Crah | Crob | Cash | Grah | Core | Crab | Come | Care Care | Com | Cash | Angre | Swan Point | North | Sourn | Straightening |
| | CANE | MFLE I I FE: | UIAU | Giao | 0140 | Giao | Giao | 2 | 1 | 2 | 2 | Cule | Grao | Giao | Core | Core | Core | Core |
| | SAMI | LE SIZE (N): | | | 0 | | | 4 | | | | - | 4 | 3 | 0 | 0 | | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | | | | | | | |
| 1,2,4-TRICHLOROBENZENE | UG/KG | 74 - 180 | 76.00 U | 129.75 U | 104.00 U | 136.00 U | 101.00 U | 82.50 U | 90.00 U | 120.00 U | 130.00 U | 130.00 U | 75.50 U | 136.67 U | 118.33 U | 98.33 U | 102.75 U | 85.00 U |
| 1,2-DICHLOROBENZENE | UG/KG | 53 - 130 | NA | 94.50 U | 75.88 U | 97.20 U | 73.00 U | 59.50 U | 64.00 U | 87.00 U | 93.33 U | 92.50 U | 54.00 U | 100.00 U | 85.83 U | 70.50 U | 74.75 U | 61.50 U |
| 1,2-DIPHENYLHYDRAZINE | UG/KG | 54 - 130 | 55.00 U | 97.50 U | 76.75 U | 100.00 U | 74.25 U | 60.50 U | 65.33 U | 88.67 U | 94.33 U | 94.00 U | 55.00 U | 103.67 U | 87.33 U | 71.83 U | 75.75 U | 62.50 U |
| 1,4-DICHLOROBENZENE | UG/KG | 72 - 170 | NA | 129.00 U | 102.00 U | 136.00 U | 99.75 U | 80.50 U | 85.33 U | 119.33 U | 123.33 U | 125.00 U | 73.50 U | 136.667 U | 115.00 U | 94.50 U | 101.25 U | 83.00 U |
| I-METHYLNAPHTHALENE | UG/KG | 12 - 180 | 12.00 U | 44.50 | 1 03.48 | 31,80 | ······································ | 13.50 U | 15.67 | 22.00 | 23.00 | 1.5+6 26.00 | A #12.75 | 22.33 U | 29.00 | 53.17 | 41.25 | 210.00 |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | UG/KG | 81 - 200 | 84.00 U | 123.25 | 117.00 U | 148.00 U | 111.00 U | 91.00 U | 98.67 U | 133.33 U | 143.33 U | 140.00 U | 83.50 U | 153.33 U | 133.33 U | 108.33 U | 115.00 U | 94.00 U |
| 2,4,6-TRICHLOROPHENOL | UG/KG | 76 - 180 | 78.00 U | 135.00 U | 107.385 U | 138.00 U | 102.00 U | 84.50 U | 91.33 U | 123.33 U | 133.33 U | 130.00 U | 77.50 U | 143.33 U | 123.33 U | 99.67 U | 106.00 U | 87.50 U |
| 2,4-DICHLOROPHENOL | UG/KG | 67 - 160 | 69.00 U | 119.75 U | 95.125 U | 126.00 U | 91.75 U | 74.50 U | 82.00 U | 110.33 U | 116.67 U | 120.00 U | 68.00 U | 126.67 U | 107.83 U | 89.00 U | 93.25 U | 77.50 U |
| 2,4-DIMETHYLPHENOL | UG/KG | 130 - 310 | 130.00 U | 227.50 U | 181.25 U | 234.00 U | 175.00 U | 145.00 U | 153.33 U | 206.67 U | 226.67 U | 220.00 U | 130.00 U | 240.00 U | 206.67 U | 170.00 U | 180.00 U | 150.00 U |
| 2,4-DINITROPHENOL | UG/KG | 620 - 1500 | 630.00 U | 1107.50 U | 883.75 U | 1160.00 U | 852.50 U | 690.00 U | 750.00 U | 1016.67 U | 1100.00 U | 1095.00 U | 630.00 U | 1176.67 U | 996.67 U | 823.33 U | 870.00 U | 720.00 U |
| 2,4-DINITROTOLUENE | UG/KG | 50 - 120 | 51.00 U | 89.00 U | 71.25 U | 92.80 U | 68.75 U | 56.00 U | 60.667 U | 82.33 U | 88.00 U | 87.50 U | 51.00 U | 94.33 U | 81.17 U | 66.67 U | 70.25 U | 58.00 U |
| 2,6-DINITROTOLUENE | UG/KG | 57 - 140 | 58.00 U | 101.50 U | 81.50 U | 106.00 U | 78.00 U | 64.00 U | 69.00 U | 92.67 U | 101.00 U | 100.25 U | 58.00 U | 108.33 U | 91.17 U | 75.67 U | 80.00 U | 66.00 U |
| 2-CHLORONAPHTHALENE | UG/KG | 73 - 180 | 75.00 U | 129.25 U | 103.63 U | 136.00 U | 100.25 U | 81.50 U | 89.33 U | 119.67 U | 126.67 U | 130.00 U | 74.50 U | 136.67 U | 118.33 U | 97.17 U | 101.75 U | 84.00 U |
| 2-CHLOROPHENOL | UG/KG | 63 - 150 | 65.00 U | 111.00 U | 89.25 U | 116.00 U | 86.25 U | 70.50 U | 76.00 U | 102.00 U | 113.33 U | 110.00 U | 64.00 U | 118.00 U | 100.33 U | 83.67 U | 88.25 U | 73.00 U |
| 2-METHYL-4,6-DINITROPHENOL | UG/KG | 63 - 150 | 65.00 U | 111.00 U | 89.25 U | 116.00 U | 86.25 U | 70.50 U | 76.00 U | 102.00 U | 113.33 U | 110.00 U | 64.00 U | 118.00 U | 100.33 U | 83.67 U | 88.25 U | 73.00 U |
| 2-METHYLNAPHTHALENE | UG/KG | 7.8 - 110 | 7.90 U | 24-178.25 | 346.38 | 61.20 | 86.25 | 11.50 | 19.93 | 27.33 | 29.00 | 59,50 | 23.25 | 32.00 | 54.67 | 135.67 | 84.50 | 490.00 |
| 2-METHYLPHENOL | UG/KG | 76 - 180 | 78.00 U | 135.00 U | 107.38 U | 138.00 U | 102.00 U | 84.50 U | 91.33 U | 123.33 U | 133.33 U | 130.00 U | 77.50 U | 143.33 U | 123.33 U | 99.6711 | 106.0011 | 87 50 11 |
| 2-NITROPHENOL | UG/KG | 54 - 130 | 55.00 U | 97.50 U | 76.75 U | 100.00 U | 74.25 U | 60.50 U | 65.33 U | 88.67 U | 94.33 U | 94.00 U | 55.00 U | 103.67 U | 87.33 U | 71.8311 | 75.7511 | 62 50 11 |
| 3.3'-DICHLOROBENZIDINE | UG/KG | 46 - 110 | 47.00 U | 82.50 U | 65.38 U | 85.20 U | 63.25 U | 51.50 U | 55.67 U | 75.67 U | 81.00 U | 80.50 U | 47.00 U | 86.33 U | 74.67 U | 61.33 U | 64.7511 | 53 50 11 |
| 3 4-METHYLPHENOL | UG/KG | 160 - 380 | 160.00 U | 342.50 | 213.38 | 308.00 | 292.50 | 290.00 | 166.67 | 336.67 | 323.33 | 362.50 | 132.25 | 210.00 | 310.00 | 295.00 | 300.00 | 205.00 |
| 3.5.5-TRIMETHYL-2-CYCLOHEXENE-1-ONE | UG/KG | 79 - 190 | 82.00 U | 142.50 U | 112.50 U | 148.00 U | 110.50 U | 89.00 U | 97.33 U | 130.00 U | 140.00 U | 140.00 11 | 81.50 11 | 153 33 11 | 128.3311 | 106 50 11 | 112 25 11 | 92 50 11 |
| 4-BROMOPHENYL PHENYL ETHER | UG/KG | 60 - 150 | 61.00 U | 107.50 U | 86.13 U | 112.00 U | 82.25 U | 67.00 U | 72.33 U | 97.33 U | 103.33 U | 103.75 U | 61.00 U | 113.33 U | 97.5011 | 79 50 11 | 84.00 11 | 69 50 11 |
| 4-CHLORO-3-METHYLPHENOL | UG/KG | 69 - 170 | 71.00 U | 123.00 U | 97.63 U | 126.00 U | 93.00 U | 77.00 U | 83.33 U | 111.33 U | 123.33 U | 120.00 11 | 70.00 11 | 126.6711 | 110.00 11 | 91.167.11 | 96 75 11 | 79 50 11 |
| 4-CHLOROPHENYL PHENYL ETHER | UG/KG | 70 - 170 | 72.00 U | 125.75 U | 98.50 U | 128.00 U | 93.75 U | 78.00 U | 84.00 U | 115.00 U | 123.33 U | 120.00 11 | 71.00 11 | 130.00 LI | 113.33.11 | 92 83 11 | 97 75 11 | 80.501/ |
| 4-NITROPHENOL | UG/KG | 51 - 120 | 52.00 U | \$3.00 | 72.38 U | 94.40 U | 70.50 U | 57.00 U | 61.67 U | 83.67 U | 89.67 U | 89.00 11 | 52.00 LI | 95.6711 | 82 83 11 | 67.8311 | 72.00 11 | 59.0011 |
| BENZOICACID | UG/KG | 1600 - 3800 | 1600.00 U | 2825.00 U | 2250.00 U | 2940.00 U | 2175.00 U | 1750.00 U | 1900.00 11 | 2533.33 U | 2800.0011 | 2775.00 11 | 600 00 11 | 3000 00 11 | 2566 67 11 | 2083 33 11 | 2225 00 11 | 1800.0011 |
| BENZYL ALCOHOL | UG/KG | 57 - 140 | 58.00 11 | 101.50 11 | 81.50 U | 106.00 U | 78.00 U | 64.00 U | 69.00 11 | 92.67 11 | 101.00.11 | 100 25 11 | 58 00 11 | 108 33 11 | 91 17 11 | 75 67 11 | 80.00 11 | 66 00 11 |
| BENZYL BUTYL PHTHALATE | LIG/KG | 55 - 130 | 56.00 U | 98.50 U | 77.88 11 | 93.40 | 75.5011 | 62.00 U | 66.6711 | 90.0011 | 95 33 11 | 95 00 11 | 56 00 11 | 104.00 11 | 89 17 11 | 73 17 11 | 77 25 11 | 63 50 11 |
| BIS(2-CHI OROFTHOXY)METHANE | LIG/KG | 68 - 160 | 70.00 U | 120 25 11 | 95.7511 | 126.00 11 | 92.2511 | 75.50 U | 82 67 11 | 111.00 11 | 120 00 11 | 120.00 11 | 69 00 11 | 126 67 11 | 108.00.11 | 90 17 11 | 01 75 11 | 78 50 11 |
| BIS(2 CHLOROFTHYL) FTHER | LIG/KG | 62 - 150 | 63.0011 | 110 75 11 | 88 38 11 | 116 00 11 | 85 25 11 | 69 00 11 | 75.00 11 | 101 67 11 | 110.00.11 | 109 50 11 | 63.00.11 | 117 67 11 | 99.67.11 | 82 33 11 | 87.00 11 | 72.00 11 |
| DIS(2-CHEOROCHTE) PHTHALATE | UG/KG | 52 - 130 | 53.00 11 | 91.00 | 75.25 | 95.00 | 66.75 | 60 50 | 63.00.11 | 135.00 | 91 33 11 | 90 75 11 | 53.00 11 | 00 33 11 | 60 17 | 60 22 11 | 72 25 11 | 60.50 11 |
| DIN BUTYL PHTHALATE | UG/KG | 45 - 110 | 46.00 11 | 77.50 | 64 50 11 | 83.40.11 | 62.2511 | 51 50 | 54 67 11 | 04.00 | 178 67 | 104.00 | 46.00 11 | 85 67 11 | 72 22 11 | 60 17 11 | 63 75 11 | 52 50 11 |
| DI N OCTVI BUTUALATE | UG/KG | 63 - 150 | 65 00 11 | 103.25 | 89 25 11 | 116.00 11 | 86 25 11 | 70 50 11 | 76.00.11 | 102.00.11 | 112 32 11 | 110.00.11 | 64.00.11 | 119 00 11 | 100 22 11 | 00.17 0 | 03.75 U | 32.30 U |
| DIPENZOELIPAN | UG/KG | 72 - 170 | 74.00.11 | 142 50 | 105 50 | 136.00 11 | 99 75 11 | 80 50 11 | 85 33 11 | 110 33 11 | 122 22 11 | 125 00 11 | 72 50 11 | 126 67 11 | 115 00 11 | 04.60.11 | 101.25 11 | 13.00 0 |
| | UG/KG | 46 - 110 | 47.00 11 | 142.00 | 65 375 11 | 85 20 11 | 63.2511 | 51 50 11 | 556711 | 75 67 11 | 91.00.11 | 90.5011 | 47.00.11 | 96 22 11 | 74 67 11 | 94.30 0 | 64.76.11 | 63.50 |
| | UG/KG | 40-110 | 47.00 0 | 07 50 11 | 76 75 11 | 100.00.11 | 74 25 11 | 60.6010 | 66 33 11 | 99 67 11 | 04.22.11 | 04.0011 | 66.00.11 | 80.33 0 | /4.0/ U | 01.33 U | 04./5 U | 53.50 U |
| | UG/KG | 54 - 150 | 69 00 11 | 117.00 11 | 04 00 11 | 122.00 11 | 00 60 11 | 72 60 11 | 78 00 11 | 110.00 11 | 94.33 U | 94.00 0 | 55.00 U | 103.67 0 | 87.33 U | /1.83 U | /5./5 U | 62.30 U |
| HEXACHLORO-1,3-BUTADIENE | UCKG | 68 140 | 60.00 1 | 104 60 11 | 94.000 | 106.00.11 | 70.30 0 | 73.30 0 | 78.00 0 | 03.00.11 | 113.33 U | 115.00 0 | 67.00 U | 126.33 U | 105.670 | 87.67 U | 92.25 0 | /6.50 U |
| HEXACHLOROBENZENE | UU/KG | 38 - 140 | 39.00 U | 104.30 U | 82.03 U | 100.00 U | 19.13 U | 05.00 0 | 70.00 0 | 93.00 0 | 102.00 U | 100.50 0 | 59.00 U | 109.00 0 | 95.33 U | 77.00 U | 81.50 U | 67.00 U |
| HEXACHLOROCYCLOPENTADIENE | UG/KG | 130 + 310 | 01.00 U | 107.50 0 | 181.25 U | 234.00 U | 175.00 0 | 145.00 0 | 155.53 U | 200.0/ U | 226.67 U | 220.00 U | 130.00 0 | 240.00 U | 206.67 U | 170.00 U | 180.00 U | 150.00 U |
| HEXACHLOROETHANE | UG/KG | 00 + 150 | 130.00 0 | 107.50 0 | 80.130 | 112.00 U | 82.25 U | 07.00 0 | 12.33 U | 97.33 U | 103.33 U | 103.75 U | 61.00 U | 113.33 U | 97.50 U | 79.50 U | 84.00 U | 69.50 U |
| M-DICHLOROBENZENE | UG/KG | /1 + 1/0 | NA | 126.25 U | 101.63 U | 128.00 0 | 99.25 U | 79.50 0 | 84.670 | 115.670 | 123.33 U | 120.00 U | 72.50 U | 136.67 U | 115.00 U | 93.67 U | 100.75 U | 82.00 U |
| METHANAMINE, N-METHYL-N-NITROSO | UG/KG | 56 - 140 | 57.00 U | 98.75 U | 80.25 U | 105.20 U | 77.00 U | 63.00 0 | 67.670 | 91.33 U | 100.00 U | 98.75 U | 57.00 U | 104.67 U | 90.00 U | 74.50 U | 78.75 U | 64.50 U |
| N-NITRUSODI-N-PROPYLAMINE | UG/KG | 84 - 200 | 87.00 U | 147.50 U | 120.75 U | 158.00 U | 117.00 U | 94.50 U | 100.67 U | 140.00 U | 146.67 U | 147.50 U | 86.50 U | 163.33 U | 135.00 U | 111.67 U | 117.50 U | 95.50 U |
| N-NITROSODIPHENYLAMINE | UG/KG | 56 - 140 | 57.00 U | 98.75 U | 80.25 U | 105.20 U | 17.00 U | 63.00 U | 67.67U | 91.33 U | 100.00 U | 98.75 U | 57.00 U | 104.67 U | 90.00 U | 74.50 U | 78.75 U | 64.50 U |
| NITROBENZENE | UG/KG | 71 - 170 | /3.00 U | 126.25 U | 101.625 U | 128.00 U | 99.25 U | 79.50 0 | 84.67 U | 115.67 U | 123.33 U | 120.00 U | 72.50 U | 136.67 U | 115.00 U | 93.67 U | 100.75 U | 82.00 U |
| PENTACHLOROPHENOL | UG/KG | 300 - 740 | 310.00 U | 542.50 U | 433.75 U | 564.00 U | 420.00 U | 340.00 U | 366.67 U | 500.00 U | 536.67 U | 530.00 U | 310.00 U | 576.67 U | 491.67 U | 406.67 U | 430.00 U | 355.00 U |
| PHENOL | UG/KG | 65 - 160 | 67.00 U | 111.00 | AC 2 85.13 | 118.00 U | 89.00 U | 125.00 | 77.33 U | 183.33 | 140.00 | 155.75 | 66.00 U | 122.33 U | 199.00 | 478.33 | 260.00 | 440.00 |

U = not detected in any sample within reach; value represents mean detection limit (DL).

NA = not analyzed

TABLE 6-6 MEAN CONCENTRATIONS OF CHLORINATED PESTICIDES (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | REACH 1D: | Ocean Reference | Outside Site 104 | inside Site 104 | Brewerton Eastern Extension | C&D Approach | C&D Approach | Craighill | Craighill Angle East | Craighili Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------|-------|----------------|--------------------|---------------------|--------------------|-----------------------------------|-----------------|-----------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | SI | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | SAM | IPLE SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | 1000 | | | | | | | | | | | | | |
| 4,4'-DDD | UG/KG | 0.42 - 1 | 0.42 U | 0.74 U | 0.59 U | 0.76 U | 0.57 U | 0.72 | 0.5 U | 0.68 U | 0.76 | 0.72 U | 0.42 U | 0.78 U | 0.67 U | 0.55 U | 0.86 | 0.51 |
| 4,4'-DDE | UG/KG | 0.4 - 0.95 | 0.4 U | 0.7 U | 0.98 | 0.72 U | 0.54 U | attai 0.99 | 0.47 U | 0.64 U | 0.69 U | 0.68 U | 0.4 U | 0.74 U | 0.63 U | 0.67 | 0.89 | 0.48 |
| 4,4'-DDT | UG/KG | 0.65 - 1.6 | 0.66 U | 1.17 U | 0.93 U | 1.18 U | 0.89 U | 0.72 U | 0.77 U | 1.06 U | 1.13 U | 1.1 U | 0.66 U | 1.22 U | 1.05 U | 0.87 U | 0.91 U | 0.75 U |
| ALDRIN | UG/KG | 0.52 - 1.2 | 0.52 U | 0.9 U | 0.73 | 0.94 U | 0.7 U | 0.57 U | 0.62 U | 0.84 U | 0.9 U | 0.89 U | 0.52 U | 0.95 U | 0.83 U | 0.68 U | 0.71 U | 0.59 U |
| ALPHA-BHC | UG/KG | 0.38 - 0.9 | 0.38 U | 0.67 U | 0.53 U | 0.69 U | 0.51 U | 0.42 U | 0.45 U | 0.62 U | 0.65 U | 0.65 U | 0.38 U | 0.71 U | 0.6 U | 0.5 U | 0.52 U | 0.43 U |
| BETA-BHC | UG/KG | 0.49 - 1.2 | 0.49 U | 0.86 U | 0.69 U | 0.89 U | 0.66 U | 0.54 U | 0.58 U | 0.79 U | 0.85 U | 0.84 U | 0.49 U | 0.92 U | 0.78 U | 0.65 U | 0.68 U | 0.56 U |
| CHLORDANE | UG/KG | 1.6 - 3.8 | 1.6 U | 2.82 U | 2.25 U | 2.94 U | 2.17 U | 1.75 U | 1.9 U | 2.6 U | 2.8 U | 2.77 U | 1.6 U | 3.0 U | 2.55 U | 2.1 U | 2.23 U | 1.85 U |
| CHLORBENSIDE | UG/KG | 3.3 - 7.9 | 3.3 U | 5.77 U | 4.67 U | 6.04 U | 4.5 U | 3.65 U | 3.93 U | 5.37 U | 5.8 U | 5.7 U | 3.33 U | 6.17 U | 5.28 U | 4.42 U | 4.6 U | 3.8 U |
| DACTHAL | UG/KG | 3.3 - 20 | 10.0 U | 17.5 U | 4.67 U | 6.04 U | 4.5 U | 3.65 U | 3.93 U | 5.37 U | 5.8 U | 5.7 U | 3.33 U | 6.17 U | 5.28 U | 4.42 U | 4.6 U | 3.8 U |
| DELTA-BHC | UG/KG | 0.49 - 1.2 | 0.49 U | 0.86 U | 0.69 U | 0.89 U | 0.66 U | 0.54 U | 0.58 U | 0.79 U | 0.85 U | 0.84 U | 0.49 U | 0.92 U | 0.78 U | 0.65 U | 0.68 U | 0.56 U |
| DIELDRIN | UG/KG | 0.43 - 1 | 0.43 U | 0.76 U | 0.6 U | 0.78 U | 0.58 U | 0.47 U | 0.51 U | 0.7 U | 0.75 U | 0.74 U | 0.43 U | 0.8 U | 0.68 U | 0.57 U | 0.59 U | 0.49 U |
| ENDOSULFAN 1 | UG/KG | 0.71 - 1.7 | 0.72 U | 1.26 U | 1.02 U | 1.28 U | 0.99 U | 0.79 U | 0.85 U | 1.16 U | 1.23 U | 1.2 U | 0.72 U | 1.37 U | 1.15 U | 0.95 U | 1.01 U | 0.82 U |
| ENDOSULFAN II | UG/KG | 0.36 - 0.86 | 0.36 U | 0.63 U | 0.60 | 0.65 U | 0.48 U | 0.4 U | 0.43 U | 0.58 U | 0.62 U | 0.62 U | 0.36 U | 0.66 U | 0.57 U | 0.47 U | 0.5 U | 0.41 U |
| ENDOSULFAN SULFATE | UG/KG | 0.83 - 2 | 0.85 U | 1.48 U | 1.18 U | 1.52 U | 1.11 U | 0.92 U | 0.99 U | 1.33 U | 1.47 U | 1.43 U | 0.84 U | 1.53 U | 1.33 U | 1.1 U | 1.15 U | 0.95 U |
| ENDRIN | UG/KG | 1.5 - 3.6 | 1.5 U | 2.65 U | 2.11 U | 2.74 U | 2.03 U | 1.65 U | 1.77 U | 2.43 U | 2.6 U | 2.58 U | 1.5 U | 2.8 U | 2.38 U | 1.98 U | 2.08 U | 1.7 U |
| ENDRIN ALDEHYDE | UG/KG | 0.93 - 2.2 | 0.95 U | 1.65 U | 1.39 | 1.68 U | 1.27 U | 1.05 U | 1.12 U | 1.53 U | 1.63 U | 1.6 U | 0.94 U | 1.73 U | 1.5 U | 1.23 U | 1.27 U | 1.05 U |
| GAMMA-BHC | UG/KG | 0.45 - 1.1 | 0.45 U | 0.79 U | 0.63 U | 0.82 U | 0.6 U | 0.5 U | 0.53 U | 0.73 U | 0.78 U | 0.77 U | 0.45 U | 0.84 U | 0.71 U | 0.59 U | 0.62 U | 0.52 U |
| HEPTACHLOR | UG/KG | 0.59 - 1.4 | 0.6 U | 1.07 U | 0.84 U | 1.08 U | 0.81 U | 0.66 U | 0.71 U | 0.97 U | 1.03 U | 1.01 U | 0.6 U | 1.13 U | 0.96 U | 0.79 U | 0.83 U | 0.69 U |
| HEPTACHLOR EPOXIDE | UG/KG | 0.8 + 1.9 | 0.82 U | 3.17 | 1.40 | 1.48 U | 1.11 U | 0.89 U | 0.97 U | 1.3 U | 1.43 U | 1.4 U | 0.81 U | 1.53 U | 1.28 U | 1.08 U | 1.12 U | 1.29 |
| METHOXYCHLOR | UG/KG | 2.6 - 6.2 | 2.6 U | 4.55 U | 3.64 U | 4.72 U | 3.5 U | 2.85 U | 3.1 U | 4.2 U | 4.47 U | 4.47 U | 2.6 U | 4.83 U | 4.I U | 3.42 U | 3.58 U | 3.0 U |
| MIREX | UG/KG | 3.3 - 7.9 | 3.3 U | 5.77 U | 4.66 U | 6.04 U | 4.5 U | 3.65 U | 3.93 U | 5.37 U | 5.8 U | 5.7 U | 3.33 U | 6.17 U | 5.28 U | 4.42 U | 4.6 U | 3.8 U |
| TOXAPHENE | UG/KG | 14 - 33 | 14.0 U | 24.5 U | 19.63 U | 25.4 U | 19.0 U | 15.5 U | 16.67 U | 22.67 U | 24.0 U | 24.0 U | 14.0 U | 26.0 U | 22.33 U | 18.5 U | 19.25 U | 16.0 U |

U = not detected in any sample within reach; value represents mean detection limit (DL).



TABLE 6-7 MEAN CONCENTRATIONS OF ORGANOPHOSPHORUS PESTICIDES (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | | | | | Brewerton | | | | | | | Cralghill | | | | | |
|----------------------------------|--------------|---------------------|----------------|-------------|-------------|-----------|----------|----------|-----------|------------|------------|-----------|-----------|---------|---------|------------|------------|---------------|
| | | | Ocean | Outside | Inside Site | Eastern | Can | Can | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolehester |
| | | REACH ID: | Reference | Site 104 | 104 | Extension | Approach | Approach | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | SAN | APLE SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | - 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | | | | | | | |
| AZINPHOS METHYL | UG/KG | 22 - 52 | 22.0 U | 38.5 U | 30.75 U | 40.0 U | 29.5 U | 24.0 U | 26.0 U | 35.67 U | 37.67 U | 37.75 U | 22.0 U | 40.67 U | 34.67 U | 29.0 U | 30.25 U | 25.0 U |
| DEMETON | UG/KG | 21 - 50 | 21.0 U | 37.0 U | 29.38 U | 38.0 U | 28.5 U | 23.0 U | 25.0 U | 34.0 U | 36.33 U | 36.0 U | 21.0 U | 39.0 U | 33.33 U | 27.83 U | 29.0 U | 24.0 U |
| ETHYL PARATHION | UG/KG | 33 - 78 | 33.0 U | 57.75 U | 46.38 U | 59.8 U | 44.25 U | 36.0 U | 39.33 U | 53.67 U | 57.0 U | 56.5 U | 33.25 U | 61.0 U | 52.33 U | 43.5 U | 45.25 U | 37.5 U |
| MALATHION | UG/KG | 16 - 37 | 16.0 U | 27.5 U | 22.0 U | 28.4 U | 21.25 U | 17.5 U | 18.67 U | 25.33 U | 27.0 U | 26.75 U | 16.0 U | 29.0 U | 24.83 U | 20.83 U | 21.75 U | 18.0 U |
| METHYL PARATHION | UG/KG | 17 - 40 | 17.0 U | 29.75 U | 23.88 U | 30.6 U | 22.75 U | 18.5 U | 20.33 U | 27.33 U | 29.0 U | 29.0 U | 17.0 U | 31.67 U | 26.83 U | 22.33 U | 23.5 U | 19.5 U |
| U = not detected in any sample w | ithin reach; | value represents me | an detection l | limit (DL). | | | | | | | | | | | | | | |

NOTE: Shaded and bolded values represent mean eoncentrations for analytes detected in at least one sample. Means calculated with ND=DL.

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TABLE 6-8 MEAN CONCENTRATIONS OF PCB AROCLORS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | REACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach | C&D Approach | Craighill | Craighili Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------------------|-----------------|--------------------|--------------------|---------------------|--------------------|-----------------------------------|-----------------|-----------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | SA | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Lore | Core | Core | Grab | Grab | Core | Core | Lore | Core |
| | SAM | PLE SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | | 4 | | 0 | 0 | 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | _ | | | | | | | | | | | | | | |
| AROCLOR 1016 | UG/KG | 5.4 - 13 | 5.4 U | 9.43 U | 7.58 U | 9.72 U | 7.25 U | 5.9 U | 6.4 U | 8.73 U | 9.33 U | 9.25 U | 5.4 U | 9.97 U | 8.57 U | 7.12 U | 7.45 U | 6.15 U |
| AROCLOR 1221 | UG/KG | 6.7 - 16 | 6.8 U | 11.98 U | 9.5 U | 12.6 U | 9.15 U | 7.45 U | 8.2 U | 11.07 U | 11.67 U | 12.0 U | 6.8 U | 12.67 U | 10.78 U | 8.97 U | 9.3 U | 7.75 U |
| AROCLOR 1232 | UG/KG | 12 - 29 | 12.0 U | 21.0 U | 17.0 U | 21.8 U | 16.5 U | 13.5 U | 14.33 U | 19.33 U | 20.67 U | 20.75 U | 12.0 U | 22.33 U | 19.0 U | 15.83 U | 16.75 U | 14.0 L |
| AROCLOR 1242 | UG/KG | 8.7 - 21 | 8.9 U | 15.75 U | 12.31 U | 15.8 U | 12.0 U | 9.7 U | 10.53 U | 14.33 U | 15.33 U | 15.0 U | 8.83 U | 16.33 U | 14.17 U | 11.83 U | 12.25 U | 10.2 L |
| AROCLOR 1248 | UG/KG | 2.6 - 6.2 | 2.6 U | 4.55 U | 6.79 | 4.72 U | 3.5 U | 2.85 U | 3.1 U | 4.2 U | 4.47 U | 4.47 U | 2.6 U | 4.83 U | 4.1 U | 3.42 U | 3.58 U | 3.0 L |
| AROCLOR 1254 | UG/KG | 7.7 - 19 | 7.9 U | 13.75 U | 228 11.43 | 13.8 U | 10.23 U | 8.55 U | 9.2 U | 12.67 U | 13.33 U | 13.0 U | 7.8 U | 14.33 U | 12.33 U | 10.12 U | 10.63 U | 8.9 L |
| AROCLOR 1260 | UG/KG | 4.9 - 12 | 4.9 U | 8.6 U | 6.9 U | 8.88 U | 6.6 U | 5.4 U | 5.8 U | 7.9 U | 8.5 U | 8.4 U | 4.9 U | 9.2 U | 7.78 U | 6.47 U | 6.75 U | 5.6 U |
| U = not detected in any sample | e within reach; | value represents m | ean detection I | imit (DL). | | | | | | | | | | | | | | |



TABLE 6-9 MEAN CONCENTRATIONS OF PCB CONGENERS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | REACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach | C&D Approach | Craighlil | Craighill Angle East | Craighill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Polnt | Tolchester North | Tolchester South | Tolchester Straightening |
|----------------------|-------|------------------|--------------------|---------------------|--|-----------------------------------|--|-----------------|-----------|-------------------------|-------------------------|-----------------------|--|-----------------|---------------|-------------------------|--|-----------------------------|
| | | SAMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | | SAMPLE SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | | | | | | | |
| BZ# 8* | UG/KG | 0.1 - 0.24 | A | 2.0.26 | 0.16 | 0.24 | 0.15 | 18 June 14 0.49 | 0.13 | 0.49 | 0.3 | 6.37 | 0.12 | 0.19 U | 0.26 | M 11 1 0.39 | 0.39 | 0.29 |
| BZ# 18* | UG/KG | 0.1 - 0.24 | 0.1 U | 0.18 U | 0.23 | Mar 0.23 | ······································ | -NO-2 0.33 | 0.12 U | F # \$2.0.29 | Pr | Croke 0.38 | 0.1 U | 0.19 U | 2 0.26 | 214 6 2 4 0.3 | | 0.26 |
| BZ# 28* | UG/KG | 0.03 - 0.13 | 0.03 U | 0.06 U | 35 9 0.18 | 1 2 Car 0.36 | 24.5.7.0.18 | 0.53 | 0.04 U | 0.35 | A | 0.31 | ang 10 me 20 0.08 | 0.15 | 0.13 | the former 0.32 | | 1 3 m to : 0.26 |
| BZ# 44* | UG/KG | 0.03 - 0.26 | 0.11 U | 0.2 | ******* 0.27 | 0.29 | Sot 1 0.21 | 0.38 | 0.13 U | 1 3 ment 0.29 | 0.19 U | | × # FE. to 0.12 | 0.2 U | 0.18 U | 0.26 | 0.28 | 2.2 |
| BZ# 49 | UG/KG | 0.17 - 0.4 | 0.17 U | 0.3 | **** * 0.37 | 14×× 10.39 | #### 0.27 | | 0.2 U | A | 0.29 U | 0.31 | * | 0.32 U | 0.27 U | 0.29 | | |
| BZ# 52* | UG/KG | 0.1 - 0.24 | 0.1 U | 清禄40.22 | Inchin 0.38 | × 0.51 | 教育部 0.27 | A | 0.12 U | .18 M. 18 | 0.18 U | 13. may 0.24 | 0.12 | 0.19 U | 0.16 U | - + + + + + + + + + 0.3 | | × **** # 0.16 |
| BZ# 66* | UG/KG | 0.05 - 0.13 | 0.05 U | 262 0.13 | 0.24 | 0.28 | 0.22 | 0.61 | 0.06 U | A | . 0.15 | 1011 | · ···································· | 0.23 | -1-0.1 | | 1.35 B.35 | 14. The officer 0.28 |
| BZ# 77• | UG/KG | 0.08 - 0.2 | 0.08 U | A.Q.19 | 1.45 | 0.2 | 0.55 | 1.65 | AN 0.13 | A 4. 0.15 | 0.14 U | | AND THE 0.39 | | 0.16 | 0.38 | 0.82 | 0.76 |
| BZ# 87 | UG/KG | 0.04 - 0.1 | 0.04 U | 0.07 U | ····· 0.20 | 0.12 | | 0.27 | 0.05 U | 0.06 U | 0.07 U | 0.13 | 0.09 | | 0.06 U | Set 1 0.1 | 21 1 met 0.11 | 1.5 |
| BZ# 101* | UG/KG | 0.05 - 0.14 | 0.05 U | 4 0.18 | 0.65 | 0.34 | mangles 0.3 | 3000 Fr 0.99 | 0.11 | | 0.16 | | T. C. W. W 0.27 | | 0.12 | 0.42 | 0.51 | 16 Date 0 0.54 |
| BZ# 105* | UG/KG | 0.18 - 0.43 | 0.18 U | 0.31 U | 0.29 | 0.33 U | | 1 | 0.21 U | 0.29 U | 0.31 U | 0.31 U | E | 0.33 U | 0.29 U | 0.24 U | 0.25 U | 0.21 U |
| BZ# 118* | UG/KG | 0.06 - 0.16 | 0.06 U | 0.16 | (1) 10.50 | 0.28 | 104510× 0.27 | HAN, 5 / 0.7 | 0.08 | Entert 0.15 | 0.14 | - # 0.36 | ***** 1 1 5 5 0.2 | 0.22 | - 0.11 | 0.22 | 6. CE | |
| BZ# 126* | UG/KG | 0.04 - 0.12 | 0.04 U | 1 0.24 | 1.02 | 0.14 | 3.04 M.O.13 | 0.39 | 0.05 U | .0.08 | 0.11 | 0.16 | 3-0.05 | W M 0.1 | | 4 O.16 | 0.2 | |
| BZ# 128* | UG/KG | 0.04 - 0.11 | 0.04 U | *** 0.09 | 0.29 | 0.09 | at 4 10 0.11 | 0.23 | 0.05 U | 0.07 U | 0.08 U | 0.17 | 0.07 | 0.09 U | 0.07 U | 0.09 | 25 0.11 | 0.08 |
| BZ# 138* | UG/KG | 0.04 - 0.11 | 0.04 U | te se 0.12 | 0.89 | | A# \$ \$ 0.27 | 0.94 | 0.06 | 0.16 | Att 18 0.09 | 0.95 | × | Mar 0.21 | . 0.11 | 22000 - 0.3 | A | 0.59 |
| BZ# 153* | UG/KG | 0.03 - 0.1 | 0.03 U | F | 1.38 | | A | 1.25 | 0.07 | See. F. 0.27 | 0.13 | WHE 1.39 | State 0.23 | 0.33 | 16 0.09 | 0.41 | 0.62 | 0.65 |
| BZ# 156 | UG/KG | 0.08 - 0.19 | 0.08 U | 0.14 U | 1 A | 0.15 U | | 0.28 | 0.09 U | 0.13 U | 0.14 U | 0.34 | 0.08 U | 0.15 U | 0.13 U | 0.12 | 0.13 | 0.16 |
| BZ# 169* | UG/KG | 0.09 - 0.23 | 0.09 U | 0.17 U | 0.13 U | 0.17 U | 0.13 U | 0.11 U | 0.11 U | 0.15 U | 0.17 U | 0.16 U | 0.09 U | 0.18 U | 0.15 U | 0.13 U | 0.13 U | 0.11 U |
| BZ# 170* | UG/KG | 0.07 - 0.17 | 0.07 U | 2 0. 0.14 | 14-55-0.41 | 0.16 | 34 Pat 0.16 | 0.42 | 0.08 U | 0.13 | 0.12 U | 0.61 | 0.08 | 0.13 U | 0.11 U | | 0.18 | 0.25 |
| BZ# 180* | UG/KG | 0.08 - 0.21 | 0.08 U | 0.18 | 1.16 | abitic 15 0.33 | C FM # 0.3 | 0.93 | 0.1 U | 0.19 | 0.17 | 1.37 | 19 | 0.19 | 0.14 U | 0.34 | 0.44 | 0.48 |
| BZ# 183 | UG/KG | 0.05 - 0.12 | 0.05 U | 0.09 U | ······································ | 0.09 U | 343003/9= 0.1 | 0.19 | 0.06 U | 0.08 U | 0.08 U | 399 Mile # 0.4 | 0.05 U | 0.09 U | 0.08 U | 0.06 U | 0.09 | |
| BZ# 184 | UG/KG | 0.05 - 0.13 | 0.05 U | 0.27 | 0.68 | 0.46 | 0.41 | ·林·金子·金子王人 | A me 0.07 | 0.28 | 0.16 | 1.98 | 0.17 | > 0.39 | 0.12 | | ······································ | |
| BZ# 187* | UG/KG | 0.05 - 0.14 | 0.06 U | Aprile \$ 0.17 | A 0.95 | 0.2 | En 18 0.2 | 1 0.5 | 0.07 U | 0.13 | ana 1 10.13 | 0.77 | 0.07 | 0.13 | aber 0.1 | 0.24 | August 1: 0.31 | La manda 4 0.29 |
| BZ# 195 | UG/KG | 0.08 - 0.21 | 0.08 U | 0.15 U | 0.17 | 0.16 U | 2 7 2 2 0.13 | St. Frank. 0.1 | 0.1 U | 0.14 U | 0.15 U | 0.23 | 0.08 U | 0.16 U | 0.14 U | 0.11 U | 0.12 U | 0.1 U |
| BZ# 206 | UG/KG | 0.13 - 0.31 | 0.13 U | 0.64 | 0.63 | + 12-2" #4 0.52 | 0.58 | ·泡水: 未来 2.1 | 0.15 U | 0.3 | 0.29 | 1.200 0.37 | 0.16 | 0.35 | 0.29 | 0.64 | 0.87 | 1.04 |
| BZ# 209 | UG/KG | 0.16 - 0.38 | 0.16 U | 13-48-12 | 0.95 | 0.89 | 0.85 | A. F. M. M. 3.8 | 0.21 | 0.53 | 0.41 | 0.65 | 0.28 | · | 0.47 | 1.04 | ······································ | 2. Historia -1.45 |
| TOTAL PCB (ND=0)* | UG/KG | • | 0.40 | 2.78 | 18.97 | 7.32 | 5.83 | 22.34 | 0.64 | 5.63 | 2.26 | 18.01 | 3.54 | 3.85 | 1.97 | 8.94 | 11.40 | 10.63 |
| TOTAL PCB (ND=1/2DL) | UG/KG | | 1.61 | 4.56 | 20.06 | 8.50 | 7.15 | 22.45 | 2.05 | 6.64 | 4.11 | 18.81 | 4.32 | 5.64 | 3.61 | 9.43 | 11.99 | 11.06 |
| TOTAL PCB (ND=DL) | UG/KG | | 2.82 | 6.35 | 21.15 | 9.69 | 8.48 | 22.55 | 3.45 | 7.65 | 5.97 | 19.61 | 5.09 | 7.44 | 5.24 | 9.91 | 12.57 | 11.49 |

U = not detected in any sample within reach; value represents mean detection limit (DL). P = >25% between two GC columns.

* = PCB congeners used for Total PCB summation, as per Table 9-3 of the 1TM (USEPA/USACE 1998). Total multiplied by a factor of 2 as per NOAA 1993

• Note that the mean of total PCB for individual samples is not equivalent to sum of mean congeners for ND=0 and ND=1/2DL.

TABLE 6-10 MEAN CONCENTRATIONS OF PAHs (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | REACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach | C&D Approach | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|-------|-----------------|--------------------|---------------------|--------------------|---------------------------------------|-----------------|-----------------|-----------------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | SAN | MPLE SIZE (N):[| 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | - 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | | | | | | | |
| ACENAPHTHENE | UG/KG | 9.5 - 23 | 9.7 U | 196.75 | 290.38 | 52.8 | 67.75 | 13.0 | 19.2 | MA X 4 38.0 | 28.0 | Sec. 64.5 | 24.75 | 41.0 | 48.67 | 114.67 | 82.25 | 265.0 |
| ACENAPHTHYLENE | UG/KG | 20 - 48 | 20.0 U | 51.75 | 28.88 | 36.0 U | 38.0 | 22.0 U | 23.67 U | 32.33 U | 34.33 U | 34.0 U | 20.0 U | 36.67 U | \$ 35.17 | 54.17 | 1 they = 41.5 | 2 34 220.0 |
| ANTHRACENE | UG/KG | 0.72 - 11 | 0.73 U | -sta 52.6 | 67.58 | 24.8 | PENA31.25 | 2.0 | 7.03 | 12.8 | 41.0 | 25.75 | 10.13 | 15.17 | 20.68 | 2 42.97 | 30.75 | 155.0 |
| BENZ[A]ANTHRACENE | UG/KG | 0.84 - 12 | 0.86 U | 98.3 | den 61.91 | 25.62 | 34.0 | 5.3 | ****** 8.6 | A Fair 14.0 | 12.8 | 1 23.48 | 14.25 | 22.6 | .15.25 | 36.67 | 26.27 | 9 |
| BENZO(A)PYRENE | UG/KG | 1.1 - 16 | 1.1 U | @117.43 | *** 80.13 | 41.94 | 56.5 | 1.0 | 15.83 | 25 017.33 | 17.57 | * 23.5 | 27.25 | 46.33 | 19.13 | 47.33 | 35.55 | 110.0 |
| BENZO(B)FLUORANTHENE | UG/KG | 1.8 - 4.3 | 1.8 U | Se 124.5 | 93.19 | 45.08 | 84.25 | ****** 8.9 | 19.1 | 31.0 | 19.43 | 42.25 | 26.5 | 36.0 | 45.2 | 110.67 | 127.00.90.3 | 230.0 |
| BENZO[G,H,I]PERYLENE | UG/KG | 2.1 - 5 | 2.1 U | 1.63 | St \$50.23 | 30.38 | 38.25 | 2400 | 12.17 | 13.8 | Anti-17.5 | 2 21.0 | 18.5 | 32.33 | 13.5 | 35.83 | .25.2 | 3575. (M. L. 1. 69.0 |
| BENZO[K]FLUORANTHENE | UG/KG | 2.1-5 | 2.1 U | faire 54.5 | 41.95 | 16.26 | 29.0 | a 19.00 2.5 | 6.7 | A 100 8.3 | 557.93 | Garage 14.5 | 10.18 | 16.8 | 10.52 | 25.83 | 18.02 | · · · · · · · · · 51.5 |
| CHRYSENE | UG/KG | 1.3 - 19 | 1.3 U | .ter 80.7 | Ac. 40.53 | 25.76 | 37.5 | 5.9 | 8.17 | Marry 201 . 12.1 | 10.57 | 20.68 | 12.32 | 19.67 | 14.37 | 36.17 | 24.52 | 71.5 |
| DIBENZ[A,H]ANTHRACENE | UG/KG | 3.4 - 8.1 | 3.4 U | 5.95 U | 9.63 | 19210 BAS 7.7 | 6.32 | 3.75 U | 4.03 U | 5.53 U | 5.9 U | 5.85 U | 3.55 | ···· 6.33 | 5.38 U | territoria to 5.8 | 5.2 | 7.55 |
| FLUORANTHENE | UG/KG | 2.6 - 38 | 2.6 U | 263.5 | 225.00 | A | 124.25 | Sec. 11.3 | 24.13 | 8.00 46.0 | 40.67 | \$7.0 | 35.5 | \$ 50.33 | 61.83 | 137.83 | 103.25 | 1 395.0 |
| FLUORENE | UG/KG | 1.3 - 3.1 | 1.3 U | # 39.38 | ₩. 4. 65.20 | · · · · · · · · · · · · · · · · · · · | 40.25 | 2.4.4.3.0 | 8.13 | apres 12.43 | 1.4.11.9 | 25.25 | State 12.1 | 15.27 | 21.37 | 46.17 | ··· · · · · 33.2 | ante 175.0 |
| INDENO[1,2,3-CD]PYRENE | UG/KG | 1.1 - 6.7 | 1.1 U | 55.8 | 30.56 | 23.06 | 27.25 | 3.55 | SHEEK 4.37 | Pater 10.87 | A | Sec. 15.32 | 11.8 | 19.97 | 10.0 | 25.67 | 18.8 | 47.5 |
| NAPHTHALENE | UG/KG | 7.3 - 18 | 7.5 U | Sec 153.0 | 237.63 | × 106.4 | 121.0 | 14.5 | 29.83 | | 70.0 | 94.75 | 39.5 | 76.0 | 99.17 | 188.17 | 134.5 | 625.0 |
| PHENANTHRENE | UG/KG | 1.3 - 19 | 1.3 U | 152.0 | 203.15 | 69.66 | 105.75 | 新进大学 6.95 | 18.8 | 33.67 | 30.67 | 1 64.25 | 28.75 | 38.33 | 54.82 | 120.0 | 85.13 | 1.445.0 |
| PYRENE | UG/KG | 0.62 - 9.3 | 0.63 U | 171.75 | 202.58 | 86.26 | 116.25 | 11.5 | Sprice 80% 24.1 | 38.33 | 36.33 | 69.25 | 38.5 | 55.67 | 48.87 | 116.83 | 85.75 | 335.0 |
| TOTAL PAH (ND=0) | UG/KG | | 0 | 1659.83 | 1699.90 | 666.56 | 951.08 | 100.10 | 202.60 | 354.63 | 321.17 | 591.48 | 311.03 | 487.67 | 487.38 | 1123.92 | 817.63 | 3294.55 |
| TOTAL PAH (ND=1/2DL) | UG/KG | | 28.76 | 1679.68 | 1714.19 | 690.25 | 954.33 | 112.98 | 218.23 | 373.57 | 341.95 | 611.40 | 322.30 | 508.07 | 505.65 | 1134.34 | 828.91 | 3294.55 |
| TOTAL PAH (ND=DL) | UG/KG | | 57.52 | 1699.53 | 1728.49 | 713.94 | 957.58 | 125.85 | 233.87 | 392.50 | 362.73 | 631.33 | 333.58 | 528.47 | 523.92 | 1144.77 | 840.20 | 3294.55 |

U = not detected in any sample within reach; value represents mean detection limit (DL).





TABLE 6-11 MEAN CONCENTRATIONS OF METALS (MG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | | Ocean | Outside | Inside | Brewerton Eastern | C&D | C&D | Castakill | Craighlil | Craighill | Craighiil | Craighill Upper | Cutoff | Swan | Tolchester | Tolchester | Tolehester |
|-----------|-------|---------------|--------------|------------|-----------|-----------------------|----------------|---------------------------------------|---------------------|--------------|----------------|---------------|--------------------|------------|--|-----------------|------------------|--|
| | | REACH ID: | Reference | Site 104 | Site 104 | Extension | Approach | Approach | Craignin | Angie Last | Angie west | Com | Conh | Angie | Core | Corr | South | Straightening |
| | S | AMPLE TYPE: | Grab | Grab | Grab | Urab | Grab | Core | Orau | Core | Core | Core | Grao | 0140 | Core | Core | Core | Core |
| | SAM | PLE SIZE (N): | | 4 | 8 | 2 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 0 | 0 | 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | _ | | | | | | | |
| ALUMINUM | MG/KG | 1.8 - 36 | 2150.0 | 15325.0 | 12801.25 | 16680.0 | 15700.0 | 15000.0 | 6173.67 | 13033.33 | ter€ 17300.0 | 15675.0 | 2625.0 | 17633.33 | 14816.67 | 14950.0 | 14950.0 | 14450.0 |
| ANTIMONY | MG/KG | 0.2 - 0.4 | 0.21 UN | at \$ 0.62 | 3.67 | 200 1 + Po 1 + 1 + 48 | Simmer the 1.1 | AN 11.15 | 0.6 | 5 1×1.23 | A. M. S. & 1:4 | ·****** 1.15 | | S \$5 1.73 | a a <1.09 | 255-55-1:12 | 10 mar 0.96 | 1.04 H & R. 1.04 |
| ARSENIC | MG/KG | 0.08 - 0.68 | 2.0 | Att 14.33 | 32.00 | 16.76 | 10.9 | 事例或第11.1 | 6.0 | state's 13.7 | 16.53 | | 1.95 | 17.57 | ·s. # 12.72 | 11.3 | Jon Mark : 12.57 | 14.55 |
| BERYLLIUM | MG/KG | 0.0097 - 0.01 | 0.1 B | 1.35 | 1.03 | 1.86 | 1.6 | · · · · · · · · · · · · · · · · · · · | 0.56 | à anti 1.43 | 55 1.57 | 145 d 1.45 | A. V 1 0.3 | 1.67 | A 1.55 | 1.68 | 1.63 | \$ 19 19 19 X 2.1 |
| CADMIUM | MG/KG | 0.01 - 0.04 | 0.02 U | 0.03 U | ANSE 1.33 | Att 1 0.26 | 6.45 2 0.33 | 0.77 | | Base: 0.26 | Mi 2 85 0.74 | Fast 14: 0,42 | 0.03 | A 0.27 | 0.34 | | ~ | S. 6 0 . 1. 0.34 |
| CHROMIUM | MG/KG | 0.08 - 0.09 | 174249E | 1 34.42 | 83.95 | 2第24 13 44.3 | 25.25 | 27.7 | 16.17 | waa/41.4 | 46.77 | 42.9 | 9.95 | 59.7 | 36.08 | 27.77 | 32.17 | 29.1 |
| COPPER | MG/KG | 0.06 - 0.7 | 1.3 3 745 14 | 38.55 | 99.50 | 240 43.18 | 35.15 | 37.35.8 | 14.03 | \$35.23 | 39.07 | ar 38.25 | 9.73 | 45.9 | 198 19: 39.5 | ANC 1. 3. 38.0 | 35.92 | 1 12 2 1 54.85 |
| IRON | MG/KG | 0.8 - 8i | \$ \$ 5640.0 | 37225.0 | 52450.00 | 40560.0 | 37100.0 | 35750.0 | 16820.0 | 36100.0 | 40400.0 | 38525.0 | P 1767.5 | 42033.33 | 35466.67 | 36716.67 | 36975.0 | 38200.0 |
| LEAD | MG/KG | 0.08 - 0.11 | 计学家法教育24 | 96-49.83 | 118.84 | 61.76 | 12 x 4 1 36.6 | 144 39.2 | AM 16.57 | 44.93 | 52.27 | 43.83 | 7 10.25 | 66.13 | 47.6 | RE- 10 42.43 | 40.95 | Energia 48.45 |
| MANGANESE | MG/KG | 0.24 - 5.2 | A | 1287.75 | 865.50 | 7214.0 | 3425.0 | The 3565.0 | \$ \$97.63 | 0.26 U | 5020.0 | 3280.0 | 527.0 | 5160.0 | 2313.33 | a. 3888.33 | 3975.0 | ······································ |
| MERCURY | MG/KG | 0.04 - 0.05 | 0.05 U | *14× 0.22 | 0.18 | A | 188400 0.19 | ##Xx 0.2 | 0.08 | 0.18 | 0.16 | Se \$2 0.16 | ine 11 0.06 | A | See. 1 0.2 | 0.18 | Maria 0.17 | 1 |
| NICKEL | MG/KG | 0.08 - 0.25 | 3.6 | At 18.35.1 | 26.A5 | | 19.02 | # # 36.15 | *******13. 8 | 38.57 | 12 Fr 42.07 | 38.1 | att \$ 8.07 | 46.67 | fan 41.82 | Arca 1 49.35 | S. 47.35 | 59.9 |
| SELENIUM | MG/KG | 0.13 - 0.72 | 0.13 U | 2.88 | 4.19 | Company 4.34 | 2.55 | | 0.88 | 2.9 | | A. 2.78 | 0.5 | 3.53 | 2.45 | 1 2.67 | 3.08 | Stating & support 3.15 |
| SILVER | MG/KG | 0.15 - 15.9 | 0.16 U | 0.76 U | 1 19 2.24 | 1 | 0,28 | 0.69 | 0.16 U | 0.16 U | 144 × 1 0.6 | AMA (# 0.17 | 0.16 U | 0.16 U | 0.18 | 135.64 5 0.58 | 0.38 | 0.16 (|
| THALLIUM | MG/KG | 0.09 - 0.37 | 0.12 U | 0.17 | 20.21 | a 10.13 | ALE 0.13 | 4.44 0.13 | 0.11 | MAR 0.12 | 6 0.12 | SC 20.13 | 0.1 U | 0.14 | Surgion 0.13 | statistic 0.12 | 0.15 | 1 |
| ZINC | MG/KG | 0.17 - 0.79 | Sec. 17.6 | 201.75 | 273.48 | att 1 304.6 | 197.25 | a 17.5 | 3*** 77.07 | 207.67 | 248.33 | Str. 208.5 | 44.73 | 294.33 | 207.5 | 207.33 | 209.75 | 238.0 |
| SEM/AVS | RATIO | | A | 0.07 | 0.42 | 1.09 | 1.73 | 0.54 | #**** 0.11 | A 11 | 0.27 | 0.09 | Star 54 0.94 | 0.05 | ······································ | -art mit - 0.18 | 0.62 | 1 |

TABLE 6-12 MEAN CONCENTRATIONS OF BUTYLTINS (UG/KG) IN SEDIMENT FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | | | | | Brewerton | | | | 1.000 | | | Cralghill | | | | | |
|----------------------------|-------------|---------------------|---------------|-----------------|---------------|---------------|---------------|----------|-----------|------------|------------|-----------|-----------|--------|--------|------------|------------|-------------------|
| | | | Ocean | Outside | Inside | Eastern | C&D | C&D | | Cralghlil | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| | | REACH ID: | Reference | Site 104 | Site 104 | Extension | Approach | Approach | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| | S | AMPLE TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | SAM | PLE SIZE (N): | 1 | 4 | 8 | 5 | 4 | 2 | 3 | 3 | 3 | 4 | 4 | 3 | 6 | 6 | 4 | 2 |
| ANALYTE | UNIT | DL (RANGE) | | | | | | | | | | | | | | | | |
| MONOBUTYLTIN | UG/KG | 1.2 - 4.8 | 7.9 U | 3.72 U | 2.59 U | 3.68 U | 3.08 U | 2.4 U | 1.97 U | 3.37 U | 3.53 U | 3.35 U | 1.45 U | 3.53 U | 3.33 U | 2.52 U | 2.95 U | 2.I U |
| DIBUTYLTIN | UG/KG | 1.6 - 6.2 | 7.1 U | 4.85 U | 3.34 U | 4.82 U | 4.05 U | 3.1 U | 2.53 U | ate 11.07 | 4.6 U | 4.4 U | 1.9 U | 4.63 U | 4.32 U | 3.28 U | 3.85 U | Shel conders 14.9 |
| TRIBUTYLTIN | UG/KG | 1.9 - 7.1 | 3.0 J | 5.65 U | 3.86 U | 5.6 U | 4.67 U | 3.55 U | 2.93 U | 5.07 U | 5.3 U | 5.05 U | 4.03 | 5.37 U | 4.98 U | 3.75 U | 4.42 U | 3.15 U |
| TETRABUTLYTIN | UG/KG | 2.1 - 8.1 | 5.6 U | 6.38 U | 4.38 U | 6.3 U | 5.3 U | 4.0 U | 3.33 U | 5.73 U | 6.03 U | 5.72 U | 2.5 U | 6.03 U | 5.65 U | 4.28 U | 5.03 U | 3.55 U |
| U = not detected in any sa | mple within | reach; value repres | ents mean det | ection limit (I | DL). J = valu | e below lowes | t calibrator. | | | | | | | | | | | |



TABLE 6-13 CONCENTRATIONS OF DIOXIN AND FURAN CONGENERS (NG/KG) IN SEDIMENT COMPOSITES^(*) FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND OCEAN REFERENCE SITE

| | | REA | CH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach | C&D Approach | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|---------------------------------------|---------------|--------------------|------------|--------------------|--|--------------------|-----------------------------------|------------------|-----------------|----------------|-------------------------|-------------------------|-----------------------|--------------------------|---------------|---------------|---------------------|---------------------|--|
| | | SAMPLE | TYPE: | Grab | Grab | Grab | Grab | Grab | Core | Grab | Core | Core | Core | Grab | Grab | Core | Core | Core | Core |
| | | SAMPLE SI | ZE (N): | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| ANALYTE | UNIT | DL (RANGE) | TEF | | | | | | | | | | | | | | | | |
| 2,3,7,8-TCDD | NG/KG | 0.09 - 1.97 | 1 | 0.1 U | 046 U | 0.31 EMPC | 1 27 U | 0.93 U | 0.31 U | 0 23 U | 0 89 U | 1.64 U | 197 U | 0.45 U | 1.07 U | 0.78 U | | 1.16 U | 0.36 U |
| 1,2,3,7,8-PECDD | NG/KG | 0.08 - 0.66 | 0.5 | 0.06 U | 043 U | 26 Tub. 2 0.16 J | 0 54 U | 0.35 U | 0.17 U | 0.12 EMPC | 8 0.86 J | 0 43 U | 199.00 0.94 J | 0.33 U | 1.07 EMPC | 0.24 U | | 0.29 U | 0.78 J |
| 1,2,3,4,7,8-HXCDD | NG/KG | 0.1 - 1.73 | 0.1 | 0120 | 0 33 U | . 16 EMPC | 1 18 U | 1.05 U | 0.42 U | 017 U | 1.27 U | 1.64 U | 160 | 0.54 U | 1.73 U | 1.1 U | 0.89 U | 1.03 U | 1441 Pre* 0.69 J |
| 1,2,3,6,7,8-HXCDD | NG/KG | 0.1 - 1.82 | 0.1 | 0.13 U | 0 34 U | mingran 0.52 J | 1.24 U | 1.11 U | 0.44 U | Steplant 0.5 J | 1.59 J | 1.72 U | 1.99 J | 0.57 U | 1.82 U | 1.15 U | 0.94 U | 1.09 U | 10 11 ANY 0.93 J |
| 1,2,3,7,8,9-HXCDD | NG/KG | 0.09 - 1.64 | 0.1 | 0.11 U | 031 U | 0.51 EMPC | 1.12 U | 0 99 U | 0.4 U | 100 Mer 0.32 J | ·*··· 1.85 J | 1.55 U | 200 2.25 J | 0 51 U | 164 U | 1.04 U | 0 84 U | 0 97 U | 1 10 113 8.81 EMPC |
| 1,2,3,4,6,7,8-HPCDD | NG/KG | 011-083 | 0.01 | -1 6.83 J | 6.95 | 12 14 20 - 7.15 | 6.77 J | 6.86 J | 3.19 J | 3. CAME 7.72 | | 13 | Be we at 33.1 | a mount 201 7.49 | A 122 | Aug 10 8.21 | 8.61 | 1- 3 44 4.78 J | na lafer an at 5.8 |
| OCDD | NG/KG | 0.23 - 6.4 | 0.001 | ******* 10.6 | 194 Mar 194 | West 4. 7 4 153 | 155 | Mary 168 | ## \$\$ 85.1 | | ALL 14: 574 | A 359 | SA. CAT 1040 | 174 45 Ether 174 | 279 | 8275 1227 | A# 1.2 . 2 167 | 143 | At 21 4 22 200 112 |
| 2,3,7,8-TCDF | NG/KG | 0.07 - 0.95 | 0.1 | 0.I U | 049U | .45 J | 0 65 U | 0.56 U | 0.29 U | | 0.9 U | 0.74 U | * 8.75 EMPC | 0.51 U | 075 U | 0 42 U | | 0.61 U | |
| 1,2,3,7,8-PECDF | NG/KG | 0.06 - 0.61 | 0.05 | 0 04 U | 0.31 U | Current 200, 0.1 J | 0.4 U | 0.26 U | 0.16 U | anie + 1.51 J | 41.18 EMPC | 0.41 U | 0 61 U | 0.36 U | 1.24 J | 0 19 U | 1.13 J | 0.21 U | |
| 2,3,4,7,8-PECDF | NG/KG | 0.05 - 0.59 | 0.5 | An 24 8.07.J | 0.3 U | 1.77 J | 0 39 U | 0.25 U | 0.16 U | | at 1.19 J | 0.4 U | 1.11 J | 0.35 U | * * 1.02 J | 1.56 J | 1.06 J | 0.21 U | ······································ |
| 1,2,3,4,7,8-HXCDF | NG/KG | 0.09 - 0 67 | 0.1 | 0.08 U | 0 32 U | | 87 3 8 8 8 AS J | 0.44 U | 0.27 U | # altre | 1.11 J | 0.52 U | 8.99 J | 8.33 EMPC | 1.09 J | 0.65 J | . SEMPC | 0.37 U | 3.18 EMPC |
| 1,2,3,6,7,8-HXCDF | NG/KG | 0 09 - 0 64 | 0.1 | 0.08 U | 0.31 U | 5 244 × 0.37 J | 4.37 EMPC | 0 42 U | 0 25 U | Waters 0.68 1 | 201 1 0.84 J | 0.5 U | 8.63 EMPC | 0.29 U | 0 64 U | 0.35 J | 0.73 EMIPC | 0.35 U | 2.63 EMPC |
| 2,3,4,6,7,8-HXCDF | NG/KG | 0.1 - 0.71 | 0.1 | 0.09 U | 0.34 U | 0.51 J | 0.44 U | 0 47 U | 0.28 U | # 197 1.15J | | 0.55 U | 0 67 U | 0.33 U | 0.71 U | 0.43 U | 0.46 U | 0.39 U | 2.78 J |
| 1,2,3,7,8,9-HXCDF | NG/KG | 0.11 - 0.78 | 0.1 | 010 | 0 38 U | 545 M (0.13 J | 0 48 U | 0 52 U | 0.31 U | (0.19 EMPC | 0 59 U | 0610 | 0.74 U | 0 36 U | 0.78 U | 047U | 0 51 U | 0.43 U | 0 55 U |
| 1,2,3,4,6,7,8-HPCDF | NG/KG | 0 07 - 0.58 | 0.01 | 4.17 B | 8.95 J | 1.94 J | - 14 181,82 J | K. 1.12 J | 6.59 J | 2.3 1 | 2.85 J | See. 2 | 3.61 J | 1.02 J | 1.73 J | 1.67 J | 1.01 J | 0.64 J | 9.02 |
| 1,2,3,4,7,8,9-HPCDF | NG/KG | 0.09 - 0.71 | 0.01 | 0.08 U | 0.42 U | Hand sout 0.24 J | 0 56 U | 0 58 U | 0210 | | 0 66 U | 0 65 U | 0.71 U | 0.5 U | 0.69 U | 0.3 U | 0 52 U | 0 31 U | . 0.61 EMPC |
| OCDF | NG/KG | 0.22 - 1.42 | 0 001 | 16 J | 188 ## 1.9 B | 3.35 J | 3.35 J | 2.56 J | - 1.14 B | 5.15 J | 7.51 J | 14 - mar 6.18 J | 8.76 J | 1. The you 3.23 J | | × × 3.4 J | 2.36 J | Sta 8 4 4 4 1.7 J | See 3.13 J |
| DIOXINS TEQ (ND=0) | NG/KG | - | 0 | 20000 0.85 | 2 | - and 1.50 | 131 A 10 4 10 128 | 1 anteria 0.25 | A.12 | | Agap tor to 2.49 | | 2.96 | ener . 0.26 | a warmen 1.10 | THE N 1.21 | 2.39 | . 8.20 | a 2 40 11 3.60 |
| DIOXINS TEQ (ND=1/2DL) | NG/KG | • | 0 | | ************************************** | 1,50 | 20 million #21.42 | Valle × 1.15 | 0.50 | 1001-116.53 | 44 SAS - 3.08 | 1.95 | 1 .12 A 4.12 | 1 1 1 2 Lapa . 0.82 | 2.05 | 1.90 | a . A | 1.17 | 3.81 |
| U = not detected. J = estimated value | ue; value les | s than lower metho | od calibra | tion limit. EMP | C = estimated | maximum possible o | concentration. B= | detected in labo | oratory blank | | | | | | | | | | |

D - Not detected. • command a ball of the new set of the set of th

(a) = one composite sample from each sampling reach submitted for dioxin/furan testing.

DL = detection limit.

TEF = Toxicity Equivalency Factor. TEQ = Toxicity Equivalency Quotient.

IEQ = foxicity Equivalency Quotient

NOTE: Shaded and bolded values represent detected concentrations

| | | | | | | Brewerton | C&D | C&D | | | | | Cralghill | | | | | T T |
|-----------------------------|-------|---------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|---------|------------|------------|---------------|
| | | TEL | Ocean | Inside | Outside | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | VALUE | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| ARSENIC | MG/KG | 7.24 | | 32.0 | 14.33 | 16.76 | 10.9 | 11.1 | | 13.7 | 16.53 | 14.53 | | 17.57 | 12.72 | 11.3 | 12.58 | 14.55 |
| CADMIUM | MG/KG | 0.676 | | 1.33 | | | | 0.77 | | | 0.74 | | | | | 0.77 | | |
| CHROMIUM | MG/KG | 52.3 | | 83.95 | | | | | | | | | | 59.7 | | | | |
| COPPER | MG/KG | 18.7 | | 99.5 | 38.55 | 43.18 | 35.15 | 35.8 | | 35.23 | 39.1 | 38.25 | | 45.9 | 39.5 | 38.0 | 35.93 | 54.85 |
| LEAD | MG/KG | 30.24 | | 118.84 | 49.83 | 61.76 | 36.6 | 39.2 | | 44.93 | 52.27 | 43.83 | | 66.13 | 47.6 | 42.43 | 40.95 | 48.45 |
| MERCURY | MG/KG | 0.13 | | 0.18 | 0.22 | 0.28 | 0.19 | 0.2 | | 0.18 | 0.16 | 0.16 | | 0.23 | 0.21 | 0.18 | 0.17 | 0.48 |
| NICKEL | MG/KG | 15.9 | | 26.45 | 35.1 | 54.48 | 49.03 | 56.15 | | 38.57 | 42.1 | 38.1 | | 46.67 | 41.82 | 49.35 | 47.35 | 59.9 |
| SILVER | MG/KG | 0.73 | | 2.24 | 0.76 | | | | | | | | | | | | | |
| ZINC | MG/KG | 124 | | 273.48 | 201.75 | 304.6 | 197.25 | 217.5 | | 207.67 | 248.33 | 208.5 | | 294.33 | 207.5 | 207.33 | 209.75 | 238.0 |
| ACENAPHTHENE | UG/KG | 6.71 | | 290.38 | 196.75 | 52.8 | 67.75 | 13.0 | 19.2 | 38.0 | 28.0 | 64.5 | 24.75 | 41.0 | 48.67 | 114.67 | 82.25 | 265.0 |
| ACENAPHTHYLENE | UG/KG | 5.87 | | 28.88 | 51.75 | 36.0 | 38.0 | 22.0 | 23.67 | 32.33 | 34.33 | 34.0 | 20.0 | 36.67 | 35.17 | 54.17 | 41.5 | 220.0 |
| ANTHRACENE | UG/KG | 46.85 | | 67.58 | 52.6 | | | | | | 1 | | | | | | | 155.0 |
| BENZ[A]ANTHRACENE | UG/KG | 74.83 | | | 98.3 | | | | | 1 | | | | | | | | 90.5 |
| BENZO[A]PYRENE | UG/KG | 88.81 | | | 117.43 | | | | | | | | | | | | | 110.0 |
| CHRYSENE | UG/KG | 107.77 | 1 | | | | | | | | Ι | | | | | 1 | | |
| DIBENZ[A,H]ANTHRACENE | UG/KG | 6.22 | | 9.63 | | 7.7 | 6.33 | | | | | | | | | 1 | | 7.5 |
| FLUORANTHENE | UG/KG | 112.82 | | 225.0 | 263.5 | | 124.25 | | | | 1 | 1 | | | | 137.83 | 1 | 395.0 |
| FLUORENE | UG/KG | 21.17 | | 65.2 | 39.4 | 27.82 | 40.25 | | | | | 25.25 | <u> </u> | | 21.37 | 46.17 | 33.2 | 175.0 |
| NAPHTHALENE | UG/KG | 34.57 | | 237.63 | 153.0 | 106.4 | 121.0 | | | 66.0 | 70.0 | 94.75 | 39.5 | 76.0 | 99.17 | 188.17 | 134.5 | 625.0 |
| PHENANTHRENE | UG/KG | 86.68 | Ĩ | 203.15 | 152.0 | | 105.75 | | | | 1 | 1 | | | | 120.0 | 1 | 445.0 |
| PYRENE | UG/KG | 152.66 | | 202.58 | 171.75 | | | | | | | | | | | | | 335.0 |
| TOTAL PAH (ND=0) | UG/KG | 1684.06 | | 1699.90 | | I | | | | | 1 | | | | | | | 3294.5 |
| TOTAL PAH (ND=1/2) | UG/KG | 1684.06 | | 1714.19 | | | | | | | | | | | | | | 3294.5 |
| TOTAL PAH (ND=DL) | UG/KG | 1684.06 | | 1728.49 | 1699.53 | | | | | | | 1 | | | | | | 3294.5 |
| 4,4'-DDD | UG/KG | 1.22 | I | İ. | Ī | T | [| | | | 1 | l l | 1 | | | 1 | T | T T |
| 4.4'-DDE | UG/KG | 2.07 | | | | 1 | | | | 1 | | 1 | 1 | | | | t | |
| 4,4'-DDT | UG/KG | 1.19 | | | | 1 | | 1 | | 1 | | | · · · · · | 1.22U | | | ····· | 1 |
| CHLORDANE | UG/KG | 2.26 | 1 | | 2.83U | 2.94U | 1 | | | 2.6U | 2.8U | 2.78U | | 3.0U | 2.55U | 1 | | |
| DIELDRIN | UG/KG | 0.715 | | | 0.76U | 0.78U | 1 | | | 1 | 0.75U | 0.74U | | 0.80U | | | | i |
| GAMMA-BHC | UG/KG | 0.32 | 0.45U | 0.63U | 0.79U | 0.82U | 0.611 | 0.50U | 0.53U | 0.73U | 0.78U | 0.77U | 0.45U | 0.84U | 0.710 | 0.59U | 0.62U | 0.521 |
| TOTAL PCB (ND=0) | UG/KG | 21.55 | | 1 | | (| r | 22.34 | | | | T | 1 | | i | 1 | | 1 |
| TOTAL PCB (ND=1/2DL) | UG/KG | 21.55 | | | | 1 | | 22.44 | | | | | 1 | | | 1 | | 1 |
| TOTAL PCB (ND=DL) | UG/KG | 21.55 | | 1 | | | | 22.55 | | | | | | | | 1 | | |
| 2-METHYLNAPHTHALENE | UG/KG | 20.21 | | 346.38 | 178.25 | 61.2 | 86.25 | | | 27.33 | 29.0 | 59.5 | 23.25 | 32.0 | 54.67 | 135.67 | 84.5 | 490.0 |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/KG | 182.16 | | 1 | T | 1 | | | | | 1 | <u> </u> | 1 | | 1 | | 1 | 1 |

TABLE 6-14 MEAN CONCENTRATIONS OF ANALYTES EXCEEDING TELs (1999)

U = not detected; mean detection limit exceed TEL value

.





TABLE 6-15 MEAN CONCENTRATIONS OF ANALYTES EXCEEDING PELs (1999)

| | | | | | | Brewerton | C&D | C&D | - | Craighill | Craighill | | Cralghill | | | | | [|
|--------------------------|-------|----------|-----------|----------|----------|-----------|-------------|----------|-----------|-----------|-----------|-----------|-----------|--------|-------|------------|--|---------------|
| | | PEL | Ocean | Inside | Outside | Eastern | Approach | Approach | | Angle | Angle | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | VALUE | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Cralghill | East | West | Entrance | Range | Angle | Point | North | South | Stralghtening |
| ARSENIC | MG/KG | 41.6 | | | | | | | | 1 | | | | | | | | |
| CADMIUM | MG/KG | 4.21 | | | | | | | | | | | | | | | | |
| CHROMIUM | MG/KG | 160.4 | | | | | - | | | | | | | | | | | |
| COPPER | MG/KG | 108.2 | | | | | | | | | | | | | | | | |
| LEAD | MG/KG | 112.18 | | 118.84 | | | | | | | | | | | | | · · · · | |
| MERCURY | MG/KG | 0.696 | | | | | | | | | | | | | | | | |
| NICKEL | MG/KG | 42.8 | | | | 54.48 | 49.03 | 56.15 | | | | | | 46.67 | | 49.35 | 47.35 | 59.95 |
| SILVER | MG/KG | 1.7 | | 2.24 | | | | | | | | | | | | | | |
| ZINC | MG/KG | 271 | | 273.48 | | 304.6 | | | | | | | | 294.33 | | | ····· | |
| ACENAPHTHENE | UG/KG | 88.9 | | 290.38 | 196.75 | | | | | 1 | | | | | | 114.67 | | 265.0 |
| ACENAPHTHYLENE | UG/KG | 127.87 | | | | | | | | | | | | | | | | 220.0 |
| ANTHRACENE | UG/KG | 245 | | | | | | | | | | | | | | | | |
| BENZ[A]ANTHRACENE | UG/KG | 692.53 | | | | | | | | | | | | | | | | |
| BENZO[A]PYRENE | UG/KG | 763.22 | | | | | | | | | | | | | | | | |
| CHRYSENE | UG/KG | 845.98 | | | | | | | | | | | | | | | | |
| DIBENZ[A,H]ANTHRACENE | UG/KG | 134.61 | | | | | | | | | | | | | | | | |
| FLUORANTHENE | UG/KG | 1493.54 | | | | | | | | | | | | | | | | |
| FLUORENE | UG/KG | 144.36 | | | | | | | | | | | | | | | | 175.0 |
| NAPHTHALENE | UG/KG | 390.64 | | | | | | | | | | | | | | | | 625.0 |
| PHENANTHRENE | UG/KG | 543.53 | | | | | | | | | | | | | | | | |
| PYRENE | UG/KG | 1397.6 | | | | | | | | | | | | | | | | |
| TOTAL PAH (ND=0) | UG/KG | 16770.4 | | | | | | | | | | | | | | [| | |
| TOTAL PAH (ND=1/2DL) | UG/KG | 16770.4 | | | | | | | | | | | | | | | · · · · | |
| TOTAL PAH (ND=DL) | UG/KG | 16770.4 | | | | | | | | | | | | | | | | |
| 4,4'-DDD | UG/KG | 7.81 | | | | | | | | | | | | | | [| | |
| 4,4'-DDE | UG/KG | 374.17 | | | | | | | | | | | | | | | | |
| 4,4'-DDT | UG/KG | 4.77 | | | | | | | | | | | | | | | | |
| CHLORDANE | UG/KG | 4.79 | | | | | | | | | 1 | | | _ | | | | |
| DIELDRIN | UG/KG | 4.3 | | | | | | | | - | | | | | | | <u> </u> | |
| GAMMA-BHC | UG/KG | 0.99 | | | | | | | | | | | | | | | t | |
| TOTAL PCB (ND=0) | UG/KG | 188.79 | | | | | | | | [| | | | | | | | |
| TOTAL PCB (ND=1/2DL) | UG/KG | 188.79 | | | | | | | | | | | | | | | <u> </u> | |
| TOTAL PCB (ND=DL) | UG/KG | 188.79 | | | | | | | | | | | | | | | l | |
| 2-METHYLNAPHTHALENE | UG/KG | 201.28 | | 346.38 | | | | | | [| | | | | | 1 | i | 490.0 |
| BIS(2-ETHYLHEXYL) PHTHAL | UG/KG | 2646.551 | | | | | | | | | | | | | | | | 4,50.0 |

TABLE 6-16 FREQUENCY OF DETECTION BY ANALYTICAL FRACTION FOR EACH APPROACH CHANNEL

| ANALYTICAL FRACTION | Brewerton Eastern Extension | C&D Approach (grabs) | C&D Approach (cores) | Craighiii | Craighlii Angie East | Craighiii Angie West | Craighili Entrance | Craighiii Upper Range | Cutoff Angle | Swan Point | Toichester North | Toichester South | Toichester Staightening |
|-----------------------------|-----------------------------------|----------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|------------|---------------------|---------------------|----------------------------|
| VOCs | 5/140 | 4/140 | 0/70 | 4/105 | 2/105 | 0/105 | 0/140 | 2/105 | 2/105 | 2/210 | 0/210 | 1/140 | 1/70 |
| SVOCs | 17/135 | 13/188 | 8/94 | 4/141 | 13/141 | 13/141 | 18/188 | 6/188 | 6/141 | 20/282 | 23/282 | 14/188 | 9/94 |
| Pesticides | 0/110 | 0/88 | 4/44 | 0/66 | 0/66 | 1/66 | 0/88 | 0/88 | 0/66 | 0/132 | 2/132 | 6/88 | 3/44 |
| Organophosphorus Pesticides | 0/25 | 0/20 | 0/10 | 0/15 | 0/15 | 0/15 | 0/20 | 0/20 | 0/15 | 0/30 | 0/30 | 0/20 | 0/10 |
| PCB Aroclors | 0/35 | 0/28 | 0/14 | 0/21 | 0/21 | 0/21 | 0/28 | 0/28 | 0/21 | 0/42 | 0/42 | 0/28 | 0/14 |
| PCB Congeners | 71/130 | 38/104 | 50/52 | 10/78 | 42/78 | 17/78 | 72/104 | 48/104 | 30/78 | 38/156 | 119/156 | 79/104 | 40/52 |
| PAHs | 69/80 | 63/64 | 28/32 | 40/48 | 42/48 | 40/48 | 56/64 | 57/64 | 43/48 | 81/96 | 88/96 | 57/64 | 32/32 |
| Metals | 79/80 | 63/64 | 31/32 | 38/48 | 41/48 | 46/48 | 61/64 | 50/64 | 44/48 | 93/96 | 92/96 | 63/64 | 30/32 |
| Butyltins | 0/20 | 0/16 | 0/8 | 0/12 | 1/12 | 0/12 | 0/16 | 2/16 | 0/12 | 0/24 | 0/24 | 0/16 | 1/8 |
| Dioxin and Furan Congeners | 6/17 | 4/17 | 4/17 | 15/17 | 12/17 | 4/17 | 11/17 | 5/17 | 8/17 | 7/17 | 11/17 | 4/17 | 15/17 |



TABLE 6-17 FREQUENCY OF DETECTION ^(a) BY ANALYTICAL FRACTION FOR SEDIMENTS FROM BALTIMORE HARBOR APPROACH CHANNELS, INSIDE SITE 104, OUTSIDE SITE 104, AND THE OCEAN REFERENCE SITE

| | FR | EQUENCY | OF DETEC | CT] | | PERCENT | DETECT | |
|-----------------------------|-------------------------------------|--------------------|---------------------|--------------------|-------------------------------------|--------------------|---------------------|--------------------|
| ANALYTICAL FRACTION | Approach Channels ^(b) | Inside Site 104 | Outside Site 104 | Ocean Reference | Approach Channels ^(b) | Inside Site 104 | Outside Site 104 | Ocean Reference |
| VOCs | 23/1540 | 8/175 | 5/140 | 0/35 | 1.5 | 4.6 | 3.6 | 0.0 |
| SVOCs | 164/2350 | 23/376 | 18/188 | 0/47 | 7.0 | 6.1 | 9.6 | 0.0 |
| Chlorinated Pesticides | 16/1100 | 8/176 | 3/88 | 0/22 | 1.5 | 4.5 | 3.4 | 0.0 |
| Organophosphorus Pesticides | 0/250 | 0/40 | 0/20 | 0/5 | 0.0 | 0.0 | 0.0 | 0.0 |
| PCB Aroclors | 0/350 | 2/56 | 0/28 | 0/7 | 0.0 | 3.6 | 0.0 | 0.0 |
| PCB Congeners | 654/1350 | 110/208 | 33/104 | 1/26 | 48.4 | 53.0 | 31.7 | 3.8 |
| PAHs | 696/800 | 116/128 | 55/64 | 0/16 | 87.0 | 90.6 | 85.9 | 0.0 |
| Metals | 731/800 | 118/128 | 58/64 | 10/16 | 91.4 | 92.2 | 90.6 | 62.5 |
| Butyltins | 4/200 | 0/32 | 0/16 | 1/4 | 2.0 | 0.0 | 0.0 | 25.0 |
| Dioxin and Furan Congeners | 106/221 | 17/17 | 4/17 | 5/17 | 48.0 | 100.0 | 23.5 | 29.4 |

(a) = total number of detected analytes / total number of analytical tests.

(b) = combined total for all approach channels.

| | | Brewerton | C&D | C&D | | | C | 0.1.1.10 | C | | | Telebook | Telebook | Tables |
|--------------------|----------------------|-----------------|--------------------|----------|------------|-------------------------|------------------|----------|-------------|--------------|------------|----------|----------|---------------|
| | Institute Cities 104 | Eastern | Approach (Crah) | Approach | Cratabili | Craighill Angle Fact | West | Entrance | Upper Raper | Culoff Angle | Swan Point | North | South | Straightening |
| Analyte (UG/KG) | Inside Site 104 | LALENSION 10 SI | (((110)) | 8.4 | 3.0 | 102 | 95 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| Mean % TOC | 1.2 | 10.5 | 9.5 | 0.4 | J.V NID | ND | an work mails of | ND | ND | ND | ND | ND | | 10.4 miles |
| 4,4'-DDD | ND | ND | ND | 50.7 | ND | ND | ND | ND | ND | ND | ND | 0.6 | 0.9 | 0.3 |
| 4,4'-DDE | 1.1 | ND | ND | 0.9 | ND | ND | ND | ND | ND | ND | ND | 0.0 | 0.0 | 0.3 |
| 4,4'-DDT | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ALDRIN | 0.8 | ND | ND | ND | ND | ND | ND | | ND | ND | ND | ND | ND | ND |
| ALPHA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BETA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DELTA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN I | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN II | 0.7 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN SULFATE | ND | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | 1.5 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| GAMMA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 1.6 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.8 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AZINPHOS METHYL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DEMETON | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ETHYL PARATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MALATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHYL PARATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | · ND |
| AROCLOR 1248 | 7.5 | ND | ND | ND | NE | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | 12.7 | ND | ND | ND | ND | NE | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | NE | NE | ND | ND | ND | ND | ND | ND | ND | ND |

TABLE 6-18A THEORETICAL BIOACCUMULATION POTENTIAL* FOR CHLORINATED PESTICIDES, ORGANOPHOSPHORUS PESTICIDES, AND PCB AROCLORS IN TISSUE: COMPARISON TO INSIDE SITE 104

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Tables 6-6 (Chlorinated Pesticides), 6-7 (Organophosphorus Pesticides), and 6-8 (PCB Aroclors). For non-detected analytes, the detection limit was used in the mean calculation.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Inside Site 104 value.









TABLE 6-18B THEORETICAL BIOACCUMULATION POTENTIAL* FOR CHLORINATED PESTICIDES, ORGANOPHOSPHORUS PESTICIDES, AND PCB AROCLORS IN TISSUE: COMPARISON TO OUTSIDE SITE 104

| | | Brewerton | C&D | C&D | | 0 | | | | | | | | |
|--------------------|--------------|----------------------|--------------------|--------|-----------|------|------|----------|-------------|--------------|------------|-------|---------------------|-----------------------------|
| Analyte (UG/KG) | Outside Site | Eastern Extension | Approach (Grab) | (Core) | Craighill | East | West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | Tolchester South | Tolchester Straightening |
| Mean % TOC | 11.4 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| 4,4'-DDD | ND | ND | ND | 0.7 | ND | ND | 0.6 | ND | ND | ND | ND | ND | .0.8 | 10.3 A 10.3 |
| 4,4'-DDE | ND | ND | ND | 0.9 | ND | ND | ND | ND | ND | ND | ND | 0.6 | 0.8 | 0.3 |
| 4.4'-DDT | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ALDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ALPHA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BETA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DELTA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN I | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN II | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN SULFATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| GAMMA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 2.2 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.8 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AZINPHOS METHYL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DEMETON | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ETHYL PARATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MALATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHYL PARATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | · ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Tables 6-6 (Chlorinated Pesticides), 6-7 (Organophosphorus Pesticides), and 6-8 (PCB Aroclors). For non-detected analytes, the detection limit was used in the mean calculation.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Outside Site 104 value.

| | | Brewerton | C&D | C&D | | | 0.1100.4 | | 0.111 | | | | 1000 C | |
|----------------------|-----|-----------|--------------------|-----------|-----------|-----------------|-------------------------|----------|---------------------------|--------------|------------|-------|-----------|------------------------------|
| A malk the (UIC/V/C) | 0 | Eastern | Approach (Grab) | Approach | Craiabili | Craighill Angle | Craighill Angle West | Entrance | Craighill Linner Range | Cutoff Anele | Swan Point | North | South | 1 olchester Straightening |
| Mean % TOC | 0.5 | 10.5 | 95 | 84 | 30 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| | ND | ND | ND | 0.4 | ND | ND | 1 30 20 20 20 6 | ND | ND | ND | ND | ND | Same 10.8 | 0.3 |
| 4,4 - DDD | ND | ND | ND | 1. 28 0.9 | ND | ND | ND | ND | ND | ND | ND | 0.6 | 8.0 4 1.2 | A |
| 4,4 DDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ALDRIN RHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DET A.BILC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CILLOPRENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAI | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DELTA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIFLORIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN I | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN II | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDOSULFAN SULFATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| GAMMA-BHC | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEPTACIILOR EPOXIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.8 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | | | | | | | | | | | | | | |
| AZINPHOS METHYL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DEMETON | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ETHYL PARATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MALATHION | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHYL PARATHION | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | | | | | | | | | 1 | | | | | |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

TABLE 6-18C THEORETICAL BIOACCUMULATION POTENTIAL* FOR CHLORINATED PESTICIDES, ORGANOPHOSPHORUS PESTICIDES, AND PCB AROCLORS IN TISSUE: COMPARISON TO OCEAN REFERENCE

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Tables 6-6 (Chlorinated Pesticides), 6-7 (Organophosphorus Pesticides), and 6-8 (PCB Aroclors). For non-detected analytes, the detection limit was used in the mean calculation.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Ocean Reference value.











TABLE 6-19A THEORETICAL BIOACCUMULATION POTENTIAL* FOR PCB CONGENERS IN TISSUE: COMPARISON TO INSIDE SITE 104

| | | Brewerton | C&D | C&D | | Cooleb III A col | C. J. M. L | C .1 1 M | C 114 | | | | | |
|------------------------|-----------------|-----------|--------|--------|-----------|------------------|------------|----------|-------------|--------------|------------|-------|---------------------|------------|
| Analyte (UG/KG) | Inside Site 104 | Extension | (Grab) | (Core) | Craighill | East | West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | Tolchester South | Tolchester |
| Mean % TOC | 7.2 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| BZ# 8 | 0.18 | 0.18 | 0.13 | 1 | 0.35 | 0.38 | 0.25 | 0.40 | 0.35 | ND | 0.24 | 0.36 | 0.35 | 0.17 |
| BZ# 18 | 0.26 | 0.17 | 0.18 | 0.31 | ND | 0.23 | 0.23 | 0,41 | ND | ND | 0.24 | 0.28 | 0.24 | 0.16 |
| BZ# 28 | 0.20 | 0.27 | 0.15 | 0.50 | ND | 0.27 | 0.12 | 0.34 | 0.23 | 0.15 | 0.12 | 6.30 | 0.36 | 0.16 |
| BZ# 44 | 0.30 | 0.22 | 0.18 | 0.36 | ND | 0.23 | ND | 0.29 | 0.35 | ND | ND | 0.24 | 0.25 | . 0.13 |
| BZ# 49 | 0.41 | 0.30 | 0.23 | 0.28 | ND | 0.23 | ND | 0.34 | 0.46 | ND | ND | 0.27 | 0.24 | 0.14 |
| BZ# 52 | 0.42 | 0.39 | 0.23 | 0.45 | ND | 0.14 | ND | 0.26 | 0.35 | ND | ND | 0.28 | 0.20 | 0.10 |
| BZ# 66 | 0.26 | 0.21 | 0.19 | 0.58 | ND | 0.15 | 0.13 | 0.29 | 0.29 | 0.23 | 0.09 | 0.26 | 0.31 | 0.17 |
| BZ# 77 | 1.60 | 0.15 | 0.46 | 1.56 | 0.35 | 0.12 | ND | 1.20 | 1.13 | 0.48 | 0.15 | 0.35 | 0.73 | 0.45 |
| BZ# 87 | 0.23 | 0.09 | 0.09 | 0.26 | ND | ND | ND | 0.14 | 0.26 | 0.09 | ND | 0.09 | 0.10 | 0.07 |
| BZ# 101 | 0.72 | 0.26 | 0.25 | 0.94 | 0.29 | 0.20 | 0.13 | 0.67 | 0.78 | 0.19 | 0.11 | 0.39 | 0.46 | 0.32 |
| BZ# 105 | 0.33 | ND | 0.22 | 0.26 | ND | ND | ND | ND | 0.52 | ND | ND | ND | ND | ND |
| BZ# 118 | 0.55 | 0.21 | 0.23 | 0.66 | 0.21 | 0.12 | 0.12 | 0.39 | 0.58 | 0.22 | 0.10 | 0.20 | 0.35 | 0.14 |
| BZ# 126 | 1.13 | 0.11 | 0.11 | 0.37 | ND | 0.06 | 0.09 | 0.17 | 0.15 | 0.10 | 0.08 | 0.15 | 0.18 | 0.05 |
| BZ# 128 | 0.32 | 0.07 | 0.09 | 0.22 | ND | ND | ND | 0.18 | 0.20 | ND | ND | 0.08 | 0.10 | 0.05 |
| BZ# 138 | 0.99 | 0.24 | 0.23 | 0.89 | 0.16 | 0.12 | 0.08 | 1.03 | 0.61 | 0.21 | 0.10 | 0.28 | 0.38 | 0.35 |
| BZ# 153 | 1.53 | 0.30 | 0.29 | 1.18 | 0.19 | 0.21 | 0.11 | 1.50 | 0.67 | 0.33 | 0.08 | 0.38 | 0.55 | 0.39 |
| BZ# 156 | 0.50 | ND | 0.13 | 0.27 | ND | ND | ND | 0.37 | ND | ND | ND | 0.11 | 0.12 | 0.10 |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 170 | 0.46 | 0.12 | 0.14 | 0.40 | ND | 0.10 | ND | 0.66 | 0.23 | ND | ND | 0.16 | 0.16 | 0.15 |
| BZ# 180 | 1.29 | 0.25 | 0.25 | 0.88 | ND | 0.15 | 0.14 | 1.48 | 0.29 | 0.19 | ND | 0.32 | 0.39 | 0.29 |
| BZ# 183 | 0.32 | ND | 0.08 | 0.18 | ND | ND | ND | 0.43 | ND | ND | ND | ND | 0.08 | 0.07 |
| BZ# 184 | 0.75 | 0.35 | 0.35 | 1.33 | 0.19 | 0.22 | 0.13 | 2.14 | 0.49 | 0.39 | 0.11 | 0.42 | 0.65 | 0.30 |
| BZ# 187 | 1.05 | 0.15 | 0.17 | 0.57 | ND | 0.10 | 0.11 | 0.83 | 0.20 | 0.13 | 0.09 | 0.22 | 0.28 | 0.17 |
| BZ# 195 | 0.18 | ND | 0.11 | 0.09 | ND | ND | ND | 0.25 | ND | ND | ND | ND | ND | ND |
| BZ# 206 | 0.70 | 0.39 | 0.49 | 1.99 | ND | 0.23 | 0.24 | 0.40 | 0.46 | 0.35 | 0.27 | 0.60 | 0.78 | 0.62 |
| BZ# 209 | 1.05 | 0.68 | 0.72 | 3.60 | 0.56 | 0.41 | 0.34 | 0.70 | 0.81 | 0.66 | 0.43 | 0.97 | 1.28 | 0.87 |
| TOTAL PCBs (ND=0) | 21.02 | 5.56 | 4.92 | 21.18 | 1.71 | 4.40 | 1.89 | 19.47 | 10.29 | 3.88 | 1.82 | 8.32 | 10.17 | 6.36 |
| TOTAL PCBs (ND=1/2 DL) | 22.22 | 6.46 | 6.04 | 21.27 | 5.46 | 5.19 | 3.45 | 20.33 | 12.54 | 5.69 | 3.32 | 8.77 | 10.69 | 6.62 |
| TOTAL PCBs (ND=DL) | 23.43 | 7.36 | 7.16 | 21.37 | 9.21 | 5.98 | 5.00 | 21.19 | 14.79 | 7.50 | 4.83 | 9.22 | 11.22 | 6.88 |

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-9. For non-detected analytes, the detection limit was used in the mean calculation.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Inside Site 104 value.

| 0 | Jutside Site | E a strenge b | | | | | | | | | | | | |
|------------------------|--------------|------------------|----------------|----------|-----------|-----------------|------------------------|-----------|-------------|--------------|------------|------------|------------|---------------|
| A Inter (IIIC/IV/C) | 104 | Eastern | Approach | Approach | | Craighill Angle | Craighiti Angle | Craighill | Cratghitt | | | Tolchester | Tolchester | Tolchester |
| Analyte (UG/KG) | 104 | Extension | (Grab) | (Core) | Craighill | East | West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | South | Straightening |
| Mean % IOC | 11.4 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| BZ# 8 | 0.18 | 0.18 | 0.13 | 0.46 | 0.35 | 0.38 | 0.25 | 9.40 | 0.35 | ND | 1. 0.24 | 0.36 | 0.35 | 0.17 |
| BZ# 18 | ND | 0.17 | 0.18 | 0.31 | ND | 0.23 | 0.23 | 0.41 | ND | ND | 0.24 | 0.28 | 0.24 | 0.16 |
| BZ# 28 | ND | 0.27 | 0.15 | .0.50 | ND | 0.27 | 0.12 | 0.34 | 0.23 | 0.15 | 0.12 | 0.30 | 0.36 | 0.16 |
| BZ# 44 | 0.14 | 0.22 | 0.18 | 0.36 | ND | 0.23 | ND | 0.29 | 0.35 | ND | ND | 0.24 | 0.25 | 0.13 |
| BZ# 49 | 0.21 | 0.30 | 0.23 | 0.28 | ND | 0.23 | ND | 0.34 | 0.46 | ND | ND | 0.27 | 0.24 | 0.14 |
| BZ# 52 | 0.15 | 0.39 | 0.23 | 0.45 | ND | 0.14 | ND | 0.26 | 0.35 | ND | ND | 0.28 | 0.20 | 0.10 |
| BZ# 66 | 0.09 | 0.21 | 0.19 | 0.58 | ND | 0.15 | 0,13 | 0.29 | 0.29 | 0.23 | 0.09 | 0.26 | 2611 031 | 0.17 |
| BZ# 77 | 0.13 | 0.15 | 0.46 | 1.56 | 0.35 | 0.12 | ND | 1.20 | 1.13 | 0.48 | 50.15 | 0.35 | 6 73 | 0.45 |
| BZ# 87 | ND | 0.09 | 0.09 | 0.26 | ND | ND | ND | 0.14 | 0.26 | 0.09 | ND | 0.09 | 0.10 | 0.07 |
| BZ# 101 | 0.13 | 6.26 | 0.25 | 0.94 | 0.29 | 0.20 | AL | 0.67 | 0.78 | 0.19 | 0.11 | 0.10 | | 0.07 |
| BZ# 105 | ND | ND | 0.22 | 0.26 | ND | ND | ND | ND | 0.52 | ND | ND | ND | ND | ND |
| BZ# 118 | 0.11 | 0.21 | 0.23 | TE 0.66 | 0.21 | 0.12 | 200 012 | 0.39 | 0 58 | 0.77 | 0.10 | 0 10 | IND O 16 | ND |
| BZ# 126 | 0.17 | 0.11 | 0.11 | 0.37 | ND | 0.06 | 90.0 | 0.17 | 0.15 | 0.10 | 0.00 | 0.15 | 0.19 | 0.06 |
| BZ# 128 | 0.06 | 0.07 | 0.09 | 0.77 | ND | ND | ND | 0.18 | 6.10 | ND | 0.08 | 0.13 | 0.18 | 0.05 |
| BZ# 138 | 0.08 | 0.24 | 0.23 | 0.89 | AL DENKS | 0.12 | 0.08 | 103 | 0.41 | 111 | 110 | 0.08 | 0.10 | 0.05 |
| BZ# 153 | 0.15 | 0.30 | 6.29 | 118 | 0.19 | | 0.11 | 1 40 | 0.01 | A 11 | 0.10 | 0.28 | 0.58 | 0.35 |
| BZ# 156 | ND | ND | 0.13 | 0.17 | ND | ND | ND | 0.17 | ND | V.J.J | 0.08 | 0.38 | 0.00 | 0.39 |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0,11 | 0.12 | 0.10 |
| BZ# 170 | 0.10 | 0.12 | 0.12 | 0.0 | ND | 0.10 | ND | 0.66 | 110 | ND | ND | ND | ND | ND |
| BZ# 180 | 0.13 | 0.75 | 0.75 | 22.0.1 | ND | nis nis | 014 | 1 40 | 6 10 | ND 10 | ND | 0.10 | 0.16 | 0.15 |
| BZ# 183 | ND | ND | 0.08 | 0.18 | ND | ND | NID | 0.41 | ND | 0.19 | ND | 0.31 | 0.39 | 0.29 |
| B7# 184 | 0.19 | 0.15 | 1 0 35 | 1 1 11 | 0.19 | 0.73 | 0.12 | | IND IO | ND | ND | ND | 0.08 | 0.07 |
| B7# 187 | 0.12 | 0.15 | 0.17 | 0.57 | ND | 0.10 | 0.15 | 0.01 | 0,49 | 0.39 | 0.11 | 0.42 | 0.65 | 0.30 |
| B7# 195 | ND | ND | 0 11 | 0.00 | ND | ND | ND. | 0.35 | 0.20 | 0.13 | 0.09 | 0.12 | 0.28 | 0.17 |
| B7# 206 | 0.45 | 0.30 | 1 | 1.00 | ND | 0.22 | 0.24 | 0.40 | ND | ND | ND | ND | ND | ND |
| B7# 209 | 0.45 | 0.68 | 0.72 | 1.50 | 0.56 | 0.23 | 0.24 | 0.40 | 0.40 | 0.33 | 0.27 | 0.60 | 0.78 | 0.62 |
| TOTAL PCBs (ND=0) | 1.04 | att a start & CA | 3.12 | 3.00 | 1.71 | 0.41 | 0.34 | 0.70 | 0.81 | 0.00 | 0.43 | 0.97 | 1.28 | 0.87 |
| TOTAL PCBs (ND=1/2 DI) | 1.94 | TATE KAK | C. S. S. S. S. | 41,10 | 1.71 | 2 10 | 1.89 | LYA/ | 10.29 | 3.88 | 1.82 | 8.32 | 10.17 | 6.36 |
| TOTAL PCPs (ND-DL) | A 44 | 7 74 | 718 | 1111111 | 2.40 | 5.19 | 3.43 | 20.33 | 12.54 | 5.69 | 3.32 | 8.77 | 10.69 | 6.62 |

TABLE 6-19B THEORETICAL BIOACCUMULATION POTENTIAL* FOR PCB CONGENERS IN TISSUE: COMPARISON TO OUTSIDE SITE 104

Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-9.
 For non-detected analytes, the detection limit was used in the mean calculation.
 Values based on 2% lipid content for either *Macoma* or *Nereis* tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Outside Site 104 value.

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| Analyte (UG/KG) | Ocean | Brewerton Extern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craighill | Craighill Angl East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|-------|----------------------------------|---------------------------|---------------------------|-----------|------------------------|-------------------------|-----------------------|--------------------------|--------------|--------------|---------------------|---------------------|-----------------------------|
| Mean % TOC | 0.5 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| BZ# 8 | 3.02 | 0.18 | 0.13 | 0.46 | 0.35 | 0.38 | 0.25 | 0.40 | 0.35 | ND | 0.24 | 0.36 | 0.35 | 0.17 |
| BZ# 18 | ND | 0.17 | 0.18 | 16.031 | ND | 0.23 | 第22月月月0.13 | 0.41 | ND | ND | 0.24 | 0.28 | 0.24 | 0.16 |
| BZ# 28 | ND | 0.17 | 0.15 | A 16 8 0.50 | ND | 0.27 | 0.12 | 0.34 | 0.23 | 0.15 | 2 25.0.12 | 0.30 | 0.36 | 0.16 |
| BZ# 44 | ND | 0.22 | 0.18 | 0.36 | ND | 0.23 | ND | 0.29 | 0.35 | ND | ND | 0.24 | 0.25 | 0.13 |
| BZ# 49 | ND | 0.30 | 0.23 | 0.28 | ND | SHE 0.23 | ND | .0.34 | 0.46 | ND | ND | 0.17 | 0.24 | 0.14 |
| BZ# 52 | ND | 0.39 | 0.23 | 11110.45 | ND | S 10.14 | ND | 0.26 | 0.35 | ND | ND | 0.78 | 0.20 | 0.10 |
| BZ# 66 | ND | 0.21 | 0.19 | Em 6.58 | ND | 0.15 | -035 | 0.29 | 0.29 | 0.23 | 0.09 | 0.25 | 0.31 | 0.17 |
| BZ# 77 | ND | 0.15 | 0.45 | AS 1.1.56 | 1055 | 0.12 | ND | 1.20 | 11.13 | 0.48 | 0.15 | 0.15 | 0.73 | 0.45 |
| BZ# 87 | ND | 0.09 | 8.09 | 0.26 | ND | ND | ND | 0.14 | 0.26 | 0.09 | ND | 0.09 | 0.10 | 0.07 |
| BZ# 101 | ND | 0.16 | 0.25 | 0.74 | 0.29 | 0.20 | 0.15 | 0.67 | 0.78 | 0.19 | 59[ABIX 0.11 | 0.39 | 6,46 | 0.32 |
| B2# 105 | ND | ND | 0.22 | 0.26 | ND | ND | ND | ND | 0.51 | ND | ND | ND | ND | ND |
| BZ# 118 | ND | 10.21 | 0.13 | 0.66 | 0.2 | 0.11 | ED-US ONE | 0.39 | 0.58 | 0.12 | 0.10 | 0.20 | 0.35 | 0.14 |
| BZ# 126 | ND | 0.11 | 0.11 | 037 | ND | 0.04 | 0.09 | 0.17 | 0.15 | | 0.08 | 0,15 | 0.18 | 0.05 |
| BZ# 128 | ND | 0.07 | | 0.12 | ND | NE | ND | 0.18 | 1 0.20 | ND | ND | 0.05 | 0.10 | 1 0.05 |
| BZ# 138 | ND | 0.24 | 0.23 | 0.89 | 0,16 | 0.11 | 1000 000 | 1.03 | 0.61 | GHAD 0.21 | 0.10 | 0.28 | 0.34 | 0.35 |
| BZ# 153 | ND | 0.30 | .0.29 | 招导的意义[]18 | 0,19 | 6114 0.2 | 0.11 | 1.50 | Sa - 0.67 | 0.33 | 80.0 | 0.38 | 0.55 | 0.39 |
| BZ# 156 | ND | ND | 0.13 | 0.27 | ND | NE | ND | 包回到20.37 | ND | ND | ND | 0.11 | 0.12 | A |
| BZ# 169 | ND | ND | ND | ND | ND | NE | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 170 | ND | 1012101-0.12 | Sub.c 0.14 | 0.40 | NE | 1110 1011 | ND | 0.66 | 0.23 | ND | ND | 0.16 | 0.16 | 0.15 |
| BZ# 180 | ND | 0.15 | 0.15 | 88.0 2.88 | ND | 0.1 | 0/14 | 1.48 | 0.29 | 1000000019 | ND | 0.31 | 0.37 | 0.29 |
| BZ# 183 | ND | ND | 0.08 | 0,18 | ND | NE | ND | 0.43 | ND | ND | ND | ND | 0.05 | 0.07 |
| BZ# 184 | ND | 6.35 | 0.35 | 10 ft 2133 | 0.19 | 0.21 | 0.15 | 214 | 0.49 | 0.39 | 0.11 | 0.42 | 0.65 | 0.10 |
| BZ# 187 | ND | 0.15 | 0.17 | 一日日二日三0.57 | ND | 0.10 | 0.11 | 0.83 | 0.20 | 0.13 | 0.09 | 0.12 | 0.28 | 0.17 |
| BZ# 195 | ND | ND | 711 0.11 | 0.09 | ND | NE | ND | 0.25 | ND | ND | ND | ND | ND | ND |
| BZ# 206 | ND | 0.39 | 0.49 | 1.99 | ND | 0.23 | 10.14 | 0.40 | 0.46 | 0.35 | 8.27 | 0.60 | 0.78 | 0.61 |
| BZ# 209 | ND | 0.68 | 0.72 | 115 - 3.60 | 0.50 | 0.41 | 0.34 | 0.70 | 0.81 | * 0.66 | Sec. 1.0.43 | 0.97 | 1.24 | 0.87 |
| TOTAL PCBs (ND=0) | 6.04 | 5.56 | 4.92 | 110-100-21.18 | 1.71 | 4.40 | 1.89 | 用油口 19.47 | 留户 10.29 | 3.88 | 1.82 | 1054 10 18.32 | Ball 10.17 | ELSold 6.36 |
| TOTAL PCBs (ND=1/2 DL) | 24.30 | 6,46 | 6.04 | 21.27 | 5.46 | 5 5.15 | 3,45 | 20.33 | 12.54 | 5.69 | 3.32 | 8.77 | 10.69 | 6.62 |
| TOTAL PCBs (ND=DL) | 42.57 | 7.36 | 7.16 | 21.37 | 9.21 | 5.91 | 5.00 | 21.19 | 14.79 | 7.50 | 4.83 | 9.22 | 11.22 | 6.88 |

Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-9.
 For non-detected analytes, the detection limit was used in the mean calculation.
 Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Ocean Reference value.

| Analyte (UG/KG) | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Toichester South | Tolchester Straightening |
|------------------------|-----------------|-----------------------------------|---------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| Mean % TOC | 7.2 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| ACENAPHTHENE | 321.6 | 40.10 | 57.25 | 12.32 | 51.20 | 29.68 | 23.46 | 69.73 | 71.92 | 41.31 | 44.84 | 106.67 | 73.39 | 158.60 |
| ACENAPHTHYLENE | 32.0 | ND | 32.11 | ND | ND | ND | ND | ND | ND | ND | 32.40 | 50.39 | 37.03 | 131.67 |
| ANTHRACENE | 74.9 | 18.84 | 26.40 | 1.90 | 18.75 | 10.00 | 9.22 | 27.84 | 29.44 | 15.28 | 19.05 | 39.97 | 27.44 | 92.77 |
| BENZ[A]ANTHRACENE | 68.6 | 19.46 | 28.73 | 5.02 | 22.94 | 10.93 | 10.73 | 25.38 | 41.41 | 22.77 | 14.05 | 34.11 | 23.44 | 54.16 |
| BENZO[A]PYRENE | 88.8 | 31.85 | 47.74 | 6.64 | 42.22 | 13.54 | 14.72 | 25.41 | 79.19 | 46.68 | 17.63 | 44.03 | 31.72 | 65.84 |
| BENZO[B]FLUORANTHENE | 103.2 | 34.24 | 71.19 | 8.44 | 50.94 | 24.21 | 16.28 | 45.68 | 77.01 | 36.27 | 41.65 | 102.95 | 80.57 | 137.66 |
| BENZO[G,H,I]PERYLENE | 55.6 | 23.07 | 32.32 | 4.45 | 32.46 | 10.78 | 14.66 | 22.70 | 53.76 | 32.57 | 12.44 | 33.33 | 22.48 | 41.30 |
| BENZO[K]FLUORANTHENE | 46.5 | 12.35 | 24.50 | 2.37 | 17.87 | 6.48 | 6.64 | 15.68 | 29.58 | 16.93 | 9.69 | 24.03 | 16.08 | 30.82 |
| CIIRYSENE | 44.9 | 19.56 | 31.69 | 5.59 | 21.79 | 9.45 | 8.86 | 22.36 | 35.80 | 19.82 | 13.24 | 33.65 | 21.88 | 43.99 |
| DIBENZ[A,H]ANTHRACENE | 10.7 | 5.85 | 5.34 | ND | ND | ND | ND | ND | 10.32 | 6.38 | ND | 5.40 | 4.64 | 4.52 |
| FLUORANTHENE | 249.2 | 71.70 | 104.99 | 10.71 | 64.35 | 35.93 | 34.08 | 94.05 | 103.16 | 50.71 | 56.97 | 128.21 | 92.13 | 236.41 |
| FLUORENE | 72.2 | 21.13 | 34.01 | 2.84 | 21.68 | 9.71 | 9.97 | 27.30 | 35.16 | 15.39 | 19.69 | 42.95 | 29.62 | 104.74 |
| INDENO[1,2,3-CD]PYRENE | 33.9 | 17.51 | 23.02 | 3.36 | 11.65 | 8.49 | 6.81 | 16.56 | 34.29 | 20.12 | 9.21 | 23.88 | 16.77 | 28.43 |
| NAPHTHALENE | 263.2 | 80.81 | 102.24 | 13.74 | 79.55 | 51.55 | 58.65 | 102.43 | 114.78 | 76.57 | 91.37 | 175.04 | 120.01 | 374.06 |
| PHENANTHRENE | 225.0 | 52.91 | 89.35 | 6.59 | 50.14 | 26.30 | 25.70 | 69.46 | 83.55 | 38.62 | 50.51 | 111.63 | 75.96 | 266.33 |
| PYRENE | 224.4 | 65.51 | 98.23 | 10.90 | 64.27 | 29.94 | 30.44 | 74.86 | 111.88 | 56.09 | 45.03 | 108.68 | 76.51 | 200.50 |
| TOTAL PAHs (ND=0) | 1882.97 | 506.25 | 803.61 | 94.88 | 540.31 | 276.99 | 269.11 | 639.43 | 903.81 | 491.35 | 449.06 | 1045.50 | 729.53 | 1971.80 |
| TOTAL PAHs (ND=1/2 DL) | 1898.80 | 524.24 | 806.36 | 107.09 | 582.00 | 291.78 | 286.53 | 660.97 | 936.58 | 511.91 | 465.89 | 1055.20 | 739.61 | 1971.80 |
| TOTAL PAHs (ND=DL) | 1914.63 | 542.23 | 809.10 | 119.29 | 623.70 | 306.57 | 303.94 | 682.51 | 969.34 | 532.46 | 482.72 | 1064.90 | 749.68 | 1971.80 |

TABLE 6-20A THEORETICAL BIOACCUMULATION POTENTIAL* FOR PAHs IN TISSUE: COMPARISON TO INSIDE SITE 104

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-10. For non-detected analytes, the detection limit was used in the mean calculation. Values based on 2% lipid content for either *Macoma* or *Nereis* tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Inside Site 104 value.





| Analyte (UG/KG) | Outside Site 104 | Brewerton Eastern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighili Angle West | Craighili Entrance | Craighilt Upper Range | Cutoff Angle | Swan Point | Tolchester North | Toichester South | Tolchester |
|------------------------|---------------------|-----------------------------------|---------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|------------|
| Mean % TOC | 11.4 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| ACENAPHTHENE | 137.5 | 40.10 | 57.25 | 12.32 | 51.20 | 29.68 | 23.46 | 69.73 | 71.92 | 41.31 | 44.84 | 106.67 | 73.39 | 158.60 |
| ACENAPHTHYLENE | 36.2 | ND | 32.11 | ND | ND | ND | ND | ND | ND | ND | 32.40 | 50.39 | 37.03 | 131.67 |
| ANTHRACENE | 36.8 | 18.84 | 26.40 | 1.90 | 18.75 | 10.00 | 9.22 | 27.84 | 29.44 | 15.28 | 19.05 | 39.97 | 27.44 | 97.77 |
| BENZ[A]ANTHRACENE | 68.7 | 19.46 | 28.73 | 5.02 | 22.94 | 10.93 | 10.73 | 25.38 | 41.41 | 22.77 | 14.05 | 34.11 | 23.44 | 54 16 |
| BENZO[A]PYRENE | 82.1 | 31.85 | 47.74 | 6.64 | 42.22 | 13.54 | 14.72 | 25.41 | 79.19 | 46.68 | 17.63 | 44.03 | 31.72 | 65.84 |
| BENZO[B]FLUORANTIIENE | 87.0 | 34.24 | 71.19 | 8.44 | 50.94 | 24.21 | 16.28 | 45.68 | 77.01 | 36.27 | 41.65 | 102.95 | 80.57 | 137.66 |
| BENZO[G,H,I]PERYLENE | 57.0 | 23.07 | 32.32 | 4.45 | 32.46 | 10.78 | 14.66 | 22.70 | 53.76 | 32.57 | 12.44 | 33.33 | 22.48 | 41 30 |
| BENZO[K]FLUORANTHENE | 38.1 | 12.35 | 24.50 | 2.37 | 17.87 | 6.48 | 6.64 | 15.68 | 29.58 | 16.93 | 9.69 | 24.03 | 16.08 | 30.82 |
| CHRYSENE | 56.4 | 19.56 | 31.69 | 5.59 | 21.79 | 9.45 | 8.86 | 22.36 | 35.80 | 19.82 | 13.24 | 33.65 | 21.88 | 43.00 |
| DIBENZ[A,H]ANTHRACENE | ND | 5.85 | 5.34 | ND | ND | ND | ND | ND | 10.32 | 6.38 | ND | 540 | 21.00 | 457 |
| FLUORANTHENE | 184.1 | 71.70 | 104.99 | 10.71 | 64.35 | 35.93 | 34.08 | 94.05 | 103.16 | 50.71 | 56.97 | 128.21 | 92.13 | 236.41 |
| FLUORENE | 27.5 | 21.13 | 34.01 | 2.84 | 21.68 | 9.71 | 9.97 | 27.30 | 35.16 | 15.39 | 19.69 | 42.95 | 29.62 | 104.74 |
| INDENO[1,2,3-CD]PYRENE | 39.0 | 17.51 | 23.02 | 3.36 | 11.65 | 8.49 | 6.81 | 16.56 | 34.29 | 20.12 | 9.21 | 23.88 | 16.77 | 28 43 |
| NAPHTHALENE | 106.9 | 80.81 | 102.24 | 13.74 | 79.55 | 51.55 | 58.65 | 102.43 | 114.78 | 76.57 | 91.37 | 175.04 | 120.01 | 174 66 |
| PHENANTHRENE | 106.2 | 52.91 | 89.35 | 6.59 | 50.14 | 26.30 | 25.70 | 69.46 | 83.55 | 38.62 | 50.51 | 111.63 | 75.96 | 266 33 |
| PYRENE | 120.0 | 65.51 | 98.23 | 10.90 | 64.27 | 29.94 | 30.44 | 74.86 | 111.88 | 56.09 | 45.03 | 108.68 | 76 51 | 200 50 |
| TOTAL PAHs (ND=0) | 1159.91 | 506.25 | 803.61 | 94.88 | 540.31 | 276.99 | 269.11 | 639.43 | 903.81 | 491.35 | 449.06 | 1045.50 | 729 53 | 1971 80 |
| TOTAL PAHs (ND=1/2 DL) | 1173.78 | 524.24 | 806.36 | 107.09 | 582.00 | 291.78 | 286.53 | 660.97 | 936.58 | 511.91 | 465.89 | 1055.20 | 739 61 | 1971 80 |
| TOTAL PAHs (ND=DL) | 1187.65 | 542.23 | 809.10 | 119.29 | 623.70 | 306.57 | 303.94 | 682.51 | 969.34 | 532.46 | 482.72 | 1064.90 | 749.68 | 1971-80 |

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-10.

For non-detected analytes, the detection limit was used in the mean calculation.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Outside Site 104 value.

| Analyte (UG/KG) | Ocean | Brewerton Eastern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighitt Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------|-----------------------------------|---------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| Mean % TOC | 0.5 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| ACENAPHTHENE | ND | 40.10 | 57.25 | 12.32 | 51.20 | 29.68 | 23.46 | 69.73 | 71.92 | 6 41.31 | 44.84 | 106.67 | 13.39 | 158.60 |
| ACENAPHTHYLENE | ND | ND | 32.11 | ND | ND | ND | ND | ND | ND | ND | 32.40 | 50.39 | 37.03 | 131.87 |
| ANTHRACENE | ND | 18.84 | 26.40 | 1.90 | 18.75 | 10.00 | 9.22 | 27.84 | 29.44 | 15.28 | 19.05 | 39.97 | 27.44 | 92.77 |
| BENZ[A]ANTHRACENE | ND | 19.46 | 28.73 | 5.02 | 22.94 | 10.93 | 10.73 | 25.38 | 41.41 | 22.77 | 14.05 | 34.11 | 23.44 | 54.16 |
| BENZO[A]PYRENE | ND | 31.85 | 47.74 | 6.64 | 42.22 | 13.54 | 14.72 | 25.41 | 79.19 | 46.68 | 17.63 | 44.03 | 31.72 | 65.84 |
| BENZO[B]FLUORANTHENE | ND | 34.24 | 71.19 | 8.44 | 50.94 | 24.21 | 16.28 | 45.68 | 77.01 | 36.27 | 41.65 | 102.95 | 80.57 | 137.66 |
| BENZO[G,H,I]PERYLENE | ND | 23.07 | 32.32 | 4.45 | 32.46 | 10.78 | 14.66 | 22.70 | 53.76 | 32.57 | 12.44 | 33.33 | 22.48 | 41.30 |
| BENZO[K]FLUORANTHENE | ND | 12.35 | 24.50 | 2.37 | 17.87 | 2 2 6.48 | 6.64 | 15.68 | 29.58 | 16.93 | 9.69 | 24.03 | 16.08 | 30.82 |
| CHRYSENE | ND | 19.56 | 31.69 | 5.59 | 21.79 | 9.45 | 8.86 | 22.36 | 35.80 | 19.82 | 13.24 | 33.65 | 21.88 | -43.99 |
| DIBENZ[A,H]ANTHRACENE | ND | 5.85 | 534 | ND | ND | ND | ND | ND | 10.32 | 6.38 | ND | 5.40 | 4.64 | 4.52 |
| FLUORANTHENE | ND | 71.70 | 104.99 | 10//1 | 64.35 | 35.93 | 34.08 | 94.05 | 103.16 | 50.71 | 56.97 | 128.21 | 92.13 | 236.41 |
| FLUORENE | ND | 21.13 | 34.01 | 2.84 | 21.68 | 9.71 | 9.97 | 27.30 | 35.16 | 15.39 | 19.69 | 42.95 | 29.62 | 104.74 |
| INDENO[1,2,3-CD]PYRENE | ND | 17.51 | 23.02 | 3.36 | 11.65 | 8.49 | 6.81 | 16.56 | 34.29 | 20.12 | 9.21 | 23.88 | 16.77 | 28.43 |
| NAPHTHALENE | ND | 80.81 | 102.24 | 13.74 | 79.55 | 51.55 | 58.65 | 102.43 | 114.78 | 76.57 | 91.37 | 175.04 | 120.01 | 374.06 |
| PHENANTHRENE | ND | 52.91 | 89.35 | 6.59 | 50.14 | 26.30 | 25.70 | 69.46 | 83,55 | 38.62 | 50.51 | 111.63 | 75.96 | 266.33 |
| PYRENE | ND | 65.51 | 98.23 | 10.90 | 64.27 | 29.94 | 30.44 | 74.86 | 111.88 | 56.09 | 45.03 | 108.68 | 76.51 | 200.50 |
| TOTAL PAHs (ND=0) | 0.00 | 506.25 | 803.61 | 94.88 | 540.31 | 276.99 | 269.11 | 639,43 | 903.81 | 491.35 | 449.06 | 1045.50 | 729.53 | 1971.80 |
| TOTAL PAHs (ND=1/2 DL) | 434.11 | 524.24 | 806.36 | 107.09 | 582.00 | 291.78 | 286.53 | 660.97 | 936.58 | | 465.89 | 1055.20 | 739.61 | 1971.80 |
| TOTAL PAHs (ND=DL) | 868.23 | 542.23 | 809.10 | 119.29 | 623.70 | 306.57 | 303.94 | 682.51 | 969.34 | 532.46 | 482.72 | 1064.90 | 749.68 | 1971.80 |

TABLE 6-20C THEORETICAL BIOACCUMULATION POTENTIAL* FOR PAHs IN TISSUE: COMPARISON TO OCEAN REFERENCE

* Values based on mean concentrations detected in the sediment samples (dry weight) as presented in Table 6-10.

For non-detected analytes, the detection limit was used in the mean calculation. Values based on 2% lipid content for either *Macoma* or *Nereis* tissue.

ND = analytes that were not detected in any sample in a reach.

The analytes that were not detected in any sample in a reach.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Ocean Reference value.







TABLE 6-21A THEORETICAL BIOACCUMULATION POTENTIAL* FOR DIOXIN IN TISSUE: COMPARISON TO INSIDE SITE 104

| Analyte (NG/KG) | Inside Site 104 | Brewerton Eastern Extension | C&Đ Approach (Grah) | C&D Approach (Core) | Craighill | Craighill Angie East | e Craighili Angle West | Craighill Entrance | Craighlil Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-------------------------|-----------------|-----------------------------------|---------------------------|---------------------------|-----------|-------------------------|---------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| % TOC | 7.2 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| 2,3,7,8-TCDD | 0.3 | ND | ND | ND | ND |) ND | ND | ND | ND | ND | ND | 0.83 | ND | ND |
| 1,2,3,7,8-PeCDD | 0.2 | ND | ND | ND | 0.32 | 0.67 | ND | 1.02 | ND | 1.08 | ND | 1.00 | ND | 0.47 |
| 1,2,3,4,7,8-HxCDD | 0.2 | ND | ND | ND | ND | ND ND | ND | ND | / ND | ND | ND | ND | ND | 0.41 |
| 1,2,3,6,7,8-HxCDD | 0.6 | ND | ND | ND | 1.33 | 1.24 | ND | 2.15 | ND | ND | ND | ND | ND | 0.56 |
| 1,2,3,7,8,9-HxCDD | 0.6 | ND | ND | ND | 0.85 | 1.44 | ND | 2.43 | ND | ND | ND | ND | ND | 0.48 |
| 1,2,3,4,6,7,8-HpCDD | 7.9 | 5.14 | 5.80 | 3.02 | 20.59 | / 18.59 | 10.89 | 35.78 | 21.77 | 12.29 | 7.56 | 8.01 | 4.27 | 3.47 |
| OCDD | 169.5 | 117.72 | 141.95 | 80.66 | 288.02 | 448.33 | 300.81 | 1124.32 | 505.63 | 281.11 | 209.15 | 155.35 | 127.59 | 67.03 |
| 2,3,7,8-TCDF | 0.5 | ND | ND | ND | 23.82 | ND | ND | 0.81 | ND | ND | ND | 1.06 | ND | 1.02 |
| 1,2,3,7,8-PeCDF | 0.2 | ND | ND | ND | 4.03 | 0.92 | ND | ND | ND | 1.25 | ND | 1.05 | ND | 1.39 |
| 2,3,4,7,8-PeCDF | 2.0 | ND | ND | ND | 24.32 | 0.93 | ND | 1.20 | ND | 1.03 | 1.44 | 0.99 | ND | 2.65 |
| 1,2,3,4,7,8-HxCDF | 0.9 | 0.36 | ND | ND | 10.35 | 0.87 | ND | 1.07 | 0.93 | 1.10 | 0.60 | 0.74 | ND | 1.90 |
| 1,2,3,6,7,8-HXCDF | 0.4 | 0.28 | ND | ND | 1.81 | 0.66 | ND | 0.68 | ND | ND | 0.32 | 0.68 | ND | 1.57 |
| 2,3,4,6,7,8-HxCDF | 0.6 | ND | ND | ND | 3.07 | 0.62 | ND | ND | ND | ND | ND | ND | ND | 1.66 |
| 1,2,3,7,8,9-HxCDF | 0.1 | ND | ND | ND | 0.51 | ND | ND | ND | / ND | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HpCDF | 2.1 | 0.77 | 0.95 | 0.56 | 6.13 | 2.23 | 1.58 | 3.90 | 2.96 | 1.74 | 1.54 | 0.94 | 0.57 | 5.40 |
| 1,2,3,4,7,8,9-HpCDF | 0.3 | ND | ND | ND | 2.32 | ND | ND' | ND | ND | ND | ND | ND | ND | 0.37 |
| OCDF | 3.7 | 2.54 | 2.16 | 1.08 | 13.73 | 5.87 | 5.18 | 9.47 | 9.39 | 5.66 | 3.13 | 2.20 | 1.52 | 1.87 |
| DIOXINS TEQ (ND=0) | 1.7 | 0.22 | 0.21 | 0.12 | 17.08 | 1.95 | 0.43 | 3.20 | 0.76 | 1.11 | 1.11 | 2.22 | 0.18 | 2.16 |
| DIOXINS TEQ (ND=1/2 DL) | 1.7 | 1.08 | 0.97 | 0.47 | 17.41 | 2.40 | 1.63 | 4.A5 | , 2.40 | 2.06 | 1.75 | 2.40 | 1.05 | 2.28 |

* Values based on actual concentrations detected in composited sediment samples (dry weight) as presented in Table 6-13. Values based on 2% lipid content for either *Macoma* or *Nereis* tissue.

ND = analytes that were not detected in a composite sample.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Inside Site 104 value.

| Analyte (NG/KG) | Outside Site 104 | Brewerton Eastern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-------------------------|---------------------|-----------------------------------|---------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| % TOC | 11.4 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| 2,3,7,8-TCDD | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.83 | ND | ND |
| 1,2,3,7,8-PeCDD | ND | ND | ND | ND | 0.32 | 0.67 | ND | 2011.02 | ND | 1.08 | ND | 1.00 | ND | THE 10.47 |
| 1,2,3,4,7,8-HxCDD | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.41 |
| 1,2,3,6,7,8-HxCDD | ND | ND | ND | ND | 1.33 | 1.24 | ND | 111 12115 | ND | ND | ND | ND | ND | 0.56 |
| 1,2,3,7,8,9-HxCDD | ND | ND | ND | ND | 0.85 | 1.44 | ND | 2.43 | ND | ND | ND | ND | ND | 0.48 |
| 1,2,3,4,6,7,8-HpCDD | 4.9 | 5.14 | 5.80 | 3.02 | 20.59 | 18.59 | 10.89 | 35.78 | 21.77 | 12.29 | 7.56 | 8.01 | 4.27 | 3.47 |
| OCDD | 135.6 | 117.72 | 141.95 | 80.66 | 288.02 | 448.33 | 200.81 | 1124.32 | 505.63 | 281.11 | 209.15 | 155.35 | 127.59 | 67.03 |
| 2,3,7,8-TCDF | ND | ND | ND | ND | 23.82 | ND | ND | 0.81 | ND | ND | ND | 1.06 | ND | 1.02 |
| 1,2,3,7,8-PeCDF | ND | ND | ND | ND | 4.03 | 0.92 | ND | ND | ND | 1.25 | ND | 1.05 | ND | 139 |
| 2,3,4,7,8-PeCDF | ND | ND | ND | ND | 24.32 | 0.93 | ND | 1.20 | ND | 1.03 | 1.44 | 0.99 | ND | 2.66 |
| 1,2,3,4,7,8-HxCDF | ND | 0.36 | ND | ND | 10.35 | 0.87 | ND | 1.07 | 0.93 | 1.10 | 0.60 | 0.74 | ND | 1.90 |
| 1,2,3,6,7,8-HXCDF | ND | 0.28 | ND | ND | 1.81 | 0.66 | ND | 0.68 | ND | ND | 0.32 | 0.68 | ND | 1.57 |
| 2,3,4,6,7,8-HxCDF | ND | ND | ND | ND | 3.07 | 0.62 | ND | ND | ND | ND | ND | ND | ND | 1.66 |
| 1,2,3,7,8,9-HxCDF | ND | ND | ND | ND | 0.51 | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HpCDF | 0.7 | 0.77 | 0.95 | 0.56 | 6.13 | 2.21 | 1.58 | 3.90 | 2.96 | 1.74 | 1.54 | 0.94 | 0.57 | A 2.R 5.40 |
| 1,2,3,4,7,8,9-11pCDF | ND | ND | ND | ND | 2.32 | ND | ND | ND | ND | ND | ND | ND | ND | 0.37 |
| OCDF | 1.3 | 2.54 | 2.16 | 1.08 | 13.73 | 5.87 | 5.18 | 9.47 | 9.39 | 5.66 | 3.13 | 2.20 | 1.51 | 1.87 |
| DIOXINS TEQ (ND=0) | 0.2 | 0.22 | 0.21 | 0.12 | 17.08 | 1.95 | 0.43 | 3.20 | 0.76 | 1.11 | 1.11 | 2.22 | 0.18 | 2.16 |
| DIOXINS TEQ (ND=1/2 DL) | 0.6 | 1.08 | 0.97 | 0.47 | 17.41 | 2.40 | 1.63 | 1.45 | 2.40 | 2.06 | 1.75 | 2.40 | 1.05 | 2.28 |

TABLE 6-21B THEORETICAL BIOACCUMULATION POTENTIAL* FOR DIOXIN IN TISSUE: COMPARISON TO OUTSIDE SITE 104

* Values based on actual concentrations detected in composited sediment samples (dry weight) as presented in Table 6-13.

Values based on 2% lipid content for either Macoma or Nereis tissue.

ND = analytes that were not detected in a composite sample.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Outside Site 104 value.





| Analyte (NG/KG) | Ocean | Brewerton Eastern Extension | C&D Approach (Grab) | C&D Approach (Core) | Craightil | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighiti Upper Range | Cutoff Angle | Swan Point | Tolchester North | Totchester South | Tolchester Straightening |
|-------------------------|-------|-----------------------------------|---------------------------|---------------------------|---------------|-------------------------|-------------------------|-----------------------|--------------------------|-------------------|--------------|---------------------|---------------------|-----------------------------|
| % TOC | 0.5 | 10.5 | 9.5 | 8.4 | 3.0 | 10.2 | 9.5 | 7.4 | 2.8 | 7.9 | 8.7 | 8.6 | 9.0 | 13.4 |
| 2,3,7,8-TCDD | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.83 | ND | ND |
| 1,2,3,7,8-PeCDD | ND | ND | ND | ND | Sime 0.37 | 0.67 | ND | 1.02 | ND | 1.08 | ND | 1.00 | ND | 0.47 |
| 1,2,3,4,7,8-HxCDD | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.41 |
| 1,2,3,6,7,8-HxCDD | ND | ND | ND | ND | 178113-1183 | 1.24 | ND | 2.15 | ND | ND | ND | ND | ND | C+ 11 0.56 |
| 1,2,3,7,8,9-HxCDD | ND | ND | ND | ND | 0.85 | 1.44 | ND | 一首告 2.43 | ND | ND | ND | ND | ND | 0.48 |
| 1,2,3,4,6,7,8-HpCDD | 12.5 | 5.14 | 5.80 | 3.02 | 34 20.59 | 18.59 | 10.89 | 35.78 | 21.77 | 12.29 | 7.56 | 8.01 | 4.27 | 3.47 |
| OCDD | 160.0 | 117.72 | 141.95 | 80.66 | 288.07 | 12 1448.33 | 300.81 | 1124.32 | 505.63 | 281.11 | STATE 209.15 | 155.35 | 127.59 | 67.03 |
| 2,3,7,8-TCDF | ND | ND | ND | ND | 341 ala 23.82 | ND | ND | CHE205ER 0.81 | ND | ND | ND | 1.06 | ND | 1.02 |
| 1,2,3,7,8-PeCDF | ND | ND | ND | ND | 4.03 | 0.92 | ND | ND | ND | 1.25 | ND | 1.05 | ND | + 1.39 |
| 2,3,4,7,8-PeCDF | 1.1 | ND | ND | ND | 24.32 | 0.93 | ND | 1.20 | ND | 1.03 | antise 1.44 | 0.99 | ND | 2.66 |
| 1,2,3,4,7,8-HxCDF | ND | 100 036 | ND | ND | 10.35 | 10.87 | ND | 1.07 | 0.93 | 1.10 | 0.60 | 0.74 | ND | 1.90 |
| 1,2,3,6,7,8-HXCDF | ND | - hone 0.28 | ND | ND | 1.81 | 0.66 | ND | 0.68 | ND | ND | 0.32 | 0.68 | ND | 5157 |
| 2,3,4,6,7,8-HxCDF | ND | ND | ND | ND | 3.07 | 0.62 | ND | ND | ND | ND | ND | ND | ND | 1.66 |
| 1,2,3,7,8,9-HxCDF | ND | ND | ND | ND | 0.51 | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HpCDF | 2.6 | 0.77 | 0.95 | 0.56 | E # 2 6.13 | 2.23 | 1.58 | F#1-J-3.90 | 2.96 | 1.74 | 1.54 | 0.94 | 0.57 | 5.40 |
| 1,2,3,4,7,8,9-HpCDF | ND | ND | ND | ND | 11 1 2.37 | ND | ND | ND | ND | ND | ND | ND | ND | 0.37 |
| OCDF | 2.4 | 2.54 | 2.16 | 1.08 | 11- Pra 13.73 | 5.87 | 5.18 | 1051 9.47 | 9.39 | 5.66 | 3.13 | 2.20 | 1.52 | 1.87 |
| DIOXINS TEQ (ND=0) | 0.8 | 0.22 | 0.21 | 0.12 | 17.05 | 4 Horard 1.95 | 0.43 | 1115 3.20 | 0.76 | 1.11 Section 1.11 | No. 1.11 | 2.22 | 0.18 | 4.2.16 |
| DIOXINS TEQ (ND=1/2 DL) | 2.4 | 1.08 | 0.97 | 0.47 | 17.41 | 2.40 | 1.63 | 18.00 A.45 | 2.40 | 2.06 | 1.75 | 2.40 | 1.05 | 2.28 |

* Values based on actual concentrations detected in composited sediment samples (dry weight) as presented in Table 6-13. Values based on 2% lipid content for either *Macoma* or *Nereis* tissue.

ND = analytes that were not detected in a composite sample.

NOTE: Shaded and bolded values represent TBP values for channels that exceed the Ocean Reference value.

TABLE 6-22 NUMBER OF MEAN CONCENTRATIONS IN TARGET ANALYTE FRACTIONS THAT EXCEED TELs (1999)

| | | | | Brewerton | C&D | C&D | | | | | Craighiil | | | T | | |
|--------------------------------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | Ocean | Inside | Outside | Eastern | Approach | Approach | | Craighiii | Craighili | Craighlil | Upper | Cutoff | Swan | Toichester | Toichester | Toichester |
| ANALYTE | Reference | Slte 104 | Site 104 | Extension | (Surficiai) | (Core) | Cralghill | Angle East | Angle West | Entrance | Range | Angie | Point | North | South | Straightening |
| METALS | 0 | 9 | 7 | 6 | 6 | 7 | 0 | 6 | 7 | 6 | 0 | 7 | 6 | 7 | 6 | 6 |
| PAHs | 0 | 10 | 11 | 5 | 7 | 2 | 2 | 3 | 3 | 4 | 3 | 3 | 4 | 6 | 4 | 12 |
| PESTICIDES | i | 1 | 3 | 3 | 1 | 1 | 1 | 2 | 3 | 3 | 1 | 4 | 2 | 1 | 1 | 1 |
| PCBs, TOTAL | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SEMIVOLATILE ORGANIC COMPOUNDS | 0 | 1 | 1 | 1 | 1 | 0 | 0 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| TOTAL # OF TEL EXCEEDANCES | i | 21 | 22 | 15 | 15 | 13 | 3 | 12 | 14 | 14 | 5 | 15 | 13 | 15 | 12 | 20 |







TABLE 6-23 NUMBER OF MEAN CONCENTRATIONS IN TARGET ANALYTE FRACTIONS THAT EXCEED PELs (1999)

| | | | | Brewerton | C&D | C&D | | | | | Craighili | | | | | |
|--------------------------------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | Ocean | Inside | Outside | Eastern | Approach | Approach | | Craighiii | Cralghill | CraighIII | Upper | Cutoff | Swan | Toichester | Tolchester | Toichester |
| ANALYTE | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Craighiil | Angie East | Angie West | Entrance | Range | Angie | Point | North | South | Stralghtening |
| METALS | 0 | 3 | 0 | 2 | i | i | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | i |
| PAHs | 0 | I | i | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | i | 0 | 4 |
| PESTICIDES | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | . 0 |
| PCBs, TOTAL | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SEMIVOLATILE ORGANIC COMPOUNDS | 0 | i | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | i |
| TOTAL # OF PEL EXCEEDANCES | 0 | 5 | I | 2 | i | I | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 1 | 6 |

TABLE 6-24 SUMMARY OF TBP RESULTS: NUMBER OF CHANNEL TBP VALUES THAT EXCEED^(a) PLACEMENT SITE/REFERENCE SITE TBP VALUES

| | | | | PERCE | PERCENT EXCEEDA | | | | |
|-----------------------------|--------------------|---------------------|--------------------|--------------------|---------------------|--------------------|--|--|--|
| ANALYTICAL FRACTION | Inside Site 104 | Outside Site 104 | Ocean Reference | Inside Site 104 | Outside Site 104 | Ocean Reference | | | |
| Chlorinated Pesticides | 4/286 | 8/286 | 9/286 | 1.4 | 2.8 | 3.1 | | | |
| Organophosphorus Pesticides | 0/65 | 0/65 | 0/65 | 0.0 | 0.0 | 0.0 | | | |
| PCB Aroclors | 0/91 | 0/91 | 0/91 | 0.0 | 0.0 | 0.0 | | | |
| PCB Congeners | 46/338 | 210/338 | 253/338 | 13.6 | 62.1 | 74.9 | | | |
| PAHs | 11/208 | 28/208 | 194/208 | 5.3 | 13.5 | 93.3 | | | |
| Dioxin and Furan Congeners | 67/221 | 95/221 | 74/221 | 30.3 | 4.3 | 33.5 | | | |

(a) = total number of channel concentrations (detects only) that exceed reference concentrations/total number of analytes tested.

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7. RECEIVING WATER AND ELUTRIATE CHEMISTRY

This chapter presents receiving water and elutriate chemistry data that were collected in 1999-2000 specifically for a Tier II evaluation of water column impacts. These data are similar to the 1997-1998 data that are presented in Chapter 3. Placement site water (Inside Site 104), and reference site water (Outside Site 104 and Ocean Reference) were submitted for analytical testing. These waters are referred to as "receiving waters" because the dredged material would be received into these waters when placed. Elutriate preparation water was collected from each channel proposed for dredging [as specified in the ITM (USEPA/USACE 1998)] for preparation of the individual elutriates in the analytical and toxicology laboratories. Water collection and preservation techniques are described in Chapter 4. A summary of site water and elutriate samples that were submitted for analytical testing is provided in Table 5-1B. Results of water column toxicological studies are addressed in Chapter 8.

7.1 SAMPLE RECEIPT



The elutriate preparation water was transported with the sediment cores from the project staging areas to EA's Ecotoxiocology Laboratory facility in Sparks, Maryland. Samples were chilled with ice during the transport period. Upon receipt, the water samples were visually inspected and the water containers were compared against the chain-of-custody record. Elutriate preparation water for the water column toxicity tests was stored in a walk-in refrigeration unit until testing, and the remaining elutriate preparation water was hand-delivered to STL-Baltimore for elutriate preparation and subsequent chemical analysis. Receiving water samples targeted for chemical analysis (sampling stations KI-3, KI-7, and KI-14 and equipment blanks) were collected in the field and hand-delivered on the day of collection to STL-Baltimore. The holding time for the receiving water samples and equipment blanks was initiated at the time of sample collection. The holding time for elutriates was initiated at the completion of the elutriate preparation process (Section 7.2). Copies of chain-of-custody forms for the site water, equipment blanks, and elutriate samples are provided in Attachment III.

7.2 ELUTRIATE PREPARATION

Elutriates were created by mixing dredging site water and sediment, allowing the mixture to settle, filtering, and testing the overlying water for dissolved constituents as per USEPA/USACE (1998) guidance. The purpose of elutriate testing is to simulate the potential mixing and release of dissolved organic or inorganic constituents into the water column during hydraulic placement of dredged material in open-water sites.

Elutriates were prepared following the Standard Elutriate Preparation specified in the ITM (USEPA/USACE 1998). A sediment/water mixture, at a 1:4 ratio of sediment-to-site water, was thoroughly mixed for 30 minutes. The mixture was then allowed to settle, and the supernatant was siphoned off and filtered to remove particulates. Elutriates were prepared using site water from each channel reach and a composite sediment sample collected from each sampling reach. In addition, although these sediments were not proposed for dredging, elutriates were prepared



Baltimore Harbor Approach Channels and Placement Site 104 Dredged Material Evaluation Draft Report from Inside Site 104, Outside Site 104, and the Ocean Reference. While these elutriates are not required by the ITM, they were prepared and analyzed for comparative purposes only. Results of the placement/reference site elutriate tests will be numerically compared to channel elutriate data. Elutriate samples that were created and tested for the Approach Channels and Placement Site 104 project are summarized in Table 7-1.

7.3 ANALYTICAL METHODS AND DETECTION LIMITS

Site waters, equipment blanks, and elutriates were analyzed for the target analytes identified in the Analytical Quality Assurance Project Plan (QAPP) (EA and STL–Baltimore 2000) which was approved by USACE and USEPA. Project-specific analytes, analytical methods and detection limits for aqueous samples are provided in Tables 5-2 and 5-5, respectively.

7.4 DATA ANALYSIS

7.4.1 Calculation of Mean Concentrations

Two sets of elutriates were prepared for each channel reach, three sets of elutriates were prepared for Inside Site 104, and a single elutriate was prepared for both Outside Site 104 and the Ocean Reference area. Mean elutriate concentrations were determined after substituting the analytical detection limit for non-detected analytes (ND = DL). If an analyte was not detected in any of the elutriates, the mean detection limit was reported and qualified with a "U" (following the approach discussed in section 7.4.2 below).

7.4.2 Concentrations of Total PCBs and Total PAHs

For each individual sample, the total PCB concentration was determined by summing the 18 summation congeners listed in Table 9-3 of the ITM, and then multiplying this sum by a factor of 2 as described in NOAA (1993). Total PAH concentrations were also determined for each sample by summing the concentrations of the individual PAHs. For both the total PCB and total PAH concentrations, three values are reported each representing the following methods for treating concentrations below the analytical detection limit:

- Non-detects = 0 (ND = 0)
- Non-detects = 1/2 of the detection limit (ND = 1/2DL)
- Non-detects = the detection limit (ND = DL)

Substituting the detection limit for all non-detects (ND = DL) provides the most conservative estimate of the non-detected concentration. This method, however, tends to produce results that are biased high, especially in datasets where the majority of samples are non-detects.

7.4.3 Comparisons to Receiving Water and WQC

Chemical results of the elutriate analyses were compared with the results of the receiving water samples (KI-3WAT, KI-7WAT), the Outside Site 104 reference water sample (KI-14WAT), and the Ocean Reference water sample. Analytes detected in the elutriates and receiving/reference waters were compared to acute and chronic aquatic life water quality criteria and to criteria for the protection of human health from consumption of aquatic organisms. Criteria were extracted from the U.S. EPA's National Recommended Water Quality Criteria (63 Fed. Reg. 68354-68364; 10 December 1998); and MDE's *proposed* water quality criteria (Maryland Register 27(17): 1628-1636; dated 25 August 2000) (Table 7-2). MDE's proposed criteria are much more extensive than the State's current listing, and the criteria for the listed compounds are identical the USEPA's criteria. Human health criteria are based on MDE's 10⁻⁵ risk-level. Elutriate metal concentrations, which were measured as *total* values, are (conservatively) compared to *dissolved* aquatic life criteria. Consistent with the technical basis of these criteria, maximum concentrations of detected constituents were compared to the acute criteria and mean concentrations were compared to chronic and human health criteria (conservatively assuming that ND=DL). USEPA's criteria values (for detected analytes only) are provided in Table 7-2.

7.5 RECEIVING WATER AND ELUTRIATE RESULTS

Results of the receiving water and elutriate analyses conducted in 1999-2000 are presented in the following subsections. Both maximum and mean results for elutriates are presented for each channel reach (Tables 7-3 through 7-27). Comparison to water quality criteria are provided in Tables 7-33A, 7-34A, and 35A. Results of equipment blank analyses are provided in Attachment III. Definitions of inorganic, organic, and dioxin/furan data qualifiers are presented in Tables 3-4, 3-5, and 6-1, respectively. Raw data are provided in Appendices I and II. Analytical narratives that include an evaluation of laboratory quality assurance/quality control results are provided in Attachment III. STL-Baltimore will retain and archive the results of these analyses for 7 years from the date of issuance of the final results.

7.5.1 Inorganic Non-Metals/Nutrients

Results of the cyanide, ammonia, nitrogen, TKN, total phosphorus, sulfide, BOD, COD, and TOC analyses are provided in Table 7-3 for receiving water. Maximum and mean elutriate concentrations are presented in Tables 7-4 and 7-5, respectively.

Cyanide was detected at the detection limit in one channel elutriate (0.005 mg/L), Tolchester North. Cyanide was not detected in the receiving waters or in any of the other channel elutriates. Concentrations of TOC in receiving water ranged from 1 to 4.3 mg/L. Outside Site 104 was below the detection limit of 2 mg/L. Mean TOC concentrations in channel elutriates ranged from 4.1 mg/L (Craighill Angle East) to 257 mg/L (Tolchester South). With the exception of Tolchester South, TOC concentrations in channel elutriates were generally less than the
concentrations reported in the elutriates from the placement/reference areas (Inside Site 104, Outside 104, and Ocean Reference).

Total sulfide in receiving water was below detection (0.35 mg/L) at Inside Site 104. Total sulfide concentrations in receiving water at Outside Site 104 and the Ocean Reference area were 0.76 and 0.93 mg/L, respectively. Tolchester South was the only channel elutriate with a detectable maximum concentration of total sulfide (0.44 mg/L); however, total sulfide was also reported in the equipment blank.

In the receiving waters, total phosphorus was only detected at the Ocean Reference area at a concentration of 0.09 mg/L (Table 7-3). Mean total phosphorus in channel elutriates ranged from 0.02 mg/L [C&D Approach (cores)] to 0.41 mg/L (Swan Point). Elutriates created with the Ocean Reference sediment had the highest TP concentration (1.30 mg/L) (Table 7-5).

Ammonia-nitrogen was detected in both receiving waters collected from within Inside Site 104 (KI-3 and KI-7) with an average concentration of 0.28 mg/L as NH₃. Ammonia-nitrogen was also detected in receiving waters from Outside Site 104 and the Ocean Reference at concentrations of 0.15 and 0.13 mg/L, respectively. Ammonia-nitrogen concentrations in the receiving waters were below both the USEPA acute and chronic saltwater criteria for aquatic life of 43 and 6.4 mg/L as NH₃, respectively (assuming 10 ppt salinity, 10°C and pH of 7.4). Ammonia-nitrogen was detected in all channel elutriate samples with maximum concentrations ranging from 1.2 to 18.1 mg/L, and mean concentrations ranging from 1.2 - 10.2 mg/L. Ammonia-nitrogen concentrations exceeded the chronic saltwater criteria for aquatic life in the Craighill Angle–East elutriate (Table 7-34A).

Total Kjeldahl nitrogen (TKN) (organic nitrogen + ammonia-nitrogen) in receiving water ranged from 0.55 mg/L at Outside Site 104 to 0.87 mg/L at Inside Site 104. Mean TKN in elutriates from the channel sites ranged from 0.65 mg/L (Craighill) to 7.25 mg/L (Craighill Angle East). Mean TKN in elutriates from the placement/reference sites fell within the range measured in the channel elutraites, with Outside Site 104 having the highest TKN concentration (4.00 mg/L) and the Ocean Reference having the lowest TKN concentration (0.80 mg/L).

7.5.2 Volatile Organic Compounds

Results of the volatile organic compound analysis in receiving water are presented in Table 7-6. Results for elutriates are provided in Table 7-7. VOCs were not detected in any of the tested receiving water samples. Of the 35 tested VOCs, dichloromethane (a commonly used laboratory chemical) was the only VOC detected in the channel elutriates. Dichloromethane concentrations (ranging from 16 to 680 μ g/L) did not exceed the human health criterion of 16,000 μ g/L (Table 7-35A).

7.5.3 Semivolatile Organic Compounds

Concentrations of the SVOC analyses in receiving waters and elutriates are presented in

Table 7-8. Maximum and mean concentrations of SVOCs in elutriates appear in Tables 7-9 and 7-10, respectively. Of the 44 SVOCs analyzed, bis(2-ethylyhexyl)phthalate was the only SVOC detected in receiving water and the channel elutriates (range of 2 to 28 μ g/L). There are no WQC for bis(2-ethylyhexyl)phthalate.

7.5.4 Chlorinated Pesticides, Organophosphorus Pesticides, and PCB Aroclors

Maximum concentrations of chlorinated pesticides, organophosphate pesticides, and PCB aroclors in receiving waters and elutriates are presented in Tables 7-11 and 7-12, respectively. Mean concentrations detected in the elutriates are presented in Table 7-13. Three of 22 tested chlorinated pesticides were detected in the receiving water from Inside Site 104 – beta-BHC, gamma-BHC, and heptachlor epoxide. Heptachlor epoxide was detected below the recommended TDL of 0.1 μ g/L (USEPA/USACE 1995). There are no published TDLs for beta-BHC and gamma-BHC.

Four of the 22 tested chlorinated pesticides were detected in the channel elutriates (beta-BHC, gamma-BHC, heptachlor, and heptachlor epoxide). Beta-BHC was detected in elutriates from each channel with the exception of C&D Approach (surficial sediments) and C&D Approach (core sediments). Maximum concentrations of beta-BHC ranged from 0.01 to 0.03 μ g/L, and were qualified as estimated. Beta-BHC was not detected in the elutriates from the placement/reference area sediments.

Gamma-BHC was detected in elutriates from all channels with the exception of Brewerton Eastern Extension, C&D Approach (core), and Craighill. The maximum concentrations of gamma-BHC ranged from 0.0092 to 0.02 μ g/L. Gamma-BHC was detected in the elutriate from Inside Site 104 reference sediments (0.01 μ g/L).

Heptachlor was detected in all channel elutriates with the exception of Craighill Upper Range. Maximum concentrations ranged from 0.03 to 0.19 μ g/L. All detected concentrations in the channels were below the USEPA/USACE (1995) TDL of 0.1 μ g/L, with the exception of Craighill Angle–West (0.19 μ g/L). Heptachlor was also detected in elutriates created with sediments from Outside Site 104 and Inside Site 104.

Heptachlor epoxide was detected in 11 of the 13 channel elutriates with a concentration range of 0.02 to 0.04 μ g/L. All detected concentrations were below the recommended TDL of 0.1 μ g/L (USEPA/USACE 1995).

PCB aroclors were not detected in receiving water samples from the placement/reference sites (Table 7-11), or in channel elutriates (Table 7-12).



7.5.5 PCB Congeners

Results of the PCB congener analysis in receiving water are presented in Table 7-14. Results for maximum and mean concentrations in channel elutriates are provided in Tables 7-15 and 7-16, respectively. Only 1 out of 26 tested PCB congeners was detected in receiving water, and 5 out of 26 tested PCB congeners were detected in channel elutriates. Note that the Total PCB results are affected substantially by the way measurements less than the Detection Limits are handled. As shown in Table 7-15 for example, no PCBs were detected at Tolchester South, but the Total PCB values assuming ND=0, ND-1/2DL, and ND=DL are 0.0, 0.0525 and 0.1050 μ g/L, respectively.

PCB congener #8 was the only congener detected from receiving waters from Inside Site 104 and this congener was also detected in the method blank. Total PCB (assuming ND = $\frac{1}{2}DL$) ranged from 0.02 to 0.03 µg/L in receiving waters (Table 7-14).

PCB congeners #8 and #18 were the most commonly detected congeners in the channel elutriates. Mean total PCB in channel elutriates (assuming ND = $\frac{1}{2}$ DL) ranged from 0.0525 µg/L (Craighill, Tolchester South) to 0.0716 µg/L (Tolchester North), compared to a range of 0.105 to 0.118 µg/L for mean Total PCB (assuming ND = DL). Mean Total PCBs for ND=0 ranged from 0 (zero) to 0.0253 µg/L. All detected congeners were measured at or below the recommended TDL of 0.01 µg/L for individual congeners (USEPA/USACE 1995). Total PCB concentrations exceeded the criteria for human health consumption of aquatic organisms in all channel and placement/reference site elutriates, even in reaches where no congeners were detected (ND=1/2DL; ND=DL) (Table 7-35A). These exceedances are largely driven by the substitution of the detection limits for non-detected congeners.

7.5.6 Polynuclear Aromatic Hydrocarbons

Results of the PAH analysis in receiving water are presented in Table 7-17. Results for elutriate maximum and mean concentrations in the channel elutriates are presented in Tables 7-18 and 7-19, respectively.

No PAHs were detected in any of the receiving water samples. Fluorene was the only PAH detected in channel elutriates [Craighill Angle East $(0.08 \ \mu g/L)$]. Fluorene and pyrene were detected in the Inside Site 104 elutriate, both at a concentration of $0.07 \ \mu g/L$. The florene concentration in the Craighill Angle elutriate did not exceed the human health criterion of 14,000 $\mu g/L$ (Table 7-35A). There are no WQC for pyrene.

7.5.7 Metals

Results of the metal analyses in receiving water are provided in Table 7-20. Six of 16 tested metals were detected in the receiving waters. Aluminum and iron were detected in all four

receiving water samples with concentrations ranging from 159 to 296 μ g/L and 28.4 to 289 μ g/L for aluminum and iron, respectively. Copper and manganese were detected in 3 out of 4 receiving water samples (75%) ranging from 0.93 to 6.7 μ g/L and 2.4 to 12.4 μ g/L, respectively. Arsenic and nickel were detected in receiving water from Outside Site 104 with concentrations of 2.5 and 2.1 μ g/L, respectively.

Maximum and mean concentrations for the metal analyses in channel elutriates are presented in Tables 7-21 and 7-22, respectively. Thirteen of the sixteen target metals were detected in the channel elutriate samples. Aluminum, arsenic, manganese, and nickel were detected in elutriates water from all 13 channels. Beryllium, lead, and silver were detected in the elutriate from Tolchester Straightening only.

Copper in the C&D Approach (surficial) was the only metal that exceeded acute (4.8 μ g/L) and chronic (3.1 μ g/L) water quality criterion (Table 7-33A). Two metals (arsenic and manganese) exceeded the human health criterion (Table 7-35A) in every channel elutriate.

7.5.8 Butyltins

Results of the butyltin analyses in receiving water are provided in Table 7-23. Tributyltin was detected in the Ocean Reference water with a estimated concentration of 80 ng/L. Maximum and mean concentrations of butyltins in channel elutriates are presented in Tables 7-24 and 7-25, respectively. Tributyltin was also detected in elutriate at the Ocean Reference (100 ng/L). Monobutyltin was detected in one channel elutriate water (Craighill Entrance) with a maximum concentration of 130 ng/L.

7.5.9 Dioxin and Furan Congeners

Results of the dioxin and furan congener analyses in receiving waters are provided in Table 7-26. Five of the 17 tested dioxin/furan congeners were detected in the receiving water samples (1,2,3,4,6,7,8-HpCDD, OCDD, 1,2,3,7,8-PeCDD, 1,2,3,7,8-PeCDF, and OCDF). The detected congeners are these with the highest toxicity. OCDD was also detected in the blank. The dioxin TEQ (at ND = 0) in receiving water ranged from 0 to 0.00059 ng/L while the dioxin TEQ (at ND = 1/2DL) in receiving water ranged from 0.00184 to 0.00279 ng/L. It is noteworthy that 2,3,7,8-TCDD, the most toxic congener which is specifically regulated by U.S. EPA's ambient water quality criteria (O'Hanian 2000), was never detected in receiving water samples from the placement/reference areas (Table 7-26).

Results of the dioxin and furan congener analyses in elutriates are provided in Table 7-27. Four out of 17 tested dioxin and furan congeners were detected in channel elutriates and many of these were detected in the laboratory blank ("B-qualified"). OCDD, the least toxic congener was the most frequently detected congener. Although the three reference sites will not be used for dredging, and elutriate testing is not required by USEPA's ITM, elutriates were prepared and analyzed for information purposes. It is noteworthy that concentrations in these elutriate samples are not different from the proposed dredging sites. Dioxin TEQ (ND = 0) in channel elutriates

ranged from 0.0 to 0.00014 ng/L while dioxin TEQ (ND = 1/2DL) in channel elutriates ranged from 0.00219 to 0.00412 ng/L. It is noteworthy that 2,3,7,8-TCDD, which is the congener specifically regulated by USEPA's ambient water quality criteria (O'Hanian 2000), was never detected in elutriates prepared from the placement/reference areas or from the channel elutriates (Table 7-27). Dioxins are hydrophobic, not easily dissolved in water, and elevated concentrations would not be expected in receiving water or elutriate samples.

7.5.10 Summary of Elutriate Chemistry

Overall, 16 of 192 (8.3%) of the tested organic and inorganic constituents were detected in the receiving water samples (Table 7-28). Twenty-nine of 192 (15.1%) of the tested organic and inorganic constituents were detected in the channel elutriates. Twenty-three of 192 (12%) of the tested organic were detected in elutriates created with sediments from the placement/reference areas. Metals were the most frequently detected constituents in both the receiving water and channel elutriates.

7.6 WATER QUALITY MODELING

7.6.1 Description of the STFATE Model

The dispersion of elutriate into the ambient water following the release of dredged material from a barge was modeled to evaluate compliance of the elutriate with applicable water quality criteria. The elutriate simulates the pore water within the dredged material. The modeling of the elutriate fraction of the dredged material was performed for both Site 104 and for the Norfolk Ocean Disposal Site (NODS).

The elutriate was modeled using STFATE, which is a standard USACE model used for computing the fate of material placed from either a split-hull barge or a hopper dredge (Johnson and Fong 1995). The model computes the movement of the material from the moment it is injected into the water column until the material is either deposited on the seafloor or transported out of the numerical grid. The model simulates both the solid fractions (e.g., clumps, sand, silt, and clay) and the elutriate that is present in the barge. The computations consist of three phases. The first phase is the convective descent of the dredged material cloud through the water column; during this time the cloud grows as a result of the entrainment of ambient water. Phase 2 occurs when the descending cloud of dredged material strikes the seafloor with a dynamic bottom collapse resulting in a radially expanding bottom surge. The bottom collapse phase continues until the estimated rate of spreading resulting from turbulent diffusion exceeds the rate of spreading of the cloud collapse. Phase 3 occurs at the end of the collapse phase when all remaining material is subjected to the passive transport, diffusion, and settling of the suspended material. This passive transport phase continues until the suspended material deposits on the seafloor or is transported out of the numerical grid. Basic output from the model consists of information on the amount of material in suspension and the footprint of the deposited material on the seafloor.

Results of the STFATE modeling described in the following sections were used to develop tables of dilution factors as a function of distance and time following the disposal event. An examination of the compliance with water quality criteria, taking into account available site dilution, is provided in Section 7.7. The solids fraction of the dredged material (sand, silt, and clay) was included in the model to provide an appropriate composition in the barge. The material fractions used in the model are discussed in Section 7.6.2. STFATE models both the elutriate and its chemical constituents as a conservative tracer that remains in the water column. During an actual barge disposal, a fraction of the elutriate remains as pore water trapped in clumps, which fall to the bottom, and therefore does not become available to the water column. Therefore the STFATE model over-estimates (by an unknown degree) the actual concentration of constituents in the water column. An additional fraction of the elutriate is buried along with the solid materials and is not available to the water column. The availability of elutriate in the water column is discussed in Section 7.6.3 along with other modeling procedures. STFATE modeling results for both a split hull barge and for the hydraulic pumping out of a barge are presented in Section 7.6.4

7.6.2 Site and Material Parameters

The STFATE model requires a number of site parameters including current velocities, vertical water column density gradients, barge dimensions, and material fractions in the barge. The dimensions used for a standard barge were a 170-ft length, 53-ft width, 18-ft loaded draft and 3,000 cubic yard (cy) capacity. The capacity was reduced from a maximum of 4,000 cy to ensure no spillage during clamshell dredging operations.

7.6.2.1 Site 104

Mean lower low water (MLLW) depths within Site 104 vary between 42 and 78 ft. Dredged material placement is not proposed in the northern one-third of the site where depths are less than 45 feet. A buffer zone has also been proposed for the southern 1,200 ft of the site where depths exceed 70 ft. The remaining site lies mainly within the 48-ft to 52-ft MLLW depth contours. A 50-ft depth was used for the STFATE model simulations. This is conservative since deeper depths provide greater mixing potential and consequently result in reduced ambient concentrations.

Allowed dispersion areas for dredged material disposal sites typically have dimensions associated with travel times on the order of several hours. For example, the allowed dispersion area for ocean disposal is defined as a 4-hour travel time following discharge [40 CFR 227.29(a)] and USEPA/USACE 1991). In addition, USEPA's ambient water quality criteria to protect against chronic effects are based on 4-day average exposure concentrations. Because of the time scales associated with attaining water quality criteria, a tidally averaged current velocity was selected for use in the STFATE model.

Site-specific data on current velocities were obtained during a monitoring program conducted by SAIC during summer 1999 [Science Applications International Corporation (SAIC) 1999, Appendix C]. The field program included the deployment of bottom mounted instrumentation at

four stations in Site 104. Of these four stations, Station 1 was at the northern end of the site in an area where disposal will not take place, and the data retrieval at Stations 2 and 3 was only for one deployment period. The average ebb and flood tide velocities for Station 4 at a 1.5-m height from the bottom are presented in the following table during three deployment periods (SAIC 1999, Appendix C).

| | Average Velocity (cm/sec) | | | | | | | | | | |
|------------------|---------------------------|------------|--|--|--|--|--|--|--|--|--|
| Date | Ebb Tide | Flood Tide | | | | | | | | | |
| 14-18 June | 21.3 | 34.7 | | | | | | | | | |
| 13-15 July | 19.5 | 30.6 | | | | | | | | | |
| 20-27-July | 14.2 | 20.6 | | | | | | | | | |
| Weighted Average | 17.1 | 26.4 | | | | | | | | | |

An average velocity is also included in the above table that was calculated by weighting the velocities by the number of observations in each deployment period. The resulting average flood tide velocity of 26.4 cm/sec (0.866 ft/sec) measured 1.5 m above the bottom exceeds the 17.1 cm/sec (0.560 ft/sec) ebb tide velocity. These velocities are consistent with the known net upstream bottom flow in the Chesapeake Bay that is induced by density driven circulation. Near the surface of the water column, the ebb tide velocity would exceed the flood tide velocity.

At Site 104, water density increases with depth because of the more saline bottom water. The vertical density difference is typically 0.007 gm/cc (SAIC 1999, Figure 5-1.1). In the STFATE model the surface density was assumed to be 1.005 gm/cc and the bottom density 1.012 gm/cc. The density of the elutriate in the dredged material, collected from a bottom depth shallower than Site 104, was assumed to be 1.011 gm/cc.

7.6.2.2 Norfolk Ocean Disposal Site

Conditions at the NODS were based on information contained in the Environmental Impact Statement (EIS) that was prepared by the USEPA Region III (1992). The site has a radius of 4 nautical miles and depths vary from 43 to 85 ft. Depths near the center of the site vary between 65 and 80 ft. A representative depth of 70 ft was used in the STFATE analysis.

The EIS reported current velocities to be typically 10 cm/sec (0.33 ft/sec) during the winter. During summer, surface velocities are lower and near bottom velocities approach 2 cm/sec (0.07 ft/sec). The 0.33-ft/sec winter current velocity was used in STFATE, appropriate for the allowed October to March dredging period.

The water density at the NODS was calculated from salinity and temperature. The EIS indicated that the ocean site is at the offshore edge of the lower salinity outflow that discharges from the mouth of the Chesapeake Bay with surface salinity values of approximately 25 ppt. Density profiles were calculated as a function of temperature assuming that salinity varies from 25 ppt at the surface to 30 ppt at the bottom.

| | Water Density (gm/cc) | | | | | | | | | | | | | |
|---------|-----------------------|--------|--------|--------|--|--|--|--|--|--|--|--|--|--|
| Depth | Salinity (ppt) | 5° C | 10° C | 20° C | | | | | | | | | | |
| Surface | 25 | 1.0205 | 1.0199 | 1.0179 | | | | | | | | | | |
| Bottom | 30 | 1.0247 | 1.0240 | 1.0219 | | | | | | | | | | |

The elutriate density of the dredged material in STFATE was assumed to be 1.011 gm/cc. The above table indicates that the water density variation between the dredging site and the ocean disposal area is controlled more by the general freshwater/ocean salinity difference than by vertical salinity variation or seasonal changes in temperature. For STFATE, the density profile at the NODS was assumed to increase from 1.020 gm/cc at the surface to 1.024 gm/cc at the bottom.

7.6.2.3 Material Properties

The material fractions used in the STFATE model for both Site 104 and the NODS were based on the physical characteristics data provided in Chapter 6. Table 6-2 provides physical data at 13 proposed dredging locations within the Baltimore Harbor approach channels. Sediment characteristics based on an average of the 13 locations are summarized below.

| Material | Sediment Fraction (% dry weight) | Sediment Fraction (% volume) |
|----------|-------------------------------------|---------------------------------|
| Sand | 12.3 | 2.21 |
| Silt | 47.1 | 8.48 |
| Clay | 40.6 | 7.31 |
| Water | | 82.0 |

The average sediment sample had a moisture content of 61.7% and a specific gravity of 1.33. The moisture content and the specific gravity indicate that the average sediment sample was 82% water (1.33×0.617) and 18% solids by volume.

7.6.3 Modeling Procedures

A numerical grid consisting of 45 rows and 29 columns was used for the STFATE model calculations. At Site 104, a 400-ft spacing was used in the longitudinal direction (between rows) and a 200-ft spacing in the transverse direction (between columns). The discharge location was placed in row 3, which provided 40 active cells (16,000 ft) in the longitudinal direction. Model calculations are not performed along the outer two rows and columns. The active width of the model grid was 4,800 ft. For the average ebb tide scenario (0.56 ft/sec), a 5-hour model simulation could be accommodated. However, for the higher velocity flood tide scenario (0.866 ft/sec), only a 4.5-hour simulation was possible before a portion of the released material was lost from the model grid.

At the NODS, the lower 0.33 ft/sec tidal velocity allowed a smaller 250-ft X 150 ft grid spacing to be used, while still accommodating a 5-hour model simulation.

STFATE predicts concentrations throughout the water column at each location on the model grid. In addition to the user-specified depths, the model also provides the maximum water column concentration at each grid location. A composite of the maximum water column concentrations at each time-step in the model simulation was used to develop summary tables of dilution as a function of distance. Within the modeled plume, the maximum concentration occurs along the plume centerline and concentrations decrease in the transverse direction away from the centerline following a Gaussian distribution. For compliance with water quality criteria, a plume average concentration has been allowed under Maryland regulations. An appropriate plume width to use with STFATE was determined by comparison with other MDE approved models. USEPA's CORMIX model is commonly used for National Pollutant Discharge Elimination System (NPDES) permitting and it also assumes a Gaussian distribution. In CORMIX, a plume width is defined as 1.414 standard deviations. This definition of plume width was adopted for application with the STFATE model. Integration under a Gaussian curve for 1.414 standard deviations on either side of the plume centerline indicates that a plume average concentration is 74.7 percent of the centerline value. Plume average concentrations are used in Section 7.6.4 when presenting results of the STFATE model.

The STFATE model was used to evaluate the release from both a split hull barge (Site 104, Ocean Disposal Site) and the hydraulic pumping out of a barge (Site 104 only). For the split hull release, the 3,000 cy barge capacity was divided into two layers with a total release time of 60 seconds. The hydraulic emptying of the barge is performed by pumping approximately 2 parts water to 1 part dredged material. The pumped material is released near bottom, within approximately six pipe diameters. A deflector plate is placed at the end of the pipe to disperse the material and prevent bottom scour. The hydraulic placement of the dredged material from a barge was assumed to take 45 minutes. The 3,000 cy barge load was divided into six 500 cy layers and released at uniform intervals during the 45-minute period. In STFATE, the material released from a barge quickly falls to the bottom with very little water column interaction. The elutriate then mixes up into the water column from a near-bottom location. The initial placement of the dredge material at a near bottom location by hydraulic pumping is, therefore, very similar to what takes place in STFATE.

Section 7.6.2 indicated that 82 percent of the barge volume consists of water. STFATE models the elutriate fraction as a conservative tracer with a non-varying mass in the water column. There are two potential mechanisms that reduce the available elutriate: (1) a portion of the dredged material in the barge consists of clumps that fall to the bottom without releasing the trapped pore water to the water column, and (2) a portion of the elutriate will be buried along with the dredged material as pore water as it settles to the bottom.

To examine what fraction of the dredged material may exist as clumps, the moisture content of sediment samples was compared to the liquid limit, one of the Atterberg limits. The moisture content and liquid limit were available for 48 sediment samples collected in Baltimore Harbor approach channels (EA 2000c, Table 4-12). A procedure for estimating the occurrence of

,

clumps was provided by USACE-WES (Paul Schroeder, personal communication). Using the WES method, the occurrence of clumps is assumed to be 100% if the moisture content is less than the liquid limit. The occurrence of clumps is assumed to be 0% if the moisture content is greater than 180% of the liquid limit. Between the liquid limit and the 180% value, the clumping fraction is linearly interpolated between 100% and 0%. Applying this technique to the 48 sediment samples resulted in a 52% occurrence of 0% clumps and a 29% occurrence of 20-40% clumps. The average occurrence of clumps in the sediment samples was 18%. If 18% of the dredged material formed into clumps, only 82% of the elutriate contained in the barge would be available to the water column.

An estimation of what portion of the elutriate from the barge is buried along with the dredged material as it settles to the bottom is illustrated in Table 7-29. The STFATE model output indicates that within 30 minutes, 12,764 ft³ of material, 87.5 percent of solids originally in the barge, was deposited on the seafloor. If this material settled at the same compactness (moisture content) as the original dredged material, 58,145 ft³ of water would be required to fill the voids between particles. The average (bulk) dilution of the elutriate from the barge was estimated by examining the dimensions of the elutriate cloud 10, 20, and 30 minutes after disposal. For example, after 10 minutes the near bottom cloud had a diameter of 211.4 ft and a thickness of 6.27 ft, resulting is a volume of 220,073 ft³. This represents an average bulk dilution of 3.3 from the original 66,420 ft³ elutriate volume in the barge. Applying bulk dilution factors 10, 20, and 30 minutes after disposal allowed the determination of the fraction of the pore water in the settled material that was obtained from the original elutriate. Table 7-29 indicates that between 10 and 30 minutes the fraction of elutriate from the barge that is buried along with the dredged material increases from 17.1 to 22.0 percent.

The settling of suspended material to the seafloor occurs with a larger void space than was present in the originally dredged sediments. Modeling performed by USACE to examine placement of dredged material at Site 104 assumed a bulking factor of 1.49 for placed dredged material immediately after placement (Johnson et al. 1999, Chapter 9). This means that the amount of water buried with the dredged material exceeds the amount originally present in the sediment by approximately 49 percent. The estimation of the amount of original elutriate buried with the dredged material in the previous paragraph assumed that the void space remains unchanged (bulking factor = 1.0). The following table provides the percent of original elutriate buried buried during the first 30 minutes for bulking factors of 1.0, 1.2, and 1.4.

| Minutes | | Bulking Factor | |
|---------|------|----------------|------|
| Γ | 1.0 | 1.2 | 1.4 |
| 10 | 17.1 | 20.5 | 28.7 |
| 20 | 20.9 | 25.1 | 35.1 |
| 30 | 22.0 | 26.4 | 37.0 |

Original Elutriate (%) Buried with Dredged Material

This analysis indicates that typically 20 to 30% of the elutriate will be buried with the dredged material, and will not be available to the water column. In addition, 5-18% of the dredged

material may be present as clumps, which would remove a similar portion of elutriate from the water column. In combination, these two mechanism remove 25-48% of the elutriate from the water column. Therefore, only 52-75% of the original elutriate will be present in the water column and subject to dispersive processes. The STFATE model was executed for three cases assuming that 50, 75, and 100% of the elutriate was available to the water column. To represent these cases in the model, the volume fraction of clumps was increased to offset the decrease in the elutriate fraction. The elutriate volume fraction was assumed to be 41.0%, 61.5% and 82.0% for the 50%, 75%, and 100% elutriate availability cases.

7.6.4 Discussion of Model Results

STFATE modeling results for the mixing of elutriate into the water column are provided in Tables 7.30 and 7.31 for Site 104 and Table 7-32 for the NODS. At Site 104, Table 7.30 provides for dredged material placement from a split hull barge, and Table 7-31 for the emptying of a barge by hydraulic pumping. Each Site 104 table contains results for an average ebb tide (0.56 ft/sec) and for an average flood tide (0.866 ft/sec) near-bottom current velocity. For the NODS, Table 7-32 provides results for the placement of material from a split hull barge at an average 0.33 ft/sec tidal velocity. The STFATE model result tables provide the distances and times following disposal for the elutriate plume to achieve dilution factors ranging from 50 to 500, and for the dilution achieved 1.0 hour and 4.0 hours following the initial release. For each tidal velocity scenario, the STFATE model was executed for three cases corresponding to 50%, 75%, and 100% elutriate availability to the water column. In Section 7.3.3, the fraction of elutriate bound within clumps and reburied as the solid material settled to the seafloor was examined within the initial 30 minutes after release. Based on this discussion, the 75% elutriate availability scenario was considered the most appropriate for use in determinations of compliance with ambient water quality criteria in Section 7.7.

The longitudinal distribution of the elutriate plume at Site 104 is illustrated in Figure 7-1 at 1-hour intervals after release of dredged material from a barge. The figure sequence displays the maximum water column concentration (as well as the concentration and distance distribution) within the elutriate plume for an average ebb tide scenario. In the figure, the y-axis displays the ratio of the water column concentration to the elutriate concentration in the barge (C_0), and the x-axis shows the distribution of the modeled plume at 1-, 2-, 3- and 4-hour intervals.

An examination of the Site 104 ebb and flood tide scenarios in Table 7-30 indicates that a similar dilution is obtained at a similar time following release. However, since the average flood tide velocity of 0.866 ft/sec is approximately 50% greater than the 0.56 ft/sec ebb tide velocity, the dilutions on a flood tide are obtained at an approximately 50% greater distance. For example, a dilution factor of 50 is achieved at a 2,443-ft distance for the ebb tide scenario (75% elutriate availability) and at a 3,477-ft distance for the flood tide scenario. The times associated with a dilution factor of 50 were very similar, 1.21 hours during ebb tide, and 1.12 hours during flood tide. The predominate ebb-flood current direction is aligned with the longer north-south dimension of Site 104. Therefore, the northern one-third of the site, which will not be used for disposal because MLLW depths are less than 45 ft, will act to contain the longer flood tide plume dimensions.

A comparison of dilution results between the three elutriate availability scenarios at Site 104 can be made by examining the ebb tide results in Table 7-30. For a dilution factor of 100, the water column distance increases from 2,863 ft to 3,312 ft, and to 3,712 ft for elutriate availability of 50%, 75%, and 100%, respectively.

STFATE modeling results for emptying a barge by hydraulic pumping at Site 104 are provided in Table 7-31. By increasing the disposal time from 60 seconds to 45 minutes, the instantaneous source strength is reduced, resulting in lower near-field concentrations and increased dilution factors at a given distance. For the ebb tide scenario (75% elutriate availability), the distance to the 100 fold dilution decreased from 3,312 ft using a split hull barge to 2,140 ft for hydraulic pumping. With hydraulic pumping, the distance to achieve similar dilutions did not increase as much between the ebb and flood tide scenarios as was observed with the split hull barge. At a dilution factor of 100, the flood tide scenario had an 11% longer distance, and at a dilution factor of 300, the flood tide scenario had a 20 percent longer distance than the ebb tide scenario.

STFATE modeling results of the mixing of elutriate into the water column at the NODS due to the placement of dredged material from a split hull barge are provided in Table 7-32. Because of the lower 0.33 ft/sec tidal velocity at the ocean site, a given dilution factor occurs at shorter distance than at Site 104. At the ocean site (75% elutriate availability) dilution factors of 50 and 500 occur at distances of 1,296 and 3,707 ft, while at Site 104 these dilution factors occur at distances of 2,443 and 6,588 ft (75% elutriate availability, ebb tide). At the ocean site, dilution factors occur at distances of 1.104 theorem and 4-hours following the disposal are 43 and 1,000, respectively, slightly greater than at Site 104.

7.7 DISCUSSION AND TIER II WATER QUALITY EVALUATION

The ITM (USEPA/USACE 1998) states that after consideration of the Tier II water quality data, one of two possible conclusions is reached regarding the potential water column impact of the proposed dredged material:

- 1) The available water quality requirements are met. Further information on water column toxicity must be evaluated in Tier III when there are contaminants of concern for which applicable water quality criteria or standards are not available or where interactive effects are of concern.
- 2) Concentrations of one or more of the dissolved contaminants of concern, after allowance for mixing, exceeds applicable water quality criteria or standards beyond the boundaries of the mixing zone. In this case, the proposed discharge of dredged material does not comply with the water quality criteria or standards.

Although a mixing zone will not be issued for open-water placement, the discussion which follows supports the conclusion that adverse effects in the water column are not expected based upon the data set collected to date.

7.7.1 Evaluation Approach

As discussed in Section 7.5 above, extensive analytical characterizations of elutriate samples from 13 different approach channels (plus 3 placement/reference areas) were conducted. The general approach used to evaluate these data is as follows:

- The analytical data for each of the approach channels and reference areas are tabularized.
- Elutriate analytes which were "detected" are compared to applicable ambient water quality criteria for the protection of saltwater aquatic life and human health from the consumption of contaminated organisms.
- Where detected concentrations exceeded applicable criteria, STFATE modeling results are used to estimate the time and distance required to reduce these elutriate concentrations to comply with ambient water quality criteria values.

This is a complex evaluation because of the large number of analytes, the different tidal conditions, alternate approaches to address concentrations which are less than the analytical detection limit (U-qualified), different release scenarios (hydraulic placement vs. split hull barge release), etc. For this reason, the discussion below is focused on the following assumptions and scenarios, that are believed to be reasonably conservative (i.e., overestimate any potential environmental effects):

- STFATE modeling results are based on a split hull barge release using average ebb tide velocities and 75% elutriate availability (Section 7.6).
- The highest detected concentration in the data set for each channel reach is compared to the saltwater acute (1-hour average) ambient water quality criteria; and the mean concentration is used to compare to the saltwater chronic (4-day average) ambient water quality criteria values.
- The ambient water quality criteria used for this assessment are USEPA's National Recommended Water Quality Criteria (63 Fed. Reg. 68354-68364; 10 December 1998); and MDE's *proposed* water quality criteria (Maryland Register 27(17): 1628-1636; dated 25 August 2000) (Table 7-2). MDE's proposed criteria are much more extensive than the State's current listing, and the criteria for the listed compounds are identical the USEPA's criteria. Human health criteria are based on MDE's 10⁻⁵ risk-level. Elutriate metal concentrations, which were measured as *total* values, are (conservatively) compared to *dissolved* aquatic life criteria.
- Tables 7-33B, 7-34B and 7-35B calculate mixing factors and compliance distances assuming that analytes reported to be not detected (ND) are equal the detection limit
- (ND = DL). This is a very conservative assumption, since it assumes that all undetected analytes are actually present in concentrations just barely under their detection limits,

which is highly unlikely. Note that the values for Total PCBs (which were rarely detected) are also presented as $ND = \frac{1}{2} DL$ and ND = 0 to put these measurements in perspective.

7.7.2 Discussion of the Elutriate Data

Few of the more than 150 target analytes evaluated as part of the approved QAPP were detected in the elutriate samples. The results from these analyses are presented in Tables 7-33 through 7-35. When these detected analytes are compared to applicable ambient water quality criteria for the protection of saltwater aquatic life and human health, a much smaller list of chemicals results (Table 7-36).

The Code of Federal Regulations [40 CFR 227.29(a)] and the USEPA/USACE (1991, Section 5) ocean disposal manual for dredged material allow a dispersion area defined by the disposal site boundary, or within the site after the 4-hour initial mixing period. Compliance with ambient water quality criteria is evaluated at the edge of that mixing zone area. Similar guidance is not presented in the USEPA/USACE (1998) Inland Testing Manual, although it does state clearly that "the discharge of dredged material cannot cause the water quality standards to be exceeded outside the mixing zone..." (p. 55, emphasis added).

For discussion purposes, the 13 approach channels and 3 reference areas are divided into two groups: (1) reaches where all ambient water quality criteria are predicted to be met within <u>one hour</u> of release (Section 7.7.3), and (2) channel reaches which have one or more analytes that would require more than one hour to comply with criteria (Section 7.7.4).

7.7.3 Channel Reaches Meeting Ambient Criteria Within One Hour for Placement Within

Site 104

There were only 12 chemicals detected in the full strength (undiluted) channel elutriate that exceeded aquatic life or human health criteria based simply on numerical comparisons (ignoring critical evaluation factors such as exposure duration and dilution). These chemicals are ammonia, arsenic, beryllium, copper, cyanide, heptachlor, heptachlor epoxide, manganese, total PCBs, silver, sulfide, and tributyltin. As shown in Table 7-36, elutriates from the following reaches would be in compliance with all ambient acute, chronic and human health criteria in less than one hour of release (assuming ND=0 for PCBs):

- C&D Approach Channels (cores)
- Craighill
- Craighill Angle East
- Craighill Entrance
- Craighill Upper Range
- Outside Site 104
- Swan Point
- Tolchester Channel- North

Tolchester- Straightening

Given (1) that the releases at the site will not be on a continuous basis at a given location; (2) the exposure assumptions that are the basis of ambient water quality criteria for the protection of aquatic life and human health are not met; and (3) that all of the elutriate analytes will be in compliance with ambient water quality criteria for the protection of aquatic life and human health within one hour of release, no unacceptable adverse effects to the water column would be expected from proper management of the materials.

7.7.4 Channel Reaches Requiring Greater Than One Hour to Meet Criteria

Only three analytes detected in elutriates that would require *more* than one hour to achieve compliance with ambient water quality criteria:

- USEPA's chronic aquatic life criterion for hydrogen sulfide,
- USEPA's human health criterion for manganese, and
- USEPA's and MDE's proposed human health criterion for heptachlor.

As shown in Table 7-36, elutriates from the following five channel reaches (and two reference areas) would require more than one hour to be in compliance with all ambient water quality criteria. It is also noteworthy that only one chemical in each reach requires more than one hour for compliance.

7.7.4.1. Manganese in Elutriate Samples From Brewerton Eastern Extension, C&D Approach Channels (surficial), and Cutoff Angle

Elutriates from these three channel reaches require more than one hour to comply with. USEPA's human health criterion for manganese. STFATE modeling indicates that the time for compliance is 1.39, 1.12 and 1.28 hours respectively for these three reaches with a maximum distance of 2,799 ft under ebb flow conditions (Table 7-36). There are several important points regarding the manganese criterion, and whether these concentrations could have any potential effect on receiving waters.

- Manganese is a naturally occurring trace element, "is a vital micronutrient for both plants and animals," and causes health problems if not present in sufficient quantities (U.S. EPA 1993, p. 157). Manganese was detected in elutriates from every channel reach and control area evaluated in this program, as would be expected based on U.S. EPA's determination that the mean natural "background concentration" in U.S. soils (from which sediments are derived) was 348,000 µg/kg dry weight (USEPA 2000, Exhibit 5-1).
- U.S. EPA's human health criterion for manganese is 100 µg/L, was published in the Agency's (1976) Red Book, and was intended "to protect against a possible human health hazard to humans by manganese accumulation in shellfish" (U.S. EPA 1993a, p. 158). EPA's criteria document states that the [normal] average human intake is approximately 10 mg per day, but that "very large doses of manganese can cause some

diseases and liver damage, but these are not known to occur in the United States. Only a few manganese toxicity problems have been found throughout the world, and these have occurred under unique circumstances (i.e., a [drinking] well in Japan near a deposit of buried batteries)" (U.S. EPA 1993a, p. 157).

- The transient and short duration of manganese in the water column will not allow mollusks to achieve a steady-state bioaccumulation factor, and therefore they will not accumulate manganese to high concentrations in their edible tissues. Within less than two hours of release, manganese concentrations are expected to be below U.S. EPA's 100 µg/L criterion, and the site can be managed such that frequent exposures are unlikely.
- The manganese human health criterion is based on humans consuming contaminated aquatic life on a regular basis for a long duration (e.g., 6.5 grams per day for a 70 year lifetime). Because there will not be a continuous release at a given location, and the proposed site will not be operational for more than 10 years, it is clear that no one could consume shellfish collected from the site on a daily basis throughout their lifetime.

In conclusion, the assumptions upon which USEPA's manganese criterion is based are not consistent with the water column exposure, that will occur at the proposed dredged material placement site. Furthermore, manganese is not currently regulated by MDE, nor is it contained in MDE's proposed criteria changes (Maryland Register 27(17): 1628-1636; dated 25 August 2000).

7.7.4.2 Sulfide in Elutriate Samples from Inside Site 104, Tolchester Channel- South, and the Ocean Reference Site

Total sulfide was only detected in elutriates at three locations, and there is some question about the accuracy of the analytical results. More specifically at Inside Site 104, sulfide was only determined in 1 of 3 samples and that sample was J-qualified; at Tolchester Channel, it was detected in only 1 of 2 samples (B-qualified); and at the Ocean Reference Site sulfide was determined in the one sample (B-qualified).

Assuming the total sulfide data are accurate, elutriates from each of these locations (one channel reach and two placement/reference areas) would require more than one hour to comply with U.S. EPA's chronic aquatic life criterion for hydrogen sulfide. STFATE modeling indicates that the time for compliance would be 3.65 hours for Inside Site 104, and 2.26 hours for Tolchester South, and 2.5 hours for the Ocean Reference (Table 7-36). There are several important points regarding the chemistry of hydrogen sulfide that influence its potential effect on saltwater aquatic life.

• U.S. EPA's (1998) aquatic life chronic criterion of 2.0 μ g/L is based upon the concentration of "*undissociated hydrogen sulfide* (H₂S)", whereas the analytical parameter measured in this program was <u>total sulfide</u>.

- Hydrogen sulfide is a naturally occurring compound that is commonly found in anaerobic aquatic sediments as an anaerobic degradation product of both inorganic and organic sulfur compounds. Under normal environmental conditions, only a small portion of the total sulfide would be expected to be in the form of H₂S. Further, when sediments are disturbed during dredging and release operations, H₂S will be quickly oxidized to yield sulfate compounds (SO₄). As noted by U.S. EPA (1993), "the fact that H₂S is oxidized in well-aerated water by natural biological systems to sulfates or is biologically oxidized to elemental sulfur has caused investigators to minimize the toxic effects of H₂S on fish and other aquatic life."
- Two of the three stations where elutriate concentrations exceed EPA chronic criteria are
 placement or reference areas (Inside Site 104 and the Ocean Reference Site).
- U.S. EPA's chronic aquatic life criterion is based on a 4-day average exposure period which is inconsistent with the environmental chemistry of hydrogen sulfide under the dredged material release and organism exposure scenarios expected at Site 104.

In conclusion, the assumptions embedded within the U.S. EPA's hydrogen sulfide criterion are not consistent with the water column exposure that will occur at the proposed placement site. Further, hydrogen sulfide is not currently regulated by MDE, nor is it contained in MDE's proposed criteria changes (Maryland Register 27(17): 1628-1636; dated 25 August 2000).

7.7.4.3 Heptachlor in Elutriate Samples from Craighill Angle West

Heptachlor was the only analyte in elutriate samples from Craighill Angle West that would require more than one hour to comply with USEPA's criteria. STFATE modeling indicates that the time for compliance with the heptachlor criterion is 1.21 hours and would be achieved within 2,279 feet during the average ebb tide (Table 7-36).

Heptachlor (and heptachlor epoxide) are manmade pesticides that were used to control termites and other insects. In late 1978, most uses were phased out, and the chemical is no longer available to the general public. It is, however, strongly adsorbed to soil and extremely resistant to biodegradation (USEPA's Website; www.epa.gov/OGWDW/dwh/t-soc/heptachlor.html). Thus, the heptachlor measured in elutriates from Craighill Angle West, and other Bay channel reaches has probably been in place for a long time without any significant increases from new sources.

USEPA's (1998) ambient criterion for the protection of human health from the consumption of contaminated aquatic life is $0.0021 \ \mu g/L$ (at MDE's 10^{-5} risk level), approximately 3 times lower than the instrument detection limit. As discussed in U.S. EPA water quality criteria documents (e.g., USEPA 1993, Appendix C) the Agency's criterion is based upon the following *assumptions*:

• the consumed organisms are exposed to the chemical for a sufficient duration that they reach a maximum steady state tissue concentration,

- a continuously exposed population of edible contaminated organisms from the site that is sufficient to feed a human population on a daily basis for 70 years.
- the criterion is the exposure concentration that is estimated to cause a lifetime carcinogenic risk of 10⁻⁵ (i.e., causing one additional cancer out of one hundred thousand exposed persons), and
- "continuous exposure to the compound" throughout a 70 year human lifespan, e.g., daily consumption of contaminated organisms collected from the site for 70 years.

The assumptions upon which U.S. EPA's heptachlor criterion is based are not consistent with the water column exposure that will occur at the proposed dredged material placement site. These data suggest that no adverse effect would be expected from exposure to heptachlor based upon the properly managed and temporary use of the site.

7.8 CONCLUSIONS AND OBSERVATIONS

The discussion and tables presented above indicate that ambient water quality criteria for the protection of aquatic life and human health criteria are exceeded for a small number of chemicals in full strength (undiluted) elutriate samples. However, given (1) the releases at the site will not be on a continuous basis at a given location, (2) the exposure assumptions that are the basis of ambient water quality criteria for the protection of aquatic life and human health are not met; and (3) that all of the elutriate analytes will be in compliance with ambient water quality criteria for the protection of aquatic life and human health within several hours of release, unacceptable adverse effects to the water column would not be expected from proper management of the discharges. Further, alternate release scenarios (e.g., hydraulic placement under specific tidal conditions) could further reduce water column exposure from elutriate conditions.

Other observations from the elutriate data set include:

- None of the elutriates prepared from the three placement/reference areas complies with all EPA criteria. In fact each of the three stations have undiluted elutriate concentrations that exceed U.S. EPA criteria for at least two analytes.
- Out of more than 150 target analytes, only three compounds (manganese, sulfides and heptachlor) would require more than 1-hour to achieve compliance with applicable ambient criteria. Two of these (manganese and sulfide) are naturally occurring chemicals which are present in all waterbodies, and the third (heptachlor) has not been commercially available since 1978.
- As shown in Table 7-36, <u>acute</u> criteria were rarely exceeded in the elutriate data set. In every instance except one, compliance with acute criteria would be attained within 3 minutes of release (less that 100 feet). The largest exceedance was for heptachlor at Craighill Angle West, where compliance with criteria would occur within 11 minutes at a distance of approximately 371 feet. USEPA's acute ambient water quality criteria are based upon a 1-hour average exposure concentration.

| Sampling Reach | Station | Sediment | Site Water ID and | Elutriate | | | |
|-----------------------------|-------------|--------------|-------------------|-------------|--|--|--|
| | | Aliquot (ml) | Volume (L) | Sample ID | | | |
| Craighill Entrance | CRE-001VC | 500 | CRE-SW | CRE-VC-EL | | | |
| | CRE-002VC | 500 | 8 | | | | |
| | CRE-003VC | 500 | | | | | |
| | CRE-004VC | 500 | | | | | |
| Craighill | CR1 | 666 | CR-SW | CR-EL | | | |
| | CR2 | 666 | 8 | | | | |
| | CR3 | 666 | | | | | |
| Craighill Angle-East | CRA-E-001VC | 666 | CRA-E-SW | CRA-E-VC-EL | | | |
| | CRA-E-002VC | 666 | 8 | | | | |
| | CRA-E-003VC | 666 | | | | | |
| Craighill Angle-West | CRA-W-001VC | 666 | CRA-W-SW | CRA-W-VC-EL | | | |
| | CRA-W-002VC | 666 | _ 8 | | | | |
| · | CRA-W-003VC | 666 | | | | | |
| Craighill Upper Range | CRU1 | 666 | CRU-SW | CRU-EL | | | |
| | _CRU2 | 666 | 8 | | | | |
| | CRU3 | 666 | | | | | |
| Cutoff Angle | | 666 | _ CUT-SW | CUT-EL | | | |
| | CUT2 | 666 | 8 | | | | |
| | | 666 | | | | | |
| Tolchester Channel - South | TLC-S-001VC | 500 | TLC-S-SW | TLC-S-VC-EL | | | |
| | TLC-S-002VC | 500 | _ 8 | | | | |
| | TLC-S-003VC | 500 | | | | | |
| | TLC-S-004VC | 500 | | | | | |
| Tolchester Channel - North | TLC-N-005VC | 333 | TLC-N-SW | TLC-N-VC-EL | | | |
| | TLC-N-006VC | | 8 | | | | |
| | TLC-N-007VC | 333 | 4 | | | | |
| | TLC-N-008VC | 333 | 4 1 | | | | |
| | TLC-N-009VC | 333 | - 1 | | | | |
| Tolebostos Sturichtoning | TLC-N-010VC | | TICOW | TICNOTI | | | |
| Totchester Strangmening | TLS-001VC | 1000 | | ILS-VC-EL | | | |
| Description Francisco | | | 8 | | | | |
| Brewerton Eastern Extension | BEI | 500 | BE-SW | BE-EL | | | |
| | BE2 | 500 | 8 | | | | |
| | DEA | 500 | 4 . 1 | | | | |
| Sum Doint | | 322 | CUT CUT | CWD VC EI | | | |
| Swan Folin | SWP-001VC | 333 | SWP-SW | SWP-VC-EL | | | |
| | SWP-002VC | 333 | 8 | | | | |
| | SWP-004VC | 333 | 4 | | | | |
| | SWP-00SVC | 222 | 4 | | | | |
| | SWP-006VC | 333 | f | | | | |
| C&D Annroach Channels | | 1000 | CD-VC-SW | CD-VC-FI | | | |
| Cours reprozen chamos | CD-002VC | 1000 | 8 | CD-VC-LL | | | |
| | CD003 | 500 | CD SW | | | | |
| | CD003 | 500 | CD-SW | CD-EL | | | |
| | CD004 | 500 | - ° | | | | |
| | CD005 | 500 | 4 | | | | |
| Placement Site 104 | KI-3 | 400 | KI CW | KI EI | | | |
| Placement She 104 | KI-5 | 400 | | KI-EL | | | |
| | KL-7-RFF | 400 | 4 | | | | |
| | KI-S-1 | 400 | 4 | | | | |
| | KI-S-2 | 400 | i | | | | |
| | KI-7-REF | 1000 | KICW | KL7-FI | | | |
| | KI-7-REFED | 1000 | | KI-/-EL | | | |
| Outside Placement Site 104 | KI-11 | 500 | KI-OUT-SW | KI-OUT | | | |
| Samer I meaning of the | KI-14 | 500 | | N-001 | | | |
| | KI-15 | 500 | 1 I | | | | |
| | KI-16 | 500 | 1 1 | | | | |

TABLE 7-1 SEDIMENT COMPOSITES FOR ELUTRIATE TESTING

TABLE 7-2 APPLICABLE FEDERAL AND PROPOSED MDE WATER QUALITY CRITERIA

| Г — — — — — — — — — — — — — — — — — — — | SALTWATER CRITERIA | | | | | | | | | | |
|---|--------------------|--------------------|----------------------|---------------------|--|--|--|--|--|--|--|
| | | | | HUMAN | | | | | | | |
| Analyte | Units | ACUTE ° | CHRONIC ^b | HEALTH ' | | | | | | | |
| NON-METALS | | | | | | | | | | | |
| CYANIDE | UG/L | 1 ^d | 1 ^d | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | 43 ^s | 6.4 ^s | | | | | | | | |
| SULFIDE, TOTAL | UG/L | | 2 ^v | | | | | | | | |
| VOCs | | | | | | | | | | | |
| DICHLOROMETHANE | UG/L | | | 16000 ^y | | | | | | | |
| SVOCs | | | | | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | | | 59 ^y | | | | | | | |
| PESTICIDES | | | | | | | | | | | |
| ALPHA-BHC | UG/L | | | 0.13 9 | | | | | | | |
| BETA-BHC | UG/L | | | 0.46 ^y | | | | | | | |
| ENDOSULFAN 1 | UG/L | 0.034 ⁿ | 0.0087 ⁿ | 240 | | | | | | | |
| GAMMA-BHC | UG/L | 0.16 | | 0.63 ^y | | | | | | | |
| HEPTACHLOR | UG/L | 0.053 | 0.0036 | 0.0021 ^y | | | | | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | 0.0036 | 0.0011 ^y | | | | | | | |
| PCBs | | | | | | | | | | | |
| TOTAL PCB | UG/L | | | 0.0017 5 | | | | | | | |
| PAHs | | | | | | | | | | | |
| FLUORENE | UG/L | | | 14,000 | | | | | | | |
| PYRENE | UG/L | | | 11000 | | | | | | | |
| METALS | | | | | | | | | | | |
| ANTIMONY | UG/L | | | 4300 | | | | | | | |
| ARSENIC | UG/L | 69 ⁸ | 36 * | 0.14 ^h | | | | | | | |
| BERYLLIUM | UG/L | | | 0.117 | | | | | | | |
| CHROMIUM | UG/L | 1100 ^k | 50 ^k | | | | | | | | |
| COPPER | UG/L | 4.8 ¹ | 3.1 ¹ | | | | | | | | |
| LEAD | UG/L | 210 ° | 8.1 ° | | | | | | | | |
| MANGANESE | UG/L | | | 100 | | | | | | | |
| MERCURY | UG/L | 1.8 ^p | 0.94 ^p | 0.051 | | | | | | | |
| NICKEL | UG/L | 74 ^r | 8.2 ^r | 4600 | | | | | | | |
| SELENTUM | UG/L | 290 ^t | 71 ^t | | | | | | | | |
| SILVER | UG/L | 1.9 ^u | | | | | | | | | |
| ZINC | UG/L | 90 ^w | 81 ^w | | | | | | | | |
| TRIBUTYLTIN | UG/L | 0.37 ^q | 0.01 9 | | | | | | | | |

Superscripts:

a = acute aquatic life criteria based on 1-hr average exposure concentrations.

b = chronic aquatic life criteria based on 4-day average exposure concentrations.

c = EPA human health criteria based on daily lifetime (70-year) average consumption of aquatic organisms; criteria based on 10⁻³ risk for carcinogens.

d = free cyanide as µg CN/L

f = applies to Total PCBs (sum of all congeners or isomer analyses).

- g = total dissolved arsenic.
- h = inorganic arsenic only.

j = from EPA 1986 Gold Book; no EPA 1998 number.

- k = dissolved chromium; hexavalent.
- l = dissolved copper.
- n = most appropriately applied to sum of alpha (1) and beta (11) endosulfan.
- o = dissolved lead.

p = dissolved total mercury.

- q = proposed criteria.
- r = dissolved nickel.

s = total ammonia as nitrogen; assumes cold weather conditions: salinity = 10 ppt, water temperature = 10 C, and pH=7.4

- t = dissolved selenium.
- u = dissolved silver.
- v = undissociated hydrogen sulfide (H₂S).
- w = dissolved zinc.
- y = carcinogen



TABLE 7-3 CONCENTRATIONS OF INORGANIC NON-METALS (MG/L) IN RECEIVINGWATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND AT THEOCEAN REFERENCE SITE

| | S | ample ID: | KI-03 | KI-07 | KI-14 | |
|-------------------------------|------|-----------|-----------------|------------------------|-------------------------|------------------------|
| | 1 | Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| CYANIDE | MG/L | 0.003 | 0.005U | 0.005U | 0.005U | 0.003U |
| NITROGEN, AMMONIA | MG/L | 0.02 | - U.32 | 0.23 | 4. 10 4 4 0.15 | 计过程 观波 0.13 |
| NITROGEN, NITRATE AND NITRITE | MG/L | 0.003 | 0.02U | 0.02U | 0.11 | 0.004U |
| NITROGEN, TOTAL KJELDAHL | MG/L | 0.19 | 0.03 | A 10.87 | 0.55 | 0,63 |
| OXYGEN DEMAND, BIOCHEMICAL | MG/L | 1 | | 2:6 | 8.23 | 2U |
| OXYGEN DEMAND, CHEMICAL | MG/L | 4.4 | 115 | | · 《户门》:"404 | 221 |
| PHOSPHORUS, TOTAL | MG/L | 0.01 | 0.01U | 0.01U | 0.05U | 0.09 |
| SULFIDE, TOTAL | MG/L | 0.35 | 0.35U | 0.35U | 0.76B | D.93B |
| TOTAL ORGANIC CARBON | MG/L | 0.08 | | | 20 | - 4.3 |

 $\overline{\mathbf{U}}$ = not detected. \mathbf{B} = found in blank.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.







TABLE 7-4 MAXIMUM CONCENTRATIONS OF INORGANIC NON-METALS (MG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| # ELU | R | EACH ID: TESTED: | Occan Reference | Outside Sitc 104 | Inside Site 104 3 | Brewerton Eastern Extension 2 | C&D Approach (Sarfleial) 2 | C&D Approach (Core) 2 | Craighill 2 | CraighШ Angle East 2 | Craighill Angle West | Craighill Entrance 2 | CraighШ Upper Range 2 | Cntoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|----------------------------|------|---------------------|---|---------------------|-------------------------|--|-------------------------------------|--|----------------|----------------------------|--|----------------------------|-----------------------------|-----------------|---------------|----------------------|--|-----------------------------|
| ANALYTE | UNIT | MIN DL | | | | | - | | | | | | | | - | | | |
| CYANIDE | MG/L | 0.0031 | 0.0031U | 0.005U | 0.005U | 0.005U | 0.005U | 0.005U | 0.005U | 0.005U | 0.005U | 0 060 | 0.005U | 0.06U | 0.005U | 1776 T10013 | 0.005U | 0.00511 |
| NITROGEN, AMMONIA | MG/L | 0.02 | 100000000000000000000000000000000000000 | 开达会 24 | 2011年1月1日月 | 和2013年1月 | 11098-0013 | の 秋田 田田 日日 | and stated a | 和你就你会让!! | STATISTICS. | 4 + 3:19,8 | 100 100 100 100 | 6 42 | 1302241 | ST GET STILL | 2422562282151 | STORE AND AND A |
| NITRITE | MG/L | 0.005 | 1444年1月月二日、144 | 0.020 | 0.02U | 0.02U | | and the second s | 0.02U | 0.02U | 0.02U | 0.020 | 0.02U | 0.02U | 0.02U | 0.02U | 0 02U | 0.02[] |
| NITROGEN, TOTAL KJELDAHL | MG/L | 0.19 | 这件III 新台湾 13 | Section 4 | STUD AND | 部できた | MACONCALL 1 | BARRADO . | 1400年間日期 | 12. 14. 14.1 | (Inclusive) | STORE ST | Caleshare 13 | 2 HERE | IRSA.7.1 | STREET, STREET, ST | Setter No. | 10100 |
| OXYGEN DEMAND, BIOCHEMICAL | MG/L | 0.37 | 20 | IU | 100.21.74 | South and the state | 10 | HANNESS I. | Statistics and | 10 | STREET, STREET | NU.52 191.30 | REDUCATION I | 100 1 00 | Look all | ou selected thirty a | and the second | 101711117 2004 |

| OATGEN DEMAND, CHEMICAL | MOL | 4.4 | 1000 1000 20101 | 1999年7月1日 1999年1月 | STATISTICS STATISTICS | CONTRACTOR OF A | Construct of the | 1296411034444 | Any: P. 201417 | 22110322010 | 100 A 100 | 11.05,74481.638 | 0048353513 | 100003459 | 45a (America 451 | 15516789249 | 100000000000000000000000000000000000000 |
|-------------------------|------|------|-----------------|-----------------------|-----------------------|------------------------|------------------|---------------|-------------------|-------------|---|-----------------|------------|-----------|------------------|-------------------|---|
| pH | pH | - | NM | 144 (T.M) 200 (T.M) | 20 A 10 A 10 A | Sector Sector | 20H 3 2.08 | Sea 349.7.6 | water and the | 出版的彩色和 | 1988 St. 7.45 | 2.284 7.73 | 1.0.1.7.8 | ×1.8.82 | TAU ART & ST | CONTRACTOR OF THE | 「「ないました」 |
| PHOSPHORUS, TOTAL | MG/L | 0.01 | A669113 | 124-24.4.14 | 0.2 | STATISTICS AND INCOMES | 20.0 Hitsel | 8.00 | Contraction (Col) | 14:00:015 | · 把包括含 0.05 | 0655020.13 | 1580 # 15 | 9.76 | 15151401 8 05 | 12:5:0.2 | ILL CULINA |
| SULFIDE, TOTAL | MG/L | 0.35 | 111111111111111 | 0 35U 7853 (3.5J | 0.350 | 0.35U | 0.35U | 0.350 | 0.350 | 0.35U | 0.350 | 0.35U | 0.350 | 035U | 0350 | ASSESSMENT OF | 0 3511 |
| TOTAL ORGANIC CARBON | MG/L | 0.05 | 345665920.9 | 10-20-21 10-14 - 44.9 | Constant and the | | Po stator | HAVE BEER | | #U | Tel 19.1 | 8U | 1.8. 1.8 | 14×3458 | AN PARANCE | | Mont Agent And |

U = not detected. B = found in blank. J = estimated value.

NM = not measured.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected eoneentrations.

TABLE 7-5 MEAN CONCENTRATIONS OF INORGANIC NON-METALS (MG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN **REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104**

| | ELUTRIATI | REACH ID ES TESTED | Occan Reference | Outside Site 104 1 ^(s) | Inside Site 104 | Brewerton Eastern Extension 2 | C&D Approach (Surficial) 2 | C&D Approach (Core) 2 | Craighill 2 | Craighill Angle East 2 | Craighili Angie West 2 | Craighill Entrance 2 | Craighill Upper Range 2 | Cutoff Angle 2 | Swan Point 2 | Tolchester North 2 | Tolchester South 2 | Toichester Straightening 2 |
|---------------------------|-----------|-----------------------|----------------------|---|--------------------|---|---------------------------------------|--------------------------------|----------------|---|------------------------------|----------------------------|-------------------------------|----------------------|--------------------|--------------------------|--------------------------|--|
| ANALYTE | UNIT | MIN DL | | | | | | | | | - | | | - | | | | |
| CYANIDE | MG/ | L 0.0031 | 0.0031U | 0.005 U | 0 005 U | 0 005 U | 0 005 U | 0 005 U | 0.005 L | 0 005 U | 0.005 U | 0 0325 U | 0 005 U | 0 0325 U | 0 005 U | 0.005 | 0 005 U | 0.005 U |
| NITROGEN AMMONIA | MGA | 0.02 | ***.** | A | Sec 1.250 | | Research THE | 178 AL 1 1 1 1 1 | | COLUMN AND | 公司 新聞書 | 6.45 | 100 100 102.235 | 264 3.7 | BARLANS | 100 Mil 445.95 | ALL CONTRACT | SMARTER OF BELLEVILLE |
| NITROGEN, NITRATE AND NIT | RITE MG/ | L 0.005 | | 0.02 U | 0.02 U | 0.02 U | | 11 | 0.02.0 | 0.02 U | 0.0125 U | 0 02 U | 0.02 U | 0.02 U | 0.0125 U | 0 02 U | 0 0125 U | 0.0125 U |
| NITROGEN, TOTAL KJELDAHL | . MG/1 | L 0.19 | (本)目前的事件 0.1 | HERE WAREA | 经济和 | CINESS AND A | 自治治治的 | C. 1. 4 100 Mar 1.40 | 2.6 | APRIL POPAL | 1 2055 19102 | REFE | 2.04 | SS 54.45 | 3-3-5.95 | 13.5 14 3.95 | A. 65 21 157 DT | 1. |
| OXYGEN DEMAND, BIOCHEM | ICAL MG/ | L 0.37 | 20 | 100 | 222.0 | SOCKORD BL | 0.685 U | 0.93 | SUSSECTION 1 | 0 685 U | 11.6 | A 100/411.2 | N 1.55 6 0.785 | Ng+12.04 | 1.7 U | 12000000000000 | CLARK AND | A 10 4 1 4 1 2 3 |
| OXYGEN DEMAND, CHEMICA | L MG/ | L 4.4 | 100 million (10) 291 | 1933219637 | 0.000 | 100 100 000 500 | 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 | 15.05.044134 | Service II | A CONTRACTOR OF A CONTRACT OF | \$34 A 98 | 14 SAL 864 | 115 45 536 | AL AL | 2.4317 | A.C. 40, 41 3.79 | NATION OF THE | NO. 00 - 2337 |
| pH | pH | | NM | 1500 000 2.89 | 4. 7.11 | 191000000000000000000000000000000000000 | 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 | 1.2.79 | SALES STATE | A STATE OF A | 742 | | 1222 05 07 12 | 400 \$ 2.73 | 2 3.51 | States TA | MINSS STAT | 1000000000000000000 |
| PHOSPHORUS TOTAL | MG/I | L 0.01 | 199454043913 | | 0.913 | 11.0 | 21 (CAN) & 0.095 | 20015 | A | 10.0 | State 8,085 | ANT 0.045 | A | 920 2.01 | SALE DIA 1 | A | 44101 2018 | A START OF A |
| SULFIDE, TOTAL | MG/I | L 0.35 | HALLANS 8.490 | 0 35 U | DESCRIPTION OF | 0 35 U | 0 35 U | 0 35 U | 0 35 U | 0 35 U | 0.35 U | See 1 5 0.195 | 0.35 U | 0 35 U | 0 35 Ü | 0.35 U | A 200 100 0.393 | 0 35 U |
| TOTAL ORGANIC CARBON | MG/I | L 0.08 | 14.000 20.31.3 | AT \$2621 | 11:247 | 240-200-19 | Sector Sector | 12111 | HANASMER D | 10.0225-255 | 4.04 U | A 4923 911 | 4.04 U | 1192 423 | 5.55 | 1-1-1-1-1-4-13 | 14 10 257 | DESCRIPTION |

U = not detected B = found in blank.

(a) = actual eoncentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

NM = not measured.

NOTE: Shaded and bolded values represent detected concentrations.





TABLE 7-6CONCENTRATIONS OF VOLATILE ORGANIC COMPOUNDS (UG/L)IN RECEIVING WATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENTSITE 104, AND AT THE OCEAN REFERENCE SITE

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| | S | ample ID: | KI-03 | KI-07 | KI-14 | |
|---------------------------|------|-----------|------------------------|------------------------|-------------------------|------------------------|
| | | Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| 1,1,1-TRICHLOROETHANE | UG/L | 1 | NT | NT | 10 | IU |
| 1,1,2,2-TETRACHLOROETHANE | UG/L | 1 | NT | NT | 10 | 10 |
| 1,1,2-TRICHLOROETHANE | UG/L | 0.7 | NT | NT | 0.7U | 1U |
| 1,1-DICHLOROETHANE | UG/L | 0.6 | NT | NT | 0.6U | 1U |
| 1,1-DICHLOROETHYLENE | UG/L | 1 | NT | NT | 10 | 10 |
| 1,2-DICHLOROETHANE | UG/L | 0.8 | NT | NT | 0.8U | 10 |
| 1,2-DICHLOROPROPANE | UG/L | 0.7 | NT | NT | 0.7U | 10 |
| 2-CHLOROETHYL VINYL ETHER | UG/L | 2 | NT | NT | 2U | 2U |
| ACROLEIN | UG/L | 6 | <u>NT</u> | NT | 6U | 6U |
| ACRYLONITRILE | UG/L | 6 | NT | NT | 6U | 6U |
| BENZENE | UG/L | 0.6 | NT | NT | 0.6U | 1U |
| BROMODICHLOROMETHANE | UG/L | 0.6 | NT | NT | 0.6U | ĪU |
| BROMOMETHANE | UG/L | 1 | NT | NT | 10 | 1U |
| CARBON DISULFIDE | UG/L | 0.7 | NT | NT | 0.7U | 1Ū |
| CARBON TETRACHLORIDE | UG/L | 1 | NT | NT | 10 | 1U |
| CFC-11 | UG/L | 1 | NT | NT | 10 | 1U |
| CFC-12 | UG/L | 2 | NT | NT | 2U | 1U |
| CHLOROBENZENE | UG/L | 0.8 | NT | NT | 0.8U | 1U |
| CHLORODIBROMOMETHANE | UG/L | 0.5 | NT | NT | 0.5U | 1U |
| CHLOROETHANE | UG/L | 0.8 | NT | NT | 0.8U | 1U |
| CHLOROFORM | UG/L | 0.6 | NT | NT | 0.6U | 1U |
| CHLOROMETHANE | UG/L | 0.5 | NT | NT | 0.5U | 2U |
| CIS-1,3-DICHLOROPROPENE | UG/L | 0.7 | NT | NT | 0.7U | 10 |
| DICHLOROMETHANE | UG/L | 1 | NT | NT | 10 | 1U |
| ETHYLBENZENE | UG/L | 0.7 | NT | NT | 0.7U | 1U |
| METHYLBENZENE | UG/L | 0.7 | NT | NT | 0.7U | 1U |
| TETRACHLOROETHENE | UG/L | 1 | NT | NT | 10 | 1U |
| TRANS-1,2-DICHLOROETHENE | UG/L | 1 | NT | NT | 10 | 1Ū |
| TRANS-1,3-DICHLOROPROPENE | UG/L | 0.7 | NT | NŤ | 0.7U | 1Ŭ |
| TRIBOMOMETHANE | UG/L | 1 | NT | NT | 10 | 10 |
| TRICHLOROETHYLENE | UG/L | 1 | NT | NT | 10 | 1U |
| VINYL CHLORIDE | UG/L | 0.9 | NT | NT | 0.9U | 1U |

U = not detected.

DL = detection limit.

NT = not tested

TABLE 7-7 CONCENTRATIONS OF VOLATILE ORGANIC COMPOUNDS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| # ELU | RE TRIATES | ACH ID: TESTED: | Ocean Reference | Outside Site | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighili | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Polnt | Tolchester North | Tolchester South | Tolchester Straightening |
|---------------------------|---------------|--------------------|--------------------|--------------|--------------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | | | | | | | | | | | | | | <u>.</u> | · · · | | | |
| ANALYTE | UNIT | MIN DL | (1997) - 19 (1997) | | | | - 115 | | | | | | | | | | | |
| ,I,1-TRICHLOROETHANE | UG/L | 1 | 10 | IU | IU | 10 | IU | 10 | 10 | 10 | IU | IU | 10 | 10 | 10 | IU | IU | 10 |
| ,1,2,2-TETRACHLOROETHANE | UG/L | 1 | 10 | IU | IU | 10 | IU | IU | 10 | 10 | IU | IU | 10 | 10 | IU | 10 | IU | 10 |
| ,I,2-TRICHLOROETHANE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| I-DICHLOROETHANE | UG/L | 0.6 | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U |
| ,1-DICHLOROETHYLENE | UG/L | 1 | 10 | IU | 10 | 10 | IU | 10 | 10 | 10 | IU IU | 10 | 1U | 10 | 10 | IU | IU | 10 |
| ,2-DICHLOROBENZENE | UG/L | 2 | 2U | 2 U | 2U | 2U | 2.1U | 2U | 2U | 2U | 2U | 2.4U | 2U | 2U | 2.2U | 2U | 2U | 2.2U |
| ,2-DICHLOROETHANE | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| ,2-DICHLOROPROPANE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| ,4-DICHLOROBENZENE | UG/L | 2 | 2U | 2 U | 2U | 2U | 2.1U | 2U | 2U | 2U | 2U | 2.4U | 2U | 2U | 2.2U | 2U | 2U | 2.2U |
| 2-CHLOROETHYL VINYL ETHER | UG/L | 2 | 20 | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U |
| ACROLEIN | UG/L | 6 | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U |
| ACRYLONITRILE | UG/L | 6 | 6U | 6U | 6U | 6U | 6U | 6U | 6U. | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U | 6U |
| BENZENE | UG/L | 0.6 | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U |
| BROMODICHLOROMETHANE | UG/L | 0.6 | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U |
| BROMOMETHANE | UG/L | 1 | 10 | IU | 10 | lU | 10 | IU | IU | 10 | IU | IU | 10 | IU | IU | IU | IU. | lu |
| CARBON DISULFIDE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| CARBON TETRACHLORIDE | UG/L | 1 | 10 | UI | 10 | IU | IU. | IU | 10 | 10 | 10 | IU | 10 | ١U | 10 | 10 | IU | 10 |
| CFC-11 | UG/L | | IU | 10 | 10 | IU | IU | IU | 10 | 10 | 10 | 10 | IU | IU | IU | IU | IU | IU |
| CFC-12 | UG/L | 2 | 2U | 20 | 2U | 2U | 2U | 2U | 2U | 2U | 20 | 2U | 2U | 2U | 2U | 2 U | 2U | 2U |
| CHLOROBENZENE | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| CHLORODIBROMOMETHANE | UG/L | 0.5 | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U |
| CHLOROETHANE | UG/L | 0.8 | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U | 0.8U |
| CHLOROFORM | UG/L | 0.6 | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U | 0.6U |
| CHLOROMETHANE | UG/L | 0.5 | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U | 0.5U |
| CIS-1,3-DICHLOROPROPENE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| DICHLOROMETHANE | UG/L | 1 | 10 | 24 130 | Swight 0 | 160 | 22 | Aug 8 1016 126 | Ar 4 110 | 63 mile # 110 | 255 F | 1. 1. 1. 170 | 160 | Anorthe 34 | & 680E | Indiana 16 | preservation 30 | and allowed to the 140 |
| ETHYLBENZENE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| M-DICHLOROBENZENE | UG/L | 2 | 2U | 20 | 2U | 2U | 2.1U | 2U | 2U | 2U | 2U | 2.4U | 2U | 2 U | 2.2U | 2 U | 2U | 2.20 |
| METHYLBENZENE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| TETRACHLOROETHENE | UG/L | 1 | 10 | 10 | 10 | IU | IU | . IU | IU | IU | IU | IU | IU | 10 | IU | IU | IU | 10 |
| TRANS-1,2-DICHLOROETHENE | UG/L | 1 | 10 | 10 | 10 | 10 | 10 | IU | 10 | 10 | 10 | 10 | 10 | 10 | IU | IU | ١U | 10 |
| TRANS-1,3-DICHLOROPROPENE | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U | 0.7U |
| TRIBOMOMETHANE | UG/L | 1 | 10 | 10 | IU | 10 | 10 | 10 | IU | 10 | IU | 10 | 10 | 10 | IU | iU | IU | IU |
| TRICHLOROETHYLENE | UG/L | 1.1 | IU | 10 | 10 | IU | 10 | IU | IU | 10 | 10 | IU | IU IU | 10 | 10 | IU | IU | 10 |
| VINYL CHLORIDE | UG/L | 0.9 | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U | 0.9U |

U = not detected, E = exceeds calibration range of the instrument.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.





TABLE 7-8CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/L) INRECEIVING WATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, ANDAT THE OCEAN REFERENCE SITE

| | Sa | mple ID: | KI-03 | KI-07 | KI-14 | |
|-------------------------------------|------|----------|-----------------|-----------------|------------------|-----------------|
| | R | each ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| 1,2,4-TRICHLOROBENZENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 1,2-DICHLOROBENZENE | UG/L | | 2U | 2U | 2U | |
| 1,2-DIPHENYLHYDRAZINE | UG/L | 3 | 3U | 3U | 3U | 3U |
| 1,4-DICHLOROBENZENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 1-METHYLNAPHTHALENE | UG/L | 0.31 | 0.31U | 0.31U | 0.31U | 0.31U |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | UG/L | 1 | 1U | 1U | 1U | 1U |
| 2,4,6-TRICHLOROPHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2,4-DICHLOROPHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2,4-DIMETHYLPHENOL | UG/L | 4 | 4U | 4U | 4U | 4U |
| 2,4-DINITROPHENOL | UG/L | 23 | 23U | 23U | 23U | 23U |
| 2,4-DINITROTOLUENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2,6-DINITROTOLUENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2-CHLORONAPHTHALENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2-CHLOROPHENOL | UG/L | 1 | 1U | 1U | 1U | 1U |
| 2-METHYL-4,6-DINITROPHENOL | UG/L | 5 | 5U | 5U | 5U | 5U |
| 2-METHYLNAPHTHALENE | UG/L | 0.21 | 0.21U | 0.21U | 0.21U | 0.21U |
| 2-METHYLPHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| 2-NITROPHENOL | UG/L | 3 | 3U | 3U | 3U | 3U |
| 3,3'-DICHLOROBENZIDINE | UG/L | 7 | 7U | 7U | 7U | 7U |
| 3,4-METHYLPHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| 3,5,5-TRIMETHYL-2-CYCLOHEXENE-1-ONE | UG/L | 2 | 2U | 2U | 2U | 2U |
| 4-BROMOPHENYL PHENYL ETHER | UG/L | 3 | 3U | 3U | 3U | 3U |
| 4-CHLORO-3-METHYLPHENOL | UG/L | 1 | IU | IU | IU | IU |
| 4-CHLOROPHENYL PHENYL ETHER | UG/L | 2 | 2U | 2U | 2U | 2U |
| 4-NITROPHENOL | UG/L | 4 | 4U | 4U | 4U | 4U |
| BENZOIC ACID | UG/L | 34 | 34U | 34U | 34U | 34U |
| BENZYL ALCOHOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| BENZYL BUTYL PHTHALATE | UG/L | 2 | 2U | 2U | 2U | 2U |
| BIS(2-CHLOROETHOXY)METHANE | UG/L | 2 | 2U | 2U | 2U | 20 |
| BIS(2-CHLOROETHYL) ETHER | UG/L | 2 | 2U | 2U | 2U | 2U |

TABLE 7-8 (CONTINUED)

| | Sai | mple ID: | KI-03 | KI-07 | KI-14 | |
|---------------------------------|------|----------|-----------------|-----------------|------------------|-----------------|
| | R | each ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 2 | 2U | 2U | 2U | 3B |
| DIBENZOFURAN | UG/L | 2 | 2U | 2U | 2U | 2U |
| DIETHYL PHTHALATE | UG/L | 3 | 3U | 3U | 3U | 3U |
| DIMETHYL PHTHALATE | UG/L | 3 | 3U | 3U | 3U | 3U |
| DI-N-BUTYL PHTHALATE | UG/L | 4 | 4U | 4U | 4U | 4U |
| DI-N-OCTYL PHTHALATE | UG/L | 3 | 3U | 3U | 3U | 3U |
| HEXACHLORO-1,3-BUTADIENE | UG/L | 2 | 2U | 2U | 2U | 2U |
| HEXACHLOROBENZENE | UG/L | 3 | 3U | 3U | 3U | 3U |
| HEXACHLOROCYCLOPENTADIENE | UG/L | 4 | 4U | 4U | 4U | 4U |
| HEXACHLOROETHANE | UG/L | 2 | 2U | 2U | 2U | 2U |
| M-DICHLOROBENZENE | UG/L | 2 | 2U | 2 U | 2U | 2 U |
| METHANAMINE, N-METHYL-N-NITROSO | UG/L | 3 | 3U | 3U | 3U | 3U |
| NITROBENZENE | UG/L | 3 | 3U | 3U | 3U | 3U |
| N-NITROSODI-N-PROPYLAMINE | UG/L | 4 | 4U | 4U | 4U | 4U |
| N-NITROSODIPHENYLAMINE | UG/L | 4 | 4U | 4U | 4U | 4U |
| PENTACHLOROPHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |
| PHENOL | UG/L | 2 | 2U | 2U | 2U | 2U |

 $\mathbf{U} =$ not detected. $\mathbf{B} =$ found in blank.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.









TABLE 7-9MAXIMUM CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/L) IN ELUTRIATES FROM BALTIMOREAPPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | RE | ACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surfielal) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester |
|------------------------------------|----------|---------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|---|
| # ELUT | RIATES 1 | TESTED: | 1(8) | l (e) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| | | | | | | | | | | | | | | | | | | |
| ANALYIE | UNIT | MINDL | 211 | 211 | 211 | 211 | 2.111 | 211 | 211 | 211 | 211 | 2.411 | 211 | 211 | 2.211 | | | 0.011 |
| 1.2 DIBUENVI UVDRAZDIE | UG/L | | 20 | 20 | 20 | 20 | 1 211 | 20 | 20 | 20 | 20 | 1 711 | 20 | 20 | 1 111 | 20 | 20 | 2.20 |
| 1 AFTINI MABUTUATENE | UGA | 0.11 | 0.111 | 0.1111 | 0 1111 | 0 3111 | 0.111 | 0 1111 | 0.111 | 0 1111 | 0.111 | 0.1111 | 0 111 | 01111 | 0.111 | 0.1111 | 30 | 3.20 |
| 1-METATLNAFATALENE | UC/L | 0.31 | 0.310 | 0.310 | 0.310 | 0.510 | 1.111 | 0.510 | 0.510 | 0.510 | 0.310 | 1.211 | 0.310 | 0.310 | 0.310 | 0.310 | 0.310 | 0.310 |
| 2.4.6 TRICHLOROBUENOL | UG/L | 1 | 211 | 211 | 211 | 211 | 2.111 | 211 | 211 | 211 | 211 | 2.411 | 211 | 211 | 2.20 | 10 | 10 | 1.10 |
| 2,4,0-1 KICHLOKOPHENOL | UG/L | 2 | 20 | 20 | 20 | 20 | 2.10 | 20 | 20 | * 211 | 20 | 2.40 | 20 | 20 | 2.20 | 20 | 20 | 2.20 |
| 2,4-DICALOROFHENOL | UG/L | 4 | 20 | 411 | 411 | 411 | 4 111 | 411 | 411 | 411 | 411 | 4 911 | 20 | 20 | 4.312 | 20 | 20 | 2 20 |
| 2.4 DINITROBLENOL | UG/L | 21 | 2311 | 2311 | 2111 | 2311 | 2411 | 2311 | 2311 | 2311 | 2310 | 2811 | 2111 | 211 | 9.30 | 2111 | 40 | 4.30 |
| 24 DINITROTOLUENE | UGA | 23 | 200 | 211 | 211 | 211 | 2 111 | 211 | 211 | 210 | 211 | 2.411 | 230 | 230 | 2 211 | 200 | 230 | 230 |
| 2.4 DINITROTOLUENE | UG/L | 2 | 20 | 20 | 20 | 211 | 2.10 | 211 | 211 | 20 | 20 | 2.40 | 20 | 20 | 2.20 | 20 | 20 | 2.20 |
| 2 CHLOBONA PHTHALENE | UG/L | 2 | 20 | 20 | 211 | 211 | 2.10 | 211 | 20 | 20 | 20 | 2.40 | 20 | 20 | 2.20 | 20 | 20 | 2.20 |
| 2 CHLOROBARNOI | 116/1 | | 111 | 111 | 111 | 111 | 1 111 | 111 | 111 | 111 | 111 | 1.90 | 111 | 20 | 1.111 | 20 | 20 | 2.20 |
| 2 METHYL 4 6 DINITROPHENOL | UG/L | - | SU | 511 | 511 | 511 | 5 11 | 511 | 50 | 511 | 511 | 6 111 | 10 (1) | (1) | 5.411 | 10 | 10 | 1.10 |
| 2 METHVI NABUTUAI ENE | UGA | 0.21 | 0.2111 | 0.2111 | 0.2111 | 0.2111 | 0.2111 | 0 2111 | 0.2111 | 0 2111 | 0.2111 | 0.10 | 0.2111 | 0.2111 | 0.2111 | 0.2111 | 0.2111 | 3.40 |
| 2 METUVI BUENOI | UGA | 2 | 211 | 211 | 211 | 211 | 2 111 | 211 | 211 | 211 | 0.210 | 2.411 | 0.410 | 0.210 | 2 210 | 0.210 | 0.210 | 0.210 |
| 2 NITROBHENOI | UG/L | 1 1 | 111 | 311 | 20 | 10 | 1 211 | 311 | 311 | 311 | 111 | 1.711 | 20 | 20 | 1 111 | 20 | 20 | 2.20 |
| 1 1 DICHLOBOBENZIDINE | UGA | 7 | 711 | 711 | 711 | 711 | 7.411 | 211 | 711 | 711 | 711 | 9.70 | 711 | 711 | 7.611 | 30 | 30 | 3.20 |
| 14 METHVI PHENOI | UGA | 2 | 211 | 211 | 211 | 211 | 2 111 | 211 | 211 | 211 | 211 | 2.411 | 211 | 211 | 2 211 | 211 | 70 | 7.30 |
| 3.5.5 TRIMETHYL 2.CVCLOHEYENE LONE | UGA | 2 | 211 | 211 | 211 | 21 | 2 111 | 211 | 211 | 20 | 211 | 2.411 | 20 | 20 | 2 20 | 20 | 20 | 2.20 |
| A-BROMOPHENYL PHENYL FTHER | UG/L | 1 | 311 | 311 | 311 | 11 | 1 211 | 311 | 311 | 311 | 311 | 3 711 | 11 | 311 | 1 11 | 111 | 20 | 1 211 |
| ACHI ORO. 3. METHYL PHENOL | UG/L | | 111 | 111 | 111 | 111 | 1 111 | 111 | 111 | 111 | 111 | 1.71 | 111 | 111 | 1.111 | | 111 | 3.20 |
| A CHI OROPHENVI PHENVI ETHER | 116/ | 2 | 211 | 211 | 211 | 211 | 2 111 | 211 | 211 | 211 | 211 | 2.41 | 211 | 211 | 2 211 | 211 | 211 | 2.20 |
| 4-NITROPHENOL | 116/ | | 411 | 40 | 411 | 40 | 4 30 | 411 | 411 | 41 | 411 | 4 911 | 411 | 411 | 4 111 | 411 | 411 | 4 311 |
| BENZOIC ACID | 1164 | 34 | 3411 | 3411 | 3411 | 3411 | 160 | 3411 | 3411 | 3411 | 3411 | 4111 | 1411 | 3411 | 1711 | 1411 | 3411 | 3711 |
| BENZYL ALCOHOL | 116/1. | 2 | 21/ | 21 | 211 | 20 | 2.10 | 211 | 211 | 211 | 211 | 2 411 | 211 | 211 | 2 211 | 211 | 211 | 2 211 |
| BENZYL BUTYL PHTHALATE | UG/L | 2 | 20 | 2U | 2U | 2U | 2.1U | 2U | 20 | 2U | 20 | 2.4U | 211 | 21 | 2 21/ | 20 | 211 | 2.20 |
| BIS(2-CHLOROETHOXY)METHANE | UG/L | 2 | 20 | 2U | 2U | 2U | 2.1U | 2U | 2U | 20 | 20 | 2.4U | 20 | 211 | 2 211 | 211 | 20 | 2.20 |
| BIS(2-CHLOROETHYL) ETHER | UG/L | 2 | 20 | 2U | 2U | 2U | 2.10 | 2U | 2U | 20 | 20 | 2.40 | 211 | 211 | 2 211 | 211 | 211 | 2.20 |
| BIS(2-ETHVI HEXYL) PHTHALATE | UG/ | 2 | 211 | 21) | AND DE LO | 211 | 2.11/ | RTTANT' | 211 | 21-CEX.08.2 | 211 | 2 411 | 211 | 20000000000 | 2 211 | 100000000000 | AND STREET | 100000000000000000000000000000000000000 |
| DIBENZOFURAN | UG/L | 2 | 21/ | 21/ | 20 | 211 | 2.10 | 21 | 21 | 211 | 211 | 2 411 | 211 | 211 | 2 211 | 211 | 211 | 2 211 |
| DIETHYL PHTHALATE | 116/1 | 3 | 311 | 311 | 30 | 311 | 3.20 | 30 | 30 | 311 | 311 | 1 711 | 311 | 311 | 1 11 | 311 | 311 | 3 211 |
| DIMETHYL PHTHALATE | 11G/L | 3 | 311 | 30 | 30 | 311 | 3.2U | 30 | 30 | 30 | 311 | 170 | 311 | 311 | 1 111 | 311 | 311 | 3 211 |
| DLN.BUTYL PHTHALATE | UG/L | 4 | 411 | 411 | 40 | 411 | 4 3 U | 4U | 411 | 40 | 411 | 4 911 | 411 | 411 | 4 311 | 411 | 411 | 4 311 |
| DI-N-OCTYL PHTHALATE | UG/L | 1 | 311 | 30 | 311 | 311 | 3.21/ | 31/ | 311 | 311 | 311 | 3 711 | 311 | 11 | 1 11 | 11 | 11 | 1.30 |
| HEXACHLORO-1 3-BUTADIENE | UG/L | 2 | 211 | 211 | 211 | 211 | 2 11/ | 211 | 211 | 211 | 211 | 2.411 | 211 | 211 | 2 211 | 211 | 20 | 2 211 |
| HEXACHLOROBENZENE | UG/L | 1 | 311 | 311 | 311 | 311 | 3 211 | 311 | 311 | 311 | 311 | 1711 | 111 | 111 | 1 111 | 10 | 20 | 1.20 |
| HEYACHLOROCYCLOPENTADIENE | UGA | | ALI | 411 | 411 | ALI A | 4 311 | 411 | 411 | 411 | 411 | 4.011 | 411 | 411 | 4 311 | 30 | 30 | 4 311 |
| HEXACHLOROFTHANE | 1164 | 2 | 711 | 211 | 211 | 211 | 2 111 | 211 | 211 | 211 | 211 | 2 411 | 211 | 211 | 9.30 | 90 | 40 | 2.30 |
| METHANAMINE N.METHYL N.NITROSO | UGA | 1 | 311 | 311 | 111 | 311 | 1 211 | 311 | 311 | 311 | 311 | 3 711 | 311 | 111 | 1 311 | 20 | 20 | 1 2120 |
| NITROBENZENE | UG/L | 1 | 111 | 111 | 111 | 30 | 1 211 | 30 | 311 | 111 | 111 | 3.70 | 111 | 11 | 1 11 | 30 | 30 | 3.20 |
| N NITROSODI N. PROPYLAMINE | UGA | - | 30 | 30 A11 | 411 | 411 | 4 311 | 411 | 411 | 411 | 411 | 3.70 | 30 | 30 | 3.30 | 30 | 30 | 3.20 |
| N.NITROSODIPHENYI AMINE | UGA | 4 | 40 | 40 | 40 | 40 | 4 30 | 40 | 40 | 40 | 40 | 4 90 | 40 | 40 | 4.30 | 40 | 40 | 4.30 |
| PENTACHI OROPHENOI | UGA | 2 | 211 | 211 | 211 | 211 | 2 111 | 211 | 211 | 211 | 211 | 2 411 | 311 | | 2.30 | 40 | 40 | 2 30 |
| PHENOL | UGA. | 2 | 211 | 21/ | 211 | 211 | 2.11 | 211 | 21 | 211 | 211 | 2 411 | 211 | 211 | 2 211 | 20 | 20 | 2 20 |

U = not detected.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

TABLE 7-10 MEAN CONCENTRATIONS OF SEMIVOLATILE ORGANIC COMPOUNDS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | | Ocean | Outside | Inside | Brewerton Eastern | C&D Approach | C&D Approach | Contacting | Craighill | Craighill | Craighill | Craightil Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
|-------------------------------------|--------|---------|-----------|----------|-----------|----------------------|--------------|-----------------|------------|------------|------------|-----------|--------------------|---------|--------|----------------|---------------|----------------|
| | RE/ | ACH ID: | Reference | Site 104 | Site 104 | Extension | (Surncial) | (Core) | Craignill | Angle Last | Angle West | Entrance | Kange | Angle | Point | North | South | Straightening |
| # ELUIK | ATES T | ESTED: | | | 3 1 | 2 | 4 | 4 | 4 | - 2 | . 4 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MIN DI | | | | | | | | | | | | | | | | |
| 1.2.4-TRICHLOROBENZENE | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 200 | 2.01 | 2 2 U | 2.0 U | 2.0 U | 2.1 U | 2011 | 2011 | 2111 |
| 1.2-DIPHENYLHYDRAZINE | UG/L | 3 | 3U | 3.0 U | 3.0 U | 3.0 U | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.35 U | 3.0 U | 3.0 U | 3.15 U | 301 | 301 | 110 |
| I-METHYLNAPHTHALENE | UG/L | 0.31 | 0.31U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0 31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0.31 U | 0 31 1/ |
| 2.2'-OXYBIS(I-CHLOROPROPANE) | UG/L | 1 | IU | 1.0 U | 1.0 U | 1.0 U | 1.05 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.1 U | 1.0 U | 1.0 U | 1.05 U | 1.0 U | 1.0 U | 1.05 [] |
| 2.4.6-TRICHLOROPHENOL | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 200 | 211 |
| 2.4-DICHLOROPHENOL | UG/L | 2 | 2U | 2.0 U | 200 | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 2.4-DIMETHYLPHENOL | UG/L | 4 | 4U | 40U | 4.0 U | 4.0 U | 4.15 U | 4.0 U | 4.0 U | 40 U | 4.0 U | 4.45 U | 4.0 U | 4.0 U | 4.15 U | 4.0 U | 4.0 U | 4.15 U |
| 2.4-DINITROPHENOL | UG/L | 23 | 23U | 23.0U | 23.0 U | 23.0 U | 23.5 U | 23 O U | 23.0 U | 23.0 U | 23 O U | 25.5 U | 23.0 U | 23.0 U | 24.0 U | 23.0 U | 23.0 U | 24 0 11 |
| 2,4-DINITROTOLUENE | UG/L | 2 | 2U | 2 O U | 200 | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 2,6-DINITROTOLUENE | UG/L | 2 | 2U | 2 O U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 20 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 2-CHLORONAPHTHALENE | UG/L | 2 | 2U | 2 O U | 2.0 U | 2.0 U | 2.05 U | 20 U | 2.0 U | 2.0 U | 2 O U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 20U | 2.1 U |
| 2-CHLOROPHENOL | UG/L | 1 | IU | 100 | 100 | 100 | 1.05 U | 1.0 U | 1.0 U | 1.0 U | 100 | 1.I U | 1.0 U | 1.0 U | 1.05 U | 1.0 U | 100 | 1.05 U |
| 2-METHYL-4,6-DINITROPHENOL | UG/L | 5 | 5U | 5 O U | 5.0 U | 5.0 U | 5.15 U | 5.0 U | 5.0 U | 5.0 U | 5.0 U | 5.55 U | 5.0 U | 5.0 U | 5.2 U | 5.0 U | 5.0 U | 5.2 U |
| 2-METHYLNAPHTHALENE | UG/L | 0.21 | 0.21U | 0.21 U | 0 21 U | 0 21 U | 0 2 I U | 0.21 U | 0.21 U | 0 21 U | 0.21 U | 0.21 U | 0.21 U | 0.21 U | 0.21 U | 0 21 U | 0 21 U | 0 2I U |
| 2-METHYLPHENOL | UG/L | 2 | 2U | 2 O U | 20U | 2.0 U | 2 05 U | 2.0 Ŭ | 2.0 U | 2.0 U | 20U | 2.2 U | 2.0 U | 2.0 U | 2.I U | 2.0 U | 2.0 U | 2.1 U |
| 2-NITROPHENOL | UG/L | 3 | 3U | 3 O U | 3 O U | 300 | 3.1 U | 3.0 U | 3 O U | 3.0 U | 3.0 U | 3 35 U | 3.0 U | 3.0 U | 3.15 U | 3.0 U | 3.0 U | 3.1 U |
| 3,3'-DICHLOROBENZIDINE | UG/L | 7 | 70 | 700 | 7.0 U | 7.0 U | 7.2 U | 7.0 U | 7.0 U | 7.0 U | 7.0 U | 7.75 U | 7.0 U | 7 O U | 7.3 U | 7.0 U | 7.0 U | 7.25 U |
| 3,4-METHYLPHENOL | UG/L | 2 | 2U | 200 | 20 U | 200 | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 20U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 3,5,5-TRIMETHYL-2-CYCLOHEXENE-1-ONE | UG/L | 2 | 2U | 2 O U | 2 O U | 200 | 2.05 U | 2 O U | 2 O U | 2.0 U | 20 U | 22 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 4-BROMOPHENYL PHENYL ETHER | UG/L | 3 | 3U | 3 O U | 30 U | 300 | 3 I U | 3 O U | 3.0 U | 3.0 U | 300 | 3 35 U | 3.0 U | 3.0 U | 3.15 U | 3.0 U | 3.0 U | 3.I U |
| 4-CHLORO-3-METHYLPHENOL | UG/L | 1 | 1U | 1.0 U | 1 0 U | 1.0 U | 1.05 U | 1.0 U | 1.0 U | 100 | 1.0 U | 1.1 U | 1.0 U | 1.0 U | 1.05 U | 100 | 100 | 1.05 U |
| 4-CHLOROPHENYL PHENYL ETHER | UG/L | 2 | 2U | 2.0 U | 2.0 U | 200 | 2 05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| 4-NITROPHENOL | UG/L | 4 | 4U | 4 0 U | 4 0 U | 400 | 415U | 40U | 4.0 U | 400 | 4 0 U | 4.45 U | 4.0 U | 4 O U | 4.15 U | 4.0 U | 4.0 U | 4.15 U |
| BENZOIC ACID | UG/L | 34 | 34U | 34.0 U | 34 O U | 34 O U | 35.0 U | 34.0 U | 34 O U | 34.0 U | 34.0 U | 37.5 U | 34.0 U | 34.0 U | 35.5 U | 34.0 U | 34.0 U | 35.5 U |
| BENZYL ALCOHOL | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 200 | 2.2 U | 2.0 U | 2.0U | 2.1 U | 2.0 U | 2.0 U | 2.I U |
| BENZYL BUTYL PHTHALATE | UG/L | 2 | 2U | 2 O U | 2.0 U | 200 | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 200 | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| BIS(2-CHLOROETHOXY)METHANE | UG/L | 2 | 2U | 2.0 U | 20 U | 200 | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 20U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| BIS(2-CHLOROETHYL) ETHER | UG/L | 2 | 2U | 200 | 20 U | 2.0 U | 2.05 U | 20 U | 2.0 U | 200 | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 2 | 2U | 2.0 U | 241-42.17 | 2.0 U | 2.05 U | 6 6 6 5 | 2.0 U | 1.1.1.2.2 | 2.0 U | 2.2 U | 2.0 U | 5. 1.35 | 2.1 U | 19/2/02/02/015 | San Diversits | 1. Alexandre 1 |
| DI-N-BUTYL PHTHALATE | UG/L | 4 | 2U | 4.0 U | 40 U | 400 | 4.15 U | 4.0 U | 4.0 U | 400 | 4.0 U | 4.45 U | 4.0 U | 40U | 4.15 U | 4.0 U | 4.0 U | 4.15 U |
| DI-N-OCTYL PHTHALATE | UG/L | 3 | 3U | 3.0 U | 3.0 U | 300 | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.35 U | 3.0 U | 3 O U | 3.15 U | 3.0 U | 3.0 U | 3.1 U |
| DIBENZOFURAN | UG/L | 2 | 3U | 2.0 U | 2.0 U | 200 | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 200 | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2 O U | 2.1 U |
| DIETHYL PHTHALATE | UG/L | 3 | 4U | 3.0 U | 3.0 U | 3.00 | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3 35 U | 3.0 U | 3 O U | 3.15 U | 30 U | 3.0 U | 3.I U |
| DIMETHYL PHTHALATE | UG/L | 3 | 3U. | 3.0 U | 3.0 U | 300 | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.35 U | 3.0 U | 3.0 U | 3.15 U | 3.0 U | 3.0 U | 31U |
| HEXACHLORO-1,3-BUTADIENE | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 20U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| HEXACHLOROBENZENE | UGAL | 3 | 3U | 3.0 U | 3.0 U | 3.0 U | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.35 U | 30U | 3.0 U | 3.15 U | 3.0 U | 3.0 U | 3.1 U |
| HEXACHLOROCYCLOPENTADIENE | UG/L | 4 | 4U | 4.0 U | 4.0 U | 4.0 U | 4.15 U | 4.0 U | 4.0 U | 4.0 U | 4.0 U | 4.45 U | 4.0 U | 4.0 U | 4.15 U | 4.0 U | 4.0 U | 4.15 U |
| HEXACHLOROETHANE | UG/L | 2 | 20 | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 2.0 U | 2.0 U | 2.1 U |
| METHANAMINE, N-METHYL-N-NITROSO | UG/L | 3 | 3U | 3.0 U | 3.0 U | 3.0 U | 3.1 U | 3.0 U | 3.0 U | 3.0 U | 3.0 U | 3.35 U | 3.0 U | 3.0 U | 3 15 U | 3.0 U | 3.0 U | 3.1 U |
| N-NITROSODI-N-PROPYLAMINE | UG/L | 4 | 3U | 4.0 U | 4.0 U | 400 | 4.15 U | 40 U | 4.0 U | 4.0 U | 4.0 U | 4 45 U | 4.0 U | 4.0 U | 4.15 U | 4.0 U | 4.0 U | 4.15 U |
| N-NITROSODIPHENYLAMINE | UG/L | 4 | 4U | 4.0 U | 40 U | 4.0 U | 4.15 U | 40 U | 4.0 U | 4.0 U | 400 | 4.45 U | 40U | 4.0 U | 4.15 U | 4.0 U | 40U | 4.15 U |
| NITROBENZENE | UG/L | 3 | 4U | 3.0 U | 3.0 U | 3.0 U | 3.1 U | 3.0 U | 3.0 U | 300 | 3.0 U | 3.35 U | 3.0 U | 3.0 U | 3.15 U | 3.0 U | 3.0 U | 3.1 U |
| PENTACHLOROPHENOL | UG/L | 2 | 2U | 2.0 U | 20 U | 200 | 2.05 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 200 | 2.0 U | 2.1 U |
| PHENOL | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.05 U | 200 | 2.0 U | 2.0 U | 2.0 U | 2.2 U | 2.0 U | 2.0 U | 2.1 U | 20 U | 2.0 U | 210 |

U = not detected.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations









TABLE 7-11CONCENTRATIONS OF CHLORINATED PESTICIDES,ORGANOPHOSPORUS PESTICIDES, AND PCB AROCLORS (UG/L) INRECEIVING WATER INSIDE PLACEMENT SITE 104, OUTSIDEPLACEMENT SITE 104, AND AT THE OCEAN REFERENCE SITE

| | | Sample ID: | KI-03 | KI-07 | KI-14 | |
|--------------------|------|------------|-----------------|-----------------|------------------|-----------------|
| | | Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| 4,4'-DDD | UG/L | 0.01 | 0.01U | 0.01UJ | 0.0IU | 0.01U |
| 4,4'-DDE | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| 4,4'-DDT | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| ALDRIN | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| ALPHA-BHC | UG/L | 0.01 | 0.0IU | 0.01UJ | 0.01U | 0.01U |
| ВЕТА-ВНС | UG/L | 0.01 | 0.01U | 0.02J | 0.01U | 0.01U |
| CHLORBENSIDE | UG/L | 0.1 | 0.1UJ | 0.1UJ | 0.IU | 0.1U |
| CHLORDANE | UG/L | 0.1 | 0.IU | 0.IUJ | 0.IU | 0.IU |
| DACTHAL | UG/L | 0.1 | 0.1U | 0.IUJ | 0.3U | 0.1U |
| DELTA-BHC | UG/L | 0.01 | 0.0IU | 0.01UJ | 0.01U | 0.01U |
| DIELDRIN | UG/L | 0.01 | 0.01U | 0.01UJ | 0.01U | 0.01U |
| ENDOSULFAN I | UG/L | 0.01 | 0.01U | 0.0IUJ | 0.0IU | 0.01U |
| ENDOSULFAN II | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| ENDOSULFAN SULFATE | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| ENDRIN | UG/L | 0.03 | 0.03U | 0.03UJ | 0.03U | 0.03U |
| ENDRIN ALDEHYDE | UG/L | 0.03 | 0.03U | 0.03UJ | 0.03U | 0.03U |
| GAMMA-BHC | UG/L | 0.0081 | 0.0081U | 0.0097J | 0.008IU | 0.0081U |
| HEPTACHLOR | UG/L | 0.02 | 0.02U | 0.02UJ | 0.02U | 0.02U |
| HEPTACHLOR EPOXIDE | UG/L | 0.01 | 0.0IU | 0,03J | 0.0IU | 0.01U |
| METHOXYCHLOR | UG/L | 0.08 | 0.08U | 0.08UJ | 0.08U | 0.08U |
| MIREX | UG/L | 0.1 | 0.IU | 0.1UJ | 0.1U | 0.10 |
| TOXAPHENE | UG/L | 0.49 | 0.49U | 0.49UJ | 0.49U | 0.49U |
| AZINPHOS METHYL | UG/L | 0.58 | 0.58U | 0.58U | 0.58U | 0.58U |
| DEMETON | UG/L | 2 | 2U | 2U | 2U | 2U |
| ETHYL PARATHION | UG/L | Ι | 10 | 1U | IU | IU |
| MALATHION | UG/L | 0.22 | 0.22U | 0.22U | 0.22U | 0.22U |
| METHYL PARATHION | UG/L | 0.24 | 0.24U | 0.24U | 0.24U | 0.24U |
| AROCLOR 1016 | UG/L | 0.33 | 0.33U | 0.33UJ | 0.33U | 0.33U |
| AROCLOR 1221 | UG/L | 0.32 | 0.32U | 0.32UJ | 0.32U | 0.32U |
| AROCLOR 1232 | UG/L | 0.29 | 0.29U | 0.29UJ | 0.29U | 0.29U |
| AROCLOR 1242 | UG/L | 0.3 | 0.3U | 0.3UJ | 0.3U | 0.3U |
| AROCLOR 1248 | UG/L | 0.09 | 0.09U | 0.09UJ | 0.09U | 0.09U |
| AROCLOR 1254 | UG/L | 0.44 | 0.44U | 0.44UJ | 0.44U | 0.44U |
| AROCLOR 1260 | UG/L | 0.41 | 0.41U | 0.4IUJ | 0.41U | 0.41U |

U = not detected. J = estimated value.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 7-12MAXIMUM CONCENTRATIONS OF CHLORINATED PESTICIDES, ORGANOPHOSPHORUS PESTICIDES, ANDPCB AROCLORS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDEPLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | | | | | Brewerton | C&D | C&D | | | | | Cralebill | | | | · · · · · · · · · · · · · · · · · · · | |
|--------------------|---------|----------|-----------|----------|-----------|---------------|-------------|----------|-----------|------------|------------|-----------|---------------|----------|--------|------------|---------------------------------------|---------------|
| | | | Ocean | Outside | Inside | Eastern | Approach | Approseh | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| | R | EACH ID: | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| # EL. | TRIATES | TESTED: | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| | | | | | | | | | | | | | | | | | | |
| ANALYTE | UNIT | MIN DL | | | | | | | - | _ | | | | | | _ | | |
| ,4'-DDD | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| 4'-DDE | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| ,4'-DDT | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02Ū |
| LDRIN | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| LPHA-BHC | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| ROCLOR 1016 | UG/L | 0.33 | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U | 0.33U |
| ROCLOR 1221 | UG/L | 0.32 | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U |
| ROCLOR 1232 | UG/L | 0.29 | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U | 0.29U |
| ROCLOR 1242 | UG/L | 0.3 | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U | 0.3U |
| AROCLOR 1248 | UG/L | 0.09 | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | U90.0 | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U | 0.09U |
| ROCLOR 1254 | UG/L | 0.44 | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U | 0.44U |
| AROCLOR 1260 | UG/L | 0.41 | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U | 0.41U |
| BETA-BHC | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 2 6 6 0.033 | 0.01U | 0.01U | 14 0;DIJ | 10:01J | 18.0×0.01J | 10101J | HA-240.92J | 0.03J | 0.03J | 0.02J | 0.025 | A |
| CHLORDANE | UG/L | 0.1 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.10 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U |
| CHLORBENSIDE | UG/L | 0,1 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U |
| DACTHAL | UG/L | 0.1 | 0.3U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U |
| DELTA-BHC | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.0IU | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| DIELDRIN | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| ENDOSULFAN I | UG/L | 0.01 | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U | 0.01U |
| ENDOSULFAN II | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| ENDOSULFAN SULFATE | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| ENDRIN | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| ENDRIN ALDEHYDE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| GAMMA-BHC | UG/L | 0.0081 | 0.0081U | 0.0081U | 19.0 1.21 | 0.0081U | A | 0.0081U | 0.0081U | 0.01 | 10.0.0.L | AKC-0.01J | Service 0.01J | 0.02 | 0.013 | 0.02 | 0.0092.1 | 44.5.0.0.02 |
| HEPTACHLOR | UG/L | 0.02 | 0.02U | | 0.05 | | 0.02 | 0.07 | 0.04 | 0.07 | 10.19 | | 0.02U | 280.05 | 1.06 | A | 15 5 0.04 | D.M. 414.0.03 |
| HEPTACHLOR EPOXIDE | UG/L | 0.01 | 0.01U | 0.01U | 0.03J | AV. 1.4.0.02J | 24 9 0.07 | 0.01U | 0.01U | 0.023 | Mar 1 0.03 | A6.10.02J | 0.03J | ## 0.03J | 10.04J | 0.03 | 4××0.03J | U.0.03J |
| METHOXYCHLOR | UG/L | 0.08 | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U | 0.08U |
| MIREX | UG/L | 0.1 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.10 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.10 | 0.1U |
| TOXAPHENE | UG/L | 0.49 | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U | 0.49U |
| AZINPHOS METHYL | UG/L | 0.58 | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U | 0.58U |
| DEMETON | UG/L | 2 | 20 | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 2U | 20 | 2U | 20 |
| ETHYL PARATHION | UG/L | i | 10 | 10 | 10 | 10 | iU | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| MALATHION | UG/L | 0.22 | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U | 0.22U |
| METHYL PARATHION | UG/L | 0.24 | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U | 0.24U |

U = not detected. J = estimated value.

MIN DL = minimum detection limit.

NOTE: Shaded and bolde





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| | RE | ACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cntoff Angle | Swan Point | Tolchester North | Tolchester South | Toichester Straightening |
|--------------------|--------|---------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|---------------------------|------------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| # ELUT | RIATES | TESTED: | 101 | 1(*) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MIN DL | | | | | | | | | | | | | | | | |
| 4,4'-DDD | UG/L | 0.01 | 0.05 | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| 4,4'-DDE | UG/L | 0.02 | 0.01U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| 4,4'-DDT | UG/L | 0.02 | 0.02U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| ALDRIN | UG/L | 0.02 | 0.02U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 Ū | 0.02 U | 0.02 U | 0.02 U |
| ALPHA-BHC | UG/L | 0.01 | 0.02U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| AROCLOR 1016 | UG/L | 0.33 | 0.01U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U | 0.33 U |
| AROCLOR 1221 | UG/L | 0.32 | 0.33U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U |
| AROCLOR 1232 | UG/L | 0.29 | 0.32U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U | 0.29 U |
| AROCLOR 1242 | UG/L | 0.3 | 0.29U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U | 0.3 U |
| AROCLOR 1248 | UG/L | 0.09 | 0.3U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U | 0.09 U |
| AROCLOR 1254 | UG/L | 0.44 | 0.09U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U | 0.44 U |
| AROCLOR 1260 | UG/L | 0.41 | 0.44U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U | 0.41 U |
| BETA-BHC | UG/L | 0.01 | 0.41U | 0.01 U | 0.01 U | Stace 0.02 | 0.01 U | 0.01 U | 10.0 4 4 4 | AL W20.01 | Bige32 0.0.01 | A | A. 0.01 | 0.02 | 227×0.02 | 14550 0.02 | 1 1 (I O.D. | |
| CHLORDANE | UG/L | 0.1 | 0.1U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U |
| CHLOROBENSIDE | UG/L | 0.1 | 0.1U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U |
| DACTHAL | UG/L | 0.1 | 0.3U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U |
| DELTA-BHC | UG/L | 0.01 | 0.01U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| DIELDRIN | UG/L | 0.01 | 0.01U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| ENDOSULFAN I | UG/L | 0.01 | 0.01U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U | 0.01 U |
| ENDOSULFAN II | UG/L | 0.02 | 0.02U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| ENDOSULFAN SULFATE | UG/L | 0.02 | 0.02U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| ENDRIN | UG/L | 0.03 | 0.03U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| ENDRIN ALDEHYDE | UG/L | 0.03 | 0.03U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| GAMMA-BHC | UG/L | 0.0081 | 0.0081U | 0.0081 U | 20.01 | 0.0081 U | | 0.0081 U | 0.0081 U | 19.0 34 4 | 115165(10.01 | 10.01 | 10.01 | ACT 0.01 | 0.01 | 0.01 | 0.01 | 0.01 |
| HEPTACHLOR | UG/L | 0.02 | 0.02U | 0.03 | 10KA 0.03 | 0.03 | 0.05 | See 140.05 | 22.0.03 | A | ····· 0.105 | 100 C.05 | 0.02 U | 0.04 | 2750.04 | A | 0.03 | Antes 0.03 |
| HEPTACHLOR EPOXIDE | UG/L | 0.01 | 0.01U | 0.01 U | LA :0.017 | 10 ANO.015 | Sen- 0.92 | 0.01 U | 0.01 U | | .02 | State 0.02 | 200.02 | × | 241.0.03 | Bing 4 8.92 | LAT | 0.02 |
| METHOXYCHLOR | UG/L | 0.08 | 0.0026U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U | 0.08 U |
| MIREX | UG/L | 0.1 | 0.08U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U |
| TOXAPHENE | UG/L | 0.49 | 0.01 U | 0.49 U | 0.49 Ü | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U | 0.49 U |
| AZINPHOS METHYL | UG/L | 0.58 | 0.58U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U | 0.58 U |
| DEMETON | UG/L | 2 | 2U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U | 2.0 U |
| ETHYL PARATHION | UG/L | i | IU | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U | 1.0 U |
| MALATHION | UG/L | 0.22 | 0.22U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U | 0.22 U |
| METHYL PARATHION | UG/L | 0.24 | 0.24U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U | 0.24 U |

U = not detected.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

TABLE 7-14CONCENTRATIONS OF PCB CONGENERS (UG/L) IN RECEIVINGWATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND ATTHE OCEAN REFERENCE SITE

| | Sa | ample ID: | KI-03 | KI-07 | KI-14 | |
|------------------------|------|-----------|-----------------|-----------------|------------------|-----------------|
| | F | Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | MIN DL | | | | |
| BZ# 8* | UG/L | 0.003 | 0.01B | 0.003U | 0.003U | 0.003U |
| BZ# 18* | UG/L | 0.0064 | 0.0064UJ | 0.0064UJ | 0.0064U | 0.0064U |
| BZ# 28* | UG/L | 0.0065 | 0.0065UJ | 0.0065UJ | 0.0065U | 0.0065U |
| BZ# 44* | UG/L | 0.0055 | 0.0055U | 0.0055U | 0.0055U | 0.0055U |
| BZ# 49 | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U |
| BZ# 52* | UG/L | 0.0022 | 0.0022U | 0.0022U | 0.0022U | 0.0022U |
| BZ# 66* | UG/L | 0.0004 | 0.0004U | 0.0004U | 0.0004U | 0.0004U |
| BZ# 77* | UG/L | 0.0025 | 0.0025U | 0.0025U | 0.0025U | 0.0025U |
| BZ# 87 | UG/L | 0.0012 | 0.0012U | 0.0012U | 0.0012U | 0.0012U |
| BZ# 101* | UG/L | 0.0026 | 0.0026U | 0.0026U | 0.0026U | 0.0026U |
| BZ# 105* | UG/L | 0.0034 | 0.0034U | 0.0034U | 0.0034U | 0.0034U |
| BZ# 118* | UG/L | 0.0018 | 0.0018U | 0.0018U | 0.0018U | 0.0018U |
| BZ# 126* | UG/L | 0.0013 | 0.0022U | 0.0022U | 0.0022U | 0.0013U |
| BZ# 128* | UG/L | 0.0022 | 0.0013U | 0.0013U | 0.0013U | 0.0022U |
| BZ# 138* | UG/L | 0.0013 | 0.0013U | 0.0013U | 0.0013U | 0.0013U |
| BZ# 153* | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U |
| BZ# 156 | UG/L | 0.0012 | 0.0012U | 0.0012U | 0.0012U | 0.0012U |
| BZ# 169* | UG/L | 0.0022 | 0.0022U | 0.0022U | 0.0022U | 0.0022U |
| BZ# 170* | UG/L | 0.0014 | 0.0014U | 0.0014U | 0.0014U | 0.0014U |
| BZ# 180* | UG/L | 0.0015 | 0.0015U | 0.0015U | 0.0015U | 0.0015U |
| BZ# 183 | UG/L | 0.0017 | 0.0017U | 0.0017U | 0.0017U | 0.0017U |
| BZ# 184 | UG/L | 0.001 | 0.001U | 0.001U | 0.001U | 0.001U |
| BZ# 187* | UG/L | 0.0053 | 0.0053U | 0.0053U | 0.0053U | 0.0053U |
| BZ# 195 | UG/L | 0.0017 | 0.0017U | 0.0017U | 0.0017U | 0.0017U |
| BZ# 206 | UG/L | 0.0024 | 0.0024U | 0.0024U | 0.0024U | 0.0024U |
| BZ# 209 | UG/L | 0.0026 | 0.0026UJ | 0.0026UJ | 0.0026U | 0.0026U |
| TOTAL PCB (ND=0)** | UG/L | - | 0.02 | 0.06 | 0 | 0 |
| TOTAL PCB (ND=1/2DL)** | UG/L | - | 0.0695 | 0.109 | 0.0525 | 0.0525 |
| TOTAL PCB (ND=DL) | UG/L | - | 0.119 | 0.159 | 0.105 | 0.105 |

U = not detected. B = found in blank. J = estimated value.

* = PCB congeners used for total PCB summation, as per Table 9-3 of the ITM (USEPA/USACE 1998). Sum multiplied by a factor of 2 as per NOAA (1993)

MIN DL = minimum detection limit.

Note: Shaded and bolded values represent mean concentrations for analytes detected

ist one sample. Means calculated with ND=DL.

**Note that the mean of the total PCBs for individual samples is not equivalent to the sum of mean individual PCBs for ND=0 and ND=1/2DL.

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TABLE 7-15MAXIMUM CONCENTRATIONS OF PCB CONGENERS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH
CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | RE | EACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------|----------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| | # ELUTRIATES | TESTED: | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | DL | | | | | | | | | | | | | | | | |
| BZ# 8* | UG/L | 0.003 | 0.003U | A. 0.0077J | 0.003U | 0.01 | . 0.0052J | 244.9.8.01 | 0.003U | 0.0066J | 31 0.0LJ | | 0.003U | 0.003U | 0.003U | . St. 4. 0.01J | 0.003U | 0.0031J |
| BZ# 18* | UG/L | 0.0064 | 0.0012U | 0.0064U | 0.0064U | 0.0064U | 2 8.0068J | · (16.0. (15) | 0.0064U | LIG.0 | 0.0064U | 0.01J | 0.0064U | | 0.0078J | 0.01 | 0.0064U | 12 - 3 0.01J |
| BZ# 28* | UG/L | 0.0065 | 0.0022U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U | 0.0065U |
| BZ# 44* | UG/L | 0.005 | 0.003U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U | 0.0055U |
| BZ# 49 | UG/L | 0.003 | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U |
| BZ# 52* | UG/L | 0.0022 | 0.0064U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U |
| BZ# 66* | UG/L | 0.0004 | 0.0018U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U | 0.0004U |
| BZ# 77* | UG/L | 0.0025 | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U | 0.0025U |
| BZ# 87 | UG/L | 0.0012 | 0.0055U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U |
| BZ# 101* | UG/L | 0.0026 | 0.0022U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U |
| BZ# 105* | UG/L | 0.0034 | 0.0065U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U | 0.0034U |
| BZ# 118* | UG/L | 0.0018 | 0.0034U | 0.0018U | 0.0018U | Le100.6 | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U | 0.0018U |
| BZ# 126* | UG/L | 0.0022 | 0.0025U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U |
| BZ# 128* | UG/L | 0.0013 | 0.0015U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U |
| BZ# 138* | UG/L | 0.0013 | 0.001U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0013U | 0.0016J | 0.0013U | 0.0013U | 0.0013U | 0.0013U | A |
| BZ# 153* | UG/L | 0.003 | 0.0026U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U | 0.003U |
| BZ# 156 | UG/L | 0.0012 | 0.0004U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U | 0.0012U |
| BZ# 169* | UG/L | 0.0022 | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U | 0.0022U |
| BZ# 170* | UG/L | 0.0014 | 0.0013U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0014U | 0.0033J |
| BZ# 180* | UG/L | 0.0015 | 0.0053U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U | 0.0015U |
| BZ# 183 | UG/L | 0.0017 | 0.0013U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U |
| BZ# 184 | UG/L | 0.001 | 0.0012U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.001U | 0.00IU |
| BZ# 187* | UG/L | 0.0053 | 0.0017U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U | 0.0053U |
| BZ# 195 | UG/L | 0.0017 | 0.0014U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U | 0.0017U |
| BZ# 206 | UG/L | 0.0024 | 0.0017U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U | 0.0024U |
| BZ# 209 | UG/L | 0.0026 | 0.0024U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U | 0.0026U |
| TOTAL PCB (ND=0)** | UG/L | • | 0 | 0.0154 | 0 | 0.0238 | 0.0136 | 0.04 | 0 | 0.0332 | 0.02 | 0.0334 | 0.0032 | 0.02 | 0.0156 | 0.04 | 0 | 0.0328 |
| TOTAL PCB (ND=1/2DL)** | UG/L | | 0.0525 | 0.0649 | 0.0525 | 0.0715 | 0.0599 | 0.0831 | 0.0525 | 0.0763 | 0.0695 | 0.0765 | 0.0544 | 0.0661 | 0.0617 | 0.0831 | 0.0525 | 0.0745 |
| TOTAL PCB (ND=DL) | UG/L | | 0.105 | 0.114 | 0.105 | 0.119 | 0.109 | 0.126 | 0.105 | 0.119 | 0.119 | 0.12 | 0.106 | 0.112 | 0.108 | 0.126 | 0.105 | 0.116 |

U = not detected. J = estimated value

* = PCB congeners used for total PCB summation, as per Table 9-3 of the ITM (USEPA/USACE 1998). Sum multiplied by a factor of 2 as per NOAA 1993.

MIN DL = minimum detection limit.

NOTE: Shaded and bolded values represent detected concentrations.

**Note that the mean of the total PCBs for individual samples is not equivalent to the sum of mean individual PCBs for ND=0 and ND=1/2DL.

| | R | EACH ID: | Ocean Reference | Ootside Site 104 | Inside Site | Brewerton Eastern Extension | C&D Approach (Sorficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighili Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------|----------|--------------------|---------------------|-------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|---|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | # ELUTRIATES | TESTED: | 1(1) | 1 (*) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MINDI | | | | | | | | | | | | | | | | |
| BZ# 8* | UG/L | 0.003 | 0.003U | 1.000 0077 | 0.003 U | | | | 0.003 11 | Sector 1 | ALL | No.2 A PLAN | 0.003.11 | 0.003.11 | 0.003 1 | | 0.003 [] | |
| BZ# 18* | UG/L | 0.0064 | 0.0012U | 0.0064 U | 0.0064 U | 0.0064 U | | Sec. 6. 6042 | 0.0064 U | | 0.0064 U | | 0.0064 U | 3 3.0017 | 6.0030 | 0.00103 | 0.0050 | 0.0031 |
| BZ# 28* | UG/L | 0.0065 | 0.0022U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | 0.0065 U | D.0065 U | 0.0065 U | 0.0065 U | 0.0065 U |
| BZ# 44* | UG/L | 0.005 | 0.003U | 0.0055 U | 0.0055 U | 0.00525 U | 0.0055 U | 0.0055 U | 0.0055 U | 0.0055 U | 0.0055 U | 0.0055 U | 0.0055 U | 0.0055 U | D.0055 U | 0.0055 U | 0.0055 U | 0.0055 U |
| BZ# 49 | UG/L | 0.003 | 0.003U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 11 |
| BZ# 52* | UG/L | 0.0022 | 0.0064U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 11 |
| BZ# 66* | UG/L | 0.0004 | 0.0018 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U | 0.0004 U |
| BZ# 77* | UG/L | 0.0025 | 0.0025U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | 0.0025 U | D.0025 U | 0.0025 U | 0.0025 U | 0.0025 U |
| BZ# 87 | UG/L | 0.0012 | 0.0055U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U |
| BZ# 101* | UG/L | 0.0026 | 0.0022U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U |
| BZ# 105* | UG/L | 0.0034 | 0.0065U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U | 0.0034 U |
| BZ# 118* | UG/L | 0.0018 | 0.0034U | 0.0018 U | 0.0018 U | A | 0.0018 U | 0.0018 U | 0.0018 U | 0.0018 U | 0.0018 U | 0.0018 U | 0.0018 U | 0.0018 U | D.0018 U | 0.0018 U | 0.0018 U | 0.0018 U |
| BZ# 126* | UG/L | 0.0022 | 0.0025U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 11 |
| BZ# 128* | UG/L | 0.0013 | 0.0015U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U |
| BZ# 138* | UG/L | 0.0013 | 0.001U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.0013 U | 0.00145 | 0.0013 U | D.0013 U | 0.0013 U | 0.0013 U | 0.0015 |
| BZ# 153* | UG/L | 0.003 | 0.0026U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U | 0.003 U |
| BZ# 156 | UG/L | 0.0012 | 0.0004U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U | 0.0012 U |
| BZ# 169* | UG/L | 0.0022 | 0.0022U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U | 0.0022 U |
| BZ# 170* | UG/L | 0.0014 | 0.0013U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | 0.0014 U | D.0014 U | 0.0014 U | 0.0014 U | 0.0024 |
| BZ# 180* | UG/L | 0.0015 | 0.0053U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U | 0.0015 U |
| BZ# 183 | UG/L | 0.0017 | 0.0013U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U |
| BZ# 184 | UG/L | 0.001 | 0.0012U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U | 0.001 U |
| BZ# 187* | UG/L | 0.0053 | 0.0017U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | 0.0053 U | D.0053 U | 0.0053 U | 0.0053 U | 0.0053 U |
| BZ# 195 | UG/L | 0.0017 | 0.0014U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 U | 0.0017 L | J D.0017 U | 0.0017 U | 0.0017 U | 0.0017 U |
| BZ# 206 | UG/L | 0.0024 | 0.0017U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 L | 0.0024 U | 0.0024 U | 0.0024 U | 0.0024 U |
| BZ# 209 | UG/L | 0.0026 | 0.0024U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 L | 0.0026 U | 0.0026 U | 0.0026 U | 0.0026 U |
| TOTAL PCB (ND=0)** | UG/L | • | 0 | 0.0154 | 0 | 0.0119 | 0.012 | 0.02 | 0 | 0.0166 | 10.0 | 0.0167 | 0.0016 | 0.01 | 0.0078 | 0.0253 | 0 | 0.0181 |
| TOTAL PCB (ND=1/2DL)** | UG/L | • | 0.0525 | 0.0649 | 0.0525 | 0.0617 | 0.0598 | 0.0678 | 0.0525 | 0.0644 | 0.061 | 0.0645 | 0.0535 | 0.0593 | 0.0571 | 0.0716 | 0.0525 | 0.0645 |
| TOTAL PCB (ND=DL) | UG/L | | 0.105 | 0.114 | 0.105 | 0.112 | 0.108 | 0.116 | 0.105 | 0.112 | 0.112 | 0.112 | 0.105 | 0.109 | 0.106 | 0.118 | 0.105 | 0.111 |

TABLE 7-16 MEAN CONCENTRATIONS OF PCB CONGENERS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

U = not detected.

* = PCB congeners used for Total PCB summation, as per Table 9-3 of the ITM (USEPA/USACE 1998).

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

Note: Shaded and bolded values represent mean concentrations for analytes detected in at least one sample. Means calculated with ND=DL.

**Note that the mean of the total PCBs for iodividual samples is not equivalent to the sum of mean individual PCBs for ND=0 and ND=1/2DL.





TABLE 7-17CONCENTRATIONS OF PAHs (UG/L) IN RECEIVING WATERINSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104, AND ATTHE OCEAN REFERENCE SITE

| | Sample ID: | | | KI-07 | KI-14 | |
|------------------------|------------|---------|-----------------|-----------------|-------------------------|-----------------|
| | Re | ach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | DL | | | | _ |
| ACENAPHTHENE | UG/L | 0.39 | 0.39U | 0.39U | 0.39U | 0.39U |
| ACENAPHTHYLENE | UG/L | 0.38 | 0.38U | 0.38U | 0.38U | 0.38U |
| ANTHRACENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZ[A]ANTHRACENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZO[A]PYRENE | UG/L | 0.04 | 0.04U | 0.04U | 0.04U | 0.04U |
| BENZO[B]FLUORANTHENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZO[G,H,I]PERYLENE | UG/L | 0.06 | 0.06U | 0.06U | 0.06U | 0.06U |
| BENZO[K]FLUORANTHENE | UG/L | 0.05 | 0.05U | 0.05U | 0.05U | 0.05U |
| CHRYSENE | UG/L | 0.02 | 0.02U | 0.02U | 0.02U | 0.02U |
| DIBENZ[A,H]ANTHRACENE | UG/L | 0.06 | 0.06U | 0.06U | 0.06U | 0.06U |
| FLUORANTHENE | UG/L | 0.04 | 0.04U | 0.04U | 0.04U | 0.04U |
| FLUORENE | UG/L | 0.06 | 0.06U | 0.06U | 0.06U | 0.06U |
| INDENO[1,2,3-CD]PYRENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U |
| NAPHTHALENE | UG/L | 0.32 | 0.32U | 0.32U | 0.32U | 0.32U |
| PHENANTHRENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U |
| PYRENE | UG/L | 0.06 | 0.06U | 0.06U | 0.06U | 0.06U |
| TOTAL PAH (ND=0) | UG/L | * | 0 | 0 | 0 | 0 |
| TOTAL PAH (ND=1/2DL) | UG/L | - | 0.82 | 0.82 | 0.82 | 0.82 |
| TOTAL PAH (ND=DL) | UG/L | - | 1.63 | 1.63 | 1.63 | 0.81 |

 $\mathbf{U} = \text{not detected}.$

MIN DL = minimum detection limit.
TABLE 7-18 MAXIMUM CONCENTRATIONS OF PAHs (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | R | EACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewertou Eastern Eatension | C&D Appreach (Surficial) | C&D Approach (Core) | Creighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Catoff Angle | Swam Point | Tolchester North | Toichester South | Tolchester Straightening |
|------------------------|------------|----------|--------------------|---------------------|--------------------|-----------------------------------|-----------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | ELUTRIATES | TESTED: | 1 | 1 | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MEN DL | | | _ | | | | | | | | | | | | | |
| ACENAPHTHENE | UG/L | 0.39 | 0.39U | 0.39U | 0 39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U | 0.39U |
| ACENAPHTHYLENE | UG/L | 0.38 | 0 38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U | 0.38U |
| ANTHRACENE | UG/L | 0.03 | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZ[A]ANTHRACENE | UG/L | 0.03 | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0 03U | 0.03U | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZO[A]PYRENE | UG/L | 0.04 | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U |
| BENZO[B]FLUORANTHENE | UG/L | 0.03 | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| BENZO[G,H,I]PERYLENE | UG/L | 0.06 | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0 06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U |
| BENZO[K]FLUORANTHENE | UG/L | 0.05 | 0 05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U | 0.05U |
| CHRYSENE | UG/L | 0.02 | 0 02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U | 0.02U |
| DIBENZ[A,HJANTHRACENE | UG/L | 0.06 | 0 06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U |
| FLUORANTHENE | UG/L | 0.04 | 0.04U | 0 04U | 0 04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U | 0.04U |
| FLUORENE | UG/L | 0.06 | 0.06U | 0.06U | Distant a. OTP | 0.06U | 0.06U | 0.06U | 0.06U | 2 2 3 4 0 B | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U |
| INDENO[1,2,3-CD]PYRENE | UG/L | 0.03 | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U |
| NAPHTHALENE | UG/L | 0.32 | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U | 0.32U |
| PHENANTHRENE | UG/L | 0.03 | 0.03U | 0.03U | 0 03U | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0.03U | 0 03U | 0.03U | 0.03U | 0.03U | 0.03U |
| PYRENE | UG/L | 0.06 | 0.06U | 0.06U | 101 | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.06U | 0.060 | 0.06U | 0.06U | 0.06U | 0.06U |
| TOTAL PAH (ND=0) | UG/L | | 0 | 0 | 0.07 | 0 | 0 | 0 | 0 | 0.08 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL PAH (ND=1/2DL) | UG/L | | 0.81 | 0.81 | 0.85 | 0.81 | 0.81 | 0.81 | 0.81 | 0.86 | 0.81 | 0.81 | 0.81 | 0.81 | 0.81 | 0.81 | 0.81 | 0.81 |
| TOTAL PAH (ND=DL) | UG/L | | 1.63 | 1.63 | 1.64 | 1.63 | 1.63 | 1.63 | 1.63 | 1.65 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 |

U = not detected. P = >25% between two GC columns.

MIN DL = minimum detection limit.



TABLE 7-19 MEAN CONCENTRATIONS OF PAHs (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | RE | ACH ID: | Occan Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------------|------------|---------|--------------------|---------------------|--------------------|--------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| | ELUTRIATES | TESTED: | 1 (6) | 1(*) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MIN DL | | | | 1-2013 | | | | | | 1 | | | | | | |
| CENAPHTHENE | UG/L | 0.39 | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0 39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U | 0.39 U |
| CENAPHTHYLENE | UG/L | 0.38 | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U | 0.38 U |
| NTHRACENE | UG/L | 0.03 | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0 03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| BENZ AJANTHRACENE | UG/L | 0.03 | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0 03 U |
| BENZO(A)PYRENE | UG/L | 0.04 | 0.04 U | 0 04 U | 0 04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0 04 U | 0.04 U | 0.04 U |
| BENZO(B)FLUORANTHENT | UG/L | 0.03 | 0 03 U | 0 03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| BENZO(G,H,I)PERYLENE | UG/L | 0.06 | 0.06 U | 0 06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U |
| SENZO(K)FLUORANTHEN | E UG/L | 0.05 | 0 05 U | 0.05 U | 0.05 U | 0.05 U | 0.05 U | 0.05 U | 0 05 U | 0.05 U | 0.05 U | 0.05 U | 0.05 U | 0.05 U | 0 05 U | 0.05 U | 0.05 U | 0.05 U |
| CHRYSENE | UG/L | 0.02 | 0.02 U | 0 02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U | 0.02 U |
| DIBENZ[A,H]ANTHRACEN | E UG/L | 0.06 | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U |
| LUORANTHENE | UG/L | 0.04 | 0.04 U | 0 04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U | 0.04 U |
| LUORENE | UG/L | 0.06 | 0.06 U | 0.06 U | Million Co | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 10201 12201 | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U |
| NDENO[1,2,3-CD]PYRENE | UG/L | 0.03 | 0.03 U | 0 03 U | 0 03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| NAPHTHALENE | UG/L | 0.32 | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U | 0.32 U |
| HENANTHRENE | UG/L | 0.03 | 0.03 U | 0 03 U | 0 03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U | 0.03 U |
| YRENE | UG/L | 0.06 | 0.06 U | 0.06 U | C. C. S. LUS | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U | 0.06 U |
| TOTAL PAH (ND=0) | UG/L | • | 0 | 0 | 0.06 | 0 | 0 | 0 | 0 | 0.07 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL PAHS (ND=1/2DL) | UG/L | • | 0.82 | 0.82 | 0.84 | 0.82 | 0.82 | 0.82 | 0.82 | 0.84 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 | 0.82 |
| TOTAL PAH (ND=DL) | UG/L | • | 1.63 | 1.63 | 1.64 | 1.63 | 1.63 | 1.63 | 1.63 | 1.64 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 | 1.63 |

U = not detected.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

TABLE 7-20CONCENTRATIONS OF METALS (UG/L) IN RECEIVINGWATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104,
AND AT THE OCEAN REFERENCE SITE

| Sample ID: | KI-03 | KI-07 | KI-14 | |
|------------|-----------------|-----------------|------------------|-----------------|
| Reach ID: | Inside Site 104 | Inside Site 104 | Outside Slte 104 | Ocean Reference |

| ANALYTE | UNIT | MIN DI | 4 | | | |
|-----------|------|--------|-------|-------|-------|-------|
| ALUMINUM | UG/L | 37.9 | 169J | 1593 | 296 | 255 |
| ANTIMONY | UG/L | 1 | 10 | 1U | 1U | 1U |
| ARSENIC | UG/L | 1.7 | 1.7U | 1.7U | 2.5B | 1.7U |
| BERYLLIUM | UG/L | 0.1 | 0.1U | 0.1U | 0.1U | 0.1U |
| CADMIUM | UG/L | 0.2 | 0.2U | 0.2U | 0.2UJ | 0.2UN |
| CHROMIUM | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | 0.7UN |
| COPPER | UG/L | 0.7 | 0.98B | 1B | 0.93B | 0.7U |
| IRON | UG/L | 5.5 | 45J | 55.1J | 289 | 28.4B |
| LEAD | UG/L | 1.1 | 1.1UJ | 1.1UJ | 1U | 2.2UN |
| MANGANESE | UG/L | 2.4 | 7.53 | 12.4J | 6.7J | 2.4U |
| MERCURY | UG/L | 0.1 | 0.1U | 0.1U | 0.1U | 0.1UN |
| NICKEL | UG/L | 2.4 | 2.4U | 2.4U | 2.1J | 2.4UN |
| SELENIUM | UG/L | 1.8 | 1.8U | 1.8U | 1.8U | 1.8U |
| SILVER | UG/L | 2.2 | 2.2U | 2.2U | 3.2U | 2.2U |
| THALLIUM | UG/L | 10 | 10UJ | 10UJ | 8.5UJ | 17UNW |
| ZINC | UG/L | 1.7 | 1.7U | 1.7U | 2.3U | 1.7U |

U = not detected. **B** = value is <RL but >IDL/MDL. **J** = estimated value.

N = spiked sample not within control limits. W = AAS out of control limits.

MIN DL = minimum detection limit.

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TABLE 7-21MAXIMUM CONCENTRATIONS OF METALS (UG/L) IN ELUTRIATES FROM BALTIMORE APPROACHCHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | RI | ACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Toichester South | Tolchester Straightening |
|-----------|----------------|---------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| # ELU | TRIATES | TESTED: | 1(1) | 1(0) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| ANALYTE | UNIT | MIN DL | | | | | | | | | | | | | | | | |
| ALUMINUM | UG/L | 37.9 | 236 | SAR 17AB | 198173B | 158B | 18 1 2 2 3 4 2 8 | 8489115B | 200 144B | 103B | U | 150 1703 | 140B | 119B | 152B | 179J | 141133B | 34 HANS 176B |
| ANTIMONY | UG/L | 1 | 10 | alex378 | 19883/7B | ***** 2.9 B | 化学会会演讲18 | S. S. S. B. | 233310月 | 1. J. H | 出来的2.0 | 消息数2.5B | 10 | 3.6B | 1233.2B | 公司公主 5日 | 学生的 (28) | 111月 11日 |
| ARSENIC | UG/L | 1.7 | 1.70 | a3208.5B | 吴君影教6B | 出现的现在分月 | 计图记的 4 段 B | -1 29B | 75-12-3月 | 周2、10m | 1223年11月 | 自由的现在。7B | and sha | 2347.2B | 家庭前月28 | 注册时间5位日 | 2000138 | 26200 al 33.3 |
| BERYLLIUM | UG/L | 0.1 | 0.IU | 0.1U | 0.1U | 0.1U | 0.10 | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 2444990.533 |
| CADMIUM | UG/L | 0.2 | 0.2UN | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.20 | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U | 0.2U |
| CHROMIUM | UG/L | 0.7 | 0.7UN | 0.91B | 0.7U | and makeling | Sec. 13B | 达公式中国 | 0.7U | 國家運用 | 5. ⁵¹ 1.8B | 88.1.8B | · 资金条件日 | 0.7U | 28-1.2B | 0.70 | TEST ALB | SUSAR AAB |
| COPPER | UG/L | 0.7 | 0.7U | 0.7U | 0.7U | ·法律书》[2] B | 空口2485.8B | NOT A KIE | 0.7U | 1288年1月 | 进行上, 20.8B | 0.7U | 0.7U | 244-1.1B | 0.7U | 是自由的 3B | 0.7U | 2.9B |
| IRON | UG/L | 5.5 | Hand 0.9B | 5.5U | NR:230B | -2-Wa193B | 5.5U | 5.5U | 影响148 | 1928 | | 10428486 | 5.5U | 法的19,4B | 1208 | 1940 | 2 1020 | 3429年7.3B |
| LEAD | UG/L | 1.1 | 2.2UN | 1.IU | 2.2U | 1.10 | 1.10 | 1.10 | 2.2U | 1.10 | 1.10 | 2.2U | 2.2U | 1.10 | 2.20 | 1.10 | 1.10 | 24888233833 |
| MANGANESE | UG/L | 2.4 | 2.4U | 3420 | 4311680 | 7830 | AREA (690 | 316643310 | 892 | 250 | 3360 | 21610 | 1210 | 7,020 | A 113950 | 24980 | HAM-3540 | 2471) |
| MERCURY | UG/L | 0.1 | 0.1UN | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.IU | 0.10 | 0.IU | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U | 0.1U |
| NICKEL | UG/L | 0.8 | 2.4UN | 24B | 28 23B | 10 Att 3.9B | A-138 | SHEET SH | 26B | 197 (F) | S.S.B | 12485558 | Contrad 1B | 11 8 4.6B | 3.8B | AND 5.9.1 | 6.2B | 4947#11.4B |
| SELENIUM | UG/L | 1.8 | 1.8U | 1.8U | 1.8U | 19B | | 1.8U | 1.8U | 1.8U | 1.8U | 1.8U | 1.8U | 228 | 1.8U | 1.8U | 1.8U | 129.2 |
| SILVER | UG/L | 2.2 | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.2U | 2.20 | 2.2U | 2.2U | 2.2U | ACT 23B |
| THALLIUM | UG/L | 1 | 17UNW | 5U | 10U | 10U | 5U | 10U | 100 | 10U | 10U | 10U | 10U | 100 | 100 | 10U | 10U | 10U |
| ZINC | UG/L | 1.7 | 1.7U | 1.7U | 1.7U | 1.70 | Max ALB | 1.70 | 1.70 | 1.70 | 1.70 | 1.70 | 1.70 | 1.7U | 112-1181 | 1.70 | 1.7U | 1.7U |

U = not detected. B = value is <RL but >1DL/MDL. J = estimated value. N = spiked sample not within control limits. W = AAS out of control limits.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

TABLE 7-22MEAN CONCENTRATIONS OF METALS (MG/L) IN ELUTRIATES FROM BALTIMORE APPROACHCHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | | | Brewerton | C&D | C&D | | | | | Craighill | | | | | Tolchester |
|-----------------------------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|--------------|
| | Ocean | Outside | Inskie | Eastern | Approach | Approach | | Cralghill | Cralghill | Cralghill | Upper | Cutoff | Swan | Tolchester | Tolchester | Stralghtenin |
| REACH ID: | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Cralghlll | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | g |
| # ELUTRIATES TESTED: | 1(1) | 1 (a) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |

| ANALYTE | UNIT | MIN DL | | | | | | | | | | | | | | | | | |
|-----------|------|--------|----------|---------------|-----------|---------------------------------------|---------------------|-----------|------------|-----------|------------|------------|-----------|--------------|----------|-------------|-------------|--------------------|---------|
| ALUMINUM | UG/L | 37.9 | SN: 236 | 北梁传动74 | 3154.67 | · · · · · · · · · · · · · · · · · · · | \$\$ 7257.85 | 82.85 | 293142.5 | Weia1535 | 84850345 | A | 803 A1125 | 3; (m. 108.8 | 145 | XXXXXXXXX | 128.5 | Sand | \$170 |
| ANTIMONY | UG/L | 1 | 10 | 200007 | Farmers 2 | 1.95 | 次/济绵和2.55 | 2.2 | 645 w.L.4 | A | A. Starlin | A\$141.7. | 1.0 U | 属其约23 | 1-12-2-1 | 39:221.75 | A STATE | 180 | 3.6 |
| ARSENIC | UG/L | 1.7 | 1.7U | 202328-5 | 87 5.07 | Matthe A.1 | 北京市内32 | 245 | 2 | | 3.95 | A.05 | 4 223.55 | 25-5 6.05 | 43.6 | <u></u> | 21. 2.3.05 | C.I.S. | 12182 |
| BERYLLIUM | UG/L | 0.1 | 0.1U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 0.1 U | 10123 | 0.315 |
| CADMIUM | UG/L | 0.2 | 0.2UN | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | 0.2 U | | 0.2 U |
| CHROMIUM | UG/L | 0.7 | 0.7UN | 1.91 | 0.7 U | 0.9 | SALESSAND | 学家48.0.85 | 0.7 U | S 656-03 | 1000001/25 | 225 | 0.85 | 0.7 U | 0.95 | 0.7 U | STREETS 0.9 | 48.66 | 約2.55 |
| COPPER | UG/L | 0.7 | 0.7U | 0.7 U | 0.7 U | 1988年19月1 | · 1888 - 1883 - 125 | 字第51135 | 0.7 U | 10118 195 | 故注意。075 | 0.7 U | 0.7 U | 13190,50.9 | 0.7 U | CANCEL OF | 0.7 U | 1.1 | 02.05 |
| IRON | UG/L | 5.5 | (当我)0.9B | 5.5 U | 1.80,83 | 自动测量12.4 | 5.5 U | 5.5 U | 以不可知 | 是主要因为 | 14 F 29 85 | 24575 | 5.5 U | 251 7.45 | 7 123.65 | 112.972.75 | 1.5.670 | Sec. | 166.4 |
| LEAD | UG/L | 1.1 | 2.2UN | 1.1 U | 1.83 U | 1.1 U | 1.1 U | 1.10 | 1.65 U | 1.1 U | 1.1 U | 1.65 U | 1.65 U | 1.1 U | 2.2 U | 1.10 | 1.1 U | | 1941 |
| MANGANESE | UG/L | 2.4 | 2.4U | 3420 | \$ 961.67 | 6625 | 2465 | 2245 | Star 841 | 37.80 | 2750 | 1565 | Act 1960 | STAR 5520 | 2620 | 16814931830 | 22105 | Sec. | 1850 |
| MERCURY | UG/L | 0.1 | 0.1UN | 0.1 U | 0.1 U | 0.1 U | DAT DISSO 20 | CAUSON D | 240.01U | CARE DALL | 22 (0.1 U | 18 1 0.4 U | NO ON U | X-220.1 U | 0.1 U | ALCONTRACT | 565011 | 10000 | 20.1«U |
| NICKEL | UG/L | 0.8 | 2.4UN | 1000 C 24 | 243 | 3.3 | States,45 | SPECIA75 | ¥CCM2452.5 | 149884435 | 4.05 | A 3.95 | SX 3.25 | 371235 | 111余禄33 | 1963.082 | 37 | and a state of the | \$ 7.25 |
| SELENIUM | UG/L | 1.8 | 1.8U | 1.8 U | 1.8 U | 1.85 | 1.85 | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 1.8 U | 2 2 | 1.8 U | 1.8 U | 1.8 U | 1.1 | A155 |
| SILVER | UG/L | 2.2 | 2.2U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 2.2 U | 22.35 | 2.25 |
| THALLIUM | UG/L | 1 | 17UNW | 5.0 U | 8.33 U | 7.5 U | 3.0 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | 7.5 U | | 7.5 U |
| ZINC | UG/L | 1.7 | 1.7U | 1.7 U | 1.7 U | 1.7 U | A 1244 2.9 | 1.7 U | 1.7 U | 1.7 U | 1.7 U | 1.7 U | 1.7 U | 1.7 U | ANDA1:75 | 1.7 U | 1.7 U | | 1.7 U |

U = not detected. B = value is <RL but >IDL/MDL. N = spiked sample not within control limits. W = AAS out of control limits.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.



TABLE 7-23CONCENTRATIONS OF BUTYLTINS (UG/L) IN RECEIVINGWATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT SITE 104,
AND AT THE OCEAN REFERENCE SITE

| Sample ID: | KI-03 | KI-07 | KI-14 | |
|------------|-----------------|-----------------|------------------|-----------------|
| Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |

ANALYTE UNIT MIN DL

| MONOBUTYLTIN | NG/L | 32 | 32U | 32U | 31U | 310U |
|---------------|------|----|-----|-----|-----|------|
| DIBUTYLTIN | NG/L | 39 | 39U | 40U | 39U | 280U |
| TRIBUTYLTIN | NG/L | 45 | 45U | 46U | 45U | 80J |
| TETRABUTLYTIN | NG/L | 51 | 5iU | 52U | 50U | 220U |

U = not detected. J = estimated value; value below lowest calibration

MIN DL = minimum detection limit.

TABLE 7-24 MAXIMUM CONCENTRATIONS OF BUTYLTINS (NG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|----------------------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | Ocean | Outside | Inside | Eastern | Approach | Approach | | Craighiil | Craighiil | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| REACII ID: | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Core) | Cralghiil | Angle East | Angle West | Entrance | Range | Angie | Point | North | South | Straightening |
| # ELUTRIATES TESTED: | l (s) | 1 (s) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |

| ANALYTE | UNIT | MIN DL | | | | | | | | | | | | | | | | |
|---------------|------|--------|------|-----|-----|-----|-----|-----|-----|-------------|-----|-----|-----|-----|-----|-----|-----|-----|
| MONOBUTYLTIN | NG/L | 32 | 310U | 33U | 34U | 32U | 33U | 34U | 34U | 33 U | 34U | 130 | 34U | 34U | 35U | 33U | 35U | 34U |
| DIBUTYLTIN | NG/L | 40 | 280U | 41U | 42U | 40U | 41U | 42U | 43U | 4IU | 42U | 41U | 42U | 42U | 43U | 41U | 43U | 43U |
| TRIBUTYLTIN | NG/L | 46 | 100J | 47U | 48U | 46U | 47U | 49U | 49U | 48U | 48U | 47U | 49U | 48U | 50U | 47U | 50U | 49U |
| TETRABUTLYTIN | NG/L | 52 | 220U | 53U | 54U | 52U | 53U | 55U | 56U | 54U | 54U | 53U | 55U | 54U | 56U | 53U | 56U | 56U |

U = not detected. J = estimated value below lowest calibrator.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.



TABLE 7-25 MEAN CONCENTRATIONS OF BUTYLTINS (NG/L) IN ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCE SITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | | | Brewerton | C&D | C&D | | | | | Creinhill | | | | | Tatat |
|-----------------------------|-----------|------------------|----------|-----------|-------------|----------|-----------|------------|------------|----------------|-----------|--------|-------|--|------------|--------------|
| | | | | Dienerton | cab | cab | | | | and the second | Craignin | | | 1. | | Toicnester |
| | Ocean | Outside | Inside | Eastern | Approach | Approach | | Craighiil | Craighill | Craighill | Upper | Cutoff | Swan | Toichester | Tolchester | Straightenin |
| REACH ID: | Reference | Site 104 | Site 104 | Extension | (Surficiai) | (Core) | Craighiil | Angie East | Angie West | Entrance | Range | Angie | Point | North | South | g |
| # ELUTRIATES TESTED: | (a) | 1 ^(a) | 3 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |

ANALYTE UNIT MIN DL

| | | | | | | | | | | | | | | - | | the second se | | |
|---------------|------|----|-------|--------|---------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|---|--------|--------|
| MONOBUTYLTIN | NG/L | 32 | 310 U | 33.0 U | 33.33 U | 32.0 U | 33.0 U | 33.5 U | 34.0 U | 33.0 U | 33.5 U | 81.5 | 33.5 U | 33.5 U | 34.0 U | 33.0 U | 34.0 U | 33.0 U |
| DIBUTYLTIN | NG/L | 40 | 280 U | 41.0 U | 41.33 U | 40.0 U | 41.0 U | 41.5 U | 43.0 U | 41.0 U | 41.5 U | 40.5 U | 41.5 U | 41.5 U | 42.0 U | 40.5 U | 42.0 U | 41.5 U |
| TRIBUTYLTIN | NG/L | 46 | 100 J | 47.0 U | 47.67 U | 46.0 U | 47.0 U | 48.5 U | 49.0 U | 48.0 U | 47.5 U | 47.0 U | 48.0 U | 48.0 U | 49.0 U | 47.0 U | 49.0 U | 47.5 U |
| TETRABUTLYTIN | NG/L | 52 | 220 U | 53.0 U | 53.67 U | 52.0 U | 53.0 U | 54.5 U | 56.0 U | 53.5 U | 53.5 U | 53.0 U | 54.0 U | 54.0 U | 55.0 U | 53.0 U | 54.5 U | 54.0 U |

U = not detected. J = estimated value below lowest calibrator.

(a) = actual concentrations for one elutriate sample reported for Outside Site 104 and Ocean Reference.

MIN DL = minimum detection limit.

TABLE 7-26CONCENTRATIONS OF DIOXIN AND FURAN CONGENERS (NG/L) IN
RECEIVING WATER INSIDE PLACEMENT SITE 104, OUTSIDE PLACEMENT
SITE 104, AND AT THE OCEAN REFERENCE SITE

| | | | Sample ID: | KI-03 | KI-07 | KI-14 | |
|------------------------|------|-------|------------|-----------------|-----------------|------------------|-----------------|
| | | | Reach ID: | Inside Site 104 | Inside Site 104 | Outside Site 104 | Ocean Reference |
| ANALYTE | UNIT | TEF | MIN DL | | | | |
| 2,3,7,8-TCDD | NG/L | 1 | 0.00154 | NT | 0.00161 U | 0.00154 U | 0.00257 U |
| 1,2,3,7,8-PECDD | NG/L | 0.5 | 0.00104 | NT | 0.00107 U | 0.00104 U | 0.00118 J |
| 1,2,3,4,7,8-HXCDD | NG/L | 0.1 | 0.00127 | NT | 0.00127 U | 0.00172 U | 0.00235 U |
| 1,2,3,6,7,8-HXCDD | NG/L | 0.1 | 0.00133 | NT | 0.00133 U | 0.0018 U | 0.00247 U |
| 1,2,3,7,8,9-HXCDD | NG/L | 0.1 | 0.0012 | NT | 0.0012 U | 0.00163 U | 0.00223 U |
| 1,2,3,4,6,7,8-HPCDD | NG/L | 0.01 | 0.00152 | NT | 0.0042 J | 0.00454 EMPC | 0.00287 U |
| OCDD | NG/L | 0.001 | 0.00387 | NT | 0.04 B | 0.03 B | 0.01 B |
| 2,3,7,8-TCDF | NG/L | 0.1 | 0.00109 | NT | 0.00137 U | 0.00109 U | 0.00167 U |
| 1,2,3,7,8-PECDF | NG/L | 0.05 | 0.00073 | NT | 0.00092 U | 0.00083 U | 0.00096 EMPC |
| 2,3,4,7,8-PECDF | NG/L | 0.5 | 0.00072 | NT | 0.0009 U | 0.00081 U | 0.00072 U |
| 1,2,3,4,7,8-HXCDF | NG/L | 0.1 | 0.00098 | NT | 0.00098 U | 0.00122 U | 0.00125 U |
| 1,2,3,6,7,8-HXCDF | NG/L | 0.1 | 0.00094 | NT | 0.00094 U | 0.00116 U | 0.00119 U |
| 2,3,4,6,7,8-HXCDF | NG/L | 0.1 | 0.00105 | NT | 0.00105 U | 0.00129 U | 0.00133 U |
| 1,2,3,7,8,9-HXCDF | NG/L | 0.1 | 0.00115 | NT | 0.00115 U | 0.00142 U | 0.00146 U |
| 1,2,3,4,6,7,8-HPCDF | NG/L | 0.01 | 0.00105 | NT | 0.00118 U | 0.00105 U | 0.0018 U |
| 1,2,3,4,7,8,9-HPCDF | NG/L | 0.01 | 0.00128 | NT | 0.00144 U | 0.00128 U | 0.0022 U |
| OCDF | NG/L | 0.001 | 0.00269 | NT | 0.00464 EMPC | 0.00364 EMPC | 0.00459 U |
| DIOXINS TEQ (ND=0) | NG/L | | | NT | 0.000042 | 0 | 0.00059 |
| DIOXINS TEQ (ND=1/2DL) | NG/L | | | NT | 0.00184 | 0.00183 | 0.00279 |

U = not detected. J = estimated value. B = detected in laboratory blank. EMPC = estimated maximum possible concentration.

MIN DL = minimum detection limit.

NT = not tested in K1-03 water.



TABLE 7-27CONCENTRATIONS OF DIOXIN AND FURAN CONGENERS (NG/L) IN 1999/2000 ELUTRIATES FROM BALTIMORE APPROACH CHANNELS, OCEAN REFERENCESITE, OUTSIDE PLACEMENT SITE 104, AND INSIDE PLACEMENT SITE 104

| | | 1 | REACH ID: | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Sorficial) | C&D Approach (Core) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Catoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|----------------------|-------|--------|-----------|-----------------|------------------|-----------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| | # ELU | FRIATE | S TESTED: | | | | | | | | | | | | | | | | · · · · · · |
| ANALYTE | UNIT | TEF | MIN DL | | | | | | | | | | | | | | | | |
| 2,3,7,8-TCDD | NG/L | 1 | 0.00160 | 0.0016 U | 0.00396 U | 0.00486 U | 0.00461 U | 0.00379 U | 0.0019 U | 0.00211 U | 0.00343 U | 0.00183 U | 0.00177 U | 0.00352 U | 0.00435 U | 0.00319 U | 0.00297 U | 0.00165 U | 0.00262 U |
| 1,2,3,7,8-PECDD | NG/L | 0.5 | 0.00072 | 0.0013 EMPC | 0.00155 U | 0.00169 U | 0.00125 U | 0.00147 U | 0.00192 U | 0.00168 U | 0.00113 U | 0.00103 U | 0.00136 U | 0.00151 U | 0.0013 U | 0.0016 U | 0.0014 U | 0.00128 U | 0.00109 U |
| 1,2,3,4,7,8-HXCDD | NG/L | 0.1 | 0.00110 | 0.0011 U | 0.00345 U | 0.00657 U | 0.00433 U | 0.00273 U | 0.00254 U | 0.00298 U | 0.00359 U | 0.00303 U | 0.00258 U | 0.00278 U | 0.00518 U | 0.00224 U | 0.00472 U | 0.00372 U | 0.00463 U |
| 1,2,3,6,7,8-HXCDD | NG/L | 0.1 | 0.00116 | 0.00116 U | 0.00363 U | 0.00691 U | 0.00455 U | 0.00287 U | 0.00267 U | 0.00313 U | 0.00378 U | 0.00319 U | 0.00271 U | 0.00292 U | 0.00545 U | 0.00236 U | 0.00496 U | 0.00391 U | 0.00487 U |
| 1,2,3,7,8,9-HXCDD | NG/L | 0.1 | 0.00104 | 0.00174 EMPC | 0.00327 U | 0.00623 U | 0.0041 U | 0.00258 U | 0.0024 U | 0.00282 U | 0.0034 U | 0.00287 U | 0.00244 U | 0.00263 U | 0.00491 U | 0.00212 U | 0.00447 U | 0.00352 U | 0.00439 U |
| 1,2,3,4,6,7,8-HPCDD | NG/L | 0.01 | 0.00215 | 0.00215 U | 0.00285 U | 0.00349 U | 0.0094 B | 0.00343 U | 0.00368 U | 0.00297 U | 0.00355 U | 0.00269 U | 0.0031 U | 0.00238 B | 0.00308B | 0.00238 U | 0.00426 B | 0.00269 U | 0.00368 B |
| OCDD | NG/L | 0.001 | 0.00432 | 0.00956 B | # 0.01 B | et 0.01 B | 0.09000 | 0.02 B | 0.04 B | 0.01 B | 0.01 B | 1 0.01.B | 9.02 B | 0.01 B | 0.01.B | 0.00894 B | 0.03 U | 8 10.0 A | · 4 (0.03 B |
| 2,3,7,8-TCDF | NG/L | 0.1 | 0.00101 | 0.00125 U | 0.00167 U | 0.00191 U | 0.00185 U | 0.00147 U | 0.00133 U | 0.00126 U | 0.00155 U | 0.00136 U | 0.0012 U | 0.00156 U | 0.00171 U | 0.00151 U | 0.00161 U | 0.00111 U | 0.00101 U |
| 1,2,3,7,8-PECDF | NG/L | 0.05 | 0.00072 | | 0.00123 U | 0.00135 U | 0.00096 U | 0.00105 U | 0.00139 U | 0.0011 U | 0.00083 U | 0.00092 U | 0.00097 U | 0.00142 U | 0.00103 U | 0.00107 U | 0.00111 U | 0.0011 U | 0.00085 U |
| 2,3,4,7,8-PECDF | NG/L | 0.5 | 0.00071 | 0.00071 U | 0.0012 U | 0.00131 U | 0.00094 U | 0.00102 U | 0.00136 U | 0.00107 U | 0.00081 U | 0.0009 U | 0.00094 U | 0.00139 U | 0.001 U | 0.00105 U | 0.00108 U | 0.00108 U | 0.00083 U |
| 1,2,3,4,7,8-HXCDF | NG/L | 0.1 | 0.00076 | | 0.00141 U | 0.00248 U | 0.0017 U | 0.00143 U | 0.00127 U | 0.00152 U | 0.00154 U | 0.00103 U | 0.00131 U | 0.00146 U | 0.00204 U | 0.00143 U | 0.00137 U | 0.0015 U | 0.00178 U |
| 1.2.3.6.7.8-HXCDF | NG/L | 0.1 | 0.00072 | 0.00118 B | 0.00134 U | 0.00237 U | 0.00162 U | 0.00136 U | 0.00122 U | 0.00144 U | 0.00147 U | 0.00098 U | 0.00124 U | 0.00139 U | 0.00195 U | 0.00136 U | 0.0013 U | 0.00143 U | 0.0017 U |
| 2,3,4,6,7,8-HXCDF | NG/L | 0.1 | 0.00080 | 0.0008 U | 0.0015 U | 0.00264 U | 0.0018 U | 0.00151 U | 0.00135 U | 0.00161 U | 0.00163 U | 0.00109 U | 0.00139 U | 0.00155 U | 0.00217 U | 0.00152 U | 0.00145 U | 0.00159 U | 0.00189 U |
| 1.2.3.7.8.9-HXCDF | NG/L | 0.1 | 0.00088 | 0.00088 U | 0.00164 U | 0.0029 U | 0.00198 U | 0.00166 U | 0.00149 U | 0.00177 U | 0.0018 U | 0.0012 U | 0.00152 U | 0.0017 U | 0.00238 U | 0.00167 U | 0.00159 U | 0.00175 U | 0.00208 U |
| 1,2,3,4,6,7,8-HPCDF | NG/L | 0.01 | 0.00131 | 0.002183 | 0.00168 U | 0.00253 U | TURINAY AN | 0.00236 U | 0.00185 U | 0.00215 U | 0.00183 U | 0.00157 U | 0.00239 U | 0.00201 U | 0.00166 3 | 0.0018 U | 0.00191 U | 0.00151 U | 0.00183 U |
| 1,2,3,4,7,8,9-HPCDF | NG/L | 0.01 | 0.00160 | 0.0016 U | 0.00205 U | 0.00309 U | 0.00175 U | 0.00287 U | 0.00226 U | 0.00262 U | 0.00224 U | 0.00192 U | 0.00291 U | 0.00245 U | 0.00227 U | 0.00219 U | 0.00233 U | 0.00184 U | 0.00223 U |
| OCDF | NG/L | 0.001 | 0.00308 | 0.00308 U | 0.00676 U | 0.00761 U | 0.01 J | 0.00693 U | 0.00793 U | 0.00616 U | 0.007 U | 0.0051 U | 0.00663 U | 0.0066 U | 0.00816 U | 0.00704 U | 0.00795 U | 0.00716 U | 0.0082 U |
| DIOXINS TEQ (ND=0) | NG/L | | | 0.00012 | 0 | 0 | 0.000137 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0000166 | 0 | 0 | 0 | 0 |
| DIOXINS TEQ (ND=1/2) | NG/L | | | 0.00138 | 0.00363 | 0.00486 | 0.00412 | 0.00337 | 0.00256 | 0.00264 | 0.0032 | 0.00219 | 0.00225 | 0.00335 | 0.0041 | 0.00303 | 0.00325 | 0.0024 | 0.00295 |

U = not detected. J = estimated value; value below lowest calibration limit B = detected in laboratory blank. EMPC = estimated maximum possible concentration.

MIN DL = minimum detection limit.

TABLE 7-28 NUMBER OF CONSTITUENTS DETECTED IN RECEIVINGWATER, CHANNEL ELUTRIATES, AND PLACEMENT SITE ELUTRIATES

| | RECEIVING W | ATER | CHANNEL ELU | FRIATES | PLACEMENT SITE/REFERENCE ELUTRIATES | | | |
|----------------------------|---|-------------|---|----------------|---|-------------|--|--|
| ANALYTICAL FRACTION | # constituents detected/ # constituents tested | Percent (%) | # constituents detected/ # constituents tested | Percent (%) | # constituents detected/ # constituents tested | Percent (%) | | |
| VOCs | 0/32 | 0 | 1/35 | 2.9 | 1/35 | 2.9 | | |
| SVOCs | 1/47 | 2.1 | 1/44 | 2.3 | 1/44 | 2.3 | | |
| Chlorinated Pesticides | 3/22 | 13.6 | 4/22 | 18.2 | 3/22 | 13.6 | | |
| Organophophorus Pesticides | 0/5 | 0 | 0/5 | 0 | 0/5 | 0 | | |
| PCB Aroclors | 0/7 | 0 | 0/7 | 0 | 0/7 | 0 | | |
| PCB Congeners | 1/26 | 3.8 | 5/26 | 19.2 | 1/26 | 3.8 | | |
| PAHs | 0/16 | 0 | 1/16 | 6.3 | 2/16 | 12.5 | | |
| Metal | 6/16 | 37.5 | 12/16 | 75 | 7/16 | 43.8 | | |
| Butyltins | 1/4 | 25 | 1/4 | 25 | 1/4 | 25 | | |
| Dioxin and Furan Congeners | 4/17 | 23.5 | 4/17 | 23.5 | 7/17 | 41.2 | | |



TABLE 7-29 CALCULATION OF ELUTRIATE FRACTION BURIED WITH SEDIMENT DURING FIRST 30MINUTES FOLLOWING DISPOSAL

| | Solids Material | | | | Elutriate | Pore Volun | ne Supplied | |
|-------|-------------------------------|-----------------|-----------------------|--------------|---------------------------|-------------|-------------|-----------|
| Time | Settled to | Pore Volu | me (ft [°]) | Bulk | Cloud | by Elut | riate (d) | Percent |
| (min) | Bottom (ft ³) (a) | Cummulative (b) | Incremental | Dilution (c) | Volume (ft ³) | Incremental | Cummulative | Total (e) |
| 10 | 8230 (56.4%) | 37490 | 37490 | 3.3 | 220,100 | 0 | 0 | 0.0 |
| 20 | 11349 (77.8%) | 51702 | 14212 | 5.6 | 372,000 | 0 | 0 | 0.0 |
| 30 | 12764 (87.5%) | 58145 | 6443 | 9.0 | 598,000 | 0 | 0 | 0.0 |

Note: Based on STFATE model run for average ebb tide velocity (0.56 ft/sec). Barge (3000 yd^3) contained 14,580 ft3 solids and 66,420 ft3 water.

- (a) Settled material volume provided by STFATE model output (% of total solids in barge).
- (b) Pore volume in settled material assuming same ratio as in original dredged material (66,420 ft³ water to 14,580 ft³ solids, 4.55 to 1).
- (c) Bulk dilution calculated from volume of elutriate cloud divided by the initial 66,420 ft³ elutriate volume in barge.
- (d) Pore volume divided by bulk dilution
- (e) Pore volume supplied by elutriate divided by the initial 66,420 ft^3 elutriate volume.

TABLE 7-30STFATE MODELING RESULTS FORPLACEMENT AT SITE 104 FROM A SPLIT HULL BARGE,
AVERAGE EBB AND FLOOD TIDE SCENARIOS

| | Elutriate Availability to Water Column 50% 75% 100% Dist (ft) Hour Dist (ft) Hour Dist (ft) Hour 660 0.33 774 0.38 958 0.48 1015 0.50 1376 0.68 1604 0.80 1423 0.71 1761 0.87 2004 0.99 16(2) 0.92 2040 1.92 2507 1.24 | | | | | | | | | | | | | |
|----------|--|------|---------------|-----------|-----------|-----------|--|--|--|--|--|--|--|--|
| | 50 | % | 75 | % | 10 | 0% | | | | | | | | |
| Dilution | Dist (ft) | Hour | Dist (ft) | Hour | Dist (ft) | Hour | | | | | | | | |
| 10 | 660 | 0.33 | 774 | 0.38 | 958 | 0.48 | | | | | | | | |
| 20 | 1015 | 0.50 | 1376 | 0.68 | 1604 | 0.80 | | | | | | | | |
| 30 | 1423 | 0.71 | 1761 | 0.87 | 2004 | 0.99 | | | | | | | | |
| 40 | 1662 | 0.82 | 2049 | 1.02 | 2507 | 1.24 | | | | | | | | |
| 50 | 1875 | 0.93 | 2443 | 1.21 | 2787 | 1.38 | | | | | | | | |
| 75 | 2481 | 1.23 | 296 1 | 1.47 | 3294 | 1.63 | | | | | | | | |
| 100 | 2863 | 1.42 | 3312 | 1.64 | 3712 | 1.84 | | | | | | | | |
| 150 | 3378 | 1.68 | 3940 | 1.95 | 4564 | 2.26 | | | | | | | | |
| 200 | 3796 | 1.88 | 4601 | 2.28 | 5180 | 2.57 | | | | | | | | |
| 250 | 4218 | 2.09 | 5081 | 2.52 | 5660 | 2.81 | | | | | | | | |
| 300 | 4631 | 2.30 | 5467 | 2.71 | 6056 | 3.00 | | | | | | | | |
| 350 | 4973 | 2.47 | 57 9 7 | 2.88 | 6397 | 3.17 | | | | | | | | |
| 400 | 5257 | 2.61 | 6084 | 3.02 | 6714 | 3.33 | | | | | | | | |
| 450 | 5512 | 2.73 | 6343 | 3.15 | 6999 | 3.47 | | | | | | | | |
| 500 | 5741 | 2.85 | 6588 | 3.27 | 7255 | 3.60 | | | | | | | | |
| | | | | | | | | | | | | | | |
| Hour | Dilution Dist (ft) | | Dilution | Dist (ft) | Dilution | Dist (ft) | | | | | | | | |
| 1 | 59 | 2016 | 39 | 2016 | 30 | 2016 | | | | | | | | |
| 4 | 1303 | 8064 | 880 | 8064 | 668 | 8064 | | | | | | | | |

Average Ebb Tide Scenario

Average Flood Tide Scenario

| | | Elutriate Availability to Water Column 50% 75% 100% Dist (ft) Hour Dist (ft) Hour Dist (ft) Hour 1272 0.41 3477 1.12 3816 1.22 3880 1.24 5083 1.63 5566 1.79 | | | | | | | | | | | | |
|----------|-----------|--|--------------|-----------|-----------|-----------|--|--|--|--|--|--|--|--|
| | 50 | % | 75 | % | 10(|)% | | | | | | | | |
| Dilution | Dist (ft) | Hour | Dist (ft) | Hour | Dist (ft) | Hour | | | | | | | | |
| 50 | 1272 | 0.41 | 3477 | 1.12 | 3816 | 1.22 | | | | | | | | |
| 100 | 3880 | 1.24 | 5083 | 1.63 | 5566 | 1.79 | | | | | | | | |
| 150 | 5135 | 1.65 | 5906 | 1.89 | 6863 | 2.20 | | | | | | | | |
| 200 | 5654 | 1.81 | 6909 | 2.22 | 7593 | 2.44 | | | | | | | | |
| 250 | 6487 | 2.08 | 7465 | 2.39 | 8456 | 2.71 | | | | | | | | |
| 300 | 6961 | 2.23 | 8188 | 2.63 | 8986 | 2.88 | | | | | | | | |
| 350 | 7323 | 2.35 | 8637 | 2.77 | 9625 | 3.09 | | | | | | | | |
| 400 | 7782 | 2.50 | 904 1 | 2.90 | 10095 | 3.24 | | | | | | | | |
| 450 | 8249 | 2.65 | 9543 | 3.06 | 10481 | 3.36 | | | | | | | | |
| 500 | 8561 | 2.75 | 9932 | 3.19 | 10902 | 3.50 | | | | | | | | |
| | | | | | | | | | | | | | | |
| Hour | Dilution | Dist (ft) | Dilution | Dist (ft) | Dilution | Dist (ft) | | | | | | | | |
| 1 | 59 | 3118 | 40 | 3118 | 31 | 3118 | | | | | | | | |
| 4 | 1420 | 12470 | 959 | 12470 | 731 | 12470 | | | | | | | | |

TABLE 7-31 STFATE MODELING RESULTS FORPLACEMENT AT SITE 104 BY HYDRAULIC PUMPING FROMA BARGE, AVERAGE EBB AND FLOOD TIDE SCENARIOS

| | | Elutriat | te Availabili | ty to Water | Column | |
|----------|-----------|-----------|---------------|-------------|-----------|-----------|
| 4 | 50 | % | 75 | % | 10 | 0% |
| Dilution | Dist (ft) | Hour | Dist (ft) | Hour | Dist (ft) | Hour |
| 50 | 531 | 0.26 | 1086 | 0.54 | 1468 | 0.73 |
| 100 | 1518 | 0.75 | 2140 | 1.06 | 2683 | 1.33 |
| 150 | 2169 | 1.08 | 2960 | 1.47 | 3583 | 1.78 |
| 200 | 2731 | 1.35 | 3608 | 1.79 | 4267 | 2.12 |
| 250 | 3213 | 1.59 | 4139 | 2.05 | 4787 | 2.37 |
| 300 | 3638 | 1.80 | 4568 | 2.27 | 5220 | 2.59 |
| 350 | 3997 | 1.98 | 4931 | 2.45 | 5597 | 2.78 |
| 400 | 4316 | 2.14 | 5244 | 2.60 | 5955 | 2.95 |
| 450 | 4595 | 2.28 | 5536 | 2.75 | 6297 | 3.12 |
| 500 | 4837 | 2.40 | 5814 | 2.88 | 6621 | 3.28 |
| | | | | | | |
| Hour | Dilution | Dist (ft) | Dilution | Dist (ft) | Dilution | Dist (ft) |
| 1 | 139 | 2016 | 94 | 2016 | 72 | 2016 |
| 4 | 1545 | 8064 | 1041 | 8064 | 787 | 8064 |

Average Ebb Tide Scenario

Average Flood Tide Scenario

| | | Elutriate Availability to Water Column50%75%100%Dist (ft)HourDist (ft)Hour01510100100100 | | | | | | | | | | | | | |
|----------|-----------|--|-----------|-----------|-----------|-----------|--|--|--|--|--|--|--|--|--|
| | 50 | % | 75 | % | 10 |)% | | | | | | | | | |
| Dilution | Dist (ft) | Hour | Dist (ft) | Hour | Dist (ft) | Hour | | | | | | | | | |
| 50 | 945 | 0.30 | 1306 | 0.42 | 1637 | 0.53 | | | | | | | | | |
| 100 | 1721 | 0.55 | 2374 | 0.76 | 2884 | 0.92 | | | | | | | | | |
| 150 | 2412 | 0.77 | 3208 | 1.03 | 4033 | 1.29 | | | | | | | | | |
| 200 | 2971 | 0.95 | 4078 | 1.31 | 5024 | 1.61 | | | | | | | | | |
| 250 | 3552 | 1.14 | 4828 | 1.55 | 5861 | 1.88 | | | | | | | | | |
| 300 | 4120 | 1.32 | 5501 | 1.76 | 6604 | 2.12 | | | | | | | | | |
| 350 | 4630 | 1.49 | 6089 | 1.95 | 7284 | 2.34 | | | | | | | | | |
| 400 | 5105 | 1.64 | 6647 | 2.13 | 7847 | 2.52 | | | | | | | | | |
| 450 | 5538 | 1.78 | 7168 | 2.30 | 8378 | 2.69 | | | | | | | | | |
| 500 | 5935 | 1.90 | 7604 | 2.44 | 8902 | 2.86 | | | | | | | | | |
| | | | | | | | | | | | | | | | |
| Hour | Dilution | Dist (ft) | Dilution | Dist (ft) | Dilution | Dist (ft) | | | | | | | | | |
| 1 | 213 | 3118 | 144 | 3118 | 110 | 3118 | | | | | | | | | |
| 4 | 2070 | 12470 | 1392 | 12470 | 1051 | 12470 | | | | | | | | | |

TABLE 7-32 STFATE MODELING RESULTS FOR PLACEMENT FROM A SPLIT HULL BARGE, AVERAGE TIDE SCENARIO AT THE NORFOLK OCEAN DISPOSAL SITE

| | | Aver | age Tide Sce | enario | | |
|----------|-----------|-----------|---------------|-------------|-----------|-----------|
| | | Elutriat | te Availabili | ty to Water | Column | |
| | 50 | % | 75 | % | 10 | 0% |
| Dilution | Dist (ft) | Hour | Dist (ft) | Hour | Dist (ft) | Hour |
| 10 | 308 | 0.26 | 436 | 0.37 | 501 | 0.42 |
| 20 | 532 | 0.45 | 690 | 0.58 | 813 | 0.68 |
| 30 | 699 | 0.59 | 929 | 0.78 | 1127 | 0.95 |
| 40 | 857 | 0.72 | 1143 | 0.96 | 1330 | 1.12 |
| 50 | 1026 | 0.86 | 1296 | 1.09 | 1487 | 1.25 |
| 75 | 1314 | 1.11 | 1605 | 1.35 | 1838 | 1.55 |
| 100 | 1519 | 1.28 | 1857 | 1.56 | 2129 | 1.79 |
| 150 | 1875 | 1.58 | 2265 | 1.91 | 2551 | 2.15 |
| 200 | 2166 | 1.82 | 2569 | 2.16 | 2872 | 2.42 |
| 250 | 2399 | 2.02 | 2811 | 2.37 | 3144 | 2.65 |
| 300 | 2589 | 2.18 | 3021 | 2.54 | 3385 | 2.85 |
| 350 | 2746 | 2.31 | 3219 | 2.71 | 3597 | 3.03 |
| 400 | 2906 | 2.45 | 3402 | 2.86 | 3778 | 3.18 |
| 450 | 3042 | 2.56 | 3563 | 3.00 | 3945 | 3.32 |
| 500 | 3177 | 2.67 | 3707 | 3.12 | 4098 | 3.45 |
| | | | | | | |
| Hour | Dilution | Dist (ft) | Dilution | Dist (ft) | Dilution | Dist (ft) |
| 1 | 63 | 1188 | 43 | 1188 | 33 | 1188 |
| 4 | 1483 | 4752 | 1000 | 4752 | 759 | 4752 |



| | | EPA | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|-----------------------------|------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|--------|------------|------------|---|
| | | SALTWATER | Eastern | Approach | Approach | | Cralghill | Cralghill | Craighiil | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | ACUTE | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angie West | Entrance | Range | Angle | Point | North | South | Straightening |
| NON-METALS | | | | | | | | | | | | | | | |
| CYANIDE | MG/L | 0.001 | | | | | | | | | | | 0.005 | | |
| NITROGEN, AMMONIA | MG/L | 43 | | | | | | | | | | | | | |
| SULFIDE, TOTAL | MG/L | a | | | | | | | | | | | | 0.44 | |
| VOCs | | | | | | | | | | | | | | | · |
| DICHLOROMETHANE | UG/L | а | 160 | 22 | 120 | 110 | 110 | 110 | 170 | 160 | 34 | 680 | 16 | 30 | 140 |
| SVOCs | | | | | | | | | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 8 | | | 11 | | 3.2 | | | | 2 | | 15 | 28 | 2.2 |
| PESTICIDES | | | _ | | | | | | | | | | | | • |
| BETA-BHC | UG/L | а | 0.03 | | | 0.01 | 0.01 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 | 0.02 | 0.02 | 0.02 |
| GAMMA-BHC | UG/L | 0.16 | | | | | | | | | | | | | |
| HEPTACHLOR | UG/L | 0.053 | | 0.07 | 0.07 | | 0.07 | 0.19 | 0.08 | | 0.06 | 0.06 | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | | | | | _ | | | | | | | | |
| PCB CONGENERS | | | | | | | | | | | | | | | • |
| TOTAL PCB (ND=0) | UG/L | а | 0.0238 | 0.0136 | 0.04 | | 0.0332 | 0.02 | 0.0334 | 0.0032 | 0.02 | 0.0156 | 0.04 | | 0.0328 |
| TOTAL PCB (ND=1/2) | UG/L | a | 0.0715 | 0.0599 | 0.0831 | | 0.0763 | 0.0695 | 0.0765 | 0.0544 | 0.0661 | 0.0617 | 0.0831 | | 0.0745 |
| TOTAL PCB (ND=DL) | UG/L | a | 0.119 | 0.109 | 0.126 | | 0.119 | 0.119 | 0.12 | 0.106 | 0.112 | 0.108 | 0.126 | | 0.116 |
| PAHs | | | | | | | | | | | | | | | |
| FLUORENE | UG/L | a | | | | | 0.08 | | | | | | | | |
| METALS | | | | | | | | | | | | | | | • |
| ANTIMONY | UG/L | a | 2.9 | 4.1 | 3.3 | 1.8 | 2.3 | 2.4 | 2.5 | | 3.6 | 3.2 | 2.5 | 1.2 | 6.1 |
| ARSENIC | UG/L | 69 | | | | | | | | | | | | | |
| BERYLLIUM | UG/L | a | | | | | | | | | | | | | 0.53 |
| CHROMIUM | UG/L | 1100 | | | | | | | | | | | | | |
| COPPER | UG/L | 4.8 | | 5.8 | | | | | | | | | | | |
| LEAD | UG/L | 210 | | | | | | | | | | | | | |
| MANGANESE | UG/L | а | 7830 | 4690 | 3310 | 892 | 4250 | 3360 | 1610 | 2210 | 7020 | 3950 | 4980 | 3540 | 2470 |
| NICKEL | UG/L | 74 | | | | | | | | | | | | | |
| SELENIUM | UG/L | 290 | | | | | | | | | | | | | |
| SILVER | UG/L | 1.9 | | | | | | | | | | | | | 2.3 |
| ZINC | UG/L | 90 | | | | | | | | | | | | | |

a = no USEPA acute saltwater criterion.

TABLE 7-33B MIXING FACTORS REQUIRED TO COMPLY WITH USEPA ACUTE SALTWATER CRITERIA

| ſ | | USEPA | Brewerton | C&D | C&D | | | | | Cralghill | | | | | |
|-----------------------------|------|------------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|---------|-------|------------|------------|---------------|
| | | SALTWATER | Eastern | Approach | Approach | | Cralghill | Cralghlll | Cralghlll | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | ACUTE | Extension | (Surficial) | (Cores) | Cralghlll | Angle East | Angle West | Entrance | Range | Angle | Polnt | North | South | Stralghtening |
| NON-METALS | | | <u> </u> | | | | | | | | | | | | |
| CYANIDE | MG/L | 0.001 | | | | | | | | | | | 5 | · · · · · | |
| NITROGEN, AMMONIA | MG/L | 43 | | | | | | | | | | | | | |
| SULFIDE, TOTAL | MG/L | a | | | | | | | | | | | | | |
| VOCs | | | | | | | | | | | | | | | |
| DICHLOROMETHANE | UG/L | a | | | | | | | | | | | | | |
| SVOCs | | | | | | | | | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | a | | | | | | | | | | | | | |
| PESTICIDES | | | | | | | | | | | | | | | |
| BETA-BHC | UG/L | a | | | | | | | | | | | | | |
| GAMMA-BHC | UG/L | 0.16 | | | | | | | | | | _ | | | |
| HEPTACHLOR | UG/L | 0.053 | | 1 | I | | I | 4 | 2 | | I | I | | | |
| HEPTACHLOR EPOXIDE | UG/L | 0.053 | | | | ł | | | | | | | 1 | | |
| PCB CONGENERS | | | l | | | | | | | | | | | | |
| TOTAL PCB (ND=0) | UG/L | 8 | | | | | | | | | | | | | |
| TOTAL PCB (ND=1/2) | UG/L | a | | | | | | | | | | | | | |
| TOTAL PCB (ND=DL) | UG/L | a a | | | | | | l | | | | | | | |
| PAHs | | | | | | . | | | | | | | | <u>.</u> | |
| FLUORENE | UG/L | . <u>a</u> | | | | | | 1 | | | | | 1 | | |
| METALS | | | | | | | | | | = | | | | | |
| ANTIMONY | UG/L | a | | | | | | | | | | | | | |
| ARSENIC | UG/L | 69 | | | | | | | | | | | | | |
| BERYLLIUM | UG/L | a | | | | L | | | | | | | | | |
| CHROMIUM | UG/L | . 1100 | | | | | <u> </u> | | | | | | | | |
| COPPER | UG/L | 4.8 | | 1 | | | | | L | l | | | | | 1 |
| LEAD | UG/L | 210 | | | | | | | | | | 1 | | | |
| MANGANESE | UG/L | , a | | | | | | | | | | | | | |
| NICKEL | UG/L | . 74 | | | | I | | | ļ | L | L | | | | ļ |
| SELENIUM | UG/L | . 290 | | ļ | | | | l | L | L | | | | | |
| SILVER | UG/L | . 1.9 | | ļ | | | | | | 1 | ļ | | L | | 1 |
| ZINC | UG/L | . 90 | | | | | | | | | | | | | |

a = no USEPA acute saltwater criterion.

Acute criteria based on 1-hr. average exposure concentrations.

Maximum detected concentrations compared to acute criteria.



TABLE 7-34A MEAN CONCENTRATIONS OF ANALYTES DETECTED IN 1999 CHANNEL ELUTRIATES THAT EXCEED USEPA CHRONIC SALTWATER CRITERIA

| | | EPA | Brewerton | C&D | C&D | | | | | Craighili | | | | | |
|-----------------------------|------|-----------|-----------|-------------|----------|-----------|------------|---------------------------------------|---|-----------|---------|---------|------------|------------|---------------|
| | | SALTWATER | Eastern | Approach | Approach | | Cralghill | Cralghill | Craighiil | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | CHRONIC | Extension | (Surficial) | (Cores) | Cralghill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| NON-METALS | | | | | | | | | | | | | | | |
| CYANIDE | MG/L | 0.001 | | | | | | | | | | | 0.005 | | |
| NITROGEN, AMMONIA | MG/L | 6.4 | | | | | 10.2 | | 6.45 | | | | | | |
| SULFIDE, TOTAL | MG/L | 0.002 | | | | | | | | | | | | 0.395 | |
| VOCs | | | | | | | • | | | | | | | | |
| DICHLOROMETHANE | UG/L | a | 160 | 22 | 120 | 110 | 110 | 110 | 170 | 160 | 34 | 680 | 16 | 30 | 140 |
| SVOCs | | • • • • • | | | | | • | · · · · · · · · · · · · · · · · · · · | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | a | | | 6.5 | | 2.2 | | | | 1.75 | | 8.5 | 15 | 2 |
| PESTICIDES | | | | | | | | | | | | | | | |
| BETA-BHC | UG/L | а | 0.02 | | | 0.01 | 0.01 | 0.01 | 0.01 | 0.015 | 0.02 | 0.02 | 0.015 | 0.015 | 0.015 |
| GAMMA-BHC | UG/L | а | | 0.00905 | | | 0.00905 | 0.00905 | 0.00905 | 0.00905 | 0.01405 | 0.00905 | 0.01405 | 0.00865 | 0.01405 |
| HEPTACHLOR | UG/L | 0.0036 | 0.03 | 0.045 | 0.045 | 0.03 | 0.045 | 0.105 | 0.05 | | 0.04 | 0.04 | 0.035 | 0.03 | 0.025 |
| HEPTACHLOR EPOXIDE | UG/L | 0.0036 | 0.015 | 0.015 | | | 0.015 | 0.02 | 0.015 | 0.02 | 0.02 | 0.025 | 0.02 | 0.02 | 0.02 |
| PCB CONGENERS | | | | | | | | | | | | | | | |
| TOTAL PCB (ND=0) | UG/L | а | 0.0119 | 0.012 | 0.02 | | 0.0166 | 0.01 | 0.0167 | 0.0016 | 0.01 | 0.0078 | 0.0253 | | 0.0181 |
| TOTAL PCB (ND=1/2) | UG/L | а | 0.0617 | 0.0598 | 0.0678 | | 0.0644 | 0.061 | 0.0645 | 0.0535 | 0.0593 | 0.0571 | 0.0716 | | 0.0645 |
| TOTAL PCB (ND=DL) | UG/L | a | 0.112 | 0.108 | 0.116 | | 0.112 | 0.112 | 0.112 | 0.105 | 0.109 | 0.106 | 0.118 | | 0.111 |
| PAHs | | | | _ | | | | | | | | | | | |
| FLUORENE | UG/L | 8 | | | | | 0.07 | | | | | | | | |
| METALS | | | | | | | | | | | | | | | |
| ANTIMONY | UG/L | а | 1.95 | 2.55 | 2.2 | 1.4 | 1.65 | 1.7 | 1.75 | | 2.3 | 2.1 | 1.75 | 1.1 | 3.6 |
| ARSENIC | UG/L | 36 | | | | | | | | | | | | | |
| BERYLLIUM | UG/L | a | | | | | | | | | | | | | 0.315 |
| CHROMIUM | UG/L | 50 | | | | | | | | | | | | | |
| COPPER | UG/L | 3.1 | | 3.25 | | | | | | | | | | | |
| LEAD | UG/L | 8.1 | | | | | | | | | | | | | 2.1 |
| MANGANESE | UG/L | 8 | 6625 | 4465 | 2245 | 847 | 3280 | 2750 | 1565 | 1960 | 5520 | 2620 | 3830 | 2210.5 | 1850 |
| NICKEL | UG/L | 8.2 | | | | | | | - 10 - 10 - 10 - 10 - 10 - 10 - 10 - 10 | | | | | | |
| SELENIUM | UG/L | 71 | | | | | | | | | | | | | - |
| SILVER | UG/L | а | | | | | | | | | | | | | 2.25 |
| ZINC | UG/L | 81 | | | | | | | | | | | | | |

a = no USEPA chronic saltwater criterion.

TABLE 7-34B MIXING FACTORS REQUIRED TO COMPLY WITH USEPA CHRONIC SALTWATER CRITERIA

| | | EPA | Brewerton | C&D | C&D | | | | | Craighlll | | | | | |
|-----------------------------|------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|--|------------|---------------|
| | | SALTWATER | Eastern | Approach | Approach | | Cralghiil | Craighiil | Cralghlii | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | CHRONIC | Extension | (Surficial) | (Cores) | Craighlll | Angle East | Angie West | Entrance | Range | Angle | Point | North | South | Stralghtening |
| NON-METALS | | | | | | | | | | | | | | | |
| CYANIDE | MG/L | 0.001 | | | | | | | - | | | | 5 | | |
| NITROGEN, AMMONIA | MG/L | 6.4 | | | | | | | | | | | | | |
| SULFIDE, TOTAL | MG/L | 0.002 | | | | | | | | | | | | 198 | |
| VOCs | | | | | | | | | | | | | | | |
| DICHLOROMETHANE | UG/L | а | | | | | | | | | | | | | |
| SVOCs | | | | | | | | | | | | | | | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | а | | | | | | | | | | | | | |
| PESTICIDES | | | | | | | | | | | | | •••••••••••••••••••••••••••••••••••••• | | |
| ВЕТА-ВНС | UG/L | a | | | | | | | | | | | | | |
| GAMMA-BHC | UG/L | а | | | | | | | | | | | | | |
| HEPTACHLOR | UG/L | 0.0036 | 8 | 13 | 13 | 8 | 13 | 29 | 14 | | 11 | 11 | 10 | 8 | 7 |
| HEPTACHLOR EPOXIDE | UG/L | 0.0036 | 4 | 4 | | | 4 | 6 | 4 | 6 | 6 | 7 | 6 | 6 | 6 |
| PCB CONGENERS | | | | | | | | | | | | | | | |
| TOTAL PCB (ND=0) | UG/L | а | | | | | | | | | | | | | |
| TOTAL PCB (ND=1/2) | UG/L | а | | | | | | | | | | | | | |
| TOTAL PCB (ND=DL) | UG/L | а | | | | | | | | | | | | | |
| PAHs | | | | | | | | | | | | | | | |
| FLUORENE | UG/L | а | | | | | | | _ | | | | | | |
| METALS | | - | | | | | | | | | | | | | |
| ANTIMONY | UG/L | a | | | | | | | | | | | 1 | | |
| ARSENIC | UG/L | 36 | | | | | | | | | | | | | |
| BERYLLIUM | UG/L | a | | | | | | | | | | | | | |
| CHROMIUM | UG/L | 50 | | | | | | | | | | | | | |
| COPPER | UG/L | 3.i | | i | | | | | | | | | | | |
| LEAD | UG/L | 8.1 | | | | | | | | | | | | | |
| MANGANESE | UG/L | а | - | | | | | | | | | | | | |
| NICKEL | UG/L | 8.2 | | | | | | | | | | | | | |
| SELENIUM | UG/L | 71 | | | | | | | | | | | | | |
| SILVER | UG/L | a | | | | | | | | | | | | | |
| ZINC | UG/L | 8i | | | | | | | | | | | | | |

a = no USEPA chronic saltwater criterion.

Chronic criteria based on 4-day average exposure concentrations.

Mean detected concentrations compared to chronic criteria.



TABLE 7-35AMEAN CONCENTRATIONS OF ANALYTES DETECTED IN 1999 CHANNEL ELUTRIATES THAT EXCEED USEPA HUMAN
HEALTH CRITERIA FOR CONSUMPTION OF AQUATIC ORGANISMS

| | | EPA | · · · · · · · · · · · · · · · · · · · | | | | | | | | | | | | · |
|-----------------------------|-------------|-----------|---------------------------------------|-------------|----------|---------------------------------------|---------------------------------------|------------|-----------|-----------|---------|-----------------|---------------------------------------|-----------------|---------------|
| | l ! | SALTWATER | Brewerton | C&D | C&D | | | | | Craighill | i | | | ' | 1 |
| | | HUMAN | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | HEALTH | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Streightening |
| NON-METALS | | | | | | | | | | 1 | l | | | | oungineering |
| CYANIDE | MG/L | а | · · · · · · · · · · · · · · · · · · · | | | r | Г т | | | ľ | | | 0.005 | r | |
| NITROGEN, AMMONIA | MG/L | a | 2.2 | 1.15 | 2.1 | 0.705 | 10.2 | 5 56 | 6.45 | 2 25 | 37 | | 5.055 | | |
| SULFIDE, TOTAL | MG/L | 8 | | | ······ | | | | 0.10 | | | | 5.755 | 0 305 | 2.34 |
| VOCs | | | L | | | | h | I | | L | l | k | L | 0.393 | · |
| DICHLOROMETHANE | UG/L | 16000 | ر <u>ا</u> | | | | | | | | | | | | |
| SVOCs | | ······ | | | <u></u> | · | | | | | B | · | | | <u> </u> |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 59 | | | | | · · · · · · · · · · · · · · · · · · · | I | | | T | (******* | | | |
| PESTICIDES | 1 -1 | h | | ` | | | A | <u>-</u> A | | l | | <u>ل</u> ے | L | | |
| BETA-BHC | UG/L | 0.46 | | 1 | | · · · · · · · · · · · · · · · · · · · | | I | | | T | | · · · · · · · · · · · · · · · · · · · | · | |
| GAMMA-BHC | UG/L | 0.63 | | | | | | | | | | | | | |
| HEPTACHLOR | UG/L | 0.0021 | 0.03 | 0.045 | 0.045 | 0.03 | 0.045 | 0.105 | 0.05 | | 0.04 | 0.04 | 0.035 | 0.03 | 0.025 |
| HEPTACHLOR EPOXIDE | UG/L | 0.0011 | 0.015 | 0.015 | | | 0.015 | 0.02 | 0.015 | 0.02 | 0.02 | 0.025 | 0.033 | 0.05 | 0.023 |
| PCB CONGENERS | | | | | | | • | | | | | | 0.02 | 0.02 | |
| TOTAL PCB (ND=0) | UG/L | 0.0017 | 0.0119 | 0.012 | 0.02 | | 0.0166 | 0.01 | 0.0167 | | 0.01 | 0.0078 | 0.0253 | ······ | 0.0181 |
| TOTAL PCB (ND=1/2) | UG/L | 0.0017 | 0.0617 | 0.0598 | 0.0678 | | 0.0644 | 0.061 | 0.0645 | 0.0535 | 0.0593 | 0.0571 | 0.0716 | | 0.0645 |
| TOTAL PCB (ND=DL) | UG/L | 0.0017 | 0.112 | 0.108 | 0.116 | | 0.112 | 0.112 | 0.112 | 0.105 | 0.109 | 0.106 | 0.118 | | 0.0043 |
| PAHs | | | | | | | | | | | | | | | |
| FLUORENE | UG/L | 14000 | | 1 | | | | | | | T | · · · · · | [] | [] | |
| METALS | | | | | | f | | L | | | | | <u> </u> | | |
| ANTIMONY | UG/L | 4300 | | | | | | T | | _ | T | Г Т | · · · · · · · · · · · · · · · · · · · | [] | |
| ARSENIC | UG/L | 1.4 | 4.1 | 3.2 | 2.45 | 2 | 8.8 | 3.95 | 5.05 | 3.55 | 6.05 | 4.6 | 5.2 | 3.05 | 18.2 |
| BERYLLIUM | UG/L | 0.117 | | | | | | | | | | | | 2.02 | 0.315 |
| CHROMIUM | UG/L | a | 0.9 | ī | 0.85 | | 1.3 | 1.25 | 1.25 | 0.85 | | 0.95 | | 0.9 | 2.55 |
| COPPER | UG/L | a | 1 | 3.25 | 1.35 | | 1.95 | 0.75 | | | 0.9 | | 1 | | 2.05 |
| LEAD | UG/L | a | | | | | | | | -, | | | | | 2.1 |
| MANGANESE | UG/L | 100 | 6625 | 4465 | 2245 | 847 | 3280 | 2750 | 1565 | 1960 | 5520 | 2620 | 3830 | 2210.5 | 1850 |
| NICKEL | UG/L | 4600 | | | | | | | · | | | | | | |
| SELENIUM | UG/L | a | 1.85 | 1.85 | | | | | | | 2 | | | | 15.5 |
| SILVER | UG/L | a | | | | | | | | | | | | | 2.25 |
| ZINC | UG/L | a | | 2.9 | | | | | | | | 1.75 | | | |

a = no USEPA human health saltwater criterion.

TABLE 7-35BMIXING FACTORS REQUIRED TO COMPLY WITH USEPA HUMAN HEALTH SALTWATER CRITERIA FOR CONSUMPTION OF
AQUATIC ORGANISMS

| | | EPA | | | | | | | | | | | | | |
|-----------------------------|------|------------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|----------|------------|---------------------------------------|---|
| | | SALTWATER | Brewerton | C&D | C&D | | | | | Cralghill | | - | | | |
| | | HUMAN | Eastern | Approach | Approach | | Craighill | Craighili | Craighili | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | UNIT | HEALTH | Extension | (Surficial) | (Cores) | Cralghill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Stralghtening |
| NON-METALS | | | | | | | | | | | | | | | |
| CYANIDE | MG/L | a | | | | | | | | | | | | | |
| NITROGEN, AMMONIA | MG/L | a | | | | | | | | | | | | | |
| SULFIDE, TOTAL | MG/L | a | | | | | | l | | | | | | | [|
| VOCs | | | - | | | | | | | | | | - | | |
| DICHLOROMETHANE | UG/L | 16000 | | | | | | | | | | | | <1 | < 1 |
| SVOCs | | | | | | | | | | | | | | - | |
| BIS(2-ETHYLHEXYL) PHTHALATE | UG/L | 59 | | | | | | | | | | | | < 1 | <1 |
| PESTICIDES | | | | | | | | | | | | | | - | |
| ВЕТА-ВНС | UG/L | 0.46 | | | | | | | | | | | | <1 | < 1 |
| GAMMA-BHC | UG/L | 0.63 | | | | | | | | | | | | 1 > | < 1 |
| HEPTACHLOR | UG/L | 0.0021 | 14 | 21 | 21 | 14 | 21 | 50 | 24 | | 19 | 19 | 17 | 14 | 12 |
| HEPTACHLOR EPOXIDE | UG/L | 0.0011 | 14 | 14 | | | 14 | 18 | 14 | 18 | 18 | 23 | 18 | 18 | 18 |
| PCB CONGENERS | | | | | | | | | | | | | | |] |
| TOTAL PCB (ND=0) | UG/L | 0.0017 | 7 | 7 | 12 | | 10 | 6 | 10 | < 1 | 6 | 5 | 15 | | 11 |
| TOTAL PCB (ND=1/2) | UG/L | 0.0017 | 36 | 35 | 40 | | 38 | 36 | 38 | 31 | 35 | 34 | 42 | | 38 |
| TOTAL PCB (ND=DL) | UG/L | 0.0017 | 66 | 63 | 68 | | 66 | 66 | 66 | 62 | 64 | 63 | 69 | | 65 |
| PAHs | | | - | | | | | | - | | | - | | · · · · · · · · · · · · · · · · · · · | |
| FLUORENE | UG/L | . 14000 | | | | | | 1 | | | | | I | | |
| METALS | | | | | | | | | | | - | | | - | • · · · · · · · · · · · · · · · · · · · |
| ANTIMONY | UG/L | 4300 | | | | | | | | | | | | 1 > | < 1 |
| ARSENIC | UG/L | . 1.4 | 3 | 2 | 2 | 1 | 6 | 3 | 4 | 3 | 4 | 3 | 4 | 2 | 13 |
| BERYLLIUM | UG/L | . 0.117 | ļ | | | | | | | ļ | | | | | 3 |
| CHROMIUM | UG/L | , a | | | | | | | | ļ | | | ļ | ļ | |
| COPPER | UG/L | , <u>a</u> | | | | ļ | | ļ | | ļ | | | ļ | | ļ·=· |
| LEAD | UG/L | , a | | | | ļ | | | | | | | <u> </u> | | ļ |
| MANGANESE | UG/L | , 100 | 66 | 45 | 22 | 8 | 33 | 28 | 16 | 20 | 55 | 26 | 38 | 22 | 19 |
| NICKEL | UG/L | . 4600 | | | L | l | ļ | ļ | <u> </u> | ļ | | | | < 1 | <u> <1</u> |
| SELENIUM | UG/L | , <u>a</u> | ┫ | ļ | ļ | ļ | _ | ļ | | | | | | <u> </u> | |
| SILVER | UG/L | , <u>a</u> | ļ | | l | | 1 | - | | <u> </u> | ļ | | | | |
| ZINC | UG/L | , a | | I | | | | I | | 1 | l | L | | 1 | 1 |

a = no USEPA human health saltwater criterion.

Human health criteria based on daily ilfetime (70-year) average consumption of aquatic organisms.

Mean detected concentrations compared to chronic criteria.



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TABLE 7-36 MIXING FACTORS FOR ANALYTES IN CHANNEL ELUTRIATES THAT EXCEED CRITERIA

Values in parenthesis indicate the estimated time and distance required to achieve the required dilution for open-water at Site 104.

Brewerton, Eastern Ext.

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|------------------------|
| ARSENIC | | | 3 (0.16 hrs, 322 ft) |
| MANGANESE | | | 66 (1.39 hrs, 2799 ft) |
| TOTAL PCB (ND=0) | | | 7 (0.33 hrs, 666 ft) |
| TOTAL PCB (ND=1/2DL) | | | 36 (0.96 hrs, 1940 ft) |
| TOTAL PCB (ND=DL) | | | 66 (1.39 hrs, 2788 ft) |
| HEPTACHLOR | | 8 (0.36 hrs, 723 ft) | 14 (0.52 hrs, 1056 ft) |
| HEPTACHLOR EPOXIDE | | 4 (0.21 hrs, 420 ft) | 14 (0.5 hrs, 1012 ft) |

C&D Approach Channels (Cores)

| EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|------------------------|--------------------------|---|
| | | 2 (0.08 hrs, 149 ft) |
| | | 22 (0.74 hrs, 1482 ft) |
| | | 12 (0.43 hrs, 862 ft) |
| | | 40 (1.02 hrs, 2043 fl) |
| | | 68 (1.41 hrs, 2833 ft) |
| 1 (0.03 hrs, 64 ft) | 13 (0.46 hrs, 924 ft) | 21 (0.72 hrs, 1440 ft) |
| - | EPA Acute Aquatic Life | EPA Acute Aquatic Life EPA Chronic Aquatic Life |

C&D Approach Channels (Surficial)

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|-------------------------|
| ARSENIC | | | 2 (0.12 hrs, 236 ft) |
| COPPER | 1 (0.02 hrs, 41 ft) | 1 (0 hrs, 10 ft) | |
| MANGANESE | | | 45 (1.12 hrs, 2256 ft) |
| TOTAL PCB (ND=0) | | | 7 (0.33 hrs, 669 ft) |
| TOTAL PCB (ND=1/2DL) | | | 35 (0.95 hrs, 1913 ft) |
| TOTAL PCB (ND=DL) | | | 63 (1.36 hrs, 2746 fl) |
| HEPTACHLOR | 1 (0.03 hrs, 64 ft) | 13 (0.46 hrs, 924 ft) | 21 (0.72 hrs, 1440 ft) |
| HEPTACHLOR EPOXIDE | | 4 (0.21 hrs, 420 ft) | 14 (0.5 hrs, 1012 ft) |

Craighili

| Anaiyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Heaith |
|------------|------------------------|--------------------------|------------------------|
| ARSENIC | | | 1 (0.04 hrs, 85 ft) |
| MANGANESE | | | 8 (0.36 hrs, 727 ft) |
| HEPTACHLOR | | 8 (0.36 hrs, 723 ft) | 14 (0.52 hrs, 1056 ft) |

Craighili Angle East

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Heaith |
|----------------------|------------------------|--------------------------|------------------------|
| NITROGEN, AMMONIA | | 2 (0.06 hrs, 118 ft) | |
| ARSENIC | | | 6 (0.31 hrs, 623 ft) |
| MANGANESE | | | 33 (0.92 hrs, 1849 ft) |
| TOTAL PCB (ND=0) | | | 10 (0.38 hrs, 767 ft) |
| TOTAL PCB (ND=1/2DL) | | | 38 (0.98 hrs, 1975 ft) |
| TOTAL PCB (ND=DL) | | | 66 (1.39 hrs, 2794 fi) |
| HEPTACHLOR | 1 (0.03 hrs, 64 ft) | 13 (0.46 hrs, 924 ft) | 21 (0.72 hrs, 1440 ft) |
| HEPTACHLOR EPOXIDE | | 4 (0.21 hrs, 420 ft) | 14 (0.5 hrs, 1012 ft) |

Craighili Angie West

| EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|------------------------|---|---|
| | | 3 (0.15 hrs, 308 ft) |
| | | 28 (0.83 hrs, 1667 ft) |
| | | 6 (0.3 hrs, 596 ft) |
| | | 36 (0.96 hrs, 1930 ft) |
| | | 66 (1.39 hrs, 2792 ft) |
| 4 (0.18 hrs, 371 ft) | 29 (0.86 hrs, 1731 ft) | 50 (1.21 hrs, 2443 ft) |
| | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |
| | EPA Acute Aquatic Life 4 (0.18 hrs, 371 ft) | EPA Acute Aquatic Life EPA Chronic Aquatic Life 4 (0.18 hrs, 371 ft) 29 (0.86 hrs, 1731 ft) 6 (0.28 hrs, 569 ft) 6 (0.28 hrs, 569 ft) |



Craighill Entrance

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|-------------------------|
| NITROGEN, AMMONIA | | 1 (0 hrs, 2 ft) | |
| ARSENIC | | | 4 (0.18 hrs, 373 ft) |
| MANGANESE | | | 16 (0.56 hrs, 1135 ft) |
| TOTAL PCB (ND=0) | | | 10 (0.38 hrs, 769 ft) |
| TOTAL PCB (ND=1/2DL) | | | 38 (0.98 hrs, 1977 ft) |
| TOTAL PCB (ND=DL) | | | 66 (1.39 hrs, 2795 ft) |
| HEPTACHLOR | 2 (0.05 hrs, 101 ft) | 14 (0.51 hrs, 1030 ft) | 24 (0.76 hrs, 1532 ft) |
| HEPTACHLOR EPOXIDE | | 4 (0.21 hrs, 420 ft) | 14 (0.5 hrs, 1012 ft) |

Craighill Upper Range

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|-------------------------------|---------------------------------|-------------------------|
| ARSENIC | | | 3 (0.13 hrs, 270 ft) |
| MANGANESE | | | 20 (0.67 hrs, 1356 ft) |
| TOTAL PCB (ND=1/2DL) | | | 31 (0.9 hrs, 1808 ft) |
| TOTAL PCB (ND=DL) | | | 62 (1.35 hrs, 2722 ft) |
| HEPTACHLOR EPOXIDE | | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |

Cutoff Angle

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|------------------------|
| ARSENIC | | | 4 (0.22 hrs, 439 ft) |
| MANGANESE | | | 55 (1.28 hrs, 2579 ft) |
| TOTAL PCB (ND=0) | | | 6 (0.3 hrs, 596 ft) |
| TOTAL PCB (ND=1/2DL) | | | 35 (0.95 hrs, 1906 ft) |
| TOTAL PCB (ND=DL) | | | 64 (1.37 hrs, 2757 ft) |
| HEPTACHLOR | l (0.01 hrs, 26 ft) | 11 (0.4 hrs, 806 ft) | 19 (0.66 hrs, 1327 ft) |
| HEPTACHLOR EPOXIDE | | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |

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Inside Site 104

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| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|--------------------|------------------------|--------------------------|-------------------------|
| SULFIDE, TOTAL | | 700 (3.65 hrs, 7365 ft) | |
| ARSENIC | | | 4 (0.18 hrs, 373 ft) |
| MANGANESE | | | 10 (0.38 hrs, 763 ft) |
| HEPTACHLOR | | 8 (0.36 hrs, 723 ft) | 14 (0.52 hrs, 1056 ft) |
| HEPTACHLOR EPOXIDE | | 5 (0.24 hrs, 477 ft) | 15 (0.55 hrs, 1108 ft) |

Ocean Reference

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------|-------------------------------|---------------------------------|------------------|
| SULFIDE, TOTAL | | 245 (2.5 hrs, 5033 ft) | |
| TRIBUTYLTIN | | | |

Outside Site 104

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|---------------------------------|------------------------|
| ARSENIC | | | 6 (0.3 hrs, 610 ft) |
| MANGANESE | | | 34 (0.93 hrs, 1887 ft) |
| TOTAL PCB (ND=0) | | | 9 (0.37 hrs, 748 ft) |
| TOTAL PCB (ND=1/2DL) | | | 38 (0.98 hrs, 1982 ft) |
| TOTAL PCB (ND=DL) | | | 67 (1.4 hrs, 2819 ft) |
| HEPTACHLOR | | 8 (0.36 hrs, 723 ft) | 14 (0.52 hrs, 1056 ft) |

Swan Point

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|-------------------------|
| ARSENIC | | | 3 (0.17 hrs, 351 ft) |
| MANGANESE | | | 26 (0.8 hrs, 1611 ft) |
| TOTAL PCB (ND=0) | | | 5 (0.24 hrs, 472 ft) |
| TOTAL PCB (ND=1/2DL) | | | 34 (0.93 hrs, 1871 ft) |
| TOTAL PCB (ND=DL) | | | 63 (1.36 hrs, 2734 ft) |
| HEPTACHLOR | 1 (0.01 hrs, 26 ft) | 11 (0.4 hrs, 806 ft) | 19 (0.66 hrs, 1327 ft) |
| HEPTACHLOR EPOXIDE | | 7 (0.33 hrs, 663 ft) | 23 (0.74 hrs, 1493 ft) |

4 of 5

Tolchester Channel - North

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|------------------------|
| CYANIDE | 5 (0.26 hrs, 523 ft) | 5 (0.26 hrs, 523 ft) | |
| ARSENIC | | | 4 (0.19 hrs, 380 ft) |
| MANGANESE | | | 38 (0.98 hrs, 1984 ft) |
| TOTAL PCB (ND=0) | | | 15 (0.55 hrs, 1093 ft) |
| TOTAL PCB (ND=1/2DL) | | | 42 (1.06 hrs, 2149 ft) |
| TOTAL PCB (ND=DL) | | | 69 (1.42 hrs, 2860 ft) |
| HEPTACHLOR | | 10 (0.38 hrs, 766 ft) | 17 (0.59 hrs, 1187 ft) |
| HEPTACHLOR EPOXIDE | | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |

Tolchester Channel - South

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| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|--------------------|-------------------------------|---------------------------------|-------------------------|
| SULFIDE, TOTAL | | 198 (2.26 hrs, 4568 ft) | |
| ARSENIC | | | 2 (0.11 hrs, 222 ft) |
| MANGANESE | | | 22 (0.73 hrs, 1469 ft) |
| HEPTACHLOR | | 8 (0.36 hrs, 723 ft) | 14 (0.52 hrs, 1056 ft) |
| HEPTACHLOR EPOXIDE | | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |

Tolchester Straightening

| Analyte | EPA Acute Aquatic Life | EPA Chronic Aquatic Life | EPA Human Health |
|----------------------|------------------------|--------------------------|-------------------------|
| ARSENIC | | | 13 (0.48 hrs, 966 ft) |
| BERYLLIUM | | | 3 (0.14 hrs, 291 ft) |
| MANGANESE | | | 19 (0.65 hrs, 1296 ft) |
| SILVER | 1 (0.02 hrs, 42 ft) | | |
| TOTAL PCB (ND=0) | | | 11 (0.39 hrs, 789 ft) |
| TOTAL PCB (ND=1/2DL) | | | 38 (0.98 hrs, 1977 ft) |
| TOTAL PCB (ND=DL) | | | 65 (1.38 hrs, 2782 ft) |
| HEPTACHLOR | | 7 (0.33 hrs, 663 ft) | 12 (0.44 hrs, 874 ft) |
| HEPTACHLOR EPOXIDE | | 6 (0.28 hrs, 569 ft) | 18 (0.64 hrs, 1278 ft) |

CHAPTER 8 -TOXICITY TESTING

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| Ma | nger Harbor | Developme | •nt | | ATTENTION: | Mr. Nathai | nial Brown | 00707.00 |
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EA Engineering, Science, and Technology

15 Loveton Circle Sparks, MD 21152 Telephone: 410-771-4950 Fax: 410-771-4204

27 October 2000



Christina E. Correale Chief - Operations Division USACE Baltimore District P.O. Box 1715 10 North Howard Street - Room 8620 Baltimore, MD 21203-1715

RE: Draft Sediment Report - Site 104, Chapter 8

Dear Ms. Correale:

Accompanying this letter please find the draft of Chapter 8 of the above referenced report. This has been carefully reviewed by senior technical scientists and by our senior editor. We are currently in the process of completing Chapters 9 and 10 and will forward them to you as soon as they are completed. Please feel free to call either Peggy Derrick or me with any questions.

Sincerely,

199 Wornel

Peggy Derrick, Project Manager/Scientist

Frank W. Pine, Ph.D.

Senior Scientist/Project Director

Cc:

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F. Hamons C. Donovan File 60957.30 P:\Federal\DOD\ARMY\projects\6095730\CORRESP\Correale 27 October 2000.doc

8. TOXICITY TESTING

EA's Ecotoxicology Laboratory performed water column and whole sediment toxicity testing on sediment composites collected from the approach channels, from Placement Site 104, from an Outside Placement Site 104 reference area, and from the Norfolk Ocean Disposal Site reference area. The testing was in compliance with ITM requirements. The toxicity testing program consisted of acute water column bioassays with *Mysidopsis bahia* (opossum shrimp), *Cyprinodon variegatus* (sheepshead minnow), *Mytilus* sp. (blue mussel), and *Menidia beryllina* (inland silverside), and 10-day whole sediment toxicity tests with *Neanthes arenaceodentata* (estuarine polychaete) and *Leptocheirus plumulosus* (estuarine amphipod). Acute water column bioassays with *Arbacia punetulata* (purple sea urchin) were initially conducted. However, USEPA Region III did not recommend this species, and USEPA representatives indicated that this species did not accurately represent aquatic organisms that could potentially be impacted in the Bay. The results of these tests are not discussed in this chapter, but are provided in Attachment V. The acute water column bioassays and the whole sediment toxicity tests evaluated the effects of exposure to the sediment elutriates and whole-sediment, respectively, on survival of the test organisms.

8.1 METHODS

The toxicity testing program consisted of three separate sequential rounds of testing:

- Round 1: Initial water column and solid phase testing with sediment from the approach channels and from Inside Placement Site 104 (September-November 1999).
- Round 2: Additional water column and solid phase testing conducted in response to recommendations from USEPA Region III-Philadelphia. Additional testing with sediment from an Outside Placement Site 104 reference area and additional testing of sediment from approach channels and Inside Placement Site 104 with additional test species (December 1999-February 2000) was conducted.
- Round 3: Testing of Ocean Reference sediment (in conjunction with Woodrow Wilson Bridge sediment testing program) (February-March 2000).

A summary of the toxicity testing schedule is provided in Table 8-1. The water column and whole sediment toxicity testing was conducted in accordance with USEPA/USACE guidance (1998) and EA (1996). The testing procedures, acceptability criteria, and quality assurance protocols are fully documented in the Quality Assurance Project Plan (QAPP) for the ecotoxicological testing program (EA 2000c) (Appendix B). To take advantages in efficiencies from testing on two concurrent programs using the same ocean placement reference site, testing procedures for the Ocean Reference sediment followed the same methodology as the Site 104 testing and are documented in the ecotoxicology QAPP for the Woodrow Wilson Bridge project (EA 2000b).

Original data sheets, records, memoranda, notes, and computer printouts for toxicity testing components are archived at EA's Baltimore Office in Sparks, Maryland. These data will be retained for a period of 5 years unless a longer period of time is requested by USACE-Baltimore District.

8.1.1 Sample Receipt and Preparation

Approximately 20 gallons of composited sediment from each sampling reach were required for the ecotoxicological (including bioaccumulation) testing. Sediment composites for the ecotoxicological testing are described in Table 8-2. Processing and homogenization of sediment followed procedures described in Chapter 4.

After completion of processing, compositing, and homogenization, reach composites and site water samples were logged in and assigned EA laboratory accession numbers. The sediment and water samples were stored in the dark in a secured walk-in cooler at 4°C until used for testing. Prior to use in toxicity testing, large rocks and debris were manually removed and discarded from each sample. Table 8-3A summarizes the sample identifications, accession numbers, and collection and receipt information for sediment and water from the approach channels, inside Placement Site 104, and outside Placement Site 104. Table 8-3B summarizes the sample identifications, accession numbers, and collected from the Ocean Reference site. Copies of chain-of-custody records are provided in Attachments V-A and V-B for approach channels/Site 104 and Ocean Reference, respectively.

8.1.2 Water Column Testing

For the water column toxicity testing, elutriates were prepared from each composited sediment sample. Prior to elutriate preparation, the site water for each sampling reach was salinity adjusted to 30 ppt, as per USEPA/USACE (1998) guidance for the selected test species, using Forty Fathoms synthetic sea salts. Following USEPA/USACE (1998) guidance, a subsample of each homogenized sediment was combined with its respective site water in a 1:4 sediment to water ratio, on a volume/volume basis. The sediment/water combination was thoroughly mixed by vigorous aeration and manually stirring for 30 minutes at 20°C, and was then allowed to settle for one hour. After settling, the supernatant was decanted off and used for the water column acute toxicity testing. All elutriates were used for testing within 24 hours of preparation.

Static, non-renewal bioassays were conducted on the prepared elutriates using *Mysidopsis bahia* (opossum shrimp), *Cyprinodon variegatus* (sheepshead minnow), *Menidia beryllina* (inland silverside), and *Mytilus* sp. (blue mussel). The test organisms were acquired from scientific organism vendors and, when appropriate, gradually acclimated to test temperature and salinity prior to use in testing. The *Mytilus* sp. testing was performed by Northwestern Aquatic Sciences (NAS) located in Newport, Oregon. Elutriates for the *Mytilus* sp. testing were prepared by EA's Ecotoxicology Laboratory, shipped via overnight delivery on the day of preparation, and were used by NAS for the *Mytilus* sp. test within 24 hours of preparation.

8.1.2.1 Mysidopsis bahia and Cyprinodon variegatus Water Column Toxicity Testing

Approach Channels and Inside Placement Site 104 (Round 1)

The 96-hour toxicity tests with *M. bahia* and *C. variegatus* were conducted in two batches (25-29 October 1999, 26-30 October 1999). Two lots of test organisms per species were acquired from Cosper Environmental Services, Inc. (Bohemia, New York). For use in the 25 October testing, stocks of *M. bahia* (lot number MB-402) and *C. variegatus* (CV-287) were received at EA on 23 October from Cosper Environmental Services. Additional lots from Cosper Environmental Services, *M. bahia* (lot MB-403) and *C. variegatus* (lot CV-288), were received on 26 October for use in toxicity testing the same day. The opossum shrimp and sheepshead minnows were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding.

Outside Placement Site 104 (Round 2)

The 96-hour toxicity tests with *M. bahia* and *C. variegatus* were conducted 4-8 January 2000. An additional test with the Inside Site 104 sediments was conducted concurrently with the Outside Site 104 sediment to compare the consistency of results for the Inside Site 104 testing in Round 1 and Round 2. Test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The lots of organisms MB-413 (*M. bahia*) and CV-304 (*C. variegatus*) were received on 4 January 2000 from Aquatic BioSystems for use in toxicity testing the same day. The opossum shrimp and sheepshead minnows were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding.

Ocean Reference Site (Round 3)

The 96-hour toxicity test with *M. bahia* was conducted on 16-20 February 2000. The test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The organisms, *M. bahia* (lot MB-422), were received at EA on 15 February 2000. The opossum shrimp were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding. *C. variegatus* tests were not conducted under the Woodrow Wilson Bridge testing program.

Test concentrations of 100, 50, and 10 percent elutriate were prepared by measuring aliquots of elutriate in a graduated cylinder and bringing to final volume with artificial seawater. A dilution water control of artificial seawater was also prepared. The artificial seawater was prepared by mixing Forty Fathoms synthetic sea salts with laboratory water to a final salinity of 30 ppt. The source of the laboratory water was the City of Baltimore municipal tap water which was passed through a high-capacity, activated carbon filtration system. This synthetic seawater formulation has proven acceptable for aquatic toxicological studies, and has been used successfully at EA for maintaining multigeneration cultures of *M. bahia*, and for holding healthy populations of estuarine and marine species. Batches of artificial seawater were aerated and aged at least 24 hours prior to use in testing.



The opossum shrimp and sheepshead minnow testing utilized 1-L beakers as test chambers. Each beaker contained 250 ml of test solution, and there were five replicate beakers per test concentration. Ten organisms were randomly introduced into each replicate for a total of 50 organisms per concentration. The test chambers were maintained at 20±1°C with a 16-hour light/8-hour dark photoperiod. Temperature, pH, and dissolved oxygen were measured in one replicate of each concentration daily for the 96-hour exposure period. Due to the size of the water quality probe and the danger of injury to the test organisms, salinity was measured in each concentration only at test initiation (prior to introduction of the test organisms) and at termination. A summary of water quality measurements is presented in Table 8-4A and 8-4B (M. bahia-approach channels and Site 104), Table 8-4C (M. bahia- ocean reference), and Tables 8-5A and 8-5B (C. variegatus-approach channels and Inside Site 104). The number of live organisms were counted daily and recorded on the test data sheets. The opossum shrimp were fed freshly hatched Artemia sp. daily to avoid cannibalism during testing. Copies of the opossum shrimp and sheepshead minnow acute toxicity test data sheets are included in Attachments V-A and V-B, for the approach channel/Site 104 and the Ocean Reference, respectively.

8.2.1.2 Menidia beryllina Water Column Toxicity Testing

Approach Channels, Inside Placement Site 104, and Outside Placement Site 104 (Round 2)

The 96-hour *M. beryllina* acute toxicity tests were conducted in five batches (20-24, 21-25, and 25-29 January 2000, and 4-8 and 7-11 February 2000). The test methodologies followed those of the *M. bahia* and *C. variegatus* water column testing with the exception of the test volume per chamber. Each test chamber for the *M. beryllina* testing contained 200 ml of test solution.

The *M. beryllina* toxicity testing was performed with 9-14 day old fish (hatched within a 24-hour period). During the testing program, five lots of *M. beryllina* were acquired from Aquatic BioSystems (Fort Collins, Colorado), and were gradually acclimated to test conditions in EA's Culture Facility prior to use in testing. The *M. beryllina* were fed *Artemia* nauplii at the 48-hour intermediate observation period during testing.

Ocean Reference (Round 3)

The 96-hour toxicity test with *M. beryllina* was conducted on 16-20 February 2000. The test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The organisms, *M. menidia* (lot MS-080), were received at EA on 15 February 2000, and were gradually acclimated to test conditions in EA's Culture Facility prior to use in testing. The *M. beryllina* were fed *Artemia* sp, nauplii (<24 hours old) during the holding period.

A summary of water quality parameters measured during the *M. beryllina* testing is presented in Tables 8-6A (approach channels, Inside Site 104, and Outside Site 104) and 8-6B (Ocean Reference). Attachments V-A and V-B contain copies of the data sheets from the *M. beryllina* toxicity testing for the approach channels/Site 104 and ocean testing, respectively.

8.1.2.3.1 Mytilus sp. Water Column Toxicity Testing

The *Mytilus* sp. bivalve embryo-larval toxicity testing was conducted by Northwestern Aquatic Sciences (NAS) located in Newport, Oregon. The 30 ppt salinity elutriates were prepared by EA on 20 December 1999 (approach channels and Site 104) and on 14 February 2000 (Ocean Reference), packed on wet ice the same day as the elutriate preparation, and shipped by overnight courier to NAS. The bivalve toxicity tests were initiated by Northwestern upon receipt of the elutriate samples on 21 December 1999 and on 15 February 2000.

The adult mussels were acquired from Carlsbad Aquafarm (Carlsbad, California) on 21 December 1999 and 15 February 2000. The mussels were induced to spawn by gradually cycling the temperature of the holding water several times through the range of 15-23°C. Spawning animals were rinsed and isolated in small dishes containing clean filtered seawater for collection of gametes. Four females and three males were isolated for gamete collection.

Unfertilized eggs were rinsed and suspended in clean seawater at a concentration of approximately 5,000 eggs/ml. Sufficient sperm was added to the egg suspension to achieve an approximate sperm to eggs ratio of 5:1. Ten minutes after sperm addition, the suspension was filtered through a 25 μ m Nitex® screen to remove remaining sperm, and the embryos were resuspended and adjusted to achieve a stock concentration of about 2,500 embryos/ml.

Test chambers were 30 ml borosilicate glass vials containing 10 ml of test solution. Test concentrations of 100, 50, 10, and 0 percent elutriate were prepared using 30 ppt site water (Yaquina Bay, Oregon) for the dilution. A laboratory control of natural clean seawater was also utilized. At test initiation, $100 \ \mu$ l of well-mixed embryo suspension was added to each test chamber. The tests were maintained at 15 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Temperature, pH, dissolved oxygen, and salinity were measured daily on surrogate test chambers (without test organisms). A summary of water quality measurements from the *Mytilus* sp. testing is provided in Table 8-7A (approach channels, Inside Site 104, Outside Site 104) and Table 8-7B (Ocean Reference). The toxicity tests were terminated on 23 December 1999 and 17 February 2000 by adding 1 ml of 37 percent buffered formalin to each test chamber. The preserved embryos were observed microscopically, to determine the percentage of normally developed larvae. The complete reports for the *Mytilus* sp. elutriate testing are presented in Attachments V-A and V-B.

8.1.3 Whole Sediment Testing

8.1.3.1 Neanthes arenaceodentata and Leptocheirus plumulosus Whole Sediment Testing

Approach Channels and Inside Site 104 (Round 1)

Whole sediment toxicity testing was conducted with the estuarine polychaete *Neanthes* arenaceodentata and the estuarine amphipod *Leptocheirus plumulosus*. The approach channel and Inside Site 104 sediments were evaluated on 22 October-1 November with

N. arenaceodentata and on 23 October-2 November 1999 with L. plumulosus. The polychaete worms (organism lot numbers NA-005) were acquired from Dr. Donald Riesh, California State University (Long Beach, California) on 21 October 1999, and the amphipods (LP-010) were acquired from University of Maryland, Wye Research and Education Center (Queenstown, Maryland) on 21 October 1999. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C and to the appropriate test salinity [20 (±10%) ppt for Leptocheirus plumulosus and 30 ($\pm 10\%$) ppt for Neanthes arenaceodentata, as per USEPA/USACE (1998)].

Outside Site 104 (Round 2)

Additional lots of N. arenaceodentata and L. plumulosus were acquired for the later testing of the Outside Site 104 reference sediment. The Inside Site 104 sediment was re-run concurrently with the Outside Site 104 tests to compare the consistency of results for Inside Site 104 testing in Round 1 and Round 2. The N. arenaceodentata lot (NA-006) was received on 28 December from Dr. Donald Reish for use in toxicity testing on 30 December 1999. The L. plumulosus lot (LP-012) was acquired from Aquatic BioSystems on 27 January for use in toxicity testing on 28 January 2000.

Ocean Reference (Round 3)

L. plumulosus (organism lot # NA-005) were acquired from Aquatic BioSystems (Fort Collins, CO) on 23 February 2000. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C. Based on recommendations from USEPA Region III, a lower Chesapeake Bay control sediment (collected from a USEPA Region III approved location) was tested in place of the Ocean Reference Site sediment. L. plumulosus are sensitive to sediment grain size characteristics (USEPA 1993b). This species of estuarine amphipod prefers fine sediment characteristic of the Chesapeake Bay and does not typically survive well in coarsegrained sandy sediment. The lower Chesapeake Bay control sediment was tested in place of the Ocean Reference sediment to reduce the potential for adverse grain-size effects on survival.

The whole sediment toxicity tests were conducted as static, non-renewal tests with 10 days of exposure to the whole sediments and overlying water. Artificial seawater (Forty Fathoms sea salts) at 30 ppt salinity for N. arenaceodentata and 20 ppt salinity for L. plumulosus was used as the overlying water. The sediments and overlying water were added to the test chambers, and the suspended sediments were allowed to settle 1-3 days. During the settling period, ammonia was monitored in the overlying water of each sediment test chamber as directed by USEPA/USACE guidance to ensure that ammonia was not an artifact of the sediment test. No replacement of the overlying water was required due to the low measured levels of ammonia (<2 mg/L NH₃-N). The addition of the test organisms to the exposure chambers marked the initiation of the toxicity tests. A natural Chesapeake Bay sediment was used as a laboratory control for the whole sediment toxicity tests.

The N. arenaceodentata and L. plumulosus tests utilized 1-L beakers as the exposure chambers, with each beaker containing 200 ml of sediment and 700 ml of overlying water. There were five



replicate chambers for each sediment sample and control. Test organisms were randomly assigned to the test chambers. The *N. arenaceodentata* test had 5 organisms per replicate for a total of 25 organisms exposed per sample, while the *L. plumulosus* test had either 20 or 25 organisms per replicate chamber for a total of 100 organisms per sample (23 October 1999 testing) or 125 organisms per sample (28 January 2000 testing).

The tests were maintained at 20 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Water quality measurements of temperature, pH, dissolved oxygen, and salinity were recorded daily on one replicate of each sample and control. Water quality parameters measured during the *N. arenaceodentata* testing are summarized in Tables 8-8A (approach channels and Inside Site 104) and 8-8B (Outside and Inside Site 104). Water quality parameters measured during the *L. plumulosus* testing are summarized in Tables 8-9A (approach channels and Inside Site 104), 8-9B (Outside and Inside Site 104), and 8-9C (Ocean Reference). Additionally, test chambers were visually inspected daily for abnormal organism behavior/lack of burrowing. The test organisms were not fed during the 10-day exposure period.

After 10 days of exposure, the test organisms were retrieved from the samples and the number of live organisms per replicate was recorded. Copies of the original data sheets for the *N. arenaceodentata* and the *L. plumulosus* testing are included in Attachments V-A (approach channels/Site 104) and V-B (Ocean Reference), respectively.

8.1.3.2 Mysidopsis bahia Whole Sediment Testing

For the Ocean Reference sediment, *Mysidposis bahia* was tested as the second species for the whole sediment bioassays. The Ocean Reference sediment was tested in conjunction with another testing program (Woodrow Wilson Bridge) that requested *M. bahia* testing, rather than *Neanthes arenaceodentata*. Either species is acceptable to USEPA Region III for evaluating open water and ocean placement.

The opossum shrimp (lot MB-423) were acquired from Aquatic BioSystems on 18 February 2000. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C and at the required test salinity (30 ppt).

The whole sediment toxicity tests were conducted as static, non-renewal tests with 10 days of exposure to the whole sediments and overlying water. Artificial seawater (Forty Fathoms sea salts) at 30 ppt salinity for *M. bahia* was used as the overlying water. The sediments and overlying water were added to the test chambers on 17 February 2000, and the suspended sediments were allowed to settle 1-5 days. During the settling period, ammonia was monitored in the overlying water of each sediment chamber. Due to the low measured levels of ammonia ($<2 \text{ mg/L NH}_3$ -N), no replacement of the overlying water was required. The addition of the test organisms to the exposure chambers marked the initiation of the toxicity tests. The *M. bahia* whole sediment toxicity tests were initiated on 18 February 2000. A natural Chesapeake Bay sediment was used as a laboratory control for the *M. bahia* tests.

The *M. bahia* tests utilized 1-L beakers as the exposure chambers, with each beaker containing 300 ml of sediment and 650 ml of overlying water. There were five replicate chambers for each
sediment sample and control. Test organisms were randomly assigned to the test chambers. The *M. bahia* tests had 10 organisms per replicate for a total of 50 organisms exposed per sample.

The tests were maintained at 20 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Water quality measurements of temperature, pH, dissolved oxygen, and salinity were recorded daily on one replicate of each sample and control. Water quality parameters measured during the Ocean Reference *M. bahia* testing are summarized in Table 8-10. The test organisms were not fed during the 10-day exposure period. After 10 days of exposure, the test organisms were retrieved from the samples and the number of live organisms per replicate was recorded. Copies of the original data sheets for the *M. bahia* testing are included in Attachment V-B.

8.1.4 Data Analysis / Statistics

Statistical analyses were performed on the water column and whole sediment test data according to USEPA/USACE (1998) guidance. Survival (or larval development) of the organisms exposed to the test material for the prescribed time period was statistically compared (p=0.05) to either the laboratory control (elutriate tests) or the reference sediment (whole-sediment tests) as appropriate for each test using the student t-test.

For the elutriate testing, a 96-hour LC50 (median lethal concentration), or EC50 (median effective concentration for *Mytlius* sp.), was calculated for each test species using either the linear interpolation, trimmed Spearman-Karber method (Hamilton et al., 1977), or probit method (as described by Stephan 1977). The LC50 is an estimate of the elutriate concentration that is lethal to 50 percent of the test organisms, or that creates a sub-lethal effect on the development of 50 percent (EC50) of the test organisms, in the time period prescribed by the test. If survival in the 100 percent elutriate concentration was at least 10 percent lower than the dilution water control, then a statistical comparison (t-Test) was performed between the 100 percent elutriate concentrations were independent and normally distributed, and that the variances of the observations were equal between the two groups. The F-Test was used to test for homogeneity of variance. The test for normality was the Shapiro-Wilk's Test. When the data did not meet the normality assumption, the nonparametric test, Wilcoxon's Rank-Sum Test, was used to analyze the data. An arc sine (square root [Y]) transformation was performed on the survival percentages, where appropriate.

For data sets in which the 100 percent concentration was statistically different from the control, an addition analysis was performed to determine the No Observed Adverse Effect Concentration (NOAEC). Based on USEPA (1993) guidance for standard multi-concentration effluent toxicity tests, a multiple mean comparison was conducted to statistically compare the 100, 50, and 10 percent concentrations to the control. This comparison utilized the dose response data from the three test concentrations and control. A concentration which had no surviving organisms was excluded from the analysis. A parametric or nonparametric statistical test was utilized based on the assumptions of normality and homogeneity of variance. The test for normality was the Shapiro-Wilk's Test, and the test for homogeneity of variance was the Bartlett's Test. For parametric data, an analysis of variance (ANOVA) and either Dunnett's Mean Comparison test or Bonferroni's T-test was used (depending on equal or unequal replicate numbers). Steel's Many-One Rank Test or the Wilcoxon Rank Sum Test were the alternative nonparametric tests.

For the whole sediment toxicity test data, statistical analyses were performed to determine if exposure to any of the sediment samples resulted in significantly lower survival of the test organisms as compared to the reference site. If survival in a test sample was at least 10 percent lower than the reference, then a t-test or Wilcoxon's Rank-Sum Test (depending on data characteristics) was performed to compare the single test sample to the reference (Inside Site 104).

For the whole sediment bioassays, only results generated within a single round of testing should be statistically compared to each other because of potential differences associated with survival in the control organisms. In this testing program, data from different rounds of whole-sediment testing could not be statistically compared against each other, but were compared qualitatively (i.e., toxic or non-toxic). For the water column bioassays, statistical comparisons are conducted against a laboratory or test control (not a reference sample). In these cases, each sample had an independent statistical comparison against a control, and the results generated by different rounds of testing stand as independent measures of toxicity.

8.1.5 Reference Toxicant Testing

In conformance with EA's QA/QC program requirements, reference toxicant testing was performed on the acquired lots of organisms utilized in the testing program or reference toxicant data were obtained from the test organism supplier. The reference toxicant tests consisted of a graded concentration series of a specific toxicant in water only tests, with no sediment present in the test chambers.

The reference toxicant for *M. bahia*, *C. variegatus*, and *M. beryllina* was potassium chloride (KCl); the reference toxicant for *Mytilus* sp. was copper sulfate (CuSO₄·5H₂0); and the reference toxicant for *N. arenaceodentata* and *L. plumulosus* was cadmium chloride (CdCl₂). Reference toxicant testing was also conducted for the species utilized in the bioaccumulation studies (Chapter 9). The results of the reference toxicant tests were compared to established control chart limits.

8.2 RESULTS

8.2.1 Water Column Testing

8.2.1.1 Mysidopsis bahia

The results of the *M. bahia* toxicity testing are presented in Tables 8-11A (approach channels and Inside Site 104), 8-11B (Inside and Outside Site 104), and 8-11C (Ocean Reference).

Approach Channels and Inside Placement Site 104 (Round 1)

The results of the initial 14 elutriate tests with *M. bahia* are provided in Table 8-11A. Five of the prepared elutriates exhibited some inhibition of survival to *M. bahia*. Samples from the Craighill Entrance (CRE), Craighill Angle-West (CRA-W), Tolchester-South (TLC-S),

Brewerton Eastern Extension (BE), and C& D Approaches (surficial CD) had 96-hour LC50s of >100 percent elutriate; however, the survival in the 100 percent elutriate concentration for each sample was statistically lower than the control. Calculation of NOAECs for these samples indicated that no effect would be expected at elutriate concentrations of >50%. Samples from the Craighill (CR), Craighill Angle-East (CRA-E), Craighill Upper Range (CRU), Cutoff Angle (CUT), Tolchester North (TLC-N), Tolchester Straightening (TLS), Swan Point (SWP), C&D Approaches (cores CD-VC), and Inside Site 104 (KI-Reference) were not acutely toxic to M. bahia, with 96-hour LC50s of >100 percent elutriate and no statistically difference in survival between the control and the 100 percent elutriate concentration.

Outside Placement Site 104 (Round 2)

The results of the *M. bahia* testing on the Outside Site 104 (KI-OUT) reference elutriate are presented in Table 8-11B. The Outside Site 104 sample was not acutely toxic to *M. bahia* with 94 percent survival in the 100 percent elutriate concentration (48-hour LC50 >100 percent elutriate). The Inside Site 104 reference elutriate was re-analyzed concurrently with the Outside Site 104 elutriate using a fresh sample of Inside Site 104 sediment and site water. The Inside Site 104 elutriate was again not acutely toxic to *M. bahia* (48-hour LC50 >100 percent elutriate), with 100 percent survival in the 100 percent elutriate concentration.

Ocean Reference (Round 3)

The results of the Ocean Reference elutriate testing with *M. bahia* are provided in Table 8-11C. The Ocean Reference sample was not acutely toxic to *M. bahia*, with a 96-hour LC50 of >100 percent elutriate and no statistical difference in survival between the control and the 100 percent elutriate concentration.

8.2.1.2 Cyprinodon variegatus

The results of the sheepshead minnow acute toxicity tests are summarized in Tables 8-12A (approach channels and Inside Site 104) and 8-12B (Inside and Outside Site 104).

Approach Channels and Inside Placement Site 104 (Round 1)

Results of the *C. variegatus* acute toxicity tests conducted on the 13 channel elutriates and Inside Site 104 are provided in Table 8-12A. None of the elutriates was acutely toxic to *C. variegatus*. Survival in the 100 percent elutriate concentrations was at least 92 percent, while survival in the controls was a minimum of 98 percent. The 96-hour *C. variegatus* LC50s for all of the tested elutriates, including the Inside Site 104 reference, were >100 percent elutriate.

Outside Placement Site 104 (Round 2)

The results of the Outside Site 104 reference elutriate testing with *C. variegatus* are summarized in Table 8-12B. The Outside Site 104 elutriate was not acutely toxic to *C. variegatus*, with 96 percent survival in the 100 percent elutriate concentration. The re-analyzed Inside Site 104 reference elutriate was also not acutely toxic (100 percent survival in the 100 percent elutriate

concentration). The 48-hour LC50s for the Inside Site 104 and Outside Site 104 elutriates were both >100 percent elutriate.

Ocean Reference (Round 3)

C. variegatus was not tested with the Ocean Reference sediment.

8.2.1.3 Menidia beryllina

Results of the inland silverside testing are summarized in Table 8-13A (approach channels, Inside Site 104, and Outside Site 104) and Table 8-13B (Ocean reference). These tests were performed as an additionally requested species during the testing program.

Approach Channels, Inside Placement Site 104 and Outside Placement Site 104 (Round 2)

The *M. beryllina* acute toxicity test results are summarized in Table 8-13A. Four of the elutriates (Inside Site 104, Outside Site 104, Swan Point, and Tolchester Straightening) were not acutely toxic to *M. beryllina*, with 48-hour LC50s of >100 percent elutriate and no statistical difference in survival between the 100 percent elutriate concentration and the control. Samples for the Craighill Upper Range and the C&D Approaches (cores) had 48-hour LC50s of >100 percent elutriate; however, there was a statistically significant decrease in survival in the 100 percent elutriate concentration of NOAECs indicated that no effects to survival would be expected in the 10% elutriate and 50% elutriate for the Craighill Upper Range and C&D Approach (cores), respectively.

The remaining nine elutriates [Craighill, Craighill Entrance, Craighill Angle-East, Craighill Angle West, Cutoff Angle, Brewerton Eastern Extension, C&D Approaches (surficial), Tolchester-South, and Tolchester-North] were all acutely toxic to *M. beryllina* with 48-hour LC50s ranging from 23.8 percent elutriate (Tolchester-South) to 70.7 percent elutriate (Craighill Entrance). Calculation of NOAECs indicated that no effects to survival would be expected at a concentration of 10% elutriate in 8 of the 9 channel elutriates. The NOAEC was <10% elutriate for Brewerton Eastern Extension.

Ocean Reference (Round 3)

The *M. beryllina* acute toxicity test results for the Ocean Reference site are presented in Table 8-13B. The Ocean Reference elutriate was not acutely toxic to *M. beryllina*, with 96-hour LC50 of >100 percent elutriate and no statistical difference in survival between the control and the 100 percent elutriate concentration.

8.2.1.4 Mytilus sp.

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Results of the blue mussel testing are summarized in Table 8-14A (approach channels, Inside Site 104, Outside Site 104) and Table 8-14B (Ocean Reference).

Approach Channels, Inside Placement Site 104 and Outside Placement Site 104 (Round 2)

| Baltimore Harbor Approach Channels a | and Placement Site 104 |
|--------------------------------------|------------------------|
| Dredged Material Evaluation | Draft Report |
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Results of the *Mytilus* sp. embryo larval toxicity tests are presented in Table 8-14A. Two samples (Craighill and Craighill Upper Range) were not acutely toxic to *Mytilus* sp., with 87 percent normal development in the 100 percent elutriate concentration and 48-hour EC50s of >100 percent elutriate. The Inside Site 104 reference elutriate (48-hour EC50>100 percent elutriate) was marginally toxic with 82 percent normal development in the 100 percent elutriate concentration which was statistically lower than the control (97 percent normal development). The Outside Site 104 reference elutriate was more toxic than the Inside Site 104 elutriate, with a 48-hour EC50 of 63.8 percent elutriate.

Elutriates for the Cutoff Angle, Tolchester-South, Tolchester-North, Tolchester Straightening, Brewerton Eastern Extension, and the C&D Approaches (surficial) had 48-hour EC50s ranging from 56.4 to 79.7 percent elutriate. Elutriates for the Craighill Angle-East, Craighill Angle-West, and Swan Point had 48-hour EC50s of 43.5, 46.8, and 47.8 percent elutriate, respectively. Elutriates for the Craighill Entrance and C&D Approaches (cores) were the most toxic with 48-hour EC50s of 22.0 and 21.2 percent elutriate, respectively. Overall, calculation of NOECs indicated that only one elutriate (C&D Approach cores) would be expected to effect larval development at concentrations of <10% elutriate.

Ocean Reference (Round 3)

The results of the elutriate testing with *Mytilus* sp. for the Ocean Reference are presented in Table 8-14B. The Ocean Reference sediment elutriate was not acutely toxic to *Mytilus* sp. (EC50 > 100 percent).

8.2.2 Whole Sediment Testing

8.2.2.1 Neanthes arenaceodentata

The results of the estuarine polychaete toxicity tests are summarized in Tables 8-15A (approach channels and Inside Site 104) and 8-15B (Outside and Inside Site 104).

Approach Channels and Inside Placement Site 104 (Round 1)

The results of the *N. arenaceodentata* toxicity testing conducted on the approach channel and Inside Site 104 sediments are summarized in Table 8-15A. After ten days of exposure, the lowest survival (92 percent) was recorded in the sediment from Craighill Angle-West. Sediments from Craighill, Cutoff Angle, Tolchester-South, Tolchester Straightening, and C&D Approaches (cores) had 96 percent survival, while sediments from Craighill Entrance, Craighill Angle-East, Craighill Upper Range, Tolchester North, Brewerton Eastern Extension, Swan Point, and C&D Approaches (surficial) had 100 percent survival. Survival in all of the sediments was within 10 percent of the Inside State 104 sediment, which had 96 percent survival, indicating that the sediments are not toxic and are not statistically different from those at the proposed placement site. The laboratory control sediment for the *N. arenaceodentata* testing had 100 percent survival.



Outside Placement Site 104 (Round 2)

As summarized in Table 8-15B, the Outside Site 104 reference sediment had 80 percent survival compared to 88 percent survival in the Inside Site 104 reference and 96 percent survival in the control. When tested statistically, survival in the Outside Site 104 sediment was not significantly different from the Inside Site 104 survival or the control survival.

Ocean Reference (Round 3)

N. arenaceodentata was not tested for the Ocean Reference site.

8.2.2.2 Leptocheirus plumulosus

The results of the estuarine amphipod toxicity tests are summarized in Tables 8-16A (approach channels and Inside Site 104), 8-16B (Outside and Inside Site 104), and 8-16C (Ocean Reference).

Approach Channels and Inside Placement Site 104 (Round 1)

The *L. plumulosus* whole sediment toxicity test results are presented in Table 8-16A. Survival in the channel sediments ranged from 86 to 96 percent, which was within the allowable 20 percent difference from the Inside Site 104 reference survival of 93 percent. Based on the survival results, the whole sediments were not acutely toxic to *L. plumulosus*. The control sediment for the *L. plumulosus* toxicity testing had 96 percent survival.

Outside Placement Site 104 (Round 2)

The Outside Site 104 reference sediment had 96 percent survival indicating that this sample was not acutely toxic to *L. plumulosus* (Table 8-16B). Survival for Inside Site 104 and the control were 87 and 98 percent, respectively.

Ocean Reference (Round 3)

The *L. plumulosus* whole sediment toxicity test results are presented in Table 8-16C. The lower Chesapeake Bay control sediment had 94 percent survival to estuarine amphipods.

8.2.2.3 Mysidopsis bahia

Ocean Reference (Round 3 only)

The results of the opossum shrimp whole sediment toxicity testing for the Ocean Reference are summarized in Table 8-17. The Ocean Reference sediment had 94 percent survival, indicating that the sediment was not acutely toxic to opossum shrimp.



8.2.3 Reference Toxicant Tests

The results of the reference toxicant tests are summarized in Table 8-18A (Round 1 and Round 2) and Table 8-18B (Round 3). The LC50s from the *M. bahia*, *C. variegatus*, *Mytilus* sp., *M. beryllina*, and *L. plumulosus* (96-hours; Round 1 and Round 2) reference toxicant tests all fell within the established laboratory control chart limits.

For *N. arenaceodentata*, control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 6.4 mg/L Cd was similar to the LC50 of 5.7 mg/L Cd for a previous lot of *N. arenaceodentata* from September 1999.

For *L. plumulosus* (reference toxicant tests for Ocean Reference-Round 3), control chart limits have not yet been established (48-hour LC50), because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 5.9 mg/L Cd (48-hour) falls within the range of LC50's from previous *L. plumulosus* reference toxicant tests (2.2-9.0 mg/L Cd).

8.3 DISCUSSION AND TIER III TOXICITY EVALUATION

The Tier III toxicity evaluation requires an analysis of water column and benthic toxicity test data. The following sections discuss the results of the water column and whole-sediment bioassays and the potential for impacts to the aquatic environment.

8.3.1 Water Column Bioassays

According to the ITM (USEPA/USACE 1998), after considering water column test results and expected mixing at the placement site, one of the following conclusions is reached:

- 1) The 100% dredged material elutriate toxicity is not statistically higher than the dilution water (laboratory control). Therefore, the dredged material is not predicted to be acutely toxic to water column organisms. However, benthic impact must also be evaluated.
- 2) The concentration of dissolved plus suspended contaminants, after allowance for mixing, does not exceed 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material is not predicted to be acutely toxic to water column organisms. However, benthic impact must also be evaluated.
- 3) The concentration of dissolved plus suspended contaminants, after allowance for mixing, exceeds 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material may have the potential to be acutely toxic to water column organisms.

- DO THIS INCLUDE THE NEW ITOT?

The evaluation guidelines in the ITM assume that state regulatory agencies will issue or permit an allowable mixing zone for open water placement events. It is important to recognize that the evaluation protocols in the ITM are guidelines, not regulations. More specifically, 404 CFR Part 230.10 (c) states that dredged material placement may not result in unacceptable adverse impact; however, the guidance in the ITM *suggests* the 0.01 (1%) LC50/EC50 concentrations as a quantitative method for assessing whether unacceptable adverse impacts might occur in the water column.

The ITM explicitly states that regional modifications of the national guidelines may be required or may be appropriate based on project-specific requirements or circumstances (ITM, p.1-1). In a situation where a mixing zone is not issued, the evaluation of the water column impacts requires alternative methods to demonstrate whether an unacceptable adverse impact is expected during placement or as a result of placement. Therefore, the evaluation of water column impacts for open-water placement at Site 104 is based on an assessment of elutriate concentration and duration of exposure to aquatic organisms. Evaluation of dredged material placement at the NODS is based on the Ocean Testing Manual (USEPA/USACE 1991) guidance which specifies that the 0.01 (1%) of LC50/EC50 must occur within a 4-hour time period inside the boundaries of the placement site (NODS).

Results of the water column toxicity testing are summarized in Table 8-19A and 8-19B. LC50/EC50 and NOAEC/NOEC values are provided in Table 8-19A, and mixing factors that would be required to determine compliance for ocean placement are provided in Table 8-19B.

In the water column tests, survival was the endpoint for the opposum shrimp, sheepshead minnow, and inland silverside tests. The endpoint of the blue mussel test was normal hinge development. All water column tests were conducted with larval or juvenile tests organisms within the age range specified by the USEPA/USACE (1998) testing guidelines: opposum shrimp (1-5 days old); sheepshead minnow (1-14 days old); inland silverside (9-14 days old); and blue mussels (< 4 hours old). In water column tests, results for 100% test elutriates are statistically compared (single-point comparison) to results of the laboratory controls as per ITM evaluation protocols, not to the results for the placement site or reference area.

Results of the water column tests indicated that the blue mussel and inland silverside were the most sensitive water column species to the project elutriates. None of the tested 100% elutriates was acutely toxic to sheepshead minnow. Although the LC50 values were >100% elutriate for all of the opposum shrimp tests, mean survival in 5 of the 13 elutriates was statistically lower than the mean laboratory control survival.

Normal development in the blue mussel is defined as transformation to the fully shelled, straight hinged, D-Shaped prodissoconch I stage. In the blue mussel tests, the 48-hr EC50 (median effect concentration) for the channel sediments ranged from 21.2 to >100% elutriate. Eleven of the 13 reaches had EC50 values for 100% elutriate that were statistically lower than the laboratory controls, indicating that eleven of the 100% elutriates affected normal development in the larval (<4 hrs old) organisms. Craighill Channel and Craighill Upper Range were the only channels where development was not statistically lower than the laboratory control. The lowest EC50 values were reported for the C&D Approach Channel-cores (21.2% elutriate) and the Craighill Entrance (22% elutriate).



Results of the inland silverside bioassays indicated that 11 of the 13 channel elutriates elicited some level of acute toxicity to juvenile (9-14 day old) inland silversides when exposed to undiluted elutriate. Tolchester Straightening and Swan Point were the only channel elutriates (100%) that were not acutely toxic to the juvenile inland silverside (LC50 >100% and not significantly different than the laboratory control). LC50 values for inland silverside ranged from 23.8% elutriate to >100% elutriate. The lowest LC50 value was reported for Tolchester-South (23.8% elutriate).

Calculation of NOAECs for oppossum shrimp, sheepshead minnow, and inland silverside and NOECs for blue mussel water column tests indicated that 11 of the 13 test reaches had NOAECs or NOECs of $\geq 10\%$ elutriate for all of the test species data. Brewerton Eastern Extension (inland silverside test) and C&D Approaches-cores (blue mussel test) yielded an NOAEC of <10% and an NOEC of <10%, respectively. These results indicate that no effect to survival or larval development would be expected to occur at elutriate concentrations below 10% based on a continuous 96-hour exposure period (oppossum shrimp, sheepshead minnow, and inland silverside) or based on a 48-hour continuous exposure period (blue mussel), with the exception of the two specified channel/species combinations with an NOAEC or NOEC of <10% elutriate. Based on the results of the STFATE modeling, an elutriate concentration of 10% would be expected at the placement site within approximately one-hour after placement occurs. The NOAECs indicate that no effect is expected for the majority of test organisms/channels at a concentration of 10% over either a 96-hour (oppossum shrimp, sheepshead minnow, and inland silverside) or a 48-hour (blue mussel) exposure period. For each placement event at Site 104, the duration of organism exposure to elutriate constituents in the water column would be expected to be short (acute), not a long-term continuous chronic exposure. Therefore, assuming that material would not be placed successively or consecutively at the same location within Site 104, no unacceptable adverse acute impact to water column organisms would be expected for a single placement event.

Evaluation of the elutriate data using methodology for whole effluent toxicity (USEPA 1991) yields similar results. The most restrictive acute toxicity test value presented in Table 8-19A is a 48-hour EC50 value for the blue mussel (21.2 percent elutriate). Per USEPA guidance, converting this value to acute toxic units (TUa) yields a value of 4.7 TUa (100/21.2). USEPA (1991) guidance in the Technical Support Document for Water Quality-Based Toxics Control requires compliance with the Agency's 0.3 TUa acute toxicity criterion at the edge of an acute mixing zone. Further, the 0.3 TUa criterion is generally interpreted as a 1-hour average concentration (p. 35), which "is expected to be fully protective for the fast-acting toxicants [e.g., chlorine, ammonia], and even more protective for slower-acting toxicants" (p. 35). Based on this guidance, the most restrictive of the elutriate test results would require 15.7 fold dilution [4.7/0.3] within one-hour to comply with USEPA's Technical Support Document guidance. STFATE modeling results presented in Chapter 7 indicates that a 16:1 dilution factor would be achieved within one hour under conservative modeling conditions. Using this approach, the other acute toxicity results presented in Table 8-19 would also not be expected to result in acute toxicity in the water column. Furthermore, this evaluation approach is believed to be , conservative because it compares 48- and 96-hour continuous exposure acute toxicity test results to a one-hour exposure duration criterion, and "fast-acting toxicants" are not expected to be present in the elutriate samples in meaningful concentrations (e.g., free chlorine, ammonia).



Ocean Placement

Based on the results of the water column toxicity testing, a maximum mixing factor of 472-fold would be required for all reaches to comply with the 0.01 LC50/EC50 requirement at the edge of the allowable mixing zone for ocean placement (Table 8-19B). This value is based on the lowest EC50 of 21.2% (C&D approach –core elutriate for blue mussel) in combination with a very conservative acute to chronic conversion factor of 0.01. Modeling of conditions at the ocean placement site indicated that a 1000-fold dilution would occur within the disposal site boundary during the allowable 4-hour ocean placement mixing period (see Chapter 7). Therefore, none of the channel elutriates is expected to be acutely toxic to aquatic organisms during ocean placement.

8.3.2 Whole-Sediment Bioassays

According to Tier III of the ITM (USEPA/USACE 1998), benthic toxicity testing of contaminants in the dredged material in Tier III will result in one of the following possible conclusions:

- Mean test organism mortality in the dredged material is not statistically greater than in the reference sediment, or does not exceed mean mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). Therefore, the dredged material is predicted not to be acutely toxic to benthic organisms. However, bioaccumulation of contaminants must also be considered.
- 2) Mean test organism mortality in the dredged material is statistically greater than in the reference sediment *and* exceeds mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). In this case, the dredged material has the potential to be acutely toxic to benthic organisms.

Results of the whole-sediment bioassays are summarized in Table 8-20. Results of the first round of the whole-sediment bioassays indicated that none of the mean survival values in the channel sediments was statistically lower than survival in Inside Site 104 sediments. None of the channel sediments was acutely toxic to either the estuarine polychaete (*Neanthes arenaceodentata*) or estuarine amphipod (*Leptocheirus plumulosus*). Therefore, the dredged material is not predicted to be acutely toxic to benthic organisms after placement occurs.

The whole sediment test results demonstrate that the sediment proposed for dredging is not predicted to be toxic to benthic organisms post-placement. The evaluation of benthic-effects for whole sediment bioassays is based on the Limiting Permissible Concentration (LPC). The LPC is defined as "...that concentration which will not cause unreasonable acute or chronic toxicity or sublethal adverse effects based on bioassay results using...appropriate sensitive marine organisms..." (USEPA/USACE 1991 and USEPA/USACE 1998). Based on the results of the whole sediment bioassays, the proposed dredged material from the channels is not significantly toxic to the tested benthic organisms. The statistical comparisons of the channel sediments to Inside Site 104 indicate that all of the channel sediments comply with Conclusion 1, above.

Because the whole-sediment test results indicate that the channel sediments are not acutely toxic to aquatic tests organisms, the channel sediments are expected to be suitable for ocean placement

8.3.3 Conclusions

In summary, conditions that would be expected to produce adverse effects in the water column at a placement site exist for a short-duration (minutes to a few hours following placement). Laboratory elutriate tests represent continuous exposure periods (48-96 hours) that greatly exceed the exposure periods that would be expected in the field. The laboratory elutriate tests provide conservative estimates of the potential for adverse water column effects. The conservatism is compounded by multiplying the LC50 concentration by a factor 0.01 for use in the mixing model. Overall, results of the whole-sediment toxicity tests are considered to be much more significant measures of the potential for adverse effects as a result of dredged material placement. Post-placement, benthic organisms and communities will be exposed to dredged material for weeks, months, or years, in comparison to the minutes or few hours of exposure experienced by organisms in the water column during the placement event. No toxicity was observed in the whole-sediment tests for the channel sediments, indicating that there would be little potential for long-term, adverse effects following open-water placement. The water column toxicity that could occur during the placement event will be short-term and localized. In addition, placement would occur during a time period when larval organisms (such as those testes in the laboratory tests) would not be expected to occur in the water column.

TABLE 8-1 SUMMARY OF TOXICITY TESTING SCHEDULE

| | | | WATER COLUMN TESTING | | | WHOLE S | EDIMENT TEST | ГING | |
|------------------|--------------------------|----------------------|----------------------|--------------------------|-----------------------|----------------------|-----------------------------|----------------------------|------------------------------------|
| | | | opossum shrimp | sheepshead minnow | blue mussel | inland silverside | estuarine polychaete | estuarine amphipod | opossum shrimp |
| TEST ROUND | DATES | TEST SEDIMENT | Mysidopsis bahia | Cyprinodon variegatus | <i>Mytilus</i> sp. | Menidia beryllina | Neanthes arenaceodentata | Leptocheirus plumulosus | Mysidopsis bahia ^(a) |
| 1 | September – | Inside Site 104 | х | Х | | | Х | Х | |
| November 1999 | November 1999 | Approach Channels | Х | Х | | | Х | Х | |
| | December 1999– | Inside Site 104 | х | х | х | х | Х | Х | |
| .2 | February 2000 | Outside Site 104 | х | x | х | X | х | Х | |
| | | Approach Channels | | | х | Х | | | |
| 3 | February – March 2000 | Ocean Reference | Х | | х | х | | X ^(b) | X |

(a) Mysidopsis bahia tested in conjunction with Woodrow Wilson Bridge testing, Neanthes arenaceodentata not tested for Ocean Reference. (b) Lower Chesapeake Bay control sediment substituted for Ocean Reference Site sediment (as requested by USEPA Region III Philadelphia) in Leptocheirus tests to minimize potential grain-size effects

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TABLE 8-2 SAMPLE COMPOSITES FOR ECOTOXICOLOGICAL TESTING

| Sampling Reach | Station | Sediment Volume | Composite |
|----------------------------|------------------|-----------------|---|
| | | (gallons) | Sample |
| Inside Placement Site 104 | KI-3 | 4 | KI-TOX |
| | KI-5 | 4 | |
| | KI-7-Ref | 4 | |
| | KI-S-1 | 4 | |
| | KI-S-2 | 4 | |
| Outside Placement Site 104 | KI-11 | 5 | KI-OUT-TOX |
| | KI-14 | 5 | |
| | KI-15 | 5 | |
| | KI-16 | 5 | |
| Brewerton Eastern | BE-1 | 5 | BE-TOX |
| Extension | BE-2 | 5 | |
| | BE-3 | 5 | |
| | BE-4 | 5 | |
| C&D Approach Channel | CD-001VC | 10 | CD-TOX |
| | CD-002VC | 10 | |
| | CD003 | 5 | |
| | CD004 | 5 | |
| | CD005 | 5 | |
| | CD006 | 3 | |
| Craighill | CR1 | 7 | CR-TOX |
| | CR2 | 7 | |
| | CR3 | 7 | |
| Craighill Angle–East | CRA-E-001VC | 7 | CRA-E-TOX |
| | CRA-E-002VC | 7 | |
| | CRA-E-003VC | / | |
| Craighill Angle–West | CRA-W-001VC | 7 | CRA-W-TOX |
| | CRA-W-002VC | 7 | |
| | CRA-W-003VC | | ODE TOY |
| Craighill Entrance | CRE-001VC | 5 | CRE-TUX |
| | <u>CRE-002VC</u> | 5 | |
| | <u>CRE-003VC</u> | 5 | |
| Cutoff Angle | | 7 | CUTTOY |
| Cuton Angle | | | COPION |
| | CUT3 | 7 | |
| Swan Point | SWP-001VC | 3.5 | SWP-TOX |
| | SWP-002VC | 3.5 | 5 |
| | SWP-003VC | 3.5 | |
| | SWP-004VC | 3.5 | |
| | SWP-005VC | 3.5 | |
| | SWP-006VC | 3.5 | |
| Tolchester Channel – | TLC-005VC | 3.5 | TLC-N-TOX |
| North | TLC-006VC | 3.5 | |
| | TLC-007VC | 3.5 | |
| | TLC-008VC | 3.5 | |
| | TLC-009VC | 3.5 | |
| | TLC-010VC | 3.5 | 74 - 49,445,9410,020 - 41 - 41 - 41 - 41 - 41 - 41 - 41 - 4 |
| Tolchester Channel – South | TLC-001VC | 5 | TLC-S-TOX |
| | TLC-002VC | 5 | ι. |
| | TLC-003VC | 5 | |
| | TLC-004VC | 5 | |
| Tolchester Straightening | TLS-001VC | 10 | TLS-TOX |
| | TLS-002VC | 10 | |

TABLE 8-3ASUMMARY OF COLLECTION AND RECEIPT INFORMATION FOR SAMPLES FROM BALTIMORE HARBORAPPROACH CHANNELS AND PLACEMENT SITE 104

| Sample <u>Identification</u> | Sample <u>Description</u> | Sample <u>Type</u> | EA Accession <u>Number</u> | Collection <u>Time and Date</u> | Receipt <u>Time and Date</u> |
|---------------------------------|----------------------------------|-----------------------|-------------------------------|------------------------------------|---------------------------------|
| CRE | Craighill Entrance | Sediment | AT9-1430 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CR | Craighill | Sediment | AT9-1431 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRA-E | Craighill Angle East | Sediment | AT9-1432 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRA-W | Craighill Angle West | Sediment | AT9-1433 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRU | Craighill Upper Range | Sediment | AT9-1434 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CUT | Cutoff Angle | Sediment | AT9-1435 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLC-S | Tolchester Channel South | Sediment | AT9-1436 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLC-N | Tolchester Channel North | Sediment | AT9-1437 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLS | Tolchester Straightening | Sediment | AT9-1438 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| BE | Brewerton Eastern Extension | Sediment | AT9-1439 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| SWP | Swan Point | Sediment | AT9-1440 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CD-VC | C&D Approaches - Cores | Sediment | AT9-1441 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CD | C&D Approaches - Surficial Grabs | Sediment | AT9-1442 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| K1 - Reference | Inside Site 104 | Sediment | AT9-1443 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1444 | 1052, 19 September 1999 | 1500, 13 October 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1611 | 1118, 22 November 1999 | 1515, 22 November 1999 |
| CR | Craighill | Site Water | AT9-1445 | 1205, 28 September 1999 | 1500, 13 October 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1446 | 1315, 18 September 1999 | 1500, 13 October 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1609 | 1029, 22 November 1999 | 1515, 22 November 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1447 | 1645, 18 September 1999 | 1500, 13 October 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1610 | 1041, 22 November 1999 | 1515, 22 November 1999 |
| CRU | Craighill Upper Range | Site Water | AT9-1448 | 1020, 28 September 1999 | 1500, 13 October 1999 |

TABLE 8-3A (CONTINUED)

| Sample | Sample | Sample | EA Accession | Collection | Receipt |
|----------------|------------------------------------|------------|--------------|-------------------------|------------------------|
| Identification | Description | Type | Number | <u>Time and Date</u> | Time and Date |
| CUT | Cutoff Angle | Site Water | AT9-1449 | 0835, 28 September 1999 | 1500, 13 October 1999 |
| CUT | Cutoff Angle | Site Water | AT9-1608 | 0958, 22 November 1999 | 1515, 22 November 1999 |
| TLC-S | Tolchester Channel South | Site Water | AT9-1450 | 0850, 20 September 1999 | 1500, 13 October 1999 |
| TLC-N | Tolchester Channel North | Site Water | AT9-1451 | 1130, 20 September 1999 | 1500, 13 October 1999 |
| TLS | Tolchester Straightening | Site Water | AT9-1452 | 1520, 27 September 1999 | 1500, 13 October 1999 |
| BE | Brewerton Eastern Extension | Site Water | AT9-1453 | 1645, 27 September 1999 | 1500, 13 October 1999 |
| SWP | Swan Point | Site Water | AT9-1454 | 1530, 19 September 1999 | 1500, 13 October 1999 |
| CD-VC | C&D Approaches - Cores | Site Water | AT9-1455 | 1230, 21 September 1999 | 1500, 13 October 1999 |
| CD-SW | C&D Approaches - Surficial Grabs | Site Water | AT9-1456 | 1420, 27 September 1999 | 1500, 13 October 1999 |
| KI – Reference | Inside Site 104 | Site Water | AT9-1457 | 1540, 28 September 1999 | 1500, 13 October 1999 |
| KI – Reference | Inside Site 104 | Site Water | AT9-1612 | 1147, 22 November 1999 | 1515, 22 November 1999 |
| CRE | Craighill Entrance | Sediment | AT9-1702 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1703 | 1300, 15 December 1999 | 0830, 20 December 1999 |
| CR | Craighill | Sediment | AT9-1704 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CR | Craighill | Site Water | AT9-1705 | 1600, 13 December 1999 | 0830, 20 December 1999 |
| CRA-E | Craighill Angle East | Sediment | AT9-1706 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1707 | 0930, 15 December 1999 | 0830, 20 December 1999 |
| CRA-W | Craighill Angle West | Sediment | AT9-1708 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1709 | 0950, 14 December 1999 | 0830, 20 December 1999 |
| CRU | Craighill Upper Range | Sediment | AT9-1710 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CRU | Craighill Upper Range | Site Water | AT9-1711 | 0915, 15 December 1999 | 0830, 20 December 1999 |
| CUT | Cutoff Angle | Sediment | AT9-1712 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CUT | Cutoff Angle | Site Water | AT9-1713 | 1035, 15 December 1999 | 0830, 20 December 1999 |

 $\overline{(a)}$ Time of collection not provided by sampler.

TABLE 8-3A (CONTINUED)

| Sample | Sample Description | Sample Type | EA Accession Number | Collection Time and Date | Receipt Time and Date |
|--------------------|----------------------------------|----------------|------------------------|-----------------------------|--------------------------|
| TLC-S | Tolchester Channel South | Sediment | AT9-1714 | 1310, 17 December 1999 | 0830, 20 December 1999 |
| TLC-S | Tolchester Channel South | Site Water | AT9-1715 | 0950, 9 December 1999 | 0830, 20 December 1999 |
| TLC-N | Tolchester Channel North | Sediment | AT9-1716 | 1040, 17 December 1999 | 0830, 20 December 1999 |
| TLC-N | Tolchester Channel North | Site Water | AT9-1717 | 1315, 8 December 1999 | 0830, 20 December 1999 |
| TLS | Tolchester Straightening | Sediment | AT9-1718 | 1310, 17 December 1999 | 0830, 20 December 1999 |
| TLS | Tolchester Straightening | Site Water | AT9-1719 | 0830, 9 December 1999 | 0830, 20 December 1999 |
| BE | Brewerton Eastern Extension | Sediment | AT9-1720 | (a), 17 December 1999 | 0830, 20 December 1999 |
| BE | Brewerton Eastern Extension | Site Water | AT9-1721 | 1420, 14 December 1999 | 0830, 20 December 1999 |
| SWP | Swan Point | Sediment | AT9-1722 | 1420, 17 December 1999 | 0830, 20 December 1999 |
| SWP | Swan Point | Site Water | AT9-1723 | 1145, 9 December 1999 | 0830, 20 December 1999 |
| CD-VC | C&D Approaches - Cores | Sediment | AT9-1724 | 1040, 17 December 1999 | 0830, 20 December 1999 |
| CD-VC | C&D Approaches - Cores | Site Water | AT9-1725 | 0845, 8 December 1999 | 0830, 20 December 1999 |
| CD | C&D Approaches - Surficial Grabs | Sediment | AT9-1726 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CD | C&D Approaches - Surficial Grabs | Site Water | AT9-1727 | 1210, 14 December 1999 | 0830, 20 December 1999 |
| KI – Reference | Inside Site 104 | Sediment | AT9-1728 | (a), 17 December 1999 | 0830, 20 December 1999 |
| KI – Reference | Inside Site 104 | Site Water | AT9-1729 | 1430, 13 December 1999 | 0830, 20 December 1999 |
| KI-OUT - Reference | Outside Site 104 | Sediment | AT9-1730 | (a), 17 December 1999 | 0830, 20 December 1999 |
| KI-OUT - Reference | Outside Site 104 | Site Water | AT9-1731 | 1235, 13 December 1999 | 0830, 20 December 1999 |

1.1.1.1.1.1

(a) Time of collection not provided by sampler.

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TABLE 8-3BSUMMARY OF COLLECTION AND RECEIPT INFORMATION FOR SAMPLESFROM THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| Sample <u>Identification</u> | Sample <u>Type</u> | EA Accession <u>Number</u> | Collection <u>Time and Date</u> | Receipt <u>Time and Date</u> |
|---------------------------------|-----------------------|-------------------------------|------------------------------------|---------------------------------|
| Lower Chesapeake Bay Control | Sediment | AT0-050 | 1200, 12 January 2000 | 0910, 19 January 2000 |
| Ocean Reference | Sediment | AT0-097 | 1130, 1 February 2000 | 0903, 3 February 2000 |
| Ocean Reference | Water | AT0-098 | 1130, 1 February 2000 | 0903, 3 February 2000 |

TABLE 8-4AWATER QUALITY PARAMETERS MEASURED DURING Mysidopsis bahia
(OPOSSUM SHRIMP) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | |
|--------------------------------|----------------------------|------------|------------------|-------------|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | |
| | (°C) | pН | (mg/L) | (ppt) | |
| Inside Site 104 | 19.8 (±0.6) | 8.2 (±0.1) | 6.9 (±0.5) | 29.0 (±0.7) | |
| Brewerton Eastern Extension | 20.1 (±0.4) | 8.2 (±0.1) | 6.8 (±0.4) | 29.0 (±0.6) | |
| C&D Approaches - Cores | 20.4 (±0.5) | 8.1 (±0.2) | 6.7 (±0.5) | 29.6 (±1.2) | |
| C&D Appr Surf. Grabs | 20.6 (±0.7) | 8.2 (±0.1) | 7.0 (±0.4) | 29.1 (±0.7) | |
| Craighill | 20.0 (±0.6) | 8.1 (±0.1) | 6.6 (±0.4) | 29.2 (±0.6) | |
| Craighill Angle East | 20.2 (±0.6) | 8.1 (±0.2) | 6.9 (±0.3) | 29.3 (±0.7) | |
| Craighill Angle West | 19.9 (±0.3) | 8.2 (±0.1) | 6.7 (±0.4) | 29.4 (±0.7) | |
| Craighill Entrance | 20.2 (±0.4) | 8.1 (±0.1) | 6.7 (±0.6) | 29.1 (±0.8) | |
| Craighill Upper Range | 19.6 (±0.4) | 8.1 (±0.1) | 6.8 (±0.4) | 29.3 (±0.7) | |
| Cutoff Angle | 20.4 (±0.6) | 8.1 (±0.1) | 6.8 (±0.4) | 29.2 (±0.5) | |
| Swan Point | 20.0 (±0.5) | 8.2 (±0.1) | 6.9 (±0.8) | 29.4 (±0.5) | |
| Tolchester North | 19.4 (±0.5) | 8.1 (±0.2) | 6.9 (±0.5) | 28.9 (±0.3) | |
| Tolchester South | 20.0 (±0.6) | 8.1 (±0.1) | 6.9 (±0.4) | 29.2 (±0.8) | |
| Tolchester Straightening | 20.0 (±0.7) | 8.1 (±0.1) | 6.8 (±0.4) | 29.0 (±0.6) | |



TABLE 8-4BWATER QUALITY PARAMETERS MEASURED DURING Mysidopsis bahia
(OPOSSUM SHRIMP) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT
SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|---------------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity <u>(</u> ppt) | | |
| Inside Site 104 | 20.2 (±0.6) | 8.3 (±0.1) | 7.0 (±0.3) | 30.5 (±1.7) | | |
| Outside Site 104 | 19.8 (±1.0) | 8.3 (±0.1) | 7.0 (±0.3) | 30.8 (±1.6) | | |
| Laboratory Control Sediment | 19.5 (±1.0) | 8.3 (±0.1) | 7.1 (±0.4) | 30.4 (±1.8) | | |

TABLE 8-4C SUMMARY OF WATER QUALITY PARAMETERS MEASUREDDURING Mysidopsis bahia (OPOSSUM SHRIMP) ELUTRIATETOXICITY TESTING ON SAMPLES FROM THE NORFOLK OCEANDISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000)

| | Water Quality Parameters – Mean (±Standard Deviation) | | | | | |
|--------------------|---|------------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Ocean Reference | 19.3 (±0.5) | 8.0 (±0.2) | 6.7 (±0.9) | 31.6 (±1.6) | | |

TABLE 8-5AWATER QUALITY PARAMETERS MEASURED DURING Cyprinodon
variegatus (SHEEPSHEAD MINNOW) ELUTRIATE TOXICITY TESTING ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| Test Elization | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| 1 est Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Inside Site 104 | 20.8 (±0.8) | 8.2 (±0.1) | 7.1 (±0.3) | 29.0 (±0.5) | | |
| Brewerton Eastern Extension | 20.2 (±0.7) | 8.2 (±0.1) | 6.9 (±0.4) | 28.9 (±0.5) | | |
| C&D Approaches - Cores | 20.5 (±0.5) | 8.2 (±0.2) | 7.0 (±0.2) | 28.9 (±0.5) | | |
| C&D Appr Surf. Grabs | 20.4 (±0.4) | 8.1 (±0.1) | 7.0 (±0.3) | 28.8 (±0.4) | | |
| Craighill | 20.6 (±0.6) | 8.1 (±0.1) | 6.9 (±0.4) | 29.3 (±0.8) | | |
| Craighill Angle East | 20.2 (±0.7) | 8.2 (±0.1) | 7.1 (±0.3) | 28.9 (±0.3) | | |
| Craighill Angle West | 20.1 (±0.6) | 8.2 (±0.1) | 7.2 (±0.2) | 29.0 (±0.3) | | |
| Craighill Entrance | 20.0 (±0.5) | 8.2 (±0.1) | 7.0 (±0.2) | 28.9 (±0.6) | | |
| Craighill Upper Range | 20.2 (±0.5) | 8.1 (±0.1) | 7.0 (±0.2) | 29.0 (±0.4) | | |
| Cutoff Angle | 20.5 (±0.5) | 8.2 (±0.1) | 7.0 (±0.2) | 29.0 (±0.6) | | |
| Swan Point | 20.3 (±0.3) | 8.2 (±0.1) | 7.1 (±0.2) | 29.6 (±0.7) | | |
| Tolchester North | 20.7 (±0.5) | 8.2 (±0.2) | 7.0 (±0.2) | 29.2 (±0.7) | | |
| Tolchester South | 20.3 (±0.6) | 8.1 (±0.1) | 7.1 (±0.3) | 29.4 (±0.9) | | |
| Tolchester Straightening | 20.8 (±0.5) | 8.1 (±0.2) | 6.9 (±0.4) | 29.1 (±0.4) | | |

TABLE 8-5BWATER QUALITY PARAMETERS MEASURED DURING Cyprinodon
variegatus (SHEEPSHEAD MINNOW) ELUTRIATE TOXICITY TESTING ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Inside Site 104 | 19.6 (±1.0) | 8.3 (±0.1) | 7.2 (±0.3) | 30.7 (±1.9) | | |
| Outside Site 104 | 20.0 (±0.9) | 8.3 (±0.1) | 7.1 (±0.2) | 30.8 (±1.6) | | |
| Laboratory Control Sediment | 19.7 (±1.1) | 8.3 (±0.1) | 7.2 (±0.3) | 30.4 (±1.9) | | |

TABLE 8-6A WATER QUALITY PARAMETERS MEASURED DURING *Menidia beryllina* (INLAND SILVERSIDE) ELUTRIATE TOXICITY TESTING ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (±Standard Deviation) | | | | |
|--------------------------------|----------------------------|------------|------------------|-------------|--|
| Test Elutriate | Temperature | TT | Dissolved Oxygen | Salinity | |
| | (°C) | рн | (mg/L) | (ppt) | |
| Inside Site 104 | 20.2 (±0.8) | 8.2 (±0.1) | 5.7 (±1.3) | 33.2 (±2.9) | |
| Outside Site 104 | 20.6 (±0.7) | 8.3 (±0.1) | 5.7 (±1.1) | 32.3 (±2.7) | |
| Brewerton Eastern Extension | 20.4 (±1.1) | 8.3 (±0.1) | 6.6 (±0.6) | 33.8 (±3.7) | |
| C&D Approaches - Cores | 20.9 (±0.8) | 8.1 (±0.1) | 6.3 (±1.0) | 29.8 (±1.1) | |
| C&D Appr Surf. Grabs | 20.4 (±1.1) | 8.3 (±0.1) | 6.7 (±0.6) | 32.1 (±3.2) | |
| Craighill | 20.2 (±0.5) | 8.2 (±0.1) | 6.6 (±0.5) | 32.1 (±1.6) | |
| Craighill Angle East | 20.6 (±0.8) | 8.2 (±0.1) | 6.5 (±0.9) | 29.8 (±1.0) | |
| Craighill Angle West | 20.7 (±1.1) | 8.2 (±0.2) | 6.5 (±0.7) | 30.7 (±1.9) | |
| Craighill Entrance | 19.9 (±1.7) | 8.3 (±0.2) | 5.9 (±1.4) | 33.0 (±3.0) | |
| Craighill Upper Range | 20.5 (±0.6) | 8.3 (±0.1) | 6.6 (±0.6) | 33.1 (±3.2) | |
| Cutoff Angle | 19.0 (±0.7) | 8.3 (±0.1) | 6.7 (±0.6) | 29.3 (±1.1) | |
| Swan Point | 20.9 (±0.9) | 8.2 (±0.1) | 5.8 (±1.3) | 35.1 (±5.6) | |
| Tolchester North | 19.3 (±0.6) | 8.1 (±0.1) | 6.6 (±0.7) | 29.3 (±0.9) | |
| Tolchester South | 21.1 (±0.8) | 8.2 (±0.1) | 6.7 (±0.5) | 29.5 (±1.7) | |
| Tolchester Straightening | 20.3 (±0.7) | 8.0 (±0.1) | 6.3 (±1.2) | 29.8 (±0.4) | |

TABLE 8-6BSUMMARY OF WATER QUALITY PARAMETERS MEASURED
DURING Menidia beryllina (INLAND SILVERSIDE) ELUTRIATE
TOXICITY TESTING ON SAMPLES FROM THE NORFOLK OCEAN
DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000)

| · · · · · · · · · · · · · · · · · · · | Water Quality Parameters – Mean (±Standard Deviation) | | | | | | |
|---------------------------------------|---|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | pН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Ocean Reference | 19.8 (±0.3) | 8.1 (±0.1) | 6.7 (±0.9) | 31.7 (±1.5) | | | |

TABLE 8-7AWATER QUALITY PARAMETERS MEASURED DURING Mytilus sp.
(BLUE MUSSEL) ELUTRIATE TOXICITY TESTING ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE
104 (Round 2: Dec. 1999-Feb. 2000) ^(a)

| | Mean (±Standard Deviation) | | | | |
|--------------------------|----------------------------|-----------------|------------------|---------------|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | |
| | (°C) | pH | (mg/L) | (ppt) | |
| | | | | | |
| Inside Site 104 | 16.0 (±0.7) | 8.4 (±0.1) | 8.2 (±0.2) | 30.2 (±0.3) | |
| Outside Site 104 | 160(.04) | 84(101) | 82(02) | 30.2 (, 0.2) | |
| | 10.0 (±0.4) | 8.4 (±0.1) | 8.2 (±0.2) | JU.2 (±0.2) | |
| Brewerton Eastern | 16.0 (±0.7) | 8.4 (±0.2) | 8.2 (±0.2) | 29.8 (±0.7) | |
| Extension | | | | | |
| | 160600 | | | 2(2(47) | |
| C&D Approacnes - Cores | 16.3 (±0.2) | 8.3 (±0.2) | 8.2 (±0.2) | 20.2 (±4.7) | |
| C&D Appr - Surf Grabs | 15.9 (+0.7) | 84(+02) | 82(+02) | 30.2 (+0.3) | |
| CGD Appl Sull. Glabs | 13.5 (±0.7) | 0.4 (±0.2) | 0.2 (10.2) | 50.2 (10.5) | |
| | 162606 | | 82(02) | 201(02) | |
| | 16.2 (±0.6) | 8.1 (±0.1) | 8.2 (±0.2) | 30.1 (±0.2) | |
| Craighill Angle East | 16.2 (+0.7) | 8.2 (+0.1) | 8.1 (+0.2) | 30.0 (±0.2) | |
| 88 | | | | | |
| Craighill Angle West | 15.7 (±0.6) | 8.2 (±0.2) | 8.2 (±0.2) | 30.0 (±0.2) | |
| | | | | | |
| Craighill Entrance | 16.3 (±0.5) | 8.2 (±0.2) | 8.1 (±0.1) | 30.0 (±0.2) | |
| Creighill Upper Dange | 161(.07) | 82 (.01) | 81(.01) | 30.0 (+0.4) | |
| Craiginii Opper Kange | 10.1 (±0.7) | 8.2 (±0.1) | 8.1 (±0.1) | 50.0 (±0.4) | |
| Cutoff Angle | 16.1 (±0.7) | 8.3 (±0.2) | 8.2 (±0.2) | 30.2 (±0.3) | |
| | | . , | | | |
| Swan Point | 16.5 (±0.1) | 8.2 (±0.2) | 8.2 (±0.2) | 29.0 (±2.0) | |
| | | | | | |
| Tolchester North | 16.7 (±0.2) | 8.2 (±0.2) | 8.1 (±0.2) | 29.0 (±1.5) | |
| Tolchester South | 16.0 (+0.8) | 82(+01) | 8 2 (+0 2) | 30.0 (+0.2) | |
| | 10.0 (10.0) | 0.2 (10.1) | 0.2 (10.2) | JU.U (10.2) | |
| Tolchester Straightening | 16.0 (±0.8) | 8.3 (±0.1) | 8.2 (±0.1) | 30.2 (±0.3) | |
| | | - | | • | |

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-7BWATER QUALITY PARAMETERS MEASURED DURING Mytilus sp.
(BLUE MUSSEL) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000) ^(a)

| | Mean (±Standard Deviation) | | | | | |
|--------------------|----------------------------|-----------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Ocean Reference | 15.8(±0.2) | 8.0(±0.1) | 8.1 (±0.2) | 30.9 (±0.3) | | |

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-8AWATER QUALITY PARAMETERS MEASURED DURING 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Neanthes arenaceodentata
(ESTUARINE POLYCHAETE) ON SAMPLES FROM BALTIMORE HARBOR
APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | | |
|---------------------------------------|----------------------------|--------------------|------------------|-------------|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | |
| <u></u> | (°C) | pH | (mg/L) | (ppt) | | |
| Laboratory Control | | | | | | |
| Sediment | 19.2 (±0.4) | 7.8 (±0.3) | 6.2 (±0.6) | 30.9 (±1.0) | | |
| | | | | | | |
| Inside Site 104 | 192(+0.6) | 79(+02) | 56(+04) | 310(+07) | | |
| | 1912 (1010) | ,, (±0. .) | | 0110 (2017) | | |
| | | | | | | |
| Brewerton Eastern | 19.1 (±0.4) | 7.9 (±0.2) | 5.9 (±0.7) | 31.0 (±0.8) | | |
| | | | | | | |
| C&D Approaches - Cores | | | | | | |
| | 19.0 (±0.5) | 7.9 (±0.2) | 6.2 (±0.5) | 30.8 (±0.8) | | |
| C&D Appr - Surf Crahe | | | | | | |
| Call Appl Sull. Glabs | 19.0 (±0.6) | 7.7 (±0.4) | 6.1 (±0.7) | 31.1 (±0.8) | | |
| · · · · · · · · · · · · · · · · · · · | | | | | | |
| Craighill | 19.1 (±0.5) | 7.8 (±0.2) | 6.2 (±0.4) | 31.7 (±1.3) | | |
| | | | | | | |
| Craighill Angle East | 19.1 (±0.4) | 7.8 (±0.3) | 5.7 (±0.7) | 31.4 (±1.0) | | |
| | | | | | | |
| Craighill Angle West | 191(+04) | 78(+02) | 57(+0.6) | 313(+11) | | |
| | 19.1 (10.1) | 7.0 (10.2) | 5.7 (10.0) | 51.5 (11.1) | | |
| Craighill Entrance | 101(04) | 70(+03) | 57(.05) | 315(10) | | |
| | 19.1 (±0.4) | 7.9 (±0.3) | 5.7 (±0.5) | 51.5 (±1.0) | | |
| Craighill Upper Range | 101(.04) | 70(.02) | 61(.06) | 216(11) | | |
| | 19.1 (±0.4) | 7.9 (±0.2) | 0.1 (±0.0) | 51.0 (±1.1) | | |
| Cutoff Angle | | | | | | |
| | 19.1 (±0.4) | 7.8 (±0.2) | 5.7 (±0.6) | 31.3 (±1.0) | | |
| Swan Point | | | | | | |
| | 19.1 (±0.4) | 7.9 (±0.3) | 5.9 (±0.7) | 31.1 (±0.7) | | |
| Tolchester North | | | | | | |
| | 19.1 (±0.4) | 7.7 (±0.3) | 5.6 (±0.5) | 31.2 (±0.9) | | |
| Talahostor South | | i | | | | |
| | 19.1 (±0.5) | 7.8 (±0.3) | 6.0 (±0.8) | 31.2 (±1.0) | | |
| Talahan St! | | | | | | |
| 1 oicnester Straightening | 19.1 (±0.5) | 7.8 (±0.3) | 6.3 (±0.7) | 31.3 (±1.0) | | |

TABLE 8-8BWATER QUALITY PARAMETERS MEASURED DURING ADDITIONAL
10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Neanthes
arenaceodentata (ESTUARINE POLYCHAETE) ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT
SITE 104 (Round 2: Dec. 1999-Feb.2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Inside Site 104 | 20.4 (±0.6) | 8.3 (±0.1) | 7.2 (±0.3) | 29.2 (±0.4) | | |
| Outside Site 104 | 20.3 (±0.6) | 8.2 (±0.2) | 7.2 (±0.3) | 28.8 (±0.7) | | |
| Laboratory Control Sediment | 20.5 (±0.4) | 8.2 (±0.2) | 7.2 (±0.3) | 29.6 (±1.9) | | |

TABLE 8-9AWATER QUALITY PARAMETERS MEASURED DURING 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Leptocheirus plumulosus (ESTUARIN
AMPHIPOD) ON SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|------------------|-------------|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | |
| | (°C) | pH | (mg/L) | (ppt) | | |
| Laboratory Control Sediment | 19.5 (±0.5) | 7.6 (±0.2) | 6.2 (±0.7) | 20.9 (±0.6) | | |
| Inside Site 104 | 19.4 (±0.4) | 7.7 (±0.3) | 5.8 (±0.8) | 21.6 (±0.8) | | |
| Brewerton Eastern Extension | 19.4 (±0.4) | 7.7 (±0.2) | 6.0 (±0.8) | 21.3 (±0.7) | | |
| C&D Approaches - Cores | 19.3 (±0.4) | 7.6 (±0.2) | 5.9 (±0.7) | 20.9 (±0.6) | | |
| C&D Appr Surf. Grabs | 19.3 (±0.4) | 7.7 (±0.2) | 6.3 (±1.0) | 21.0 (±0.7) | | |
| Craighill | 19.6 (±0.5) | 7.7 (±0.2) | 6.0 (±0.6) | 21.3 (±0.7) | | |
| Craighill Angle East | 19.5 (±0.5) | 7.5 (±0.2) | 5.1 (±1.0) | 21.2 (±0.6) | | |
| Craighill Angle West | 19.5 (±0.5) | 7.6 (±0.2) | 5.0 (±0.9) | 21.2 (±0.6) | | |
| Craighill Entrance | 19.5 (±0.4) | 7.6 (±0.2) | 5.5 (±1.1) | 21.5 (±0.6) | | |
| Craighill Upper Range | 19.4 (±0.5) | 7.6 (±0.2) | 6.3 (±0.7) | 21.5 (±0.7) | | |
| Cutoff Angle | 19.4 (±0.5) | 7.6 (±0.2) | 5.7 (±0.9) | 21.3 (±0.8) | | |
| Swan Point | 19.3 (±0.3) | 7.6 (±0.2) | 5.8 (±0.9) | 21.3 (±0.7) | | |
| Tolchester North | 19.3 (±0.4) | 7.5 (±0.2) | 5.3 (±0.9) | 21.1 (±0.6) | | |
| Tolchester South | 19.4 (±0.3) | 7.6 (±0.2) | 5.7 (±0.6) | 21.3 (±0.7) | | |
| Tolchester Straightening | 19.4 (±0.4) | 7.7 (±0.2) | 6.2 (±0.7) | 21.1 (±0.5) | | |

TABLE 8-9BWATER QUALITY PARAMETERS MEASURED DURING ADDITIONAL
10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Leptocheirus
plumulosus (ESTUARINE AMPHIPOD) ON SAMPLES FROM BALTIMORE
HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| I est Elutriate | Temperature(°C)pH | | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Inside Site 104 | 21.8 (±3.2) | 8.0 (±0.1) | 5.5 (±1.0) | 22.6 (±1.8) | | |
| Outside Site 104 | 22.1 (±2.8) | 8.0 (±0.1) | 3.9 (±1.1) | 21.0 (±1.6) | | |
| Laboratory Control Sediment | 21.9 (±3.1) | 7.9 (±0.2) | 5.7 (±0.7) | 20.0 (±2.2) | | |

TABLE 8-9CSUMMARY OF WATER QUALITY PARAMETERS FROM 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Leptocheirus plumulosus (ESTUARINE
AMPHIPOD) FOR THE NORFOLK OCEAN DISPOSAL SITE REFERENCE
AREA (Round 3: Feb. 2000)

| | Mean (±Standard Deviation) | | | | | |
|--|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Lower Chesapeake Bay Control ^(a) | 20.3 (±0.4) | 8.1 (±0.1) | 5.2 (±0.6) | 21.5 (±1.2) | | |

(a) = Tested in place of Ocean Reference sediment due to potential for adverse grain size effects.

TABLE 8-10SUMMARY OF WATER QUALITY PARAMETERS FROM 10-DAY
WHOLE SEDIMENT TOXICITY TESTING WITH Mysidopsis bahia FOR
THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000)

| | Mean (±Standard Deviation) | | | | | |
|-----------------|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Elutriate | Temperature (°C) | pН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Ocean Reference | 20.5 (±0.6) | 8.0 (±0.2) | 6.2 (±0.5) | 31.5 (±1.3) | | |

TABLE 8-11A RESULTS OF ELUTRIATE TOXICITY TESTING WITH Mysidopsis bahia (OPOSSUM SHRIMP) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | 96-Hour Survival (%)* | | | | 96-Hour | 96-Hour |
|-----------------------------------|-----------------------|------|-------------|-------------------|-----------------------------------|------------------------|
| Test Elutriate | Lab | Pei | rcent Elutr | iate | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 100 | 98 | 98 | 100 | > 100 | 100 |
| Brewerton Eastern Extension | 98 | 92 | 96 | 78 ^(a) | > 100 | 50 |
| C&D Approaches – Cores | 94 | 98 | 96 | 86 | > 100 | 100 |
| C&D Approaches – Surface Grabs | 100 | 92 | 98 | 84 ^(a) | > 100 | 50 |
| Craighill | 98 | .94 | 98 | 94 | > 100 | 100 |
| Craighill Angle East | 96 | 92 | 92 | 92 | > 100 | 100 |
| Craighill Angle West | 98 | 98 | 98 | 76 ^(a) | > 100 | 50 |
| Craighill Entrance | 98 | 98 | 98 | 54 ^(a) | > 100 | 50 |
| Craighill Upper Range | 100 | 98 | 98 | 98 | > 100 | 100 |
| Cutoff Angle | 98 | 90 | 98 | 92 | > 100 | 100 |
| Swan Point | 96 | 94 | 90 | 84 | > 100 | 100 |
| Tolchester North | 98 | 96 | 98 | 94 | > 100 | 100 |
| Tolchester South | 100 | 98 | 92 | 83 ^(a) | > 100 | 100 ^(b) |
| Tolchester Straightening | 96 | 92 | 98 | 94 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

(b) Multiple comparison (100, 50, 10 vs. control) shows no significance; single comparison (50 vs. control) shows significance (NOAEC = 10).

TABLE 8-11BRESULTS OF ADDITIONAL ELUTRIATE TOXICITY
TESTING WITH Mysidopsis bahia (OPOSSUM SHRIMP) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| | 96-Hour Survival (%)* | | | | 96-Hour | 96-Hour |
|--------------------------------|-----------------------|------|-----------------|------|------------------|---------------|
| Test Elutriate | Lab Pe | | rcent Elutriate | | LC ₅₀ | NOAEC |
| | Control | 10 % | 50 % | 100% | (% elutriate) | (% elutriate) |
| Inside 104 | 98 | 100 | 96 | 100 | > 100 | 100 |
| Outside 104 | 96 | 98 | 94 | 94 | > 100 | 100 |
| Laboratory Control Sediment | 94 | 92 | 90 | 88 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

TABLE 8-11CRESULTS OF ELUTRIATE TOXICITY TESTING WITH
Mysidopsis bahia (OPOSSUM SHRIMP) ON SAMPLES FROM
THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000)

| Test Elutriate | 9 | 6-Hour Su | rvival (%) | 96-Hour | 96-Hour | |
|-----------------|----------------|-------------|-------------|--------------|-----------------------------------|------------------------|
| | Lab Control | Per 10 % | cent Elutri | late 100% | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| Ocean Reference | 98 | 100 | 98 | 100 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

TABLE 8-12A RESULTS OF ELUTRIATE TOXICITY TESTING WITH *Cyprinodon variegatus* (SHEEPSHEAD MINNOW) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | 9 | 96-Hour St | urvival (%) | 96-Hour LC ₅₀ (% elutriate) | 96-Hour NOAEC (% elutriate) | |
|-----------------------------------|---------|------------|-------------|--|-----------------------------------|-----|
| Test Elutriate | Lab | Pe | rcent Eluti | | | |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 100 | 100 | 100 | 100 | > 100 | 100 |
| Brewerton Eastern Extension | 100 | 100 | 100 | 96 | > 100 | 100 |
| C&D Approaches – Cores | 100 | 100 | 96 | 98 | > 100 | 100 |
| C&D Approaches – Surface Grabs | 100 | 98 | 100 | 98 | > 100 | 100 |
| Craighill | 98 | 98 | 94 | 100 | > 100 | 100 |
| Craighill Angle East | 100 | 98 | 96 | 100 | > 100 | 100 |
| Craighill Angle West | 100 | 94 | 96 | 96 | > 100 | 100 |
| Craighill Entrance | 100 | 100 | 98 | 98 | > 100 | 100 |
| Craighill Upper Range | 100 | 98 | 98 | 98 | > 100 | 100 |
| Cutoff Angle | 100 | 92 | 98 | 92 | > 100 | 100 |
| Swan Point | 100 | 98 | 100 | 94 | > 100 | 100 |
| Tolchester North | 100 | 100 | 100 | 98 | > 100 | 100 |
| Tolchester South | 98 | 98 | 98 | 98 | > 100 | 100 |
| Tolchester Straightening | 100 | 100 | 98 | 98 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.
TABLE 8-12BRESULTS OF ADDITIONAL ELUTRIATE TOXICITY TESTING
WITH Cyprinodon variegatus (SHEEPSHEAD MINNOW) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| · · · · · · · · · · · · · · · · · · · | 96-Hour Survival (%)* | | | | 96-Hour | |
|---------------------------------------|-----------------------|------|-------------|------|------------------|------------------------|
| Test Elutriate | Lab | Per | rcent Elutr | iate | LC ₅₀ | 96-Hour |
| | Control | 10 % | 50 % | 100% | (% elutriate) | NOAEC (% elutriate) |
| Inside 104 | 100 | 100 | 100 | 100 | > 100 | 100 |
| Outside 104 | 98 | 100 | 100 | 96 | > 100 | 100 |
| Laboratory Control Sediment | 100 | 100 | 100 | 100 | > 100 | 100 |

*Survival based on a mean of 5 repliccate tests.

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TABLE 8-13A RESULTS OF ELUTRIATE TOXICITY TESTING WITH
Menidia beryllina (INLAND SILVERSIDE) ON SAMPLES FROM BALTIMORE
HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| | 96-Hour Survival (%)* | | | 96-Hour | 96-Hour | |
|-----------------------------------|-----------------------|------|-------------|-------------------|-----------------------------------|------------------------|
| Test Elutriate | Lab | Pe | rcent Eluti | riate | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 94 | 92 | 90 | 86 | > 100 | 100 |
| Outside 104 | 90 | 94 | 92 | 74 | > 100 | 100 |
| Brewerton Eastern Extension | 100 | 80 | 56 | 4 ^(a) | 53.4 | <10 |
| C&D Approaches – Cores | 91 | 98 | 90 | 76 ^(a) | > 100 | 50 |
| C&D Approaches – Surface Grabs | 94 | 78 | 54 | 10 ^(a) | 52.9 | 10 |
| Craighill | 96 | 98 | 60 | 38 ^(a) | 69.7 | 10 |
| Craighill Angle East | 92 | 97 | 60 | 0 ^(b) | 54.4 | 10 |
| Craighill Angle West | 92 | 97 | 77 | 0 ^(b) | 60.7 | 10 |
| Craighill Entrance | 92 | 96 | 72 | 22 ^(a) | 70.7 | 10 |
| Craighill Upper Range | 90 | 84 | 56 | 64 ^(a) | > 100 | 10 |
| Cutoff Angle | 92 | 94 | 50 | 0 ^(b) | 41.5 | 10 |
| Swan Point | 92 | 94 | 80 | 86 | > 100 | 100 |
| Tolchester North | 91 | 88 | 60 | 6 ^(a) | 50.8 | 10 |
| Tolchester South | 100 | 100 | 2 | 0 ^(b) | 23.8 | 10 |
| Tolchester Straightening | 91 | 90 | 92 | 80 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

TABLE 8-13BRESULTS OF ELUTRIATE TOXICITY TESTING WITH
Menidia beryllina (INLAND SILVERSIDE) ON SAMPLES FROM
THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000)

| Test Flutriate | 9 | 6-Hour Su | rvival (%) | * | 96-Hour | 96-Hour |
|-----------------|---------|-----------------------|------------|------|---------------|---------------|
| I est Liutiate | Lab | Lab Percent Elutriate | | iate | LCso | NOAEC |
| | Control | 10 % | 50 % | 100% | (% elutriate) | (% elutriate) |
| Ocean Reference | 98 | 98 | 100 | 98 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

TABLE 8-14A RESULTS OF ELUTRIATE TOXICITY TESTING WITH *Mytilus* sp. (BLUE MUSSEL) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)^(a)

| | 48-Hour Normal Development (%)* | | | 48-Hour | | |
|-----------------------------------|---------------------------------|------|-------------|--------------------|-----------------------------------|-----------------------|
| Test Elutriate | Lab | Pei | rcent Elutr | iate | EC ₅₀ (% elutriate) | NOEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 97 | 95 | 91 | 82 ^(b) | > 100 | 50 |
| Outside 104 | 96 | 92 | 85 | 2 ^(b) | 63.8 | 10 |
| Brewerton Eastern Extension | 94 | 91 | 81 | 0.1 ^(b) | 60.9 | 10 |
| C&D Approaches – Cores | 98 | 92 | 0 | 0 ^(c) | 21.2 | <10 |
| C&D Approaches – Surface Grabs | 99 | 98 | 87 | 2 ^(b) | 64.2 | 10 |
| Craighill | 93 | 89 | 91 | 87 | > 100 | 100 |
| Craighill Angle East | 92 | 94 | 54 | 0 ^(c) | 43.5 | 10 |
| Craighill Angle West | 91 | 90 | 59 | 0 ^(c) | 46.8 | 10 |
| Craighill Entrance | 95 | 93 | 0 | · 0 ^(c) | 22.0 | 10 |
| Craighill Upper Range | 90 | 89 | 86 | 87 | > 100 | 100 |
| Cutoff Angle | 96 | 98 | 92 | 27 ^(b) | 79.7 | 50 |
| Swan Point | 96 | 94 | 63 | 0 ^(c) | 47.8 | 10 |
| Tolchester North | 99 | 95 | 78 | 0.1 ^(b) | 56.4 | 10 |
| Tolchester South | 96 | 93 | 83 | 0 (c) | 61.7 | 10 |
| Tolchester Straightening | 95 | 95 | 84 | 0 ^(c) | 61.7 | 10 |

*Survival based on a mean of 5 replicate tests.

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

(b) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

(c) A concentration which had no normally developed organisms was not statistically compared to the laboratory control.

TABLE 8-14B RESULTS OF ELUTRIATE TOXICITY TESTING WITH Mytilus sp.(BLUE MUSSEL) ON SAMPLES FROM THE NORFOLK OCEANDISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) ^(a)

| | 48-Hour Normal Development (%)* | | | | 48-Hour | |
|-----------------|---------------------------------|-------------------|------|------|------------------|-----------------------|
| Test Elutriate | Lab | Percent Elutriate | | | EC ₅₀ | 96-Hour |
| | Control | 10 % | 50 % | 100% | (% elutriate) | NOEC (% elutriate) |
| Ocean Reference | 98.2 | 99.5 | 98.8 | 98.5 | > 100 | 100 |

*Survival based on a mean of 5 replicate per tests.

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-15ARESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Neanthes arenaceodentata (ESTUARINE POLYCHAETE) ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 1: Sept.-Nov. 1999)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|--------------------------|----------------------------|-----------------------------------|
| Laboratory Control | | (,) |
| Sediment | 25 / 25 | 100 |
| Inside Site 104 | 24 / 25 | 96 |
| Brewerton Eastern Ext. | 25 / 25 | 100 |
| C&D Approaches - Cores | 24 / 25 | 96 |
| C&D Appr Surf. Grabs | 25 / 25 | 100 |
| Craighill | 24 / 25 | 96 |
| Craighill Angle East | 25 / 25 | 100 |
| Craighill Angle West | 23 / 25 | 92 |
| Craighill Entrance | 25 / 25 | 100 |
| Craighill Upper Range | 25 / 25 | 100 |
| Cutoff Angle | 24 / 25 | 96 |
| Swan Point | 25 / 25 | 100 |
| Tolchester North | 25 / 25 | 100 |
| Tolchester South | 24 / 25 | 96 |
| Tolchester Straightening | 24 / 25 | 96 |

*Based on 5 replicates of five animals each

TABLE 8-15BRESULTS OF ADDITIONAL 10-DAY WHOLE SEDIMENT TOXICITY
TESTING WITH Neanthes arenaceodentata (ESTUARINE POLYCHAETE) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|--------------------------------|----------------------------|--------------------------------------|
| Inside Site 104 | 22 / 25 | 88 |
| Outside Site 104 | 20 / 25 | 80 |
| Laboratory Control Sediment | 24 / 25 | 96 |

*Based on 5 replicates of five animals each

TABLE 8-16ARESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Leptocheirus plumulosus (ESTUARINE AMPHIPOD) ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 1: Sept.-Nov. 1999)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|--------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 96 / 100 | 96 |
| Inside Site 104 | 93 / 100 | 93 |
| Brewerton Eastern Ext. | 88 / 100 | 88 |
| C&D Approaches - Cores | 88 / 100 | 88 |
| C&D Appr Surf. Grabs | 94 / 100 | 94 |
| Craighill | 93 / 100 | 93 |
| Craighill Angle East | 91 / 100 | 91 |
| Craighill Angle West | 93 / 100 | 93 |
| Craighill Entrance | 96 / 100 | 96 |
| Craighill Upper Range | 86 / 100 | 86 |
| Cutoff Angle | 92 / 100 | 92 |
| Swan Point | 89 / 100 | 89 |
| Tolchester North | 93 / 100 | 93 |
| Tolchester South | 92 / 100 | 92 |
| Tolchester Straightening | 89 / 100 | 89 |

*Based on 5 replicates of 20 animals each

TABLE 8-16BRESULTS OF ADDITIONAL 10-DAY WHOLE SEDIMENT TOXICITY
TESTING WITH Leptocheirus plumulosus (ESTUARINE AMPHIPOD) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|--------------------------------|----------------------------|--------------------------------------|
| Inside Site 104 | 87 / 100 | 87 ^(b) |
| Outside Site 104 | 96 / 100 | 96 |
| Laboratory Control Sediment | 98 / 100 | 98 |

(a) Survival in sample Inside Site 104 was statistically less (P=0.05) than the laboratory control sediment.

*Based on 5 replicates of 20 animals each

TABLE 8-16CRESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Leptocheirus plumulosus (ESTUARINE AMPHIPOD) FOR THE NORFOLK
OCEAN DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) ^(a)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|---------------------------------|----------------------------|--------------------------------------|
| Lower Chesapeake Bay Control | 94 / 100 | 94 |

(a) Lower Chesapeake Bay control sediment tested in place of Norfolk Ocean Disposal Site sediment to minimize potential grain size effect.

*Based on 5 replicates of 20 animals each



TABLE 8-17RESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Mysidopsis bahia (OPOSSUM SHRIMP) FOR THE NORFOLK OCEAN
DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) (a)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | |
|-----------------|----------------------------|--------------------------------------|--|
| Ocean Reference | 47 / 50 | 94 | |

(a) Mysidopsis bahia tested in place of Neanthes arenaceodentata for Ocean Reference testing.

*Based on 5 replicates of 10 animals each



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TABLE 8-18A RESULTS OF REFERENCE TOXICANT TESTING – BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104

| <u>Test Species</u> | <u>Reference Toxicant</u> | Testing Date | Endpoint | Acceptable Control Chart |
|--------------------------|--|---------------------|---------------------------------|--------------------------|
| Mysidopsis bahia | Potassium chloride | October 1999 | 48-hour LC50: 0.57 g/L KCl | 0.12-0.65 g/L KCl |
| | (KCl) | January 2000 | 48-hour LC50: 0.71 g/L KCl | 0.32-0.86 g/L KCl |
| Cyprinodon variegatus | Potassium chloride | October 1999 | 48-hour LC50: 1.65 g/L KCl | 1.20-1.87 g/L KCl |
| | (KCl) | January 2000 | 48-hour LC50: 1.41 g/L KCl | 1.25-1.60 g/L KCl |
| Menidia beryllina | Potassium chloride (KCl) | January 2000 | 48-hour LC50: 1.18 g/L KCl | 1.10-1.49 g/L KCl |
| Mytilus sp. | Copper sulfate (CuSO ₄ •5H ₂ O) | December 1999 | 48-hour EC50: 10.3 μ g/L Cu | 8.05-12.5 μg/L Cu |
| Neanthes arenaceodentata | Cadmium chloride | October 1999 | 96-hour LC50: 6.4 mg/L Cd | (a) |
| | $(CdCl_2)$ | December 1999 | 96-hour LC50: 5.7 mg/L Cd | (a) |
| Leptocheirus plumulosus | Cadmium chloride | October 1999 | 96-hour LC50: 0.22 mg/L Cd | 0.21-0.26 mg/L Cd |
| | $(CdCl_2)$ | January 2000 | 96-hour LC50: 0.24 mg/L Cd | 0.21-0.25 mg/L Cd |

(a) Control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50s of 6.4 and 5.7 mg/L Cd were similar to the LC50 of 5.7 mg/L Cd on a previous lot of *N. arenaceodentata* from September 1999.

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TABLE 8-18B RESULTS OF REFERENCE TOXICANT TESTING - NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| Test Species | <u>Reference Toxicant</u> | Endpoint | Acceptable Control Chart <u>Limits</u> |
|-------------------------|--|-----------------------------|---|
| Mysidopsis bahia | Potassium chloride (KCl) | 96-hour LC50: 0.65 g/L KCl | 0.59-0.72 g/L KCl |
| Menidia beryllina | Potassium chloride (KCl) | 96-hour LC50: 0.969 g/L KCl | 0.82-1.15 g/L KCl |
| Leptocheirus plumulosus | Cadmium chloride (CdCl ₂) | 48-hour LC50: 5.9 mg/L Cd | (a) |

(a) Control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 5.9 mg/L Cd falls within the range of LC50's from previous L. plumulosus reference toxicant tests (2.2-9.0 mg/l Cd).



TABLE 8-19A SUMMARY OF RESULTS FOR WATER COLUMN BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS, PLACEMENT SITE 104, OUTSIDE SITE 104 REFERENCE, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| | Opossum Shrimp | | Sheepshead Minnow | | | Blue Mussel | | | Inland Silverside | | | |
|---|------------------|----------------------------|-----------------------|---------------|----------------------------|---------------|---------------|----------------------------|-------------------|---------------|----------------------------|---------------|
| | Mysidopsis bahia | | Cyprinodon variegatus | | Mytilus sp. | | | Menidia bervilina | | | | |
| | | Statistical | | | Statistical | | | Statistical | | | Statistical | |
| | 96-hour LC50 | Difference 100% | 96-Hour NOAEC | 96-hour LC50 | Difference 100% | 96-Hour NOAEC | 48-hour EC50 | Difference 100% | 96-Hour NOEC | 96-hour LC50 | Difference 100% | 96-Hour NOAEC |
| Sample Identification | (% elutriate) | vs. Control ^(a) | (% elutriate) | (% elutriate) | vs. Control ^(a) | (% elutriate) | (% elutriate) | vs. Control ^(a) | (% elutriate) | (% elutriate) | vs. Control ^(a) | (% elutriate) |
| | | | | | | | | | | | | |
| Inside Placement Site 104 | >100 | No | 100 | >100 | No | 100 | >100 | Yes | 50 | >100 | No | 100 |
| Brewerton Eastern Extension | >100 | Yes | 50 | >100 | No | 100 | 60.9 | Yes | 10 | 53.4 | Yes | <10 |
| C&D Approaches (Cores) | >100 | No | 100 | >100 | No | 100 | 21.2 | Yes | <10 | >100 | Yes | 50 |
| C&D Approaches (Surficial) | >100 | Yes | 50 | >100 | No | 100 | 64.2 | Yes | 10 | 52.9 | Yes | 10 |
| Craighill | >100 | No | 100 | >100 | No | 100 | >100 | No | 100 | 69.7 | Yes | 10 |
| Craighill Entrance | >100 | Yes | 50 | >100 | No | 100 | 22.0 | Yes | 10 | 70.7 | Yes | 10 |
| Craighill Angle - East | >100 | No | 100 | >100 | No | 100 | 43.5 | Yes | 10 | 54.4 | Yes | 10 |
| Craighill Angle - West | >100 | Yes | 50 | >100 | No | 100 | 46.8 | Yes | 10 | 60.7 | Yes | 10 |
| Craighill Upper Range | >100 | No | 100 | >100 | No | 100 | >100 | No | 100 | >100 | Yes | 10 |
| Cutoff Angle | >100 | No | 100 . | >100 | No | 100 | 79.7 | Yes | 50 | 41.5 | Ycs | 10 |
| Swan Point | >100 | No | 100 | >100 | No | 100 | 47.8 | Yes | 10 | >100 | No | 100 |
| Tolchester Channel - North | >100 | No | 100 | >100 | No | 100 | 56.4 | Yes | 10 | 50.B | Yes | 10 |
| Tolchester Channel - South | >100 | Yes | 100 | >100 | No | 100 | 61.7 | Yes | 10 | 23.B | Yes | 10 |
| Tolchester Straightening | >100 | No | 100 | >100 | No | 100 | 61.7 | Yes | 10 | >100 | No | 100 |
| Inside Placement Site 104 ^(b) | >100 | No | 100 | >100 | No | 100 | - | - | • | • | • | - |
| Outside Placement Site 104 (c) | >100 | No | 100 | >100 | No | 100 | 63.8 | Yes | 10 | >100 | No | 100 |
| Norfolk Ocean Reference Site ^(d) | >100 | No | 100 | NT | NT | NT | >100 | No | 100 | >100 | No | 100 |

NT = not tested; water column bioassays with Norfolk Ocean Reference elutriate conducted with opossum shrimp, blue mussel, and inland silverside only

(a) Statistical significance analyzed at P=0.05; survival (LC50) or normal development (EC50) in 100% elutriate significantly lower than the control.

(b) Inside Placement Site 104 elutriate re-tested concurrently with Outside Placement Site 104 water column bioassays (re-tested with opossum shrimp and sheepshead minnow only)

(c) Outside Placement Site 104 water column bioassays with opossum shrimp and sheepshead minnow conducted separately from channel elutriate bioassays

(d) Norfolk Ocean Reference testing conducted independently of Inside/Outside Placement Site 104 and channel elutriate tests

TABLE 8-19B SUMMARY OF RESULTS FOR WATER COLUMN BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS, PLACEMENT SITE 104, OUTSIDE SITE 104 REFERENCE, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| | Opossum Shrimp | | Sheepshead Minnow | | Blue Mussel | | | Inland Silverside | | | | |
|---|----------------|------------------------|-------------------|---------------|----------------------------|------------------------------|---------------|----------------------------|------------------------------|-------------------|----------------------------|---------------------|
| | | Mysidopsis bahia | | | Cyprinodon variegatus | | Mytilus sp. | | | Menidia beryllina | | |
| | | Statistical | Mixing Factor | | Statistical | Mixing Factor Required to | | Statistical | Mixing Factor Required to | | Statistical | Mixing Factor |
| | 96-hour LC50 | Difference 100% vs. | Required to | 96-hour LC50 | Difference 100% | Achieve 0.01 | 48-hour EC50 | Difference 100% | Achieve 0.01 | 96-hour LC50 | Difference 100% | Required to Achieve |
| Sample Identification | (% elutriate) | Control ^(*) | Achieve 0.01 LC50 | (% elutriate) | vs. Control ^(a) | LC50 | (% elutriate) | vs. Control ^(a) | EC50 | (% elutriate) | vs. Control ^(a) | 0.01 LC50 |
| Inside Placement Site 104 | >100 | No | NA | >100 | No | NA | >100 | Yes | <100 | >100 | No | NA |
| Brewerton Eastern Extension | >100 | Yes | <100 | >100 | No | NA | 60.9 | Yes | 164 | 53.4 | Yes | 188 |
| C&D Approaches (Cores) | >100 | No | NA | >100 | No | NA | 21.2 | Yes | 472 | >100 | Yes | <100 |
| C&D Approaches (Surficial) | >100 | Yes | <100 | >100 | No | NA | 64.2 | Yes | 156 | 52.9 | Yes | 189 |
| Craighill | >100 | No | NA | >100 | No | NA | >100 | No | NA | 69.7 | Yes | 144 |
| Craighill Entrance | >100 | Yes | <100 | >100 | No | NA | 22.0 | Yes | 454 | 70.7 | Yes | 141 |
| Craighill Angle - East | >100 | No | NA | >100 | No | NA | 43.5 | Yes | 230 | 54.4 | Yes | 184 |
| Craighill Angle - West | >100 | Yes | <100 | >100 | No | NA | 46.8 | Yes | 214 | 60.7 | Yes | 165 |
| Craighill Upper Range | >100 | No | NA | >100 | No | NA | >100 | No | NA | >100 | Yes | NA |
| Cutoff Angle | >100 | No | NA | >100 | No | NA | 79.7 | Yes | 126 | 41.5 | Yes | 241 |
| Swan Point | >100 | No | NA | >100 | No | NA | 47.8 | Yes | 209 | >100 | No | NA |
| Tolchester Channel - North | >100 | No | NA | >100 | No | NA | 56.4 | Yes | 177 | 50.8 | Yes | 197 |
| Tolchester Channel - South | >100 | Yes | <100 | >100 | No | NA | 61.7 | Yes | 162 | 23.8 | Yes | 420 |
| Tolchester Straightening | >100 | No | NA | >100 | No | NA | 61.7 | Yes | 162 | >100 | No | NA |
| Inside Placement Site 104 ^(b) | >100 | No | NA | >100 | No | NA | - | - | | • | | - |
| Outside Placement Site 104 ^(c) | >100 | No | NA | >100 | No | NA | 63.8 | Yes | 157 | >100 | No | NA |
| Norfolk Ocean Reference Site ^(d) | >100 | No | NA | NT | NT | - | >100 | No | NA | >100 | No | NA |

NT = not tested; water column bioassays with Norfolk Ocean Reference elutriate conducted with opossum shrimp, blue mussel, and inland silverside only

(a) Statistical significance analyzed at P=0.05; survival (LC50) or normal development (EC50) in 100% elutriate significantly lower than the control.

(b) Inside Placement Site 104 elutriate re-tested concurrently with Outside Placement Site 104 water column bioassays (re-tested with opossum shrimp and sheepshead minnow only)

(c) Outside Placement Site 104 water column bioassays with opossum shrimp and sheepshead minnow conducted separately from channel elutriate bioassays

(d) Norfolk Ocean Reference testing conducted independently of Inside/Outside Placement Site 104 and channel elutriate tests

NA = mixing calculation is not applicable if mean survival in 100% elutriate is not statistically lower than the mean survival in the laboratory control.



TABLE 8-20SUMMARY OF RESULTS FOR WHOLE SEDIMENT BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS,
INSIDE SITE 104, OUTSIDE SITE 104, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| | | Estuarine Polychaet | e | Estuarine Amphipod | | | | |
|--|-----------------------|----------------------------|--------------------|-------------------------|-------------------------|----------------------------------|--|--|
| | Neanthes arenaceodata | | | Leptocheirus plumulosus | | | | |
| | | Statistical Difference vs. | 10% Difference | | Statistical Difference | 20% Difference | | |
| | Mean Survival | Inside Placement Site | vs. Inside | Mean Survival | vs. Inside Placement | vs. Inside Placement Site 104 | | |
| Sample Identification | (%) | 104 ^(a) | Placement Site 104 | (%) | Site 104 ^(a) | | | |
| Laboratory Control Sediment ^(b) | 100 | NA | NA | 96 | NA | NA | | |
| Inside Placement Site 104 | 96 | NA | NA | 93 | NA | NA | | |
| Brewerton Fastern Extension | 100 | No | No | 88 | No | No | | |
| C&D Approaches (Cores) | 96 | No | No | 88 | No | No | | |
| C&D Approaches (Surficial) | 100 | No | No | 94 | No | No | | |
| Craighill | 96 | No | No | 93 | No | No | | |
| Craighill Entrance | 100 | No | No | 96 | No | No | | |
| Craighill Angle - East | 100 | No | No | 91 | No | No | | |
| Craighill Angle - West | 92 | No | No | 93 | No | No | | |
| Craighill Upper Range | 100 | No | No | 86 | No | No | | |
| Cutoff Angle | 96 | No | No | 92 | No | No | | |
| Swan Point | 100 | No | No | 89 | No | No | | |
| Tolchester Channel - North | 100 | No | No | 93 | No | No . | | |
| Tolchester Channel - South | 96 | No · | No | 92 | No | No | | |
| Tolchester Straightening | 96 | No | No | 89 | No | No | | |
| Laboratory Control Sediment (b) | 96 | NA | NA | 98 | NA | NA | | |
| Inside Placement Site 104 | 88 | NA | NA | 87 | NA | NA | | |
| Outside Placement Site 104 | 80 | NA | NA | 96 | NA | NA | | |
| Norfolk Ocean Reference Site (c)(d) | 94 | NA | NA | - | - | - · | | |
| Southern Chesapeake Bay Control Sediment ^{(d)(e)} | - | - | - | 94 | NA | NA | | |

NA = not applicable; control and reference (Inside Site 104) survival not statistically compared as per USEPA/USACE guidelines (1998)

(a) Statistical significance analyzed at P=0.05; channel sediments statistically compared to Inside Placement Site 104

(b) Control serves as indicator for test acceptability/validity

(c) Mysidoposis bahia tested in place of Neanthes arenaceodentata for Norfolk Ocean Disposal Site reference area sediment

(d) Norfolk Ocean Reference tests conducted independently of Inside/Outside Placement Site 104 and channel whole-sediment tests

(e) Southern Chesapeake Bay Control Sediment tested with Leptocheirus in place of Norfolk Ocean Reference sediment as requested by EPA Region III-Philadelphia

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11/7/00

Draft - Chapter 9 Bioaccumulation Studies

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9. BIOACCUMULATION STUDIES

Sediments from the approach channels, placement site, and reference areas were evaluated in 28-day bioaccumulation studies with *Nereis virens* (sand worm) and *Macoma nasuta* (blunt-nose clam). The studies measured survival of the test organisms and the potential for bioaccumulation of contaminants in organism tissue as a result of exposure to the channel, placement site, and reference area sediments. The design of the bioaccumulation studies followed guidance from USEPA/USACE (1991 and 1998) with input from USEPA Region III–Philadelphia.

9.1 BIOACCUMULATION EXPOSURE METHODS

The bioaccumulation testing program consisted of three separate rounds of testing (Table 9-1):

- *Round 1*: Initial bioaccumulation testing of sediment from the approach channels and from Inside Placement Site 104 (November–December 1999).
- Round 2: Additional bioaccumulation testing conducted in response to recommendations from USEPA Region III-Philadelphia. Testing of sediment from an Outside Placement Site 104 reference area for tissue analysis; additional testing of sediment from Inside Placement Site 104 for tissue analysis; and additional testing of sediment from approach channels for tissue analysis of dioxin and furan congeners (January-February 2000).
- *Round 3:* Bioaccumulation testing of Ocean Reference sediment (in conjunction with Woodrow Wilson Bridge sediment testing program) (February-March 2000) to evaluate the potential for ocean placement of dredged material.

A summary of the bioaccumulation testing schedule is provided in Table 9-1. Methodology for the bioaccumulation studies followed guidance in the ITM (USEPA/USACE 1998) and the Ocean Testing Manual (USEPA/USACE 1991). Bioaccumulation testing protocols are thoroughly described in the Quality Assurance Project Plan for the ecotoxicological testing program (EA 2000d) (Appendix B). Testing procedures for the Ocean Reference sediment followed the same guidance as the Site 104 testing and are documented in the Ecotoxicology QAPP for the Woodrow Wilson Bridge Project (EA 2000b). Original data sheets, records, notes, memoranda, and computer printouts for the bioaccumulation exposures are archived at EA's Baltimore Office in Sparks, Maryland. These data will be retained for a period of 5 years unless a longer period is requested by USACE–Baltimore District.

9.1.1 Test Set-Up and Procedures

Bioaccumulation testing was conducted with the sand worm (*Nereis virens*) and the blunt-nose clam (*Macoma nasuta*). The 13 test sediments from the approach channel and one sediment sample from Inside Site 104 were evaluated on 5 November–3 December 1999 with *N. virens* and on 4 November–2 December 1999 with *M. nasuta*. The adult worms (organism lot NV-017)

were obtained from Aquatic Research Organisms (Hampton, New Hampshire) on 5 November 1999, and the adult clams (lot MA-011) were acquired from Aquatic Research Organisms on 2 November 1999.

A second round of bioaccumulation testing with *N. virens* and *M. nasuta* was conducted in January 2000 to evaluate additional contaminants of concern. The 13 test sediments, Inside Site 104 sediments, and Outside Site 104 sediments were tested during the period 13 January through 10 February 2000 with *N. virens* and during the period 12 January through 9 February with *M. nasuta*. This second round of testing utilized organism lot numbers NV-019 (*N. virens*) and MA-012 (*M. nasuta*). Organism lot NV-019 was received from Aquatic Research Organisms on 13 January 2000, and organism lot MA-012 was received from Brezina & Associates (Dillon Beach, California) on 12 January 2000. Both organism lots were used for testing on the day of receipt.

Bioaccumulation testing was also conducted for the Ocean Reference sediment (Round 3) with the sand worm (*Nereis virens*) and the blunt-nose clam (*Macoma nasuta*). The *N. virens* testing with the Ocean Reference sediment (Round 3) was initiated on 10 February 2000 and completed on 08 March 2000. The *M. nasuta* testing with the Ocean Reference sediment (Round 3) was initiated on 08 February 2000 and completed on 07 March 2000. The adult worms (organism lots NV-020 and NV-021) and adult clams (organism lots MA-013 and MA-014) were obtained from Aquatic Research Organisms (Hampton, New Hampshire) on 8 February 2000 for the Ocean Reference sediment testing.

For the bioaccumulation testing, the sediment samples and overlying water were added to the test chambers the day prior to test initiation. The overlying water was artificial seawater (Forty Fathoms sea salts). Natural sediments from the organism collection sites were used as laboratory controls in the bioaccumulation testing. The bioaccumulation tests were 28 days in duration and were conducted as static, renewal assays. The overlying water was replaced three times a week by siphoning approximately 80 percent of the overlying water from the aquaria and replacing with new overlying water taking care not to disturb the sediment surface.

The bioaccumulation tests were conducted in 10-gallon aquaria with a layer of approximately 3-5 cm (3.8 L) of sediment and 19 L of overlying water per aquarium. There were five replicates per sediment sample and three replicates per control sediment. Based on the analytical tissue biomass requirements, the first round of bioaccumulation testing utilized 26 organisms in each replicate chamber for the *N. virens* testing and 50 organisms per chamber for the *M. nasuta* testing. The second round of bioaccumulation testing (conducted in January 2000) used 20 (*N. virens*) and 30 (*M. nasuta*) organisms per test chamber.

During the 28-day exposure period, the test chambers were maintained at a target temperature of 20°C for the *N. virens* and 13°C for the *M. nasuta*, with a 16-hour light/8-hour dark photoperiod. Gentle aeration was provided to each aquarium throughout the test period. Observations of mortality and abnormal organism behavior were recorded daily, and dead organisms were removed from the test chambers. Measurements of temperature, pH, dissolved oxygen, and salinity of the overlying water were recorded daily on one replicate of each sample and control. The water quality measurements for *N. virens* are summarized in Tables 9-2A (Round 1), 9-2B

(Round 2), and 9-2C (Round 3). Water quality measurements for M. nasuta are summarized in Tables 9-3A (Rounds 1), Table 9-3B (Round 2), and 9-3C (Round 3). The organisms were not fed during the exposure period. At the end of the 28-day test periods, the test organisms were retrieved from the test chambers and the number of surviving organisms was recorded. Copies of the original data sheets from the N. virens and M. nasuta testing are included in Attachments V-A (Round 1 and Round 2) and V-B (Round 3).

9.1.2 Reference Toxicant Testing

In conformance with EA's QA/QC program requirements, reference toxicant testing was performed on the acquired lots of organisms utilized in the testing program or reference toxicant data were obtained from the test organism supplier. The reference toxicant tests consisted of a graded concentration series of a specific toxicant in water only tests, with no sediment present in the test chambers.

The reference toxicant for the bioaccumulation species, *Nereis virens* and *Macoma nasuta*, was sodium dodecyl sulfate (SDS). The results of the reference toxicant tests were compared to established laboratory control chart limits.

9.1.3 Tissue Preparation and Homogenization

After 28 days of exposure, surviving organisms were recovered and placed in holding tanks containing 20 ppt artificial seawater and no sediment to purge their digestive tracts. The organisms were not fed during this period. At the end of the 24-hour purging period, the shells of the clams were rinsed with de-ionized (DI) water, the clams were shucked, and the soft tissues and liquids inside the shell were placed into pre-cleaned glass jars. Worms were rinsed with DI water to remove the external salts (originating from the purge chambers) and were placed directly into pre-cleaned glass jars. Tissues for each replicate were placed into separate jars, labeled, and frozen until delivered to the analytical laboratory. Required containers, preservation techniques, and holding time requirements for tissue samples are provided in Table 9-4. Tissues were processed on 04 December 1999 (worms – Round 1), 05 December 1999 (clams – Round 1), 11 February 2000 (worms – Round 2), 10 February 2000 (clams – Round 2), 09 March 2000 (worms – Round 3), and 08 March (clams – Round 3).

In addition to analyzing test tissues, pre-test tissue and tissue from control organisms were also submitted for chemical analysis. Pre-test tissue represents organism tissue upon delivery at the ecotoxicology laboratory (prior to test initiation). These tissues originate from organisms that are sacrificed from each shipment and subsequently frozen. These organisms are not exposed to test sediments, but contaminants in their tissues would represent baseline contaminants that could accumulate from exposure to their natural environment. Control tissue originates from organisms exposed to natural sediment (that they were collected from and shipped in) after a 28-day exposure period. These organisms are exposed to the same 28-day laboratory environment as the test sediments and organisms. Worm and clam tissues were hand-delivered to the analytical laboratory on 08 December 1999 (Round 1), 11 February 2000 (Round 2), and 09 March 2000 (Round 3). The chain-of-custody forms are provided in Attachments V-A and V-B. Tissues were held frozen until analysis. Prior to analysis, tissues for each replicate were separately thawed, homogenized, and weighed to the nearest gram. Aliquots from each replicate were removed for analysis of target fractions. Prior to analytical testing, a 2-gram sub-sample from each replicate was removed for determination of percent moisture by EA's Ecotoxicology Laboratory.

9.1.4 Analytical Methods and Detection Limits

Tissue samples were analyzed for the following analyte fractions: SVOCs, metals, chlorinated pesticides, PAHs, PCB congeners, dioxin and furan congeners, lipids, and moisture content. Target fractions were selected based on the results of the sediment analyses and TBP calculations (as discussed in Chapter 6) and discussions with USEPA Region III–Philadelphia. The project-specific analytical methods for tissue analyses are provided in Table 5-2. Detection limits for tissue analyses are provided in Table 5-5. With the exception of the percent moisture and lipid determinations, Gel Permeation Chromatography (GPC) clean-up was conducted after the extraction process to reduce lipid and other tissue-related matrix interferences that could impact results of the chemical analyses.

9.1.5 Data Analysis and Statistics

The effect of channel sediments on test organism survival and chemical accumulation in tissues were evaluated by comparison to tests on reference sediments. Statistical analyses of survival data and tissue chemistry data were performed according to procedures outlined in the ITM (USEPA/USACE 1998).

9.1.5.1 Test Organism Survival

Overall per cent survival of test organisms in the five replicate tests for each sediment sample was calculated to determine whether organism exposure to the channel sediments resulted in reduced survival of the test organisms as compared the placement site or reference sediment.

If survival in the channel test sediments was not more than ten percent less than survival in the reference sediment, no reduction in survival is indicated. If survival in a channel sediment was more than ten percent lower than in a reference sediment, a statistical test to determine whether survival rates were significantly different was performed.

The data were characterized with respect to distribution and equality of variance to determine the appropriate statistical test. The null hypothesis that survival in channel sediment and the reference sediment were equal was tested using a *t*-test or Wilcoxon's Rank-Sum Test. If the test showed that mean survival was not equal in the two sediments, it was concluded that the channel sediment was more toxic to test organisms than the reference or placement site sediment.



9.1.5.2 Tissue Chemical Residue Data

Before performing the statistical protocols to evaluate whether organisms exposed to channel sediments have greater concentrations of analytes in tissue than those exposed to reference or placement site sediments, several steps were taken to prepare the tissue data for evaluation. These steps were:

- Treatment of data reported below the detection limit; and
- Characterizing the distribution (normal, lognormal, nonparametric) and variance of tissue residue data.

Each of these steps is described below. The entire bioaccumulation data analysis process is illustrated in Figure 9-1.

9.1.5.3 Treatment of Tissue Residues Below the Analytical Detection Limit

When the tissue concentration of an analyte is not greater than a value that reflects a reliable quantity, the analytical laboratory either reports the data as an estimated concentration (qualified by "J") or as not detected (ND or U). Where data were reported as estimated, the reported concentration was assumed to be a true value in the statistical analyses. In cases where a chemical was detected in the laboratory blank (qualified by a "B"), the reported concentration was also assumed to be a true value in the statistical analyses; however, mean concentrations calculated using "B" qualified data are flagged with a "B" in the summary tables.

Data that were reported as not detected (censored) were treated in accordance with guidelines presented by Clarke (1995) for small samples. The actual concentrations of these data are unknown and are presumed to fall between zero and the detection limit (DL). Clark (1995) examined various methods for handling small data sets that include non-detects, and found that, in most cases, substitution of DL/2 when more than 40 percent of the data are non-detects was an appropriate conservative approach. In general, statistical power (i.e., the probability of correctly rejecting the null hypothesis) will decline as the amount of censoring increases. In cases where the data are more than 60 to 80 percent non-detects, it is unlikely that any censoring method will result in acceptable statistical power. Following these guidelines, tissue residue data were censored to replace the non-detects with either the analyte's detection limit (\leq 40% of data below detection), or one-half the detection limit (> 40% of data below detection). Where all (100%) bioaccumulation test replicates were reported as non-detected, the analyte was assumed to be absent from the tissue (ND=0).

9.1.5.4 Characterization of Data Distribution and Variance

In order to determine the appropriate statistical test protocol by which to compare channel sediments to reference or placement site sediment, the data for each case need to be characterized with respect to distribution and variance.

Tests for normality for each analyte were conducted by combining tissue-residue concentrations from all channels and computing the Shapiro Wilk's W statistic for the residuals as described in Conover (1980). The W statistic was computed for both original and log-transformed data. Following USEPA/USACE (1998) guidance for N > 20, a normal distribution was rejected if $W < W_{crit}$ at the 1% confidence level ($\alpha = 0.01$). If the distribution deviated significantly from normality and a lognormal distribution could not be rejected (i.e., would result in a normalized distribution of the data) ($\alpha = 0.01$), then a lognormal distribution was assumed. In cases in which both normal and lognormal distributions were rejected, data were expressed as rankits for the purposes of site comparisons.

Levene's test was used to determine if variances among sites were significantly different ($\alpha = 0.1$) (USEPA/USACE 1998 for n = 2 to 9). Results of normality tests and associated data transformation information are provided in Appendix G.

9.1.5.5 Comparisons of Channel Tissue-Residues to Placement/Reference Tissue-Residues

The sample concentrations for each channel, placement site, or reference area were compared to test the null hypothesis that bioaccumulation in organisms exposed to channel sediment did not exceed the bioaccumulation in organisms exposed to reference sediment. There were five replicate data points for each analyte for each site except that some analytes were tested in two rounds for Inside Site 104 resulting in 10 replicates for those cases. For channel reaches, only chemical analytes that were detected in the tissue (i.e., detected in at least one replicate) were evaluated. If a chemical analyte was not detected in any of the replicates for a channel, no comparison of means was conducted. A flow diagram depicting the decision tree for statistical comparisons of channel tissue-residues to the placement site/reference tissue-residues is depicted in Figure 9-1.

The test used to compare tissue residue concentrations is dependent upon whether variances were equal or not equal among placement/reference. In cases where variances were equal, *t*-tests were performed on the MSE determined from an analysis of variance (ANOVA) using channel reaches as the single factor. If an analyte was not detected in all five replicates of the placement/reference tissues, data were censored as described in Section 9.1.4.2 and the MSE among all placement/reference tissues was used as an estimate of the sample variance.

Where variances were not equal, *t*-tests were performed using the Satterhwaite approximation for separate variances (USEPA/USACE 1998). Three sets of one-way *t*-tests were used to determine if the mean tissue-residue concentration for each channels significantly exceeded the mean tissue-residue concentrations of either Inside Site 104, Outside Site 104, or the Ocean Reference

at the 95% confidence level. The relevant null-hypothesis depends upon the sample distributions as determined in Section 9.1.4.2. The null hypothesis are

| Analyte Distribution | H ₀ |
|----------------------|---|
| normal | $\overline{X}_{site} \leq \overline{X}_{ref}$ |
| lognormal | $\overline{Y}_{site} \leq \overline{Y}_{ref}$; $y_i = \log_{10} x_i$ |
| rankits | $Median_{site} \leq Median_{ref}$ |

All statistical analyses were completed using modified programs from USEPA/USACE 1998 for SAS[©], Version 7.

9.1.5.6 Determination of Total PCBs, Total PAHs, and TEQs for Dioxin

Statistical comparisons for individual PCB and dioxin congeners provide little meaningful information. Because the majority of criteria and effect data are based on total PCBs and dioxin TEQs (not individual congeners), statistical analyses were only conducted for total PCB and TEQ concentrations in the tissue. For PCBs and dioxins, statistical comparisons were only conducted for Total PCBs and dioxin TEQs.

For each individual tissue replicate, PCB concentrations were determined by summing the 18 summation congeners (as specified in Table 9-3 of the ITM). The total summed concentration was then multiplied by a factor of 2 following the NOAA (1993) approach for total PCB determinations.

Total PAHs were determined by summing the concentrations of PAHs in each sample. One PAH, acenaphthylene, was omitted from the total PAH calculation for tissues. The detected concentrations of acenaphthylene varied widely among the tissue replicates. The USEPA analytical method for analyzing PAHs (method SW8310) uses two types of detectors: an ultraviolet detector (UV) and a flourescence detector. The UV detector is a universal detector that responds to a variety of compounds (not just PAHs); the fluorescence detector responds very specifically to individual PAHs, with the exception of acenaphthylene. False positives for acenaphthalene by SW8310 are common in tissue matrices because of the lack of confirmation by the more selective fluorescence detector (Miller 2000, personal communication). Therefore, the detected concentrations of acenaphthylene may be biased high or may be false positives and are based on instrumentation that is non-specific for acenaphthylene.

In the summation calculations for both total PCBs and total PAHs, three total values are presented in the data tables:

- Non-detects = zero (ND=0);
- Non-detects = $\frac{1}{2}$ of the detection limit (ND= $\frac{1}{2}$ DL); and
- Non-detects = the detection limit (ND=DL).

The substitution of the detection limit (ND=DL) provides the most conservative approach to calculating and evaluating the data. However, in cases where few PCB congeners or PAHs are detected, the detection limit drives the total value and overestimates the actual expected concentration.

The TEQs for dioxin were calculated following the approach in USEPA (1989). Each congener was multiplied by the TEF and then the values for each congener were summed. Concentrations that were flagged with a "B" (detected in blank) or "EMPC" (estimated maximum possible concentration) were not included in the TEQ calculation as per the USEPA Region III dioxin validation guidance (USEPA Region III 1999). The TEQs were calculated using both ND=0 and ND=1/2DL.

Mean total PCB, total PAH, and TEQ concentrations that were used in the means comparisons statistics were determined by averaging the total or TEQ that was calculated for each individual replicate. In the PCB, PAH, and dioxin tables, note that the average of the sums does not equal the sum of the average concentrations for each congener or analyte for calculations using ND=0 and ND=1/2DL.

9.1.5.7 Uptake Ratios

In order to evaluate the magnitude of chemical uptake and to assist with the evaluation of the statistically significant tissue-residue results, an Uptake Ratio (UR) was calculated to assess the magnitude of contaminant uptake by organisms during the 28-day period. The 28-day uptake ratio was derived as the ratio of the day 28 mean tissue-residue concentration to the day 0 (PRETEST tissue) mean tissue-residue concentration for each chemical analyte. Because each lot of organisms had a separate set of pre-test analyses, URs were calculated separately for each lot of organisms using the corresponding pre-test tissue-residue results:

$$U_r = \frac{\overline{X}_{r=28}}{\overline{X}_{r=0}}$$

where:

 U_r = Uptake Ratio $\overline{X}_{i=0}$ = day 0 (PRETEST) mean tissue-residue concentration $\overline{X}_{i=28}$ =day 28 mean tissue-residue concentration

The chemical data for the three lots of pre-test tissues are presented in Appendix G.



9.2 **RESULTS**

9.2.1 Survival

9.2.1.1 Nereis virens

The survival results of the first round of *N. virens* bioaccumulation testing are summarized in Table 9-5A. None of the sediments was toxic to *N. virens*. Mean survival in the Inside Site 104 sediment was 89%. Mean survival in all the channel sediments ranged 79 to 98%. With the exception of the C&D Approach (surficial sediment), all channels with mean survival that fell below 89% were within the 10% allowable difference. The C&D Approach (surficial sediment) had the lowest survival (79 percent); however, the mean survival was not statistically lower than Inside Site 104 survival. The control sediment had 95 percent survival.

The results of the second round of bioaccumulation testing with *N. virens* are presented in Table 9-5B. None of the tested sediments was toxic to *N. virens*. The Outside Site 104 reference sediment had the lowest mean survival (83%) of all of the tested sediments. Mean survival in the Inside Site 104 sediment (89%) and in Outside Site 104 sediment (83%) was not statistically different. Mean survival in the channel sediments ranged from 88 to 98%. Survival in the control sediment was 97%.

Results of the third round of bioaccumulation testing with *N. virens* are provided in Table 9-5C. Mean survival rates in the Ocean Reference sediment and the control sediment was 96 percent.

9.2.1.2 Macoma nasuta

Table 9-6A presents the survival results of the first round of *M. nasuta* bioaccumulation testing. None of the sediments were toxic to *M. nasuta*. Mean survival in the channel sediments ranged from 93 to 98 percent, while the sediment from Inside Site 104 had 96 percent survival. The control sediment had 93 percent survival.

The results of the second round of bioaccumulation testing with *M. nasuta* are summarized in Table 9-6B. The Inside Site 104 and Outside Site 104 sediments had 93 and 92 percent survival, respectively. Mean survival rates in sediments from Craighill Entrance (81 percent survival) and Craighill Angle-East (79 percent survival) was statistically lower than mean survival in both the Inside and Outside Site 104 sediments. The remaining channel sediments had mean survival rates ranging from 86 to 97 percent, which were all within 10 percent of mean survival for both Inside and Outside Site 104 sediments. The control sediment had 92 percent survival.

Results of the third round of bioaccumulation testing with *M. nasuta* are provided in Table 9-6C. Mean survival rates in the Ocean Reference sediment and the control sediment were 99 percent.

9.2.2 Reference Toxicant Tests

The results of the reference toxicant tests are summarized in Table 9-7A (Round 1 and Round 2) and Table 9-7B (Round 3). The LC50s from the *N. virens*, and *M. nasuta* reference toxicant tests all fell within the established laboratory control chart limits.

9.2.3 Tissue Contaminant Analysis

Results of the tissue chemistry analyses are presented in the following subsections. The mean analyte concentration for five replicate analyses was used for statistical comparisons to Inside Site 104, Outside Site, and the Ocean Reference. The raw data for each tissue replicate are provided in Appendix D. In several cases where tissue mass was limited (due to mortality in the test chambers during the 28-day test), four replicates were tested for chemical constituents and used in statistical comparisons, rather than five.

Analytical results for tissue-residues are reported on a wet weight basis. Lipid content is reported as a percent of total wet weight. Data qualifiers for the organic data are provided in Table 3-3. Inorganic data qualifiers are provided in Table 3-4. Dioxin and furan qualifiers are provided in Table 6-1. Prior to data analysis organic data were validated using USPEA Region III protocols as described in Section 5.4). Analytical narratives, which include a synopsis of laboratory QA/QC results for Laboratory Control Samples and Matrix Spike/Matrix Spike Duplicate Recoveries, are provided in Attachment VI. The analytical laboratory will retain and archive the results of these analyses for 7 years from the date of issuance of the final results.

Statistical comparisons of placement sites and reference area tissue residues to each are provided for comparative purposes. These statistical comparisons are not required by the ITM (USEPA/USACE 1998) or the Ocean Testing Manual (USEPA/USACE 1991).

9.2.3.1 Metals

Nereis virens

Table 9-8A presents mean concentrations of metals in worm tissues exposed to channel sediments and highlights concentrations that are statistically higher than concentrations reported for Inside Site 104. Table 9-8B presents mean concentrations of metals in worm tissue and highlights concentrations that are statistically higher than concentrations reported for Outside Site 104. Table 9-8C presents mean concentrations of metals in worm tissue and highlights concentrations that are statistically higher than concentrations reported for Outside Site 104. Table 9-8C presents mean concentrations of metals in worm tissue and highlights concentrations that are statistically higher than concentrations reported for the Ocean Reference. Uptake Ratios (UR) for metals in worm tissue are provided in Table 9-9.

<u>Summary</u>

Mercury and thallium were not detected in the worm tissues exposed to channel test sediments. Arsenic, cadmium, iron, lead, and selenium tissue residue concentrations in the channel tests were not found to be significantly different from those exposed to any of the three reference/placement site sediments. While aluminum and silver tissue residues were significantly higher than reference site concentrations in some cases, the actual concentrations were less than in the pretest tissues as indicated by the Uptake Ratios that were less than one or less than in the control tissues. Mean beryllium concentrations in the worm tissue were below the recommended TDL of 0.1 mg/kg in tissue (USEPA/USACE 1995). Antimony, beryllium, chromium, copper, manganese, nickel, and zinc were present in test organism tissues at levels that were significantly higher than references for some cases as discussed below.

Comparisons to Inside Site 104

Antimony, beryllium, chromium, copper, manganese, and nickel had mean tissue-residues concentrations that were statistically higher than the Inside Site 104 tissue-residues for at least one channel. Zinc concentrations were not different than in Inside Site 104 tissues. Beryllium and chromium were statistically higher than the Inside Site 104 tissue for nine of the thirteen channels. None of the metals detected in tissue-residues for Craighill Angle–East and Craighill Entrance was statistically higher than Inside Site 104. Tissues from the remaining channels contained at least one target metal with a mean concentration that was statistically higher than the Inside Site 104.

Comparisons to Outside Site 104

Antimony, beryllium, chromium, copper, manganese, nickel, and zinc had mean tissue concentrations that were statistically higher than the Outside Site 104 tissue. Chromium, copper, and nickel were statistically higher than the Outside Site 104 tissue for all thirteen channel reaches. Tissues from each channel contained at least four target metals with mean concentrations that were statistically higher than the Outside Site 104 tissues.

Comparisons to Ocean Reference

Beryllium, manganese, silver, and zinc had mean concentrations that were statistically higher than the Ocean Reference tissue in at least one channel. Beryllium was the metal that most frequently exceeded the Ocean Reference tissue concentration (9 of 13 channel reaches). However, mean beryllium concentrations in the worm tissue were below the recommended TDL of 0.1 mg/kg in tissue (USEPA/USACE 1995). Tissues from each station, with the exception of Craighill Angle–East and Craighill Entrance, contained one target metal with a mean concentration that was statistically higher than the Ocean Reference.

Macoma nasuta

Table 9-10A presents mean concentrations of metals in clam tissue and highlights concentrations that are statistically higher than Inside Site 104 tissue-residues. Table 9-10B presents mean concentrations of metals in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue-residues. Table 9-10C presents mean concentrations of metals in clam tissue and highlights concentrations of metals in clam tissue and highlights concentrations of metals in clam tissue are statistically higher than Outside Site 104 tissue-residues. Table 9-10C presents mean concentrations of metals in clam tissue and highlights concentrations that are statistically higher than the Ocean Reference tissue. URs for metals in clam tissue are provided in Table 9-11.

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of metals were statistically higher than the Inside Site 104 tissue in 61 of 208 cases (29%) (Table 9-10A). Eleven of the sixteen tested metals in clam tissue (aluminum, antimony, arsenic, beryllium, chromium, iron, manganese, mercury, nickel, selenium, and zinc) had mean concentrations that were statistically higher than the Inside Site 104 for at least one channel. Selenium and manganese were statistically higher than the Inside Site 104 reference tissue at twelve and eleven of the thirteen stations, respectively. Mean mercury concentrations were statistically higher than Inside Site 104 at two of the thirteen stations; however, uptake ratios for the significant mercury concentrations were ≤ 1 , indicating that these concentrations were actually lower than the concentrations detected in pre-test tissues. Tissues from each channel reach contained at least two target metals with mean concentrations that were statistically higher than Inside Site 104 tissues.

Comparisons to Outside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of metals were statistically higher than the Outside Site 104 reference tissue in 104 of 208 cases (50%) (Table 9-10B). Twelve of the sixteen tested metals in clam tissue (aluminum, antimony, arsenic, beryllium, cadmium, chromium, iron, manganese, mercury, nickel, selenium, and zinc) had mean concentrations that were statistically higher than the Outside Site 104 tissue for at least one channel. Mean concentrations of chromium and nickel in tissues for all of the channel reaches were statistically higher than the Outside Site 104 tissue. Iron, manganese, and selenium were statistically higher than the Outside Site 104 tissue for 12 of the 13 channel reaches. Uptake ratios for the significant beryllium, cadmium, mercury, and zinc concentrations were frequently ≤ 1 , indicating that these concentrations were actually lower than the concentrations detected in pre-test tissues. Tissues from each channel contained at least five target metals with mean concentrations that were statistically higher than Outside Site 104.

Comparisons to Ocean Reference

In the thirteen sets of study area clam tissue tests, mean concentrations of metals were statistically higher than the Ocean Reference tissue in 49 of 208 cases (24%) (Table 9-10C). Ten of the sixteen detected metals in clam tissue (aluminum, antimony, beryllium, chromium, iron, manganese, mercury, nickel, selenium, and silver) had mean concentrations that were statistically higher than the Ocean Reference tissue in at least one channel. Manganese was statistically higher than the Ocean Reference for all thirteen channel reaches. Iron and beryllium were statistically higher than the Ocean Reference tissue at twelve and ten of the thirteen channel reaches, respectively. Concentrations factors for the significant beryllium and mercury concentrations detected in pre-test tissues. In addition, mean beryllium concentrations in the clam tissue were below the recommended TDL of 0.1 mg/kg in tissue (USEPA/USACE 1995). Tissues from each channel reach contained at least two target metals with mean concentrations that were statistically higher than the Ocean Reference.



9.2.3.2 Pesticides

Nereis virens

Table 9-12A presents mean concentrations of chlorinated pesticides in worm tissue and highlights concentrations that are statistically higher than Inside Site 104. Table 9-12B presents mean concentrations of chlorinated pesticides in worm tissue and highlights concentrations that are statistically higher than Outside Site 104. Table 9-12C presents mean concentrations of chlorinated pesticides in worm tissue and highlights concentrations of chlorinated pesticides in worm tissue and highlights concentrations of chlorinated pesticides in worm tissue and highlights concentrations of chlorinated pesticides in worm tissue and highlights concentrations that are statistically higher than the Ocean Reference. URs for chlorinated pesticides in worm tissue are provided in Table 9-13.

Comparisons to Inside Site 104

In the thirteen sets of channel worm tissue tests, mean concentrations of pesticides were statistically higher than the Inside Site 104 tissue in 34 of 286 cases (12%). Mean concentrations of thirteen chlorinated pesticides detected throughout the channel test tissues were statistically higher than Inside Site 104. 4,4'-DDD, dacthal, delta-BHC, endosulfan II, heptachlor, and heptachlor epoxide were the most frequently detected pesticides with mean concentrations that were statistically higher than Inside Site 104 tissue. Tissue from 11 of the 13 channel reaches contained at least one pesticide compound with a concentration that was statistically higher than the Inside Site 104. Craighill Channel and Craighill Angle East were the only two channel reaches with no statistical exceedances. Uptake ratios for the significant delta-BHC, endrin, and gamma-BHC concentrations were ≤ 1 , indicating that these concentrations of heptachlor and heptachlor epoxide were below the recommended TDL of 10 µg/kg in tissue (USEPA/USACE 1995).

Comparisons to Outside Site 104

In the thirteen sets of channel worm tissue tests, mean concentrations of pesticides were statistically higher than the Outside Site 104 tissue in 43 of 286 cases (15%). Mean concentrations of eleven chlorinated pesticides were statistically higher than the Outside Site 104 tissue for at least one of the thirteen channel reaches. 4,4'-DDD, aldrin, chlorbenside, dacthal, delta-BHC, gamma-BHC, and heptachlor epoxide were the most frequently detected pesticides with mean concentrations that were statistically higher than the Outside Site 104 tissue from the each of the thirteen channel reaches contained at least one pesticide compound with a concentration that was statistically higher than Outside Site 104, with the exception of Craighill Angle East (no exceedances). Uptake ratios for many of the significant concentrations of aldrin, chlorbenside, delta-BHC, endrin, and gamma–BHC concentrations were ≤ 1 (with the exception of tissue-residues in C&D Approach–surficial), indicating that these concentrations were actually lower than the concentrations detected in pre-test tissues. Significant concentrations of gamma-BHC, heptachlor, and heptachlor epoxide were below the recommended TDLs of 10 μ g/kg in tissue (USEPA/USACE 1995).

Comparisons to the Ocean Reference

In the thirteen sets of channel worm tissue tests, mean concentrations of chlorinated pesticides were statistically higher than the Ocean Reference tissue in 21 of 286 cases (7%). Mean concentrations of seven pesticides were statistically higher than the Ocean Reference tissue for at least one of the thirteen channel reaches. 4,4'-DDD, chlorbenside, dacthal, and heptachlor were the pesticides with mean concentrations that most frequently exceeded the mean concentrations in the Ocean Reference tissue. Significant concentrations of heptachlor were below the recommended TDL of 10 μ g/kg in tissue (USEPA/USACE 1995). None of the pesticides detected in tissues for Craighill Channel, Craighill Angle–East, or Craighill Upper Range were statistically higher than the Ocean Reference tissue.

Macoma nasuta

Table 9-14A presents mean concentrations of chlorinated pesticides in clam tissue and highlights concentrations that are statistically higher than the Inside Site 104 tissue. Table 9-14B presents mean concentrations of chlorinated pesticides in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-14C presents mean concentrations of chlorinated pesticides in clam tissue and highlights concentrations of chlorinated pesticides in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-14C presents mean concentrations of chlorinated pesticides in clam tissue and highlights concentrations that are statistically higher than the Ocean Reference tissue. URs for chlorinated pesticides in clam tissue are provided in Table 9-15.

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of pesticides were statistically higher than the Inside Site 104 tissue in 19 of 286 cases (7%). Mean concentrations of seven pesticides (4,4'-DDD, 4,4'-DDT, alpha-BHC, dieldrin, endosulfan II, gamma-BHC, and heptachlor) were statistically higher than the Inside Site 104 tissue for at least one of the thirteen channel reaches. Uptake ratios for the significant 4,4'-DDD, gamma-BHC, and heptachlor concentrations were ≤ 1 , indicating that these concentrations were lower than the concentrations detected in pre-test tissues. In addition, mean concentrations of 4,4'-DDD, 4,4'-DDT, dieldrin, endosulfan II, gamma-BHC, and heptachlor were below the recommended TDL of 10 μ g/kg in tissue (USEPA/USACE 1995). There were no statistically significant exceedances for pesticides detected in C&D approach (surficial), Craighill Channel, Craighill Angle–East, Craighill Entrance, or Tolchester North.

Comparisons to Outside Site 104

In the thirteen sets of study area clam tissue tests, mean concentrations of pesticides and PCB Aroclors were statistically higher than the Outside Site 104 reference tissue in 21 of 286 cases (7%). Mean concentrations of seven pesticides (4,4'-DDD, 4,4'-DDT, alpha-BHC, dieldrin, endosulfan II, gamma-BHC, and heptachlor) were statistically higher than the Outside Site 104 tissue for at least one of the thirteen channel reaches. Uptake ratios for the significant 4,4'-DDD, gamma-BHC, and heptachlor concentrations were ≤ 1 , indicating that these concentrations were lower than the concentrations detected in pre-test tissues. In addition, mean concentrations of 4,4'-DDD, 4,4'-DDD, 4,4'-DDT, dieldrin, endosulfan II, gamma-BHC and heptachlor were below the recommended TDL of 10 μ g/kg in tissue (USEPA/USACE 1995). There were no statistically



significant exceedances for pesticides detected in the Craighill Channel, Craighill Angle-East, Craighill Entrance, or Tolchester North.

Comparisons to the Ocean Reference

In the thirteen sets of channel clam tissue tests, mean concentrations of pesticides were statistically higher than the Ocean Reference tissue in 22 of 286 cases (8%). Mean concentrations of eight pesticides (4,4'-DDD, 4,4'-DDT, aldrin, alpha-BHC, beta-BHC, endosulfan I, endosulfan II, and gamma-BHC) were statistically higher than the Ocean Reference tissue for at least one of the thirteen channel reaches. Tissue from the thirteen stations contained at least one pesticide compound with a concentration that was statistically higher than the reference with the exception of Tolchester North, Tolchester South, and Tolchester Straightening. Uptake ratios for the significant 4,4'-DDD and heptachlor concentrations were ≤ 1 , indicating that these concentrations were actually lower than the concentrations detected in pre-test tissues. In addition, mean concentrations of 4,4'-DDD, 4,4'-DDT, endosulfan II, and gamma-BHC were below the recommended TDL of 10 μ g/kg in tissue (USEPA/USACE 1995). There were no statistically significant exceedances for pesticides detected in Tolchester North, Tolchester South, or the Tolchester Straightening.

9.2.3.3 PAHs

Nereis virens

Table 9-16A presents mean concentrations of PAHs in worm tissue and highlights concentrations that are statistically higher than Inside Site 104 tissue. Table 9-16B presents mean concentrations of PAHs in worm tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-16C presents mean concentrations of PAHs in worm tissue and highlights concentrations of PAHs in worm tissue for PAHs in worm tissue are statistically higher than the Ocean Reference tissue. URs for PAHs in worm tissue are provided in Table 9-17.

Comparisons to Inside Site 104

In the thirteen sets of channel worm tissue tests, mean concentrations of PAHs were statistically higher than the Inside Site 104 tissue in 0 of 208 cases (0%). None of the mean individual PAH concentrations or mean total PAH concentrations measured in channel tissues were statistically higher than mean concentrations measured in Inside Site 104 tissues.

Comparisons to Outside Site 104

In the thirteen sets of channel worm tissue tests, mean concentrations of PAHs were statistically higher than the Outside Site 104 tissue in 0 of 208 cases (0%). None of the mean individual PAH concentration or mean total PAH concentrations measured in channel tissues were statistically higher than mean concentrations measured in Outside Site 104 tissues.

Comparisons to the Ocean Reference

In the thirteen sets of channel worm tissue tests, mean concentrations of PAHs were statistically higher than the Ocean Reference tissue in 2 of 208 cases (<1%). The mean worm tissue concentration for chrysene was statistically higher in tissues for C&D Approach (surficial);

however, the uptake ratio was <1, indicating that the mean concentration was actually lower than the mean concentration detected in pre-test tissues. The mean worm tissue concentration for fluoranthene was statistically higher in tissues for Tolchester North; however, the uptake ratio was <1, indicating that indicating that the mean concentration was actually lower than the mean concentration detected in pre-test tissues. The significant concentrations for both chrysene and fluoranthene were both below the recommended TDL of 20 μ g/kg in tissue (USEPA/USACE 1995). None of the mean total PAH concentrations measured in the channels were statistically higher than the Ocean Reference concentration, with the exception of Tolchester North; however, the uptake ratio was <1, indicating that the mean concentration was actually lower than the mean concentration detected in pre-test tissues.

Macoma nasuta

Table 9-18A presents mean concentrations of PAHs in clam tissue and highlights concentrations that are statistically higher than Inside Site 104 tissue. Table 9-18B presents mean concentrations of PAHs in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-18C presents mean concentrations of PAHs in clam tissue and highlights concentrations of PAHs in clam tissue for PAHs in clam tissue are provided in Table 9-19.

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of PAHs were statistically higher than the Inside Site 104 tissue in 14 of 208 cases (7%). Mean concentrations of six PAHs [acenaphthylene, benz(a)anthracene, benzo(b)fluoranthene, fluorene, naphthalene, and phenanthrene] were statistically higher than the Inside Site 104 clam tissue for at least one of the thirteen channel reaches. Benz(a)anthracene was the PAH with mean tissue concentrations that most frequently exceeded the Inside Site 104 tissue-residue (7 of 13 channel reaches). Mean concentrations of six PAHs for Tolchester Straightening tissue statistically exceeded Inside Site 104 tissue-residue. Significant concentrations of benz(a)anthracene, benzo(b)fluoranthene, fluorene, and phenanthrene were below the recommended TDL of 20 μ g/kg in tissue (USEPA/USACE 1995). There were no statistically significant exceedances for mean PAH concentrations in tissues for Brewerton Eastern Extension, C&D Approach (cores), Craighill Channel, Craighill Angle-East, Craighill Entrance, and Swan Point. None of the mean total PAH concentrations were statistically higher than the tissue-residues measured for Inside Site 104.

Comparisons to Outside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of PAHs were statistically higher than the Outside Site 104 tissue in 21 of 208 cases (10%). Mean concentrations of six PAHs [acenaphthylene, anthracene, benzo(b)fluoranthene, chrysene, fluoranthene, and naphthalene] were statistically higher than the reference tissue for at least one of the thirteen channel reaches. Acenaphthylene and chrysene were the PAHs with mean tissue concentrations that most frequently exceeded the Outside Site 104 tissue-residues. The uptake ratio for the significant acenaphthylene concentration at C&D Approach (surficial) was <1, indicating that this concentration was actually lower than the concentrations detected in pre-test tissues. Significant concentrations of anthracene, benzo(b)fluoranthene, chrysene, and fluoranthene were

below the recommended TDL of 20 μ g/kg in tissue (USEPA/USACE 1995). There were no statistically significant exceedances for mean PAH concentrations in clam tissue for Craighill Channel, Craighill Angle–East, Craighill Entrance, and Swan Point. None of the mean total PAH concentrations were statistically higher than the tissue-residues measured for Outside Site 104.

Comparisons to Ocean Reference

In the thirteen sets of channel clam tissue tests, mean concentrations of PAHs were statistically higher than the Ocean Reference tissue in 87 of 208 cases (42%). Mean concentrations of eleven PAHs were statistically higher than the reference tissue for at least one of the thirteen sites. Mean tissue concentrations of benzo(b)fluoranthene and fluorene were statistically higher than the Ocean Reference for all 13 channel reaches. With the exception of acenaphthylene and naphthalene, all other significant PAH concentrations were below the recommended target detection limit of 20 μ g/kg in tissue (USEPA/USACE 1995). Clam tissue for each of the thirteen channel reaches contained at least one PAH compound with a mean concentration that was statistically higher than the Ocean Reference. Mean total PAH concentrations measured in all channel tissues (ND=1/2 DL and ND=DL) were statistically higher than the total PAH concentration measured in the Ocean Reference tissue. Uptake ratios for mean total PAHs ranged from 1.07 to 2.65. The highest uptake ratios (>2) were reported for Craighill Upper Range, Tolchester South, and Tolchester Straightening.

9.2.3.4 PCB Arociors

Nereis virens

Table 9-20A presents mean concentrations of PCB aroclors in worm tissue and highlights concentrations that are statistically higher than the Inside Site 104 tissue. Table 9-20B presents mean concentrations of PCB aroclors in worm tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-20C presents mean concentrations of PCB aroclors in worm tissue and highlights concentrations of PCB aroclors in worm tissue and highlights concentrations of PCB aroclors in worm tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-20C presents mean concentrations of PCB aroclors in worm tissue and highlights concentrations that are statistically higher than the Ocean Reference. URs for PCB aroclors in worm tissue are provided in Table 9-21.

Comparisons to Inside Site 104, Outside Site 104, and Ocean Reference

In the thirteen sets of channel worm tissue tests, none of the tested PCB congeners were detected in either the channel, placement site, or reference worm tissues. There were no statistical exceedances for Inside Site 104, Outside Site 104, or Ocean Reference tissue-residues (0 of 273 cases).

Macoma nasuta

Table 9-22A presents mean concentrations of PCB aroclors in clam tissue and highlights concentrations that are statistically higher than the Inside Site 104 tissue. Table 9-22B presents mean concentrations of PCB aroclors in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-22C presents mean concentrations of PCB aroclors in clam tissue and highlights concentrations of PCB aroclors in clam tissue and highlights concentrations of PCB aroclors in clam tissue and highlights concentrations that are statistically higher than the Ocean Reference. URs for PCB aroclors in clam tissue are provided in Table 9-23.



Comparisons to Inside Site 104, Outside Site 104, and Ocean Reference

In the thirteen sets of channel worm tissue tests, none of the tested PCB aroclors were detected in either the channel, placement site, or reference worm tissues. There were no statistical exceedances for Inside Site 104, Outside Site 104, or Ocean Reference tissue-residues (0 of 273 cases).

9.2.3.5 PCB Congeners

Nereis virens

Table 9-24A presents mean concentrations of PCB congeners in worm tissue and highlights mean concentrations for Total PCBs that are statistically higher than Inside Site 104 tissue. Table 9-24B presents mean concentrations of PCB congeners in worm tissue and highlights mean concentrations for Total PCBs that are statistically higher than Outside Site 104 tissue. Table 9-24C presents mean concentrations of PCB congeners in worm tissue and highlights mean concentrations for Total PCBs that are statistically higher than Outside Site 104 tissue. Table 9-24C presents mean concentrations of PCB congeners in worm tissue and highlights mean concentrations for Total PCBs that are statistically higher than the Ocean Reference tissue. URs for PCB congeners and Total PCBs in worm tissue are provided in Table 9-25.

Because many of the congeners were non-detects or were detected at concentrations that were substantially below the TDL for individual congeners in tissue $(2 \mu g/kg)$, the total PCB comparisons using ND=0 are the values that will most appropriately show statistical exceedances of the placement site/reference concentrations.

Comparisons to Inside Site 104

In the thirteen sets of channel worm tissue tests, none of the mean concentrations for Total PCBs was statistically higher than the Inside Site 104 tissue-residue (0 of 13 cases).

Comparisons to Outside Site 104

In the thirteen sets of channel worm tissue tests, none of the mean concentrations for Total PCBs was statistically higher than the Outside Site 104 tissue in (0 of 13 cases).

Comparisons to Ocean Reference

In the thirteen sets of channel worm tissue tests, mean concentrations for Total PCBs were statistically higher than the Ocean Reference tissue in 3 of 13 cases (23%) (ND=0). Total PCB concentrations in worm tissue for Craighill Channel, Craighill Entrance, and the Cutoff Angle were statistically higher than the Ocean Reference tissue.

Macoma nasuta

Table 9-26A presents mean concentrations of PCB congeners in clam tissue and highlights mean Total PCB concentrations that are statistically higher than Inside Site 104 tissue. Table 9-26B presents mean concentrations of PCB congeners in clam tissue and highlights mean Total PCB concentrations that are statistically higher than Outside Site 104 tissue. Table 9-26C presents mean concentrations of PCB congeners in clam tissue and highlights mean Total PCB


concentrations that are statistically higher than the Ocean Reference. URs for PCB congeners and Total PCBs in clam tissue are provided in Table 9-27.

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, none of the mean concentrations for Total PCBs was statistically higher than the Inside Site 104 tissue-residue (0 of 13 cases).

Comparisons to Outside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations for Total PCBs were statistically higher than the Outside Site 104 tissue in 4 of 13 cases (31%) (ND=0). Total PCB concentrations (ND=0) in clam tissue for Brewerton Eastern Extension, C&D Approach (cores), Craighill Entrance, and Swan Point were statistically higher than Outside Site 104 tissue-residues. Uptake ratios for significant Total PCB tissue-residues were <1, indicating that the mean total concentrations were lower than the mean concentrations detected in pre-test tissues.

Comparisons to Ocean Reference

In the thirteen sets of channel clam tissue tests, mean concentrations for Total PCBs were statistically higher than the Ocean Reference tissue in 7 of 13 cases (54%). Total PCB concentrations (ND=0) in clam tissue for Brewerton Eastern Extension, C&D Approach (surficial), C&D Approach (cores), Craighill Channel, Craighill Angle East, Craighill Entrance, and Swan Point were statistically higher than Outside Site 104 tissue-residues. Uptake ratios for significant Total PCB tissue-residues were <1, however, indicating that the mean total concentrations were lower than the mean concentrations detected in pre-test tissues.

9.2.3.6 Semivolatile Organic Compounds (SVOCs)

Nereis virens

Table 9-28A presents mean concentrations of semivolatile organic compounds in worm tissue and highlights concentrations that are statistically higher than Inside Site 104 tissue. Mean concentrations of semivolatile organic compounds in worm tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue are presented in Table 9-28B. Table 9-28C presents mean concentrations of semivolatile organic compounds in worm tissue and highlights concentrations that are statistically higher than the Ocean Reference tissue. URs for semivolatile organic compounds in worm tissue are provided in Table 9-29.

Comparisons to Inside Site 104 and Ocean Reference

In the thirteen sets of channel worm tissue tests, mean concentrations of SVOCs were statistically higher than the Inside Site 104 and Ocean Reference tissues in 8 of 611 cases (1%) for each site comparison. Five of the forty-seven targeted SVOCs (2-methylphenol, di-n-butyl phthalate, di-n-octyl phthalate, pentachlorophenol, and phenol) were statistically higher than the Inside Site 104 and Ocean Reference worm tissues. 2-Methylphenol was the only constituent with a UR that exceeded a value of 2; URs for the other four constituents were <1.4.

Comparisons to Outside Site 104

In the thirteen sets of channel worm tissue tests, mean concentrations of SVOCs were statistically higher than the Outside Site 104 tissues in 7 of 611 cases (1%) for each site comparison. Four of the forty-seven targeted SVOCs (2-methylphenol, di-n-octyl phthalate, pentachlorophenol, and phenol) were statistically higher than the Outside Site 104 worm tissues. 2-Methylphenol was the only constituent with a UR that exceeded a value of 2; URs for the other three constituents were <1.4.

Macoma nasuta

Table 9-30A presents mean concentrations of SVOCs in clam tissue and highlights concentrations that are statistically higher than Inside Site 104 tissue-residues. Table 9-30B presents mean concentrations of SVOCs in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-30C presents mean concentrations of SVOCs in clam tissue and highlights concentrations of SVOCs in clam tissue and highlights concentrations of SVOCs in clam tissue and highlights concentrations that are statistically higher than Outside Site 104 tissue. Table 9-30C presents mean concentrations of SVOCs in clam tissue and highlights concentrations that are statistically higher than the Ocean Reference tissue. URs for semivolatile organic compounds in clam tissue are provided in Table 9-31.

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of SVOCs were statistically higher than the Inside Site 104 tissue in 11 of 611 cases (2%). Mean concentrations of four of the forty-seven tested SVOCs (2-methylphenol, 3,4-methylphenol, 4-nitrophenol, and benzoic acid) were statistically higher than the Inside Site 104 clam tissue for at least one of the thirteen channel reaches. Uptake ratios for the two of the four significant 3,4-methylphenol concentrations were ≤ 1 , indicating that these concentrations were lower than the concentrations detected in pre-test tissues.

Comparisons to Outside Site 104

In the thirteen sets of channel clam tissue tests, mean concentrations of SVOCs were statistically higher than the Outside Site 104 tissue in 15 of 611 cases (2%). Mean concentrations of five of the forty-seven tested SVOCs (1-methylnaphthalene, 2-methylphenol, 3,4-methylphenol, 4-nitrophenol, and benzoic acid) were statistically higher than the Outside Site 104 clam tissue. Uptake ratio for the two of five the significant benzoic acid concentrations were ≤ 1 , indicating that these concentrations were lower than the concentrations detected in pre-test tissues.

Comparisons to Ocean Reference

In the thirteen sets of channel clam tissue tests, mean concentrations of SVOCs were statistically higher than the Ocean Reference tissue in 23 of 611 cases (4%). Mean concentrations of six of the forty-seven tested SVOCs (1-methylnaphthalene, 2-methylphenol, 3,4-methylphenol, 4-nitrophenol, benzyl alcohol, and phenol) were statistically higher than the Ocean Reference clam tissue for at least one channel reach. Clam tissue for all thirteen channel reaches contained at least one SVOC with a concentration that was statistically higher than the Outside Site 104 clam tissue.



9.2.3.7 Dioxin and Furan Congeners

Nereis virens

Table 9-32A presents mean concentrations of dioxins and furans in worm tissue and highlights mean TEQ concentrations that are statistically higher than Inside Site 104. Table 9-32B presents mean concentrations of dioxins and furans in worm tissue and highlights mean TEQ concentrations that are statistically higher than Outside Site 104. Table 9-32C presents mean concentrations of dioxins and furans in worm tissue and highlights mean TEQ concentrations of dioxins and furans in worm tissue and highlights mean TEQ concentrations that are statistically higher than Outside Site 104. Table 9-32C presents mean concentrations of dioxins and furans in worm tissue and highlights mean TEQ concentrations that are statistically higher than the Ocean Reference. URs for dioxins and furans in worm tissue are provided in Table 9-33.

2,3,7,8-TCDD, the most toxic dioxin congener was detected in worm tissue for 5 of the 13 tested channel reaches, with concentrations ranging from 0.0602 to 0.132 ng/kg. These concentrations are substantially below the recommended TDL of 1 ng/kg in tissue (USEPA/USACE 1995). The highest mean concentrations were measured for the least toxic congener, OCDD, ranging from 5.5 to 10.4 ng/kg. The TDL for OCDD in tissue is 10 ng/kg (USEPA/USACE 1995).

Comparisons to Inside Site 104

In the thirteen sets of channel worm tissue tests, none of the mean TEQ concentrations for dioxins and furans in worm tissue was statistically higher than the Inside Site 104 tissue-residues (0 of 13 cases).

Comparisons to Outside Site 104

In the thirteen sets of channel worm tissue tests, mean TEQ concentrations for dioxins and furans in worm tissue were statistically higher than the Outside Site 104 in 9 of 13 cases (69%) (ND=1/2 DL). When zero was substituted for the detection limit in the TEQ calculations, mean TEQ concentrations for only two channels (Brewerton Eastern Extension and Craighill Entrance) were statistically higher than Outside Site 104.

Comparisons to the Ocean Reference

In the thirteen sets of channel worm tissue tests, mean TEQ concentrations for dioxins and furans were statistically higher than the Ocean Reference tissue in 3 of 13 cases (23%) (ND=1/2 DL). When zero was substituted for the detection limit, none of the mean TEQ values were statistically higher than the Outside Site 104 tissue-residue.

Macoma nasuta

Table 9-34A presents mean concentrations of dioxins and furans in clam tissue and highlights mean TEQ concentrations that are statistically higher than Inside Site 104 tissue. Table 9-34B presents mean concentrations of dioxins and furans in clam tissue and highlights mean TEQ concentrations that are statistically higher than Outside Site 104. Table 9-34C presents mean concentrations of dioxins and furans in clam tissue and highlights mean TEQ concentrations of dioxins and furans in clam tissue and highlights mean TEQ concentrations that are statistically higher than Outside Site 104. Table 9-34C presents mean concentrations of dioxins and furans in clam tissue and highlights mean TEQ concentrations that are statistically higher than the Ocean Reference tissue. URs for dioxin and furan congeners in clam tissue are provided in Table 9-35.

2,3,7,8-TCDD, the most toxic dioxin congener, was not detected in the channel clam tissues. The highest congener concentrations were reported for OCDD (the least toxic congener).

Comparisons to Inside Site 104

In the thirteen sets of channel clam tissue tests, none of the mean TEQ concentrations for dioxins and furans in worm tissue was statistically higher than the Inside Site 104 tissue-residues (0 of 13 cases).

Comparisons to Outside Site 104

In the thirteen sets of channel clam tissue tests, mean TEQ concentrations for dioxins and furans were statistically higher than the Outside Site 104 tissue in 1 of 13 cases (8%) (ND=1/2DL). The mean TEQ in clam tissue for the C&D Approach (surficial) was the only concentration that was statistically higher than Outside Site 104. When zero was substituted for the detection limit, the mean TEQ value was not statistically higher than the Outside Site 104 tissue.

Comparisons to the Ocean Reference

In the thirteen sets of channel clam tissue tests, mean TEQ concentrations for dioxins and furans were statistically higher than the Ocean Reference tissue in 7 of 13 cases (54%) (ND=1/2DL). When zero was substituted for the detection limit, the mean TEQ value was not statistically higher than the Ocean Reference Site 104 tissue-residue.

9.2.3.8 Lipids and Percent Moisture

Lipid and percent moisture values for worm and clam tissue are provided in Tables 9-36 and 9-37, respectively. Mean lipid values ranged from 0.27% to 1.39% of total wet body weight for worms and from 0.05% to 0.17% of total wet body weight for clams. Mean percent moisture in worm tissue ranged from 84.5% to 87.1%. Percent moisture in clam tissue ranged from 85.7% to 87.5%.

9.2.3.9 Summary

Of the 151 chemical constituents that were tested in the channel tissues, 95 constituents (63%) were detected in the channel worm tissues and 91 constituents (60%) were detected in the channel clam tissues. For all channels combined, 32 constituents in worm tissue statistically exceeded at least one of the placement site/reference area tissue-residues, and 44 constituents in clam tissue statistically exceeded at least one of the placement site/reference area tissue-residues. Overall, metals and PAHs were more frequency detected in clam tissue, and pesticides and dioxins were more frequently detected in worm tissue.

A total of 1,430 statistical comparisons (13 channel reaches x 110 statistically compared constituents) of channel tissue-residues were conducted for each test species at each placement/reference area (i.e., Inside Site 104, Outside Site 104, Ocean Reference), resulting in an overall total of 8,580 statistical comparisons. The 8,580 statistical comparisons resulted in a combined total of 728 statistical exceedances of the placement site/reference tissues (262 worm

exceedances and 466 clam exceedances). A summary of statistical exceedances for each placement/reference area is provided in Table 9-38A (worms) and Table 9-38B (clams).

The statistical comparisons of the mean concentrations were conducted at an alpha level of 0.05, which indicates that out of every 100 statistical comparisons, at least 5 statistically significant differences will be due to chance alone. Therefore, in the 8,580 statistical comparisons, approximately 429 of the 728 statistical exceedances may have occurred to chance alone.

The number of statistical tissue-residue exceedances for each channel versus Inside Site 104 is provided in Table 9-39A (worms) and 9-39B (clams). The number of statistical tissue-residue exceedances for each channel versus Outside Site 104 is provided in Tables 9-40A (worms) and 9-40B (clams). The number of statistical exceedances for each channel versus the Ocean Placement Site is provided in Tables 9-41A (worms) and 9-41B (clams).

Based on the results of the means comparisons, a master list of COPCs was developed, consisting of every analyte in channel tissue that statistically exceeded at least one placement site/reference tissue-residue. COPCs for worm and clam tissue are provided in Table 9-42. COPCs were further evaluated to determine if the statistical exceedances were ecologically relevant.

9.3 TISSUE-RESIDUE EVALUATION PROCESS

The purpose of the bioaccumulation testing is to predict the potential for uptake of chemical contaminants in the dredged material by aquatic organisms. When tissue concentrations of contaminants of concern in dredged material statistically exceed those of organisms exposed to the reference material, the ITM (USEPA/USACE 1998) and Ocean Testing Manual (USEPA/USACE 1991) recommend coordination with the USEPA Regional Representatives, the USACE District Engineer, or other regional/state regulatory authorities to develop and agree upon case-specific evaluation criteria. These criteria should be based on technical evaluations made with local input and should emphasize factors deemed appropriate for each regional or geographic area with respect to open water or ocean placement.

An evaluation process for this project was developed based on guidance in the ITM (USEPA/USACE 1998), guidance in the Ocean Testing Manual (USEPA/USACE 1991), and input from USEPA Region III–Philadelphia (William Muir, USEPA–Region III, personal communications, 2000). The bioaccumulation evaluation process is depicted in Figure 9-2. Prior to initiating the evaluation, the chemical concentrations achieved after 28 days of exposure were converted to steady-state (Css) concentrations. The initial evaluation consisted of comparing steady-state concentrations of chemical constituents that statistically exceeded the placement site or reference concentrations (COPCs) to available fish tissue screening criteria for human consumption [i.e., USFDA Action Levels (USFDA 1998) and USEPA Tolerance/Guidance Levels (USFDA 1998)]. Based on guidance in the ITM and Ocean Testing Manual, if tissue-residue concentrations are not statistically lower than an FDA Action Level or USEPA Tolerance Value/Guidance Level, the dredged material is not suitable for open-water or ocean placement.

The remaining steps in this evaluation process provide additional weight-of-evidence to determine if placement of the dredged material has the potential to cause unacceptable adverse impacts. Following comparisons to FDA Action Levels and USEPA Tolerance/Guidance Levels, the Css for the COPCs were statistically compared to USEPA Fish Tissue Screening Values (USEPA Office of Water 1995), residue-effect data (ERED, Jarvinen and Ankley 1999; USEPA 2000d), and USEPA Region III fish tissue Risk-Based Concentrations (RBCs) (2000a). In addition, Critical Body Residue (CBR) was calculated to assess the impact of PAH and pesticide body burden in aquatic organisms.

Following the statistical comparisons to screening criteria, the following factors were considered when evaluating the integrated results of the bioaccumulation testing (as recommended by the ITM (USEPA/USACE 1998):

- What is the toxicological importance of the contaminants whose bioaccumulation statistically exceeds that from the reference? (e.g., Do they biomagnify? Do they have effects at low concentrations?)
- By what magnitude does the bioaccumulation from the dredged material exceed bioaccumulation from the reference material?
- What is the propensity for the contaminants, with statistically significant bioaccumulation, to biomagnify within aquatic foodwebs? Contaminants that biomagnify include DDT, PCBs, methylmercury, and dioxin/furans.
- What is the magnitude by which contaminants whose bioaccumulation from the dredged material exceeds that from the reference material also exceeds the concentrations found in comparable species living in the vicinity of the proposed placement location?
- For how many contaminants is bioaccumulation from the dredged material statistically greater than bioaccumulation from the reference material?

For this project, the above evaluation process was used to assess the integrated effect of bioaccumulation and the ecological relevance of the results. The purpose of the evaluation was to provide decision-makers with scientifically valid information to facilitate dredged material placement determinations.

Each step of the COPC tissue-residue evaluation process is described in the following sections. The results of the evaluation are discussed in detail in Section 9.4.

9.3.1 Conversion of Tissue-Residues to Steady-State Concentrations

Uptake of individual contaminants from water or sediment into tissue tends to reach a steadystate concentration, after which continued accumulation is minimal. Most constituents, if they bioaccumulate, will be detectable in tissue after a 28-day exposure period, even if the steadystate has not been reached (USEPA/USACE 1998). Steady-state tissue residues (Css) were estimated from day 28 residues using the relationship described in USEPA/USACE (1998), that relates the proportion of Css reached in 28-day laboratory exposures as a function of the octanolwater coefficient (K_{OW}) for neutral organics. Log K_{ows} for neutral organics are provided in Table 9-43 (also see Figure 9-3). The log K_{ow} indicates the proportion of the steady-state concentration that is expected within 28 days. The steady-state concentration of each chemical constituent was determined by applying a steady-state correction factor that was equivalent to the reciprocal of the decimal fraction of the expected Css at day 28 (USEPA/USACE 1998). Metals concentrations were assumed to reach steady-state during the 28-day exposure period.

9.3.2 Determination of 95% Upper Confidence Levels of the Mean (UCLM)

The UCLM was calculated using censored data as described in Section 9.1.5.3. The 95% UCLM was determined using the t-statistic as described in Sokal and Rohlf (1981):

$$UCLM = \overline{x} + t_{[0.95,df]} \sqrt{S^2 / n}$$
 for normally distributed analytes, and
$$UCLM = \exp(\overline{x} + t_{[0.95,df]} \sqrt{S^2 / n})$$
 for log-normally distributed analytes

where

 \overline{x} = sample mean of normal log-transformed data S^2 = sample variance of normal log-transformed data n = number of sample replicates. $t_{[0.95,df]}$ = one-tailed Student's *t* statistic for $\alpha = 0.95$ and *df* degrees of freedom

If variances were not significantly different (Levene's test with $\alpha = 0.10$), the MSE was used in place of the sample variance S^2 . The *t*-statistic was then evaluated for df = N - k degrees of

freedom, where N is the total number of observations $(\sum_{i=1}^{k} n_i)$ and k is the number of sites. If

variances were unequal, then the actual sample variance was used for S^2 and t was evaluated for df = n - 1 degrees of freedom.

9.3.3 Tissue Contaminant Concentrations Compared to USFDA Action Levels and USEPA Tolerance / Guidance Levels

The initial evaluation consisted of comparing steady-state concentrations of chemical constituents that statistically exceeded the placement site or reference concentrations (COPCs) to available fish tissue screening criteria for human consumption (i.e., USFDA Action Levels (USFDA 1998) and USEPA Tolerance/Guidance Levels (USFDA 1998). These values are derived from risk assessment evaluations for application as critical limits for determining the acceptability of aquatic organisms as food sources to humans. Food lots that exceed the USFDA Action Levels or USEPA Tolerance/Guidance Levels are removed from the market place, and are not considered safe for human consumption. The USFDA Action Levels and USEPA

Tolerance/Guidance Levels (Table 9-44) are generally applicable to shellfish, as well as finfish. If two values were provided, the most conservative value was used in this evaluation.

For substances with FDA action levels (USFDA 1998), USEPA Tolerance / Guidance Levels (USFDA 1998), the criteria values were compared to the one-tailed 95% UCLM tissue-residue concentration for each channel. If the UCLM was below the criterion value (indicating a 95% probability that the population mean tissue-residue concentration for the channel is below the criterion value), it was concluded that the criterion value was not exceeded. As per USEPA/USACE (1998) guidance, all tissue-residue UCLMs for all channels were compared to FDA Action Levels and EPA Tolerance/Guidance Values. UCLM comparisons to FDA Action Levels and USEPA Tolerance/Guidance Levels are provided in Tables 9-45A/B and 9-46A/B, respectively.

9.3.4 Weight-of Evidence Comparisons

For chemical constituents that did not have USFDA Action Levels and USEPA Tolerance/Guidance Levels, more conservatively derived values for screening and evaluation of ecological health were reviewed in a weight-of-evidence assessment. Each parameter in the weight-of-evidence assessment is described in the following sections.

9.3.4.1 Tissue Contaminant Concentrations Compared to USEPA Fish Tissue Screening Values

The USEPA Office of Water has published Screening Values (SVs) for fish tissue in *Guidance* for Assessing Chemical Contaminant Data for Use in Fish Advisories (1995). The USEPA SVs for fish tissue were developed based on average consumption rates and body weight for the general adult population. The published SVs reflect exposure as the oral reference dose (RfD) for non-carcinogens and as the 10^{-5} risk level for carcinogens. The SVs assume that an average adult (with a body weight of 70 kg) consumes 6.5 grams of fish per day over a lifetime. Application of the USEPA fish tissue SVs to the 95% UCLM steady-state channel tissue concentrations is inherently conservative.

For substances with USEPA Fish Tissue Screening Values (USEPA Office of Water 1995), the criteria values were compared to the one-tailed 95% UCLM tissue-residue concentration for each channel. If the UCLM was below the criterion value (indicating a 95% probability that the population mean tissue-residue concentration for the channel is below the criterion value), it was concluded that the criterion value was not exceeded. USEPA fish tissue SVs are provided in Table 9-47. UCLM comparisons were conducted only for individual channel concentrations that statistically exceeded at least one of the reference areas. UCLM comparisons to USEPA fish tissue SVs are provided in Tables 9-48A (worms) and 9-48B (clams).



9.3.4.2 Comparisons to Residue-Effects Data

To evaluate the ecological effects of bioaccumulation to organisms, several residue-effects databases and references were consulted to identify empirical residue-effects data in published literature:

- The Environmental Residue-Effects Database (ERED), maintained by USEPA and USACE-WES (USACE-WES 1999 and www.wes.army.mil/el/ered/index.html);
- Linkage of Effects to Tissue Residues: Development of a Comprehensive Database for Aquatic Organisms Exposed to Inorganic and Organic Chemicals (Jarvinen and Ankley 1999); and
- Appendix to Bioaccumulation Testing and Interpretation for the Purpose of Sediment Quality Assessment: Status and Needs. Chemical-Specific Summary Tables (USEPA 2000b).

ERED (USACE-WES 1999), Jarvinen and Ankley (1999), and USEPA (2000b) are compilations of data relating bioaccumulation of individual chemicals to measurable biological effects in particular aquatic organisms. ERED is maintained by USEPA and USACE and is updated regularly. The biological changes or endpoints in ERED, Jarvinen and Ankley (1999), and USEPA (2000b) include any endpoints reported in the peer-reviewed scientific literature in conjunction with appropriate bioaccumulation data. Some of these studies involve important physiological processes, but measure specific biological endpoints whose consequences, if any, at the organism or ecosystem level are not at all clear (e.g., reduced glucose content of the coelomic fluid). Other endpoints are of clear importance at the level of the organism (e.g., survival, growth) or ecosystem (e.g., various measures of reproduction). Data from ERED, Jarvinen and Ankley (1999), and USEPA (2000) only provide information regarding effects of individual chemicals and do not consider synergistic effects related to accumulation of multiple constituents.

Because of the diversity of species and chemical constituents reported in the literature, the data sources may contain relatively few data points for the exact same species and chemicals tested in this evaluation. However, data for related species and chemicals are useful in evaluating potential bioaccumulation effects. For evaluation of this project, if data were not available for *Nereis* species, data for any annelid were considered potentially useful substitutes. If data were not available for available for *Macoma* species, data for any bivalve mollusk were considered potentially useful substitutes.

PAHs were evaluated by the critical body residue approach discussed in Section 9.3.4. The laboratory measurement considered most useful for this evaluation was "No Observed Effect Dose" (NOED), because this indicates that effects were not observed at a bioaccumulation concentration at least that high; and, therefore, the NOED is presumably a "safe" concentration

in terms of those effects. If NOED data were not available, "Lowest Observed Effects Dose" (LOED) data were used. LOED is the lowest bioaccumulation level (concentration) studied at which the effects were observed, and implies that the effects may occur at a level lower by some unknown margin. The literature values that were most relevant for evaluation of bioaccumulation in *Nereis* (worms) and *Macoma* (clams) are provided in Appendix G.

Steady-state concentrations of COPCs were statistically compared to available residue-effects data from the literature following the procedure described in Section 9.3.2. Only those COPCs for which no USFDA Action Levels, USEPA Tolerance/Guidance Values, or USEPA Fish Tissue Screening values existed were statistically compared to residue-effects data (if available for individual COPCs). Specific residue-effect data (and the associated primary references) that were used in the statistical evaluation are provided in Table 9-49. UCLM comparisons for COPCs (for all channels) to relevant residue-effects data are provided in Tables 9-50A (worms) and 9-50B (clams).

9.3.4.3 Critical Body Residue

In addition to using the literature values established in the literature, Critical Body Residue (CBR) was used to evaluate the potential impact of neutral organic compound (PAHs and pesticides) bioaccumulation in the organism. CBR is useful for evaluating potential cumulative effects of multiple neutral organic contaminants. The CBR approach is based on the PAH (neutral oragnic) primary mode of lethality, which is narcosis (causing unconsciousness, immobility, or death). Studies have shown that narcosis occurs when the concentration of total PAH in tissues exceeds a critical threshold (McCarty and Mackay 1993).

The CBR is the sum of the tissue concentrations of neutral organics (PAHs or PAHs and pesticides) on a μ mol/g wet weight basis. The mean concentration of each detected PAH and pesticide was converted to μ mol/g, a total CBR for each channel was calculated by summing the μ mol/g concentrations of neutral organics in each set of channel tissues. The CBR threshold for chronic narcosis is in the range of 0.2 to 0.8 umol/g wet weight for aquatic invertebrates (McCarty and Mackay 1993). That is, if the CBR is less than the threshold of 0.2 to 0.8 μ mol/g wet weight, chronic narcosis is not expected from the total neutral organics body burden in the organism. The CBR threshold for acute narcosis is about ten times higher, in the range of 2 to 8 μ mol/g wet weight for aquatic invertebrates (McCarty and Mackay 1993).

Critical body residues of neutral organics (PAHs and pesticides) for both *Macoma nasuta* and *Nereis virens* were compared to acute and chronic effect levels. Results of the calculations are provided in Table 9-51A (worms) and 9-51B (clams).

9.3.4.4 Comparisons with USEPA Region III Human Health RBCs for Fish Consumption

Analytes that were detected in tissue-residues and that statistically exceeded one of the placement or reference areas were compared to the USEPA Region III's human health RBCs for fish tissue consumption (USEPA 2000a). Like the fish tissue advisory guidelines, RBCs are based on risk assessment evaluations. The RBCs are developed as highly protective screening

limits for managing Superfund sites. Consequently, the assumptions included in the risk evaluations (used to derive the RBCs) tend to be even more conservative than the fish tissue advisory guidelines. Thus, the RBC comparisons to channel tissue-residues are a conservative assessment.

If an RBC did not exist for a compound, the RBC for a closely related compound was substituted; these substitutions are footnoted on relevant tables. If a closely related compound could not be identified for substitution, the constituent was not screened against a numerical criterion. The upper 95% confidence level of the estimated mean tissue-residue concentration at steady-state was compared to the RBC (lifetime cancer risk of 10⁻⁶⁾ for carcinogenic constituents and one-tenth of the RBC (a hazard quotient of 0.1) for non-carcinogenic constituents, following recommended methodology described in USEPA (1993X). The 95% UCLM comparisons for COPCs (for all channels) to USEPA Region III fish tissue RBCs are provided in Tables 9-52A (worms) and 9-52B (clams).

9.4 DISCUSSION AND TIER III BIOACCUMULATION EVALUATION

9.4.1 Bioaccumulation Survival Rates

Survival rates in the three rounds of bioaccumulation testing with *Nereis virens* and *Macoma nasuta* indicated that all of the sediments were of sufficient quality to support test organisms throughout the 28-day test period (Table 9-53). These results, in combination with the results for the whole-sediment toxicity testing (Chapter 8), indicate that the sediments are of sufficient quality to support benthic communities post-placement.

9.4.2 Tissue-Residue Concentrations

Tissue-residues were evaluated in two phases. The first phase involved statistically comparing all tissue concentrations to USFDA Action Levels and USEPA Tolerance/Guidance Levels. The second phase involved statistical comparisons of chemical concentrations in channel test tissues to chemical concentrations in tissues exposed to placement site/reference sediments and comparisons to other ecological benchmarks.

9.4.2.1 USFDA Action Levels and USEPA Tolerance/Guidance Levels

According to the ITM (USEPA/USACE 1998), after tissue residues are compared to FDA levels, one of the following conclusions is reached:

- 1) Tissue concentrations of one or more contaminants are not statistically less than the FDA levels. Therefore, the dredged material is predicted to result in unacceptable benthic bioaccumulation of contaminants.
- 2) Tissue concentrations of all contaminants are either statistically less than FDA levels or there are no FDA levels for the contaminants. In this case, the information is insufficient to make a factual determination with respect to benthic bioaccumulation of contaminants. The

dredged material requires further evaluation under Tier III to make a factual determination under the Guidelines.

Mercury, total PCBs, aldrin+dieldrin, chlordane, DDD+DDE+DDT, mirex, and total heptachlor steady-state tissue-residue concentrations were statistically compared to USFDA Action Levels (Tables 9-45A and 9-45B). Results indicated that steady-state tissue-residue for all of the constituents were significantly lower than USFDA Action Levels. Concentrations of arsenic, cadmium, chromium, lead, and nickel were statistically compared to USEPA Tolerance/Guidance Levels (Tables 9-46A and 9-46B). Results indicated that the steady-state tissue-residue for all of the constituents were statistically lower than the criteria. This finding leads to conclusion #2 above, indicating that the material may be acceptable for open water or ocean placement pending further analyses. The additional analyses are:

- Comparison of channel tissue-residues to placement site/reference tissue-residues (Section 9.4.2.2) and
- Comparison of channel tissue-residues that statistically exceed placement site/reference tissue-residues to other ecological benchmarks.

9.4.2.2 Comparison of Channel Tissue-Residues to Placement Site/Reference Tissue-Residues

According to the guidance in the ITM (USEPA/USACE 1998), contaminant concentrations in tissues exposed to dredged material, that are statistically lower than FDA levels (or for which no FDA levels exist), are compared to tissue contaminant concentrations for organisms exposed to reference sediment. One of the following conclusions is reached:

- 1) Tissue concentrations of contaminants of concern in organisms exposed to dredged material do not statistically exceed those of organisms exposed to reference sediment. Therefore, the dredged material is not predicted to result in unacceptable benthic bioaccumulation of contaminants. However, benthic toxicity tests must also be evaluated.
- 2) Tissue concentrations of contaminants of concern in organisms exposed to dredged material statistically exceed those of organisms exposed to reference sediment. The final conclusion regarding benthic bioaccumulation of contaminants requires region-specific technical evaluation. Additional testing (Tier IV) may be required and benthic toxicity must be evaluated.

Statistical comparisons to reference tissue concentrations indicated that, for all channels combined, 32 constituents in worm tissue and 44 constituents in clam tissue exceeded tissue-residues for at least one of the placement/reference sites. Overall, for clams and worms combined, 41, 44, and 45 Contaminants Of Potential Concern (COPCs) were identified for Inside Site 104, Outside Site 104, and Ocean Reference, respectively (Table 9-42).



9.4.2.3 Comparison of Channel Tissue-Residues to Other Ecological Benchmarks

Following guidance provided by USEPA Region III-Philadelphia, COPCs were statistically compared to other available fish tissue criteria and residue-effects data to determine the ecological significance and relevance of the detected concentrations. A channel-by-channel summary of statistical comparisons for COPCs to relevant criteria/effect data is provided in Tables 9-52A (worms) and 9-52B (clams).

UCLM (95%) steady-state concentrations of dioxin (TEQ), arsenic, mercury, selenium, DDT, chlordane, dieldrin, endosulfan I and II, gamma-BHC, and total heptachlor were compared to USEPA Fish Tissue Screening Values (USEPA 1995) (see Tables 9-52A and 9-52B). Results revealed that the 95% UCLM steady-state concentrations for dioxin (TEQ) and arsenic in clams exceeded the criteria for several channels, and the 95% UCLM steady-state concentration for dioxin and total heptachlor in worm tissue exceeded the criteria for several channels.

Comparisons to residue-effects data identified only one constituent, benzo(a)pyrene in clam tissue, with a 95%UCLM that exceeded relevant residue-effect data (Table 9-50B).

Calculation of CBR for PAHs and PAHs + pesticides indicated that the total body burdens for neutral organics in all channels were substantially below the concentrations that would be expected to cause either acute or chronic effects to aquatic organisms (Tables 9-51A and 9-51B).

When compared to USEPA Region III RBCs for fish tissue, the 95% UCLM for 19 constituents in clam tissue and 14 constituents in worm tissue exceeded the RBC (Tables 9-52A and 9-52B). UCLMs were compared to the whole RBC value for carcinogenic constituents and one-tenth of the RBC value for non-carcinogenic constituents.

9.4.3 Integrated Evaluation for Channel/Placement Options

Three placement options are evaluated for each of the thirteen channel reaches. COPCs for each channel/placement options are summarized in Table 9-54. When evaluating tissue-residue data, it is important to remember that bioaccumulation is a phenomenon, and does not necessary produce an adverse effect to organism viability or ecological resources. The effects of bioaccumulation are dependent upon exposure (concentration and duration). Statistical exceedance of a placement site/reference tissue-residue does not imply ecological relevance or adverse effect. In the COPC evaluation process, fish tissue criteria were conservatively applied to tissue-residues for benthic invertebrates. In some cases there are significant differences between channel and placement site/reference tissue concentrations; however, the detected concentration varies little from that which is reported in the baseline pre-test tissue. In addition, in some cases, the tissue-residues of the COPCs are either below the recommended detection limits in *QA/QC Guidance for Dredged Material Evaluations* (USEPA/USACE 1995) or are detected in only one of five tested tissue replicates.

An integrated evaluation of all of the tissue-residue information is necessary to make an informed decision regarding the relevance of the statistical exceedances against both the placement site/reference tissues and against conservative criteria. In the following sections, the COPCs for each channel are assessed based on statistical analyses, numerical criteria, properties, and toxicological importance. COPCs were successively screened against available criteria: USFDA Action Levels, USEPA Tolerance/Guidance Levels, USEPA Fish Tissue Screening Values, Residue-Effect Data, and USEPA Region III RBCs. The screening started with the FDA Action Levels and ended with the RBCs. If the UCLM (95%) of a chemical constituent was less than the criterion in the first tier, the constituent was eliminated as a COPC. Only those constituents with no USFDA Action Levels, USEPA Tolerance/Guidance Levels, or USEPA Screening Values were screened against either available residue-effect data or the RBCs. In addition, if a constituent had an uptake ratio (UR) of less than 1 (i.e., concentration less than the pre-test tissue-residue) or if a constituent was detected in the laboratory method blank, it was eliminated as a COPC. A summary of the integrated evaluation for each channel is provided in Tables 9-55 through 9-67.

9.4.3.1 Brewerton Eastern Extension

The integrated evaluation of COPCs for Brewerton Eastern Extension is summarized in Tables 9-55A (worms) and 9-55B (clams).

Twelve COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean Reference tissue-residues. Of those 12 COPCs, the dioxin TEQ (ND=0 and ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. Both of the values are very conservative benchmarks used in the screening phase of risk analysis.

Twenty-three COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 23 COPCs, the dioxin TEQ (ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. Both of the values are very conservative benchmarks used in the screening phase of risk analysis.

Dioxin is a relevant COPC for evaluating Brewerton Eastern Extension sediments for Outside Site 104 and ocean placement.

9.4.3.2 C&D Approach (Surficial)

The integrated evaluation of COPCs for the C&D Approach channels (surficial sediment) is summarized in Tables 9-56A (worms) and 9-56B (clams).

Thirteen COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 13

COPCs, alpha-BHC and chlorbenside were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for alpha-BHC exceeded the USEPA Region III RBC. There are no published criteria for screening chlorbenside in fish tissues. Although detected in the tissue-residues, neither alpha-BHC nor chlorbenside was detected in the sediments collected in C&D Approach surficial sediments.

Twenty-four COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 24 COPCs, the dioxin TEQ (ND=1/2DL), benzo[a]anthracene, and benzo[b]flouranthene were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for dioxin exceeded the USEPA fish tissue SV and it also exceeded the RBC. Benzo[a]anthracene and benzo[b]flouranthene both exceeded the USEPA Region III RBC.

Dioxin, benz[a]anthracene, benzo[b]flouranthene, alpha-BHC, and chlorbenside are relevant COPCs for evaluating C&D Approach (surficial sediments) for placement at either Inside Site 104, Outside Site 104, or the Ocean placement site. Benzo[a]anthracene is only a relevant COPC for Inside Site 104 and Ocean placement. Dioxin is only a relevant COPC for Outside Site 104 and ocean placement.

9.4.3.3 C&D Approach (Cores)

The integrated evaluation of COPCs for C&D Approach channels (core sediment) is summarized in Tables 9-57A (worms) and 9-57B (clams).

Thirteen COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 13 COPCs, the dioxin TEQ (ND=0 and ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. Both of the values are very conservative benchmarks used in the screening phase of risk analysis.

Twenty-three COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 23 COPCs, alpha-BHC was the only constituent that was not eliminated as a COPC in the integrated evaluation process. Alpha-BHC exceeded the USEPA Region III RBC; however, it was not detected in the core sediments tested for the C&D Approach Channel.

Dioxin is a relevant COPC for C&D Approach (Cores) for placement Outside Site 104. IN addition, alpha-BHC is a relevant COPC for Inside Site 104, Outside Site 104, and ocean placement.

9.4.3.4 Craighill Channel

The integrated evaluation of COPCs for the Craighill Channel is summarized in Tables 9-58A (worms) and 9-58B (clams).

Seven COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 7 COPCs, the dioxin TEQ (ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. Both of the values are very conservative benchmarks used in the screening phase of risk analysis. The dioxin TEQ was not a COPC when ND=0 in the TEQ calculations.

Twenty COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 20 COPCs, benzo[b]fluoranthene was the only constituent that was not eliminated as a COPC in the integrated evaluation process. Benzo[b]fluoranthene exceeded the USEPA Region III RBC; however, it was detected below the recommended TDL (USEPA/USACE 1995).

Dioxin is a relevant COPC for evaluating Craighill Channel sediments for Outside Site 104 and ocean placement. Benzo[b]fluoranthene is a relevant COPC for ocean placement only.



9.4.3.5 Craighill Angle East

The integrated evaluation of COPCs for Craighill Angle East is summarized in Tables 9-59A (worms) and 9-59B (clams).

Five COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 5 COPCs, the dioxin TEQ (ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. The dioxin TEQ was not a COPC when ND=0 in the TEQ calculations.

Nineteen COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 19 COPCs, the dioxin TEQ (ND=1/2 DL) and beta-BHC were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. Beta-BHC exceeded the USEPA Region III RBC; however, beta-BHC was not detected in sediment tested from Craighill Angle East. The dioxin TEQ was not a COPC when ND=0 in the TEQ calculations.

Dioxin is a relevant COPC for evaluating Craighill Angle East sediment for Outside Site 104 and ocean placement. Beta-BHC is a relevant COPC for evaluating ocean placement only.



9.4.3.6 Craighill Angle West

The integrated evaluation of COPCs for Craighill Angle West is summarized in Tables 9-60A (worms) and 9-60B (clams).

Thirteen COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 13 COPCs, the dioxin TEQ (ND=1/2DL) and beta-BHC were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for the dioxin TEQ exceeded the USEPA fish tissue SV and it also exceeded the RBC. Beta-BHC exceeded the USEPA Region III RBC; however, beta-BHC was not detected in sediment tested from Craighill Angle West. The dioxin TEQ was not a COPC when ND=0 in the TEQ calculations.

Twenty-two COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 22 COPCs, benzo[a]anthracene, benzo[b]fluoranthene, and alpha-BHC were the only constituents that were not eliminated as COPCs in the integrated evaluation process. Tissue-residues of Benzo[a]anthracene and benzo[b]fluoranthene, and alpha-BHC exceeded the USEPA Region III RBC. In addition, both benzo[a]anthracene and benzo[b]fluoranthene were detected below the recommended TDL (USEPA/USACE 1995). Although detected in the tissue, alpha-BHC was not detected in sediment tested for Craighill Angle West.

Dioxin, benz[a]anthracene, benzo[b]flouranthene, alpha-BHC, and beta-BHC are relevant COPCs for evaluating Craighill Angle West sediments for either Inside Site 104, Outside Site 104, or the Ocean placement. Benz[a]anthracene is a relevant COPC only for Inside Site 104 and ocean placement; benzo[b]fluoranthene is a relevant COPC for ocean placement only; and alpha-BHC is a relevant COPC for placement Inside Site 104 only; dioxin is a relevant COPC for placement Outside Site 104; and beta-BHC is a relevant COPC for Inside Site 104, Outside Site 104, and ocean placement.

9.4.3.7 Craighill Entrance

The integrated evaluation of COPCs for Craighill Entrance is summarized in Tables 9-61A (worms) and 9-61B (clams).

Twelve COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 12 COPCs, the dioxin TEQ (ND=1/2DL) and chlorbenside were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for the dioxin TEQ exceeded the USEPA fish tissue SV and it also exceeded the RBCThe dioxin TEQ for ND=0 did not exceed the USEPA fish tissue screening value. Although detected in the worm tissue, chlorbenside was not detected in the sediment from Craighill Entrance. There are no fish tissue screening criteria for chlorbenside.

Twenty COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 20 COPCs, benzo[b]fluoranthene was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA Region III RBC; however, benzo[b]fluoranthene was detected below the recommended TDL (USEPA/USACE 1995).

Dioxin is a relevant COPC for evaluating Craighill Entrance sediment for Outside Site 104 placement. Chlorbenside is a relevant COPC for evaluating sediment for Outside Site 104 and ocean placement. Benzo[b]fluoranthene is a relevant COPC for ocean placement only.

9.4.3.8 Craighill Upper Range

The integrated evaluation of COPCs for Craighill Upper Range is summarized in Tables 9-62A (worms) and 9-62B (clams).

Ten COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Each of the ten constituents was eliminated as a COPC in the integrated evaluation process.



TEQ was not a COPC, when ND=0 in the TEQ calculations. The 95% UCLM for benzo(a)pyrene exceeded residue-effect data and the RBC. Benzo[b]flouranthene exceeded the USEPA Region III RBC. Mean concentrations of both benzo(a)pyrene and benzo[b]flouranthene were less than the recommended TDL (USEPA/USACE 1995).

Dioxin, benzo[a]pyrene, and benzo[b]flouranthene are relevant COPCs for evaluating Craighill Upper Range sediments for ocean placement only.

9.4.3.9 Cutoff Angle

The integrated evaluation of COPCs for the Cutoff Angle is summarized in Tables 9-63A (worms) and 9-63B (clams).

Twelve COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 12 COPCs, the dioxin TEQ (ND=1/2DL) and chlorbenside were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for the dioxin TEQ exceeded the USEPA fish tissue SV and it also exceeded the RBCAlthough detected in the worm tissue, chlorbenside was not detected in the sediment from Craighill Entrance. There are no fish tissue screening criteria for chlorbenside.

Twenty-five COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 25 COPCs, benz[a]anthracene, benzo[a]pyrene, and benzo[b]flouranthene were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for benzo(a)pyrene exceeded residue-effect data and the RBC. Benz[a]anthracene and benzo[b]flouranthene both exceeded the USEPA Region III RBC. Mean concentrations of benz[a]anthracene, benzo[a]pyrene, and benzo[b]flouranthene were less than the recommended TDL (USEPA/USACE 1995).

Benz[a]anthracene, benzo[a]pyrene, benzo[b]flouranthene, and chlorbenside are relevant COPCs for evaluating Cutoff Angle sediments for ocean placement. Benz[a]anthracene is a relevant COPC for evaluating placement Inside Site 104. Dioxin and chlorbenside are relevant COPCs for evaluating Outside Site 104 placement.

9.4.3.10 Swan Point

The integrated evaluation of COPCs for the Swan Point Channel is summarized in Tables 9-64A (worms) and 9-64B (clams).

Fifteen COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 15 COPCs, the dioxin TEQ (ND=1/2DL) and beta-BHC were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for the dioxin TEQ exceeded the USEPA fish tissue SV and it also exceeded the RBC. The dioxin TEQ was not a COPC when ND=0 was used in TEQ calculation. Beta-BHC exceeded the USEPA Region III RBC; however, beta-BHC was not detected in sediment tested from Swan Point. In addition, it was only detected in two of the four replicate tissue samples.

Twenty COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Each of the 20 constituents was eliminated as a COPC in the integrated evaluation process.

Dioxin and beta-BHC are relevant COPC for evaluating Swan Point sediments for Outside Site 104 and ocean placement, respectively.

9.4.3.11 Tolchester North

The integrated evaluation of COPCs for Tolchester Channel North is summarized in Tables 9-65A (worms) and 9-65B (clams).

Fourteen COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Each of the 14 constituents was eliminated as a COPC in the integrated evaluation process.

Twenty COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 20 COPCs, benz[a]anthracene and benzo[b]flouranthene were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for benz[a]anthracene and benzo[b]flouranthene both exceeded the USEPA Region III RBC. Mean concentrations of both benz[a]anthracene and benzo[b]flouranthene were less than the recommended TDL (USEPA/USACE 1995).

Benz[a]anthracene and benzo[b]flouranthene are relevant COPCs for evaluating Tolchester North sediments for ocean placement. Benz[a]anthracene is also a relevant COPC for evaluating sediment proposed for placement Inside Site 104.

9.4.3.12 Tolchester South

The integrated evaluation of COPCs for Tolchester Channel South is summarized in Tables 9-66A (worms) and 9-66B (clams).

Ten COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Each of the 10 constituents was eliminated as a COPC in the integrated evaluation process.

Twenty COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 20 COPCs, benz[a]anthracene and benzo[b]flouranthene were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for benz(a)anthracene and benzo[b]flouranthene both exceeded the USEPA Region III RBC. Mean concentrations of both benz[a]anthracene and benzo[b]flouranthene were less than the recommended TDL (USEPA/USACE 1995).

Benz[a]anthracene and benzo[b]flouranthene are relevant COPCs for evaluating Tolchester South sediment for ocean placement. Benz[a]anthracene is also a relevant COPC for evaluating placement of Tolchester South sediment Inside Site 104.

9.4.3.13 Tolchester Straightening

The integrated evaluation of COPCs for the Tolchester Straightening is summarized in Tables 9-67A (worms) and 9-67B (clams).

Ten COPCs were identified in worm tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 10 COPCs, the dioxin TEQ (ND=1/2DL) was the only constituent that was not eliminated as a COPC in the integrated evaluation process. The 95% UCLM exceeded the USEPA fish tissue SV and it also exceeded the RBC. The dioxin TEQ was not a COPC, when ND=0 in the TEQ calculations.



Twenty-eight COPCs were identified in clam tissue based on the results of the statistical comparisons against Inside Site 104, Outside Site 104, and Ocean tissue-residues. Of those 28 COPCs, benz[a]anthracene, benzo[a]pyrene, and benzo[b]flouranthene were the only constituents that were not eliminated as COPCs in the integrated evaluation process. The 95% UCLM for benzo(a)pyrene exceeded residue-effect data and the RBC. Benz[a]anthracene and benzo[b]flouranthene both exceeded the USEPA Region III RBC. Mean concentrations of benz[a]anthracene, benzo[a]pyrene, and benzo[b]flouranthene were less than the recommended TDL (USEPA/USACE 1995).

Dioxin, benz[a]anthracene, benzo[b]flouranthene, benzo(a)pyrene, and chlorbenside are relevant COPCs for evaluating Tolchester Straightening sediments for open water or ocean placement. Benzo[b]flouranthene is a relevant COPC for evaluating Inside Site 104, Outside Site 104, and ocean placement. Dioxin is a relevant COPC for evaluating placement Outside Site 104. Benz[a]anthracene is a relevant COPC for evaluating placement Inside Site 104 and ocean placement. Benzo(a) pyrene is a relevant COPC for evaluating ocean placement only.

9.5 SUMMARY OF BIOACCUMULATION STUDIES

The results of the integrated bioaccumulation evaluation yield a total of seven COPCs in the 13 channel reaches (Table 9-68). Dioxin and benzo[b]fluoranthene are COPCs in 9 of 13 channel reaches; benz[a]anthracene is a COPC in 6 of 13 channel reaches; chlorbenside is a COPC in 4 of 13 channel reaches; and benzo(a)pyrene, alpha-BHC, and beta-BHC are COPCs in 3 of 13 channel reaches. Benzo(a)pyrene is a relevant COPC for ocean placement only. Benz[a]anthracene is a relevant COPC for Inside Site 104 and ocean placement only.

Eight of the 13 channel reaches contain at least one COPC for placement Inside Site 104. Brewerton Eastern Extension, Craighill, Craighill Angle East, Craighill Entrance, and Craighill Upper Range contain no COPCs that are relevant to placement Inside Site 104.

Nine of the 13 channel reaches contain at least one COPC for placement outside Site 104. Dioxin is one of the COPCs relevant to placement Outside Site 104 for each of those nine of the channels. Craighill Upper Range, Swan Point, Tolchester North, and Tolchester South have no COPCs that are relevant to placement Outside Site 104.

All 13 channels contain at least one COPC for ocean placement.

The results of the statistical comparisons to placement site/reference tissues and the results of the integrated evaluation for each channel are summarized as follows:

• Although a total of 53 COPCs was identified as a result of the statistical comparisons against the placement site/reference tissue-residues, the integrated evaluation revealed only seven COPCs that warrant further consideration (Table 9-68).

- Although not detected in many of the sediments, pesticides, such as alpha-BHC, beta-BHC, and chlorbenside were detected in the channel tissues at concentrations that statistically exceeded one or more placement site/reference tissue-residues.
- Dioxin tended to be a COPC only when the TEQ was calculated using ND=1/2 DL. Dioxin was a COPC in both worm and clam tissue.
- PAHs (particularly benz[a]anthracene, benzo(a)pyrene, and benzo[b]fluoranthene) were detected in the channel clam tissues at concentrations that statistically exceeded one or more placement site/reference tissue-residues; however, the mean detected concentrations were less than the recommended TDL (USEPA/USACE 1995). CBR for Total PAHs was not exceeded, however, in any of the channel tissues.
- Chlorbenside was retained as a COPC for worm tissue because there are no fish-tissue criteria to screen the tissue-residues.

The remaining COPCs were retained based in comparison to conservative screening values which indicate that a potential for risk cannot be ruled out. Further evaluation with respect to more realistic exposure scenarios at each placement site is necessary to determine whether the potential for risk is significant.











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Source: USEPA/USACE 1998; figured adapted from MacFarland (1994)

Figure 9-3. Proportion of steady-state concentration (Css) of neutral organic compounds expected to be reached in 28-day laboratory exposure. Log K_{ow} values for tested constituents are provided in Table 9-43.



TABLE 9-1 SUMMARY OF BIOACCUMULATION TESTING SCHEDULE

| | | | TEST SPECIES | | ANALYTES TESTED | | | | | | |
|---------------|-----------------------------|----------------------|------------------|--------------------|-----------------|------------|------|------------------|-------|------------------------|---------------------|
| | | | Sand Worm | Blunt-Nose Clam | | | | | | Diaxin | I inide and |
| TEST ROUND | DATES | TEST SEDIMENT | Nereis virens | Macoma nasuta | Metals | Pesticides | PAHs | PCB Congeners | SVOCs | and Furan Congeners | Percent Moisture |
| 1 | November– December | Inside Site 104 | X | x | X | X | X | х | x | | X |
| | 1999 | Approach Channels | Х | Х | Х | х | Х | Х | X | | x |
| 2 | January February 2000 | Inside Site 104 | х | Х | x | х | х | Х | х | x | x |
| | | Outside Site 104 | х | x | x | Х | x | x | x | x | X |
| | | Approach Channels | X ^(a) | X ^(a) | | | | | | x | x |
| 3 | February – March 2000 | Ocean Reference | Х | х | x | Х | х | x | x | x | x |

(a) tested for dioxin/furan congeners, lipids, and percent moisture only.

TABLE 9-2A WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Nereis virens* (SAND WORM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: 5 November – 3 December 1999)

| | Mean (±Standard Deviation) | | | |
|-----------------------------------|----------------------------|------------|------------------|-------------|
| Test Sediment | Temperature | | Dissolved Oxygen | Salinity |
| | (°C) | pH | (mg/L) | (ppt) |
| Laboratory Control | 19.6 (±0.5) | 7.9 (±0.1) | 6.2 (±0.9) | 30.5 (±0.4) |
| Inside Site 104 | 19.2 (±0.6) | 7.9 (±0.1) | 6.3 (±0.7) | 30.3 (±0.7) |
| Brewerton Eastern Ext | 19.5 (±0.4) | 7.9 (±0.1) | 6.3 (±0.7) | 30.2 (±0.8) |
| C&D Approaches – Surface Grabs | 18.6 (±0.5) | 8.0 (±0.1) | 6.6 (±0.5) | 30.1 (±0.8) |
| C&D Approaches - Cores | 19.1 (±0.6) | 7.9 (±0.1) | 6.5 (±0.6) | 30.1 (±0.9) |
| Craighill | 19.5 (±0.5) | 8.0 (±0.1) | 6.4 (±0.7) | 30.4 (±0.6) |
| Craighill Angle East | 19.5 (±0.5) | 7.9 (±0.1) | 6.4 (±0.7) | 30.2 (±0.7) |
| Craighill Angle West | 19.4 (±0.4) | 7.9 (±0.1) | 6.2 (±0.9) | 30.2 (±0.5) |
| Craighill Entrance | 19.5 (±0.6) | 7.9 (±0.1) | 6.2 (±0.8) | 30.3 (±0.7) |
| Craighill Upper Range | 19.3 (±0.5) | 8.0 (±0.1) | 6.6 (±0.6) | 30.3 (±0.5) |
| Cutoff Angle | 18.7 (±0.6) | 7.9 (±0.1) | 6.6 (±0.5) | 30.1 (±0.8) |
| Swan Point | 19.3 (±0.5) | 7.9 (±0.1) | 6.4 (±0.8) | 30.3 (±0.7) |
| Tolchester North | 18.9 (±0.5) | 7.9 (±0.1) | 6.6 (±0.6) | 30.5 (±0.7) |
| Tolchester South | 18.6 (±0.4) | 7.9 (±0.1) | 6.6 (±0.6) | 30.2 (±0.7) |
| Tolchester Straightening | 18.8 (±0.4) | 7.9 (±0.1) | 6.6 (±0.5) | 30.2 (±0.8) |

TABLE 9-2B WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Nereis virens* (SAND WORM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: 13 January – 10 February 2000)

| | Mean (±Standard Deviation) | | | | | | |
|---------------------------|----------------------------|----------------|------------------|-------------|--|--|--|
| Test Sediment | Temperature | | Dissolved Oxygen | Salinity | | | |
| | (°C) | pН | (mg/L) | (ppt) | | | |
| Laboratory Control | 17.9 (±0.7) | 8.0 (±0.1) | 6.8 (±0.9) | 30.8 (±0.6) | | | |
| | | | | | | | |
| Incide Side 104 | 181(.00) | 80(.01) | 66(.0.8) | 207(.08) | | | |
| | 18.1 (±0.6) | 8.0 (±0.1) | 0.0 (±0.8) | 30.7 (±0.8) | | | |
| | | | | | | | |
| Outside Site 104 | 18.0 (±0.7) | 8.0 (±0.1) | 6.5 (±0.7) | 30.5 (±0.9) | | | |
| | | | | | | | |
| Brewerton Eastern Ext. | 19.9 (±0.5) | 8.0 (±0.1) | 6.5 (±0.6) | 30.5 (±1.0) | | | |
| | , , , | | | | | | |
| CPD Anna Ala | 10.0 (0.0) | 00(01) | | 20 ((0 8) | | | |
| C&D Approacnes – | 19.0 (±0.6) | 8.0 (±0.1) | 6.4 (±1.0) | 30.6 (±0.8) | | | |
| Surface Grabs | | | | | | | |
| | | | | | | | |
| C&D Approaches - Cores | 18.8 (±0.6) | 7.9 (±0.1) | 6.4 (±0.6) | 30.6 (±0.9) | | | |
| | | | | | | | |
| Craighill | 19.8 (±0.4) | 8.1 (±0.2) | 6.6 (±0.7) | 30.6 (±1.1) | | | |
| | | | | | | | |
| Craighill Angle Fast | 105(106) | 79(.01) | 64(+0.8) | 30.4(+1.2) | | | |
| Craigini Angie East | 19.5 (±0.0) | 7.9 (±0.1) | 0.4 (±0.8) | JU.4 (±1.2) | | | |
| | | | | | | | |
| Craighill Angle West | 18.9 (±0.4) | 8.0 (±0.2) | 6.5 (±0.9) | 30.4 (±1.2) | | | |
| | | | | | | | |
| Craighill Entrance | 19.9 (±0.5) | 7.9 (±0.1) | 6.3 (±1.0) | 30.5 (±1.1) | | | |
| | | | | | | | |
| Craighill Upper Range | 18.6 (±0.4) | $8.0(\pm 0.1)$ | 6.5 (±1.1) | 30.4 (±1.3) | | | |
| | · · · · | | | | | | |
| Cutoff Angle | 180(105) | 79(+01) | 66(+07) | 304(+0.0) | | | |
| | 18.0 (±0.5) | 7.9 (±0.1) | 0.0 (±0.7) | 50.4 (±0.5) | | | |
| | | | | | | | |
| Swan Point | 19.1 (±0.6) | 8.0 (±0.1) | 6.5 (±0.7) | 30.7 (±0.8) | | | |
| | | | | | | | |
| Tolchester North | 18.6 (±1.4) | 7.9 (±0.1) | 6.6 (±0.7) | 30.4 (±1.0) | | | |
| | | | | | | | |
| Tolchester South | 17.6 (+0.6) | 8.0 (+0.1) | 7.1 (±0.6) | 30.5 (±0.9) | | | |
| | | | (2000) | | | | |
| Talahastan Start Literat | 20.0 (0.5) | 90(01) | (5(07) | 2060.00 | | | |
| 1 olchester Straightening | 20.0 (±0.5) | 8.0(±0.1) | 0.3 (±0./) | 30.0 (±0.0) | | | |

TABLE 9-2C WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Nereis virens* (SAND WORM) ON SAMPLES FROM THE NORFOLK OCEAN DISPOSAL REFERENCE SITE (Round 3: 10 February – 8 March 2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Sediment | Temperature (°C) | pН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Laboratory Control | 19.8 (±0.5) | 8.0 (±0.2) | 7.2 (±0.9) | 30.5 (±0.7) | | |
| Ocean Reference | 19.7 (±0.3) | 8.2 (±0.2) | 7.0 (±1.0) | 30.4 (±0.6) | | |

TABLE 9-3A WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Macoma nasuta* (BLUNT-NOSE CLAM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: 4 November – 2 December 1999)

| | | Mean (±Standard Deviation) | | | | |
|----------------------------------|-------------|----------------------------|------------------|-------------|--|--|
| Test Sediment | Temperature | | Dissolved Oxygen | Salinity | | |
| | (°C) | рН | (mg/L) | (ppt) | | |
| Laboratory Control | 11.4 (±0.4) | 8.0 (±0.1) | 8.1 (±0.8) | 30.6 (±0.6) | | |
| Inside Site 104 | 11.9 (±0.3) | 8.0 (±0.1) | 8.0 (±0.6) | 30.2 (±0.6) | | |
| Brewerton Eastern Ext | 11.5 (±0.4) | 8.0 (±0.1) | 8.0 (±0.6) | 30.2 (±0.6) | | |
| C&D Approaches- Surface Grabs | 11.7 (±0.4) | 7.9 (±0.1) | 7.9 (±1.0) | 30.1 (±0.7) | | |
| C&D Approaches - Cores | 11.4 (±0.4) | 7.9 (±0.1) | 7.9 (±0.8) | 30.1 (±0.7) | | |
| Craighill | 11.5 (±0.4) | 8.0 (±0.1) | 8.1 (±0.7) | 30.4 (±0.6) | | |
| Craighill Angle East | 11.5 (±0.4) | 7.9 (±0.1) | 7.9 (±0.7) | 30.1 (±0.4) | | |
| Craighill Angle West | 11.5 (±0.4) | 7.9 (±0.1) | 7.9 (±0.8) | 30.1 (±0.4) | | |
| Craighill Entrance | 10.9 (±0.6) | 8.0 (±0.1) | 8.3 (±0.8) | 30.5 (±0.7) | | |
| Craighill Upper Range | 11.5 (±0.4) | 8.0 (±0.1) | 7.9 (±1.0) | 30.4 (±0.6) | | |
| Cutoff Angle | 12.1 (±0.4) | 8.0 (±0.1) | 8.0 (±0.7) | 30.3 (±0.6) | | |
| Swan Point | 11.3 (±0.3) | 7.9 (±0.1) | 8.1 (±0.7) | 30.2 (±0.4) | | |
| Tolchester North | 11.8 (±0.4) | 8.0 (±0.1) | 8.0 (±0.7) | 30.2 (±0.7) | | |
| Tolchester South | 12.3 (±0.3) | 8.0 (±0.1) | 8.1 (±0.6) | 30.2 (±0.6) | | |
| Tolchester Straightening | 11.3 (±0.4) | 7.9 (±0.1) | 8.1 (±0.7) | 30.2 (±0.7) | | |

TABLE 9-3B WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Macoma nasuta* (BLUNT NOSE CLAM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: 12 January – 9 February 2000)

| | Mean (± Standard Deviation) | | | | | | |
|--------------------------|-----------------------------|------------|------------------|---------------|--|--|--|
| Test Sediment | Temperature | | Dissolved Oxygen | Salinity | | | |
| | (°C) | pН | (mg/L) | (ppt) | | | |
| Laboratory Control | 13.5 (±0.2) | 8.1 (±0.1) | 8.2 (±0.7) | 30.9 (±0.4) | | | |
| | | | | | | | |
| Inside Site 104 | 13.0 (±0.2) | 8.1 (±0.1) | 8.2 (±0.8) | 30.3 (±0.4) | | | |
| | | | | | | | |
| Outside Site 104 | 13.0 (±0.2) | 8.1 (±0.1) | 8.0 (±0.8) | 30.3 (±0.5) | | | |
| | | | | | | | |
| Brewerton Eastern Ext. | 12.8 (±0.4) | 8.1 (±0.1) | 8.3 (±0.7) | 30.4 (±0.6) | | | |
| | | | | | | | |
| C&D Approaches – | 12.9 (±0.2) | 8.1 (±0.1) | 8.4 (±0.7) | 30.2 (±0.5) | | | |
| Surface Grabs | | | | | | | |
| | | | | | | | |
| C&D Approaches - Cores | 13.0 (±0.2) | 8.0 (±0.1) | 8.3 (±0.6) | 30.1 (±0.6) | | | |
| | | | | | | | |
| Craighill | 12.3 (±0.3) | 8.1 (±0.1) | 8.5 (±0.6) | 30.6 (±0.4) | | | |
| | | | | | | | |
| Craighill Angle East | 12.9 (±0.3) | 8.0 (±0.1) | 8.1 (±0.7) | 30.3 (±0.6) | | | |
| | | | | | | | |
| Craighill Angle West | 12.6 (±0.2) | 8.0 (±0.1) | 8.1 (±0.9) | 30.4 (±0.5) | | | |
| | | | | | | | |
| Craighill Entrance | 12.8 (±0.2) | 8.0 (±0.1) | 8.1 (±0.7) | 30.5 (±0.5) | | | |
| | | | | | | | |
| Craighill Upper Range | 12.6 (±0.2) | 8.1 (±0.1) | 8.0 (±1.2) | 30.4 (±0.5) | | | |
| | | | | | | | |
| Cutoff Angle | 13.0 (±0.2) | 8.0 (±0.1) | 8.0 (±0.6) | 30.4 (±0.5) | | | |
| | | 00400 | | 205(05) | | | |
| Swan Point | 13.1 (±0.3) | 8.0 (±0.1) | 8.1 (±0./) | 30.3 (±0.3) | | | |
| 100 J. J. (| | 00(01) | | 202(00) | | | |
| Tolchester North | 12.7 (±0.3) | 8.0 (±0.1) | 8.0 (±1.0) | 30.3 (±0.6) | | | |
| | | 00(01) | | 20 4 (- 0.5) | | | |
| Tolchester South | 12.8 (±0.2) | 8.0 (±0.1) | 8.1 (±0.9) | 30.4 (±0.5) | | | |
| | | | | 20.2 (0.0) | | | |
| Tolchester Straightening | 13.0 (±0.2) | 8.1 (±0.1) | 8.2 (±1.1) | 30.3 (±0.5) | | | |



TABLE 9-3C WATER QUALITY PARAMETERS MEASURED DURING 28-DAY BIOACCUMULATION TESTING WITH *Macoma nasuta* (BLUNT NOSE CLAM) ON SAMPLES FROM THE NORFOLK OCEAN DISPOSAL REFERENCE SITE (Round 3: 8 February – 7 March 2000)

| | Mean (±Standard Deviation) | | | | | |
|--------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| Test Sediment | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Laboratory Control | 12.1 (±0.9) | 8.0 (±0.2) | 8.2 (±1.3) | 30.2 (±0.6) | | |
| Ocean Reference | 11.5 (±0.7) | 8.1 (±0.1) | 8.5 (±0.5) | 30.5 (±0.5) | | |

TABLE 9-4 REQUIRED CONTAINERS, PRESERVATION TECHNIQUE, ANDHOLDING TIMES FOR TISSUE SAMPLES

| Parameter | Mass Required (grams) | Container ^(a) | Preservative | Holding Time ^(b) |
|--|-----------------------------|--------------------------|------------------------------|--|
| Inorganics | | | | |
| Mercury | 5 | G | Frozen, $\leq -20^{\circ}$ C | 28 days |
| Other Metals | 5 | G | Frozen, $\leq -20^{\circ}$ C | 6 months |
| Organics | | | | |
| Lipids | 5 | G | Frozen, ≤ -20°C | Up to 1 year if frozen (14 days after thaw) to analysis |
| Organotins | 10 | G | Frozen, ≤ -20°C | Up to 1 year if frozen (14 days after thaw) to extraction, 7 days from extraction to derivatization, 40 days after extraction |
| Dioxins/Furans | 30 | G | Frozen, ≤ -20°C | Up to 1 year if frozen (30 days after thaw) to extraction, 40 days after extraction |
| Pesticides (Organochlorine), PCBs (Aroclors and Congeners), Semivolatile Organics, Polynuclear Aromatic Hydrocarbons | 110 | G | Frozen, ≤ -20°C | Up to 1 year if frozen (14 days after thaw) to extraction, 40 days after extraction |

(a) P = plastic; G = glass.
(b) From time of sample collection.

TABLE 9-5ARESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH
Nereis virens (SAND WORM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104 (Round 1: 5 November – 3 December 1999)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) | |
|-----------------------------------|----------------------------|--------------------------------------|--|
| Laboratory Control Sediment | 74/78 | 95 | |
| Inside Site 104 | 116/130 | 89 | |
| Brewerton Eastern Ext. | 107/130 | 82 | |
| C&D Approaches - Cores | 123/130 | 95 | |
| C&D Approaches – Surface Grabs | 103/130 | 79 | |
| Craighill | 123/130 | 95 | |
| Craighill Angle East | 122/130 | 94 | |
| Craighill Angle West | 128/130 | 98 | |
| Craighill Entrance | 124/130 | 95 | |
| Craighill Upper Range | 118/130 | 91 | |
| Cutoff Angle | 121/130 | 93 | |
| Swan Point | 110/130 | 85 | |
| Tolchester North | 127/130 | 98 | |
| Tolchester South | 123/130 | 95 | |
| Tolchester Straightening | 115/130 | 88 | |

*Total for 5 replicates of 26 animals each, except for 3 replicates of 26 for the laboratory control sediment

TABLE 9-5B RESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH Nereis virens (SAND WORM) ON SAMPLES FROM BALTIMORE HARBOR APPROAC CHANNELS AND PLACEMENT SITE 104 (Round 2: 13 January – 10 February 2000)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) |
|-----------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 58/60 | 97 |
| Inside Site 104 | 111/125 | 89 |
| Outside Site 104 | 104/125 | 83 |
| Brewerton Eastern Ext. | 91/100 | 91 |
| C&D Approaches - Cores | 93/100 | 93 |
| C&D Approaches – Surface Grabs | 96/100 | 96 |
| Craighill | 95/100 | 95 |
| Craighill Angle East | 88/100 | 88 |
| Craighill Angle West | 89/100 | 89 |
| Craighill Entrance | 89/100 | 89 |
| Craighill Upper Range | 96/100 | 96 |
| Cutoff Angle | 98/100 | 98 |
| Swan Point | 94/100 | 94 |
| Tolchester North | 94/100 | 94 |
| Tolchester South | 94/100 | 94 |
| Tolchester Straightening | 95/100 | 95 |

* Total for 3 replicates of 20 animals for laboratory control sediment, 5 replicates of 25 animals each for inside & outside site 104, 5 replicates of 20 animals for all other test sediments.

TABLE 9-5CRESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH
Nereis virens (SAND WORM) ON SAMPLES FROM THE NORFOLK OCEAN DISPOSAL
REFERENCE SITE (Round 3: 10 February – 8 March 2000)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) |
|--------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 72/75 | 96 |
| Ocean Reference | 120/125 | 96 |

*Total for 3 replicates of 25 animals each for laboratory control sediment and 5 replicates of 25 animals each for ocean reference.
TABLE 9-6A RESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH Macoma nasuta (BLUNT-NOSE CLAM) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: 4 November – 3 December 1999)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) |
|-----------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 139/150 | 93 |
| Inside Site 104 | 239/250 | 96 |
| Brewerton Eastern Ext. | 240/250 | 96 |
| C&D Approaches - Cores | 242/250 | 97 |
| C&D Approaches – Surface Grabs | 242/250 | 97 |
| Craighill | 243/250 | 97 |
| Craighill Angle East | 243/250 | 97 |
| Craighill Angle West | 241/250 | 96 |
| Craighill Entrance | 239/250 | 96 |
| Craighill Upper Range | 246/250 | 98 |
| Cutoff Angle | 235/250 | 94 |
| Swan Point | 237/250 | 95 |
| Tolchester North | 235/250 | 94 |
| Tolchester South | 233/250 | 93 |
| Tolchester Straightening | 234/250 | 94 |

*Total for 5 replicates of 50 animals each; except for laboratory control sediment 3 replicates of 50 animals each.

TABLE 9-6BRESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH
Macoma nasuta (BLUNT-NOSE CLAM) ON SAMPLES FROM BALTIMORE HARBOR
APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: 12 January – 9
February 2000)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) |
|-----------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 83/90 | 92 |
| Inside Site 104 | 139/150 | 93 |
| Outside Site 104 | 138/150 | 92 |
| Brewerton Eastern Ext. | 137/150 | 91 |
| C&D Approaches - Cores | 131/150 | 87 |
| C&D Approaches – Surface Grabs | 138/150 | 92 |
| Craighill | 143/150 | 95 |
| Craighill Angle East | 118/150 | 79 ^(a) |
| Craighill Angle West | 132/150 | 88 |
| Craighill Entrance | 121/150 | 81 ^(a) |
| Craighill Upper Range | 148/152 | 97 |
| Cutoff Angle | 142/150 | 95 |
| Swan Point | 130/150 | 87 |
| Tolchester North | 129/150 | 86 |
| Tolchester South | 130/150 | 87 |
| Tolchester Straightening | 137/150 | 91 |

*Survival was statistically (P=0.05) lower than Inside and Outside Site 104

TABLE 9-6C RESULTS OF 28-DAY BIOACCUMULATION TESTING (SURVIVAL) WITH Macoma nasuta (BLUNT-NOSE CLAM) ON SAMPLES FROM THE NORFOLK OCEAN DISPOSAL REFERENCE SITE (Round 3: 8 February – 7 March 2000)

| Test Sediment | No. Alive/ No. Exposed* | 28-Day Percent Survival (mean) |
|--------------------------------|----------------------------|--------------------------------------|
| Laboratory Control Sediment | 119/120 | 99 |
| Ocean Reference | 199/200 | 99 |



| <u>Test Species</u> | <u>Reference Toxicant</u> | <u>Testing Date</u> | <u>Endpoint</u> | <u>Acceptable Control</u> <u>Chart Limits</u> |
|---------------------|---------------------------|---------------------|-----------------------------|--|
| Nereis virens | Sodium dodecyl sulfate | November 1999 | 48-hour LC50: 13.0 mg/L SDS | 2.6 – 64.1 mg/L SDS |
| | (SDS) | January 2000 | 48-hour LC50: 17.7 mg/L SDS | 0 – 61.6 mg/L SDS |
| Macoma nasuta | Sodium dodecyl sulfate | November 1999 | 48-hour LC50: 90.0 mg/L SDS | 20.4 – 97.3 mg/L SDS |
| | (SDS) | January 2000 | 48-hour LC50: 68.0 mg/L SDS | 20.5 – 101.4 mg/L SDS |

TABLE 9-7B RESULTS OF REFERENCE TOXICANT BIOACCUMULATION TESTING – NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| <u>Test Species</u> | <u>Reference Toxicant</u> | Endpoint | <u>Acceptable Control</u> <u>Chart Limits</u> |
|---------------------|---------------------------|--|--|
| Nereis virens | Sodium dodecyl sulfate | (Lot NV-020) 48-hour LC50: 19.1 mg/L SDS | 0 – 60.1 mg/L SDS |
| | (SDS) | (Lot NV-021) 48-hour LC50: 14.5 mg/L SDS | 0 – 58.6 mg/L SDS |
| Macoma nasuta | Sodium dodecyl sulfate | (Lot MA-013) 48-hour LC50: 77.3 mg/L SDS | 20.5 – 101.4 mg/L SDS |
| | (SDS) | (Lot MA-014) 48-hour LC50: 77.3 mg/L SDS | 22.1 – 101.9 mg/L SDS |



TABLE 9-8A Nereis virens (SAND WORM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO INSIDE SITE 104

| Analyte (mg/kg) | I Site 104 | Outside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------|---------------|---------------------|--------------------|-----------------------------------|-----------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 8.61 | 8.54 | 19.20** | 8.35 | 14.60* | 8.30 | 9.66 | 8.80 | 20.88* | 8.16 | 12.20* | 32.92** | 8.75 | 10.96** | 12.34** | 11.28** |
| ANTIMONY | 0.05 | 0.05 | 0.14** | 0.09 | ND | 0.13** | 0.05 | 0.05 | 0.11* | 0.06 | 0.09 | 0.09 | 0.16* | 0.15** | 0.16** | 0.08 |
| ARSENIC | 1.96 | 1.94 | 4.28** | 1.70 | 1.70 | 1.55 | 2.18 | 1.50 | 1.67 | 1.62 | 1.83 | 1.70 | 1.26 | 1.56 | 1.70 | 1.47 |
| BERYLLIUM | ND | ND | ND | 0.07** | ND | 0.07** | ND | ND | 0.03** | ND | 0.03** | 0.03** | 0.07** | 0.03** | 0.03** | 0.03** |
| CADMIUM | 0.03 | 0.04* | 0.06** | ND | ND | 0.03 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 0.15 | ND | 1.36** | 0.53** | 0.15 | 0.45** | 0.22 | 0.18 | 0.32** | 0.13 | 0.49** | 0.50** | 0.47** | 0.44** | 0.79** | 0.44** |
| COPPER | 1.36 | 0.88 | 1.64 | 1.68* | 1.86* | 1.58 | 1.54 | 1.48 | 1.52 | 1.44 | 1.58 | 1.50 | 1.55 | 1.64 | 1.54 | 1.38 |
| IRON | 72.64 | 67.66 | 102.46 | 59.22 | 82.34 | 61.74 | 69.84 | 65.86 | 57.12 | 68.08 | 55.90 | 113.68 | 56.65 | 58.26 | 64.90 | 57.46 |
| LEAD | 0.32 | 0.30 | 0.43* | ND | 0.28 | 0.17 | ND | ND | ND | ND | 0.19 | 0.20 | 0.18 | ND | ND | ND |
| MANGANESE | 1.00 | 0.66 | 1.20 | 1.18 | 3.78** | 2.90* | 1.05 | 1.46 | ND | 1.40 | ND | 8.92* | 1.25 | ND | ND | ND |
| MERCURY | ND | 0.06 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NICKEL | 0.49 | ND | 1.26** | 1.00** | 0.65 | 1.11** | 0.92** | 0.62 | 0.61 | 0.52 | 0.87** | 0.94** | 0.73 | 0.85** | 0.82* | 0.90** |
| SELENIUM | 0.42 | 0.45 | 0.52 | ND | 0.42 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| SILVER | 0.08 | 0.09 | ND | 0.13 | 0.05 | 0.09 | ND | ND | 0.10 | ND | 0.06 | 0.07 | 0.15 | 0.10 | 0.09 | 0.11 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ZINC | 20.32 | 13.10 | 18.32 | 20.77 | 26.76 | 38.98 | 33.00 | 35.50 | 12.32 | 19.00 | 25.23 | 13.12 | 16.40 | 30.86 | 41.66 | 20.44 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-8B Nereis virens (SAND WORM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO OUTSIDE SITE 104

| Analyte (mg/kg) | Ontside Site 104 | Inside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Cralghill | Craighili Angle East | Craighlii Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolc hester Straightening |
|-----------------|---------------------|--------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|------------------------------|
| ALUMINUM | 8.54 | 8.61 | 19.20** | 8.35 | 14.6 | 8.3 | 9.66 | 8.8 | 20.88 | 8.16 | 12.20* | 32.92* | 8.75 | 10.96** | 12.34** | 11.28** |
| ANTIMONY | 0.05 | 0.05 | 0.14** | 0.09 | ND | 0.13** | 0.05 | 0.05 | 0,11* | 0.06 | 0.09 | 0.09 | 0.16* | 0.15** | 0.16** | 0.08 |
| ARSENIC | 1.94 | 1.96 | 4.28** | 1.7 | 1.7 | 1.55 | 2.18 | 1.5 | 1.67 | 1.62 | 1.83 | 1.7 | 1.26 | 1.56 | 1.7 | 1.47 |
| BERYLLIUM | ND | ND | ND | 0.07** | ND | 0.07** | ND | ND | 0.03** | ND | 0.03** | 0.03** | 0.07** | 0.03** | 0.03** | 0.03** |
| CADMIUM | 0.04 | 0.03 | 0.06 | ND | ND | 0.03 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | ND | 0.15* | 1.36** | 0.53** | 0.15** | 0,45** | 0.22* | 0.18** | 0.32** | 0.13** | 0.49** | 0.50** | 0.47** | 0.44** | 0.79** | 0.44** |
| COPPER | 0.88 | 1.36** | 1.64** | 1.68** | 1.86** | 1.58** | 1.54** | 1.48** | 1.52** | 1.44** | 1.58** | 1.50** | 1.55* | 1.64** | 1.54** | 1.38* |
| IRON | 67.66 | 72.64 | 102.46 | 59.22 | 82.34 | 61.74 | 69.84 | 65.86 | 57.12 | 68.08 | 55.9 | 113.68 | 56.65 | 58.26 | 64.9 | 57.46 |
| LEAD | 0.3 | 0.32 | 0.43* | ND | 0.28 | 0.17 | ND | ND | ND | ND | 0.19 | 0.2 | 0.18 | ND | ND | ND |
| MANGANESE | 0.66 | I | 1.2 | 1.18 | 3.78** | 2.90** | 1.05 | 1.46* | ND | 1.40* | ND | 8.92** | 1.25* | ND | ND | ND |
| MERCURY | 0.06 | NÐ | ND | ND | ND | ND | ND | ND | ND | ND | ND | NE | ND | ND | ND | ND |
| NICKEL | ND | 0.49* | 1.26** | 1.00** | 0.65** | -1.11** | 0.92** | 0.62* | 0.61* | 0.52* | 0.87** | 0.94** | 0.73** | 0.85** | 0.82** | 0.90** |
| SELENIUM | 0.45 | 0.42 | 0.52 | ND | 0.42 | ND | ND | ND | ND | ND | ND | NE | ND | ND | ND | ND |
| SILVER | 0.09 | 0.08 | ND | 0.13 | 0.05 | 0.09 | ND | ND | 0.1 | ND | 0.06 | 0.07 | 0.15 | 0.1 | 0.09 | 0.11 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | NE | ND | NE | ND | ND | ND | ND |
| ZINC | 13.1 | 20.32* | 18.32 | 20.77 | 26.76* | 38.98 | 33.00** | 35.5 | 12.32 | 19 | 25.23 | 13.12 | 16.4 | 30.86 | 41.66* | 20.44 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were significantly higher than Outside Site 104 (p < 0.05*, p < 0.01**).

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TABLE 9-8C Nereis virens (SAND WORM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (mg/kg) | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surfielal) | C&D Approach (Cores) | Craighill | Cralghill Angle East | Craighill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------|--------------------|---------------------|--------------------|-----------------------------------|-----------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 19.20 | 8.54 | 8.61 | 8.35 | 14.60 | 8.30 | 9.66 | 8.80 | 20.88 | 8.16 | 12.20 | 32.92 | 8.75 | 10.96 | 12.34 | 11.28 |
| ANTIMONY | 0.14 | 0.05 | 0.05 | 0.09 | ND | 0.13 | 0.05 | 0.05 | 0.11 | 0.06 | 0.09 | 0.09 | 0.16 | 0.15 | 0.16 | 0.08 |
| ARSENIC | 4.28 | 1.94 | 1.96 | 1.70 | 1.70 | 1.55 | 2.18 | 1.50 | 1.67 | 1.62 | 1.83 | 1.70 | 1.26 | 1.56 | 1.70 | 1.47 |
| BERYLLIUM | ND | ND | ND | 0.07** | ND | 0.07** | ND | ND | 0.03** | ND | 0.03** | 0.03** | 0.07** | 0.03** | 0.03** | 0.03** |
| CADMIUM | 0.06 | 0.04 | 0.03 | ND | ND | 0.03 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 1.36 | ND | 0.15 | 0.53 | 0.15 | 0.45 | 0.22 | 0.18 | 0.32 | 0.13 | 0.49 | 0.50 | 0.47 | 0.44 | 0.79 | 0.44 |
| COPPER | 1.64 | 0.88 | 1.36 | 1.68 | 1.86 | 1.58 | 1.54 | 1.48 | 1.52 | 1.44 | 1.58 | 1.50 | 1.55 | 1.64 | 1.54 | 1.38 |
| IRON | 102.46 | 67.66 | 72.64 | 59.22 | 82.34 | 61.74 | 69.84 | 65.86 | 57.12 | 68.08 | 55.90 | 113.68 | 56.65 | 58.26 | 64.90 | 57.46 |
| LEAD | 0.43 | 0.30 | 0.32 | ND | 0.28 | 0.17 | ND | ND | ND | ND | 0.19 | 0.20 | 0.18 | ND | ND | ND |
| MANGANESE | 1.20 | 0.66 | 1.00 | 1.18 | 3.78* | 2.90 | 1.05 | 1.46 | ND | 1.40 | ND | 8.92 | 1.25 | ND | ND | ND |
| MERCURY | ND | 0.06 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NICKEL | 1.26 | ND | 0.49 | 1.00 | 0.65 | 1.11 | 0.92 | 0.62 | 0.61 | 0.52 | 0.87 | 0.94 | 0.73 | 0.85 | 0.82 | 0.9 |
| SELENIUM | 0.52 | 0.45 | 0.42 | ND | 0.42 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| SILVER | ND | 0.09 | 0.08 | 0.13* | 0.05 | 0.09 | ND | ND | 0.10 | ND | 0.06 | 0.07 | 0.15* | 0.10* | 0.09 | 0.11 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ZINC | 18.32 | 13.10 | 20.32 | 20.77 | 26.76 | 38.98 | 33.00* | 35.50 | 12.32 | 19.00 | 25.23 | 13.12 | 16.40 | 30.86 | 41.66 | 20.44 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Cralghill | Craighill Angle East | Craighill Angle West | CraighIII Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Toichester North | Tolchester South | Tolchester Straightening |
|-----------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 0.13 | 0.41 | 0.28 | 0.17 | 0.30 | 0.17 | 0.2 | 0.18 | 0.43 | 0.17 | 0.25 | 0.68 | 0.18 | 0.23 | 0.25 | 0.23 |
| ANTIMONY | 0.99 | 0.69 | 0.73 | 1.61 | ND | 2.22 | 0.81 | 0.88 | 1.89 | 0.97 | 1.57 | 1.52 | 2:78 | 2.65 | 2.75 | 1.41 |
| ARSENIC | 2.47 | 1.16 | 1.27 | 1.18 | 1.18 | 1.08 | 1.51 | 1.04 | 1.15 | 1.12 | 1.26 | 1.18 | 0.87 | 1.08 | 1.18 | 1.02 |
| BERYLLIUM | ND | ND | ND | 6.75 | ND | 7.00 | ND | ND | 3:00 | ND | 2.75 | 3.40 | 7.00 | 3.00 | 3.00 | 3.00 |
| CADMIUM | 2.17 | 1.65 | 1.43 | ND | ND | 1.5 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 2.15 | ND | 0.84 | 1.84 | 0.53 | 1.56 | 0.75 | 0.64 | 1.12 | \$0.46 | 1.70 | 1.74 | 1.62 | 1.53 | 2.76 | 1.54 |
| COPPER | 0.85 | 1.39 | 1.39 | 1.28 | 1.42 | 1.21 | 1,18 | 1.13 | 1.16 | 1.10 | 1.21 | 1.15 | 1.19 | 1.26 | 1.18 | 1.06 |
| IRON | 0.3 | 0.81 | 0.66 | 0.41 | 0.57 | 0.43 | 0.49 | 0.46 | 0.4 | 0.47 | 0.39 | 0.79 | 0.39 | 0.41 | 0.45 | 0.4 |
| LEAD | 0.25 | 0.87 | 1.35 | ND | 1.47 | 0.9 | ND | ND | ND | ND | 0.98 | 1.04 | 0.95 | ND | ND | ND |
| MANGANESE | 0.24 | 0.62 | 0.5 | 0.44 | 1.43 | 1.09 | 0.4 | 0,55 | ND | 0.53 | ND | 3.37 | 0.47 | ND | ND | ND |
| MERCURY | ND | 1.22 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NICKEL | 2.94 | ND | 3.63 | 8.85 | 5.70 | 9.79 | 8.10 | 5.51 | 5.36 | 4.62 | 7:70 | 8:28 | 6:44 | 7.54 | 7.25 | 7.92 |
| SELENIUM | 1.62 | 1.84 | 1.22 | ND | 0.97 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| SILVER | ND | 0.78 | 0.68 | 1.00 | 0.42 | 0.69 | ND | ND | 0.75 | ND | 0.42 | 0.57 | 1.17 | 0.75 | 0.72 | 0.83 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ZINC | 1.69 | 0.58 | 1.33 | 1.64 | 2.11 | 3.08 | 2.61 | 2.8 | 0.97 | 1.5 | 1.99 | 1.04 | 1.29 | 2.44 | 3.29 | 1.61 |

TABLE 9-9 Nereis virens (SAND WORM): UPTAKE RATIOS FOR METALS

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



TABLE 9-10A Macoma nasuta (BLUNT-NOSE CLAM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO INSIDE SITE 104

| | 1.11.01 | 0.111 | | Brewerton | C&D | C&D | | | | | | | | | | |
|-----------------|-------------|------------|------------|-----------|-------------|---------------------|-----------|-------------------------|-------------------------|-------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| Analyte (mg/kg) | Inside Site | Site 104 | Reference | Extension | (Surficial) | Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Stralghtening |
| ALUMINUM | 32.36 | 30.40 | 32.86 | 67.44* | 41.12 | 42.12 | 48.42** | 45.82** | 51.42 | 38.34* | 36.02 | 30.08 | 38.10 | 36.94 | 27.08 | 47.2 |
| ANTIMONY | 0.04 | 0.10 | | 0.11** | ND | 0.14** | | 0.18** | ND | (C) 9:13* | ND | ND | 0.10** | 0.13** | 0.16** | 10 min 0,13** |
| ARSENIC | 2.39 | 2.13 | 4.28** | 2.60 | 2.58 | 2.26 | 2.78** | 2.54 | 2.58 | 2.80* | 2.48 | 2.78 | 2.54 | 2.40 | 2.30 | 2.44 |
| BERYLLIUM | 0.05 | ND | ND | 0.03 | 0.08** | 0.03 | ND | ND | 0.03 | ND | 0.03 | 0.03 | 0.03 | 0.04 | 0.04 | 0.04 |
| CADMIUM | 0.02 | 0.01 | | ND | 0.04 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 0.45 | ND | × # 1.70** | 0.76* | 0.58 | 0.66 | 2.51* | 2.2411 | 0.84** | 1.61 | 0.82** | 0.57 | 0.60 | 0.46 | 0.40 | 0.43 |
| COPPER | 2.47 | 2.55 | 4.06** | 2.66 | 2.56 | 2.04 | 2.72 | 2.38 | 2.50 | 2.78 | 2.74 | 2.32 | 2.36 | 2.48 | 1.94 | 2.14 |
| IRON | 174.47 | 119.25 | 120.80 | 172.60 | 200.20 | 196.00 | 232.00** | 222.40* | 219.80 | 188.20 | 173.00 | 187.40 | 161.60 | 161.20 | 121.60 | 167.4 |
| LEAD | 0.65 | PERS 0.95* | 0.51 | 0.35 | 0.38 | 0.17 | ND | 0.18 | 0.28 | 0.19 | 0.29 | 0.19 | ND | ND | ND | 0.25 |
| MANGANESE | 7.39 | 6.23 | 2.48 | 22:45** | 12.30** | 10.86* | 14.907* | 21.28** | 21.10** | 16.66** | 11.36** | 1632** | 11.34** | 10.04** | 8.98 | 4.52 |
| MERCURY | 0.08 | 0.06 | ND | 0.12* | 0.12* | 0.12 | ND | ND | ND | ND | ND | ND | 0.11 | ND | ND | ND |
| NICKEL | 0.73 | 0.30 | ·***** | 1.08* | 1.14** | 0.97 | 2.54** | 2.36** | 1.06* | Sec. 1.72** | | 0.87 | 0.87 | 0.86 | 0.72 | 0.93 |
| SELENIUM | 0.42 | 0.36 | 0.71** | | 0.45 | 0.65** | A.83** | | 0.62** | 9,74** | 0.52** | 0.57* | 0.67** | 0.73** | 9.69++ | 0.83** |
| SILVER | 0.10 | 0.14 | 0.05 | 0.13 | 0.23 | 0.11 | 0.13 | 0.08 | 0.05 | 0.11 | 0.06 | 0.18 | 0.10 | 0.11 | 0.18 | 0.13 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ZINC | 14.33 | 12.43 | 22.78** | 17.14 | 15.42 | 14.30 | 16.949 | 15.04 | 18.5414 | 17.96 | 16.98* | 19.62 | 14.30 | 15.22 | 14.16 | 14.08 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and boided cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-10BMacoma nasuta (BLUNT-NOSE CLAM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO
OUTSIDE SITE 104

| Analyte (mg/kg) | Outside Site | Inside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------|--------------|--------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 30.40 | 32.36 | 32.86 | 67.44* | 41.12 | 42.12 | 48.42** | 45.82* | 51.42 | 38.34 | 36.02 | 30.08 | 38.10 | 36.94 | 27.08 | 47.20* |
| ANTIMONY | 0.10 | 0.04 | 0.17 | 0.11 | ND | 0.14 | 0.19* | 0.18* | ND | 0.13 | ND | ND | 0.10 | 0.13 | 0.16 | 0.13 |
| ARSENIC | 2.13 | 2.39* | 4.28** | 2.60** | 2.58* | 2.26 | 2.78** | 2.54** | 2.58** | 2.80** | 2.48* | 2.78* | 2.54* | 2.40 | 2.30 | 2.44** |
| BERYLLIUM | ND | 0.05 | ND | 0.03** | 0.08** | 0.03** | ND | ND | 0.03** | ND | 0.03** | 0.03** | 0.03* | 0.04** | 0.04** | 0.04** |
| CADMIUM | 0.01 | 0.02* | 0.07** | ND | 0.04** | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | ND | 0.45* | 1.70** | 0.76** | 0.58** | 0.66** | 2.51** | 2.24** | 0.84** | 1.61** | 0.82** | 0.57** | 0.60** | 0.46** | 0.40** | 0.43** |
| COPPER | 2.55 | 2.47 | 4.06** | 2.66 | 2.56 | 2.04 | 2.72 | 2.38 | 2.50 | 2.78 | 2.74 | 2.32 | 2.36 | 2.48 | 1.94 | 2 14 |
| IRON | 119.25 | 174.47* | 120.80 | 172.60* | 200.20** | 196.00** | 232.00** | 222.40** | 219.80* | 188.20** | 173.00* | 187.40** | 161.60* | 161.20* | 121.60 | 167.40* |
| LEAD | 0.95 | 0.65 | 0.51 | 0.35 | 0.38 | 0.17 | ND | 0.18 | 0.28 | 0.19 | 0.29 | 0.19 | ND | ND | ND | 0.25 |
| MANGANESE | 6.23 | 7.39 | 2.48 | 22.46** | 12.30** | 10.86** | 14.90** | 21.28** | 21.10** | 16.66** | 11.36** | 16.32** | 11.34** | 10.04** | 8.98* | 4.52 |
| MERCURY | 0.06 | 0.08 | ND | 0.12* | 0.12* | 0.12* | ND | ND | ND | ND | ND | ND | 0.11* | ND | ND | ND |
| NICKEL | 0.30 | 0.73** | 1.44** | 1.08** | 1.14** | 0.97** | 2.54** | 2.36** | 1.06** | 1.72** | 1.14** | 0.87** | 0.87** | 0.86** | 0.72** | 0.93** |
| SELENIUM | 0.36 | 0.42 | 0.71** | 0.80** | 0.45 | 0.65** | 0.83** | · 0.82** | 0.62** | 0.74** | 0.62** | 0.57* | 0.67** | 0.73** | 0.69** | 0.83** |
| SILVER | 0.14 | 0.10 | 0.05 | 0.13 | 0.23 | 0.11 | 0.13 | 0.08 | 0.05 | 0.11 | 0.06 | 0.18 | 0.10 | 0.11 | 0.18 | 0.13 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ZINC | 12.43 | 14.33* | 22.78** | 17.14** | 15.42** | : 14.30* | 16.94** | 15.04* | 18.54** | 17.06** | 16.98** | 19.62* | 14.30 | 15.22* | 14.16 | 14.08* |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).



TABLE 9-10C Macoma nasuta (BLUNT-NOSE CLAM): MEAN METAL CONCENTRATIONS (MG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (mg/kg) | Ocean Reference | Outside Site | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------|--------------------|--------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 32.86 | 30.40 | 32.36 | . 67.44* | 41.12 | 42.12 | 48.42 | 45.82 | 51.42 | 38.34 | 36.02 | 30.08 | 38.10 | 36.94 | 27.08 | 47.2 |
| ANTIMONY | 0.17 | 0.10 | 0.04 | 0.11 | ND | 0.14 | 0.19* | 0.18 | ND | 0.13 | 8 ND | ND | 0.10 | 0.13 | 0.16 | 0.13 |
| ARSENIC | 4.28 | 2.13 | 2.39 | 2 60 | 2.58 | 2 26 | 2.78 | 2.54 | 2.58 | 2.80 | 2.48 | 2.78 | 2.54 | 2.40 | 2.30 | 2.44 |
| BERYLLIUM | ND | ND | 0.05 | 0.03** | 0.08** | 0.03** | ND | ND | 0.03** | ND | 0.03** | 0.03** | 0.03* | 0.04** | 0.04** | 0.04** |
| CADMIUM | 0.07 | 0.01 | 0.02 | ND | 0.04 | ND | ND | ND | ND | NE | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 1.70 | ND | 0.45 | 0.76 | 0.58 | 0.66 | 2.51 | 2.24* | 0.84 | 1.61 | 0.82 | 0.57 | 0.60 | 0.46 | 0.40 | 0.43 |
| COPPER | 4.06 | 2.55 | 2.47 | 2.66 | 2.56 | 2.04 | 2.72 | 2.38 | 2.50 | 2.78 | 3 2.74 | 2.32 | 2.36 | 2.48 | 1.94 | 2.14 |
| IRON | 120.80 | 119.25 | 174.47* | 172.60** | 200.20** | 196.00** | 232.00** | 222.40** | 219.80* | 188.20** | 173.00* | 187.40** | 161.60* | 161.20** | 121.60 | 167.40** |
| LEAD | 0.51 | 0.95** | 0.65 | 0.35 | 0.38 | 0.17 | ND | 0.18 | 0.28 | 0.19 | 0.29 | 0.19 | ND | ND | ND | 0.25 |
| MANGANESE | 2.48 | 6.23** | 7.39** | 22.46** | 12.30** | 10.86** | 14.90*1 | 21.28** | 21.10** | 16.66** | 11.36** | 16.32** | 11.34** | 10.04** | 8.98** | 4.52* |
| MERCURY | ND | 0.06 | 0.08* | 0.12** | 0.12** | 0.12** | NE | ND | ND | NE | ND | ND | 0.11** | ND | ND | ND |
| NICKEL | 1.44 | 0.30 | 0.73 | 1.08 | 1.14 | 0.97 | 2.54 | 2.36** | 1.06 | 1.72 | 2 1.14 | 0.87 | 0.87 | 0.86 | 0.72 | 0.93 |
| SELENIUM | 0.71 | 0.36 | 0.42 | 0.80 | 0.45 | 0.65 | 0.83 | 0.82* | 0.62 | 0.74 | 0.62 | 0.57 | 0.67 | 0.73 | 0.69 | 0.83* |
| SILVER | 0.05 | 0.14 | 0.10 | 0.13 | 0.23* | 0.11* | 0.13* | 0.08 | 0.05 | 0.11 | 0.06 | 0.18 | 0.10 | 0.11 | 0.18* | 0.13 |
| THALLIUM | ND | ND | ND | ND | ND | ND | NE | ND | ND | NE | ND | ND | ND | ND | ND | ND |
| ZINC | 22.78 | 12.43 | 14.33 | 17.14 | 15.42 | 14.30 | 16.94 | 15.04 | 18.54 | 17.00 | 5 16.98 | 19.62 | 14.30 | 15.22 | 14.16 | 14.08 |

ND = not detected in any of five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Cralghill | Craighlll Angle East | Cralghill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--|-----------------|---------------|---------------------|---------------------|-----------------------------|
| ALUMINUM | 1.74 | 3.28 | 3.57 | 1 | 4.63 | 4.74 | 5.45 | 5.16 | 5.79 | 4.32 | 4.06 | 3.39 | 4.29 | 4.16 | 3.05 | 5.32 |
| ANTIMONY | 2.93 | 1.54 | 0.88 | 3.03 | ND | 4:00 | 5.49 | 5.09 | ND | 3.74 | ND | ND | 2.91 | 3.66 | 4.57 | 3.66 |
| ARSENIC | 1.48 | 0.77 | 0.93 | 1.07 | 1.07 | 0.93 | 1.15 | 1.05 | 1.07 | 1.16 | ······································ | 1.15 | * 1.05 | 0.99 | 0.95 | 1.01 |
| BERYLLIUM | ND | ND | 0.99 | 0.47 | 1.06 | 0.44 | ND | ND | 0.47 | ND | 0.47 | 0.47 | 0.42 | 0.50 | 0.53 | 0.50 |
| CADMIUM | 1.4 000 4.08 | 0.54 | 0.63 | ND | 0.76 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| CHROMIUM | 17.00 | ND | 1.80 | 1.92 | 1.48 | 1.69 | 6.38 | 5.69 | 2.13 | 4.08 | 2.08 | 144 | 1.52 | 1:16 | 1.02 | 1.09 |
| COPPER | 0.88 | 0.8 | 1.09 | 1.48 | 1.42 | 1.13 | 1.51 | 1.32 | 1.39 | 1.54 | 1.52 | 1.29 | 1.31 | 1.38 | 1.08 | 1.19 |
| IRON | 1.41 | 2.14 | 2.84 | 2.63 | 3.05 | 2.98 | 3.53 | 3.39 | 3.35 | 2.87 | 2.63 | 2.85 | 2.46 | 2.45 | 1.85 | 2.55 |
| LEAD | 1.07 | 3.02 | 2.94 | 1.96 | 2.16 | 0.99 | ND | 1.02 | 1.57 | 1.06 | 1.65 | 1.1 | ND | ND | ND | 1.44 |
| MANGANESE | 2.48 | 6.22 | 5.93 | 14.97 | 8.20 | 7.24 | 9.93 | 14.19 | 14.07 | 11.11 | 7.57 | 10.88 | 7.56 | 6.69 | 5.99 | 3.01 |
| MERCURY | ND | 1.34 | 0.99 | 0.97 | 0.95 | 0.95 | ND | ND | ND | ND | ND | ND | 0.93 | ND | ND | ND |
| NICKEL | 4.70 | 1.03 | 1.45 | 1.69 | 1.79 | 1:52 | 3.99 | 3.71 | 1.67 | 2.70 | 1.79 | 1.36 | 136 | 1.35 | 1.13 | 1.47 |
| SELENIUM | 1.58 | 0.63 | 0.9 | 1.93 | 1.09 | 1.56 | 1.99 | 1.97 | 1.50 | 1177 | 1.50 | 1.38 | 1.61 | 1.76 | 1.65 | 1.99 |
| SILVER | 0.81 | 0.69 | 0.81 | 2.63 | 4.88 | 2.25 | 2.75 | 1.63 | 1.04 | 2.33 | 1.33 | 3.67 | 2.13 | 2.29 | 3.71 | 2.67 |
| THALLIUM | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND |
| ZINC | 1.36 | 0.88 | 0.94 | 1.05 | AME 10.94 | 0.87 | Ceta 1.03 | 0.92 | 72 113 | 1.04 | 1.04 | 1.20 | 0.87 | 0.93 | 0.86 | 0.86 |

TABLE 9-11 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR METALS

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



 TABLE 9-12A
 Nereis virens (SAND WORM):
 MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG)

 COMPARED TO INSIDE SITE 104

| Analyte (ug/kg) | Inside Site 104 | Outside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Cralghilt Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| 4,4'-DDD | 2.50 | ND | 2.05 | ND | 5.88** | ND | 2.23 | 1.65 | 1.11 | 2.39 | 1.55 | 2.75 | ND | 13.74** | 10.32* | 10.73** |
| 4,4'-DDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4'-DDT | ND | ND | ND | ND | 5.72 | ND | ND | ND | 1.00 | ND | ND | ND | ND | 2.20* | 1.28 | ND |
| ALDRIN | 13.83 | ND | 4.90 | ND | 28.60* | ND | 0.43 | 0.35 | 3.35 | 0.57 | 2.87 | ND | ND | ND | ND | ND |
| ALPHA-BHC | 2.01 | 0.48 | 0.91 | ND | 3.54* | 0.87 | 0.56 | 0.51 | 0.49 | 0.48 | ND | ND | 1.90 | ND | 0.66 | ND |
| BETA-BHC | 0.87 | 0.74 | ND | 0.63 | ND | 0.69 | 0.97 | 1.55 | 4.86** | 1.07 | 1.72 | 2.31 | 1.22 | 3.71 | ND | 2.14 |
| CHLORBENSIDE | 14.92 | ND | 4.84 | 3.32 | 35.20** | ND | 9.07 | 2.37 | 8.60 | 15.40 | 8.35 | 14.80 | ND | 9.22 | 15.66 | 13.4 |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | 25.21 | ND | 25.25 | 50.60 | 116.80** | 41.00* | 4.84 | 10.46 | 45.46 | ND | 40.00* | 76.00** | 37.50 | 44.00** | 41.20* | 50.50** |
| DELTA-BIIC | ND | ND | 2.82**B | 2.02** | 1.64 | 2.70** | 1.66 | ND | 0.79 | 1.80**B | ND | ND | 2.63** | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | 1.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.55 | 0.56 |
| ENDOSULFAN I | 2.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.60 | 1.52 | ND |
| ENDOSULFAN II | 7.91 | 55.25** | 15.28** | 17.68** | 8.09 | 16.78** | ND | 3.65 | 1.52 | 2.61 | 2.12 | 3.16 | -24,25** | 9.88* | 7.20 | 6.59 |
| ENDOSULFAN SULFATE | ND | ND | 1.36 | ND | 0.70 | ND | ND | ND | ND | ND | ND | 0.70 | ND | 0.74 | ND | ND |
| ENDRIN | ND | ND | 1.54 | ND | ND | ND | 0.79 | ND | ND | ND | 1.27* | 0.65 | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | ND | 0.82 | NÐ | 0.74 | ND | ND | 0.88 | 1.00 | ND | ND | 0.93 | ND | ND | ND |
| GAMMA-BHC | 1.37 | ND | 3.00** | 0.57 | 3.16* | ND | 0.79 | ND | 1.80 | 1.04 | 1.58 | 2.00* | ND | ND | ND | ND |
| HEPTACHLOR | 1.17 | 1.73* | ND | 3.37* | ND | 4,59* | ND | ND | ND | ND | ND | ND | 5.27** | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 0.73 | 0.48 | 2.92 | 2.04** | ND | 3.52** | ND | ND | ND | ND | ND | ND | 2.45** | 2.45* | 1.36 | 1.57 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of five tested replicates.

Asterisks, shaded and boided cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-12B Nereis virens (SAND WORM): MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

| Analyte (ug/kg) | Outside Site | Inside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Cralghill Angle East | Craighill Angle West | Cralghill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester Snuth | Tolchester Straightening |
|--------------------|--------------|--------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| 4,4'-DDD | ND | 2.5 | 2.05 | ND | 5.88** | ND | 2.23 | 1.65 | 1.11 | 2.39* | 1.55 | 2.75* | ND | 13.74** | 10.32** | 10.73** |
| 4,4'-DDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4'-DDT | ND | ND | ND | ND | 5.72 | ND | ND | ND | 1 | ND | ND | ND | ND | 2.2 | 1.28 | ND |
| ALDRIN | ND | 13.83* | 4.90** | ND | 28.60** | ND | 0.43 | 0.35 | 3.35* | 0.57* | 2.87* | ND | ND | ND | ND | ND |
| ALPHA-BIIC | 0.48 | 2.01* | 0.91 | ND | 3.54** | 0.87 | 0.56 | 0.51 | 0.49 | 0.48 | ND | ND | 1.9 | ND | 0.66 | ND |
| BETA-BHC | 0.74 | 0.87 | ND | 0.63 | ND | 0.69 | 0.97 | 1.55 | 4.86** | 1.07 | 1.72 | 2.31 | 1.22 | 3.71 | ND | 2.14 |
| CHLORBENSIDE | ND | 14.92* | 4.84* | 3.32 | 35.20** | ND | 9.07* | 2.37 | *** 8.60** | 15.40** | 8.35** | 14.80** | ND | 9.22 | 15.66 | 13.40** |
| CIILORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | 25.21 | 25.25** | 50.60* | 116.80** | 41.00** | 4.84 | 10.46 | 45.46 | ND | 40.00** | 76.00** | 37.50** | 44.00** | 41.20** | 50.50** |
| DELTA-BIIC | ND | ND | 2.82**B | 2.02** | 1.64 | 2.70** | 1.66 | ND | 0.79 | 1.80**B | ND | ND | 2.63** | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | 1.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.55 | 0.56 |
| ENDOSULFAN I | ND | 2.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.6 | 1.52 | ND |
| ENDOSULFAN II | 55.25 | 7.9 i | 15.28 | 17.68 | . 8.09 | 16.78 | ND | 3.65 | 1.52 | 2.61 | 2.12 | 3.16 | 24.25 | 9.88 | 7.2 | 6.59 |
| ENDOSULFAN SULFATE | ND | ND | 1.36 | ND | 0.7 | ND | ND | ND | ND | ND | ND | 0.7 | ND | 0.74 | ND | ND |
| ENDRIN | ND | ND | 1.54 | ND | ND | ND | 0.79 | ND | ND | ND | 1.27* | 0.65 | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | NĐ | 0.82 | ND | 0.74 | ND | ND | 0.88 | 1 | ND | ND | 0.93 | ND | ND | ND |
| GAMMA-BIIC | ND | 1.37 | 3.00** | 0.57 | 3.16** | ND | 0.79 | ND | 1.80* | 1.04* | 1.58** | 2.00** | ND | ND | ND | ND |
| HEPTACIILOR | 1.73 | 1.17 | ND | 3.37 | ND | 4.59 | ND | ND | ND | ND | ND | ND | 5.27* | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 0.48 | 0.73 | 2.92 | > 2.04** | ND | 3.52** | ND | ND | ND | ND | ND | ND | 4 2.45** | alle 2.45ª | 1.36 | 1.57 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).





TABLE 9-12C Nereis virens (SAND WORM): MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (ug/kg) | Ocean Rcference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| 4.4'-DDD | 2.05 | ND | 2.50 | ND | 5,88* | ND | 2.23 | 1.65 | 1.11 | 2.39 | 1.55 | 2.75 | ND | 13.74** | 10.32* | 10.73** |
| 4.4'-DDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4.4'-DDT | ND | ND | ND | ND | 5.72 | ND | ND | ND | 1.00 | ND | ND | ND | ND | 2.20 | 1.28 | ND |
| ALDRIN | 4.90 | ND | 13.83 | ND | 28.60** | ND | 0.43 | 0.35 | 3.35 | 0.57 | 2.87 | ND | ND | ND | ND | ND |
| ALPHA-BHC | 0.91 | 0.48 | 2.01 | ND | 3.54* | 0.87 | 0.56 | 0.51 | 0.49 | 0.48 | ND | ND | 1.90 | ND | 0.66 | ND |
| BETA-BHC | ND | 0.74 | 0.87 | 0.63 | ND | 0.69 | 0.97 | 1.55 | 4.86** | 1.07 | 1.72 | 2.31 | 1.22* | 3.71 | ND | 2.14 |
| CIILORBENSIDE | 4.84 | ND | 14.92 | 3.32 | 35.20** | ND | 9.07 | 2.37 | 8.60 | 15.40* | 8.35 | 14.80* | ND | 9.22 | 15.66 | 13.40* |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | 25.25 | ND | 25.21 | 50.60 | 116.80** | 41.00* | 4.84 | 10.46 | 45.46 | ND | 40.00 | 76.00** | 37.50 | 44.00** | 41.20* | 50.50** |
| DELTA-BHC | 2.82 | ND | ND | 2.02 | 1.64 | 2.70 | 1.66 | ND | 0.79 | 1.80 | ND | ND | 2.63 | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | 1.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.55 | 0.56 |
| ENDOSULFAN I | ND | ND | 2.55 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.60 | 1.52 | ND |
| ENDOSULFAN II | 15.28 | 55.25** | 7.91 | 17.68 | 8.09 | 16.78 | ND | 3.65 | 1.52 | 2.61 | 2.12 | 3.16 | 24.25 | 9.88 | 7.20 | 6.59 |
| ENDOSULFAN SULFATE | 1.36 | ND | ND | ND | 0.70 | ND | ND | ND | ND | ND | ND | 0.70 | ND ND | 0.74 | ND | ND |
| ENDRIN | 1.54 | ND | ND | ND | ND | ND | 0.79 | ND | ND | ND | 1.27 | 0.65 | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | ND | 0.82 | ND | 0.74 | ND | ND | 0.88 | 1.00 | ND | ND | 0.93 | ND | ND | ND |
| GAMMA-BIIC | 3.00 | ND | 1.37 | 0.57 | 3.16 | ND | 0.79 | ND | 1.80 | 1.04 | 1.58 | 2.00 | ND | ND | ND | ND |
| HEPTACHLOR | ND | 1.73** | 1.17 | 3.37* | ND | 4.59 | ND | ND | ND | ND | ND | NE | 5.27** | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 2.92 | 0.48 | 0.73 | 2.04 | ND | 3.52 | ND | ND | ND | ND | ND | NE | 2.45 | 2.45 | 1.36 | 1.57 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NE | ND ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NE | ND ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NE | ND ND | ND | ND | ND |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighiil Entrance | Craighill Upper Range | Cutoff Angie | Swan Point | Toichester North | Tolchester South | Toichester Straightening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| 4,4'-DDD | 2.8 | ND | 3.33 | ND | 7.84 | ND | 2.97 | 2.2 | 1.48 | 3.19 | 2.07 | 3.67 | ND | 18.32 | 13.76 | 14.30 |
| 4,4'-DDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4'-DDT | ND | ND | ND | ND | 2.77 | ND | ND | ND | 0.48 | ND | ND | ND | ND | 1:06 | 0.62 | ND |
| ALDRIN | 4.88 | ND | 2.11 | ND | 2.25 | ND | 0.03 | 0.03 | 0.26 | 0.04 | 0.23 | ND | ND | ND | ND | ND |
| ALPHA-BHC | 1.65 | 0.12 | 4.12 | ND | 9.83 | 2.41 | 1.54 | 1.41 | 1.36 | 1.33 | ND | ND | 5.26 | ND | 1.82 | ND |
| BETA-BHC | ND | 0.54 | 1.49 | 1.63 | ND | 1.79 | 2.51 | 4.03 | 12.62 | 2.78 | 4.46 | 6 | 2.23.17 | 9.63 | ND | 5.56 |
| CHLORBENSIDE | 2.03 | ND | 2.15 | 0.34 | 3.60 | ND | 0.93 | 0.24 | 0,88 | 1.57 | 0.85 | 1.51 | ND | 0.94 | 1.6 | 1.37 |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | 0.79 | ND | 0.97 | 1.64 | 3.80 | 1:33 | 0.16 | 0.34 | 1.48 | ND | 1.30 | 2.47 | 1.22 | 1,43 | 1.34 | 1.64 |
| DELTA-BHC | 2.76 | ND | ND | 0.36 | 0.29 | 0.49 | 0.3 | ND | 0.14 | 0.32 | ND | ND | 0.47 | ND | ND | ND |
| DIELDRIN | ND | ND | ND | ND | 0.93 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.33 | 0.34 |
| ENDOSULFAN 1 | ND | ND | 8.36 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 8.53 | 4.99 | ND |
| ENDOSULFAN II | 8.93 | 64.24 | 8.72 | 3.17 | 1.45 | 3.01 | ND | 0.66 | 0.27 | 0.47 | 0.38 | 0.57 | 4.35 | 1.77 | 1.29 | 1.18 |
| ENDOSULFAN SULFATE | 2.96 | ND | ND | ND | 0.79 | ND | ND | ND | ND | ND | ND | 0.79 | ND | 0.84 | ND | ND |
| ENDRIN | 2.15 | ND | ND | ND | ND | ND | 0.11 | ND | ND | ND | 0.17 | 0.09 | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | ND | ND | 1.37 | ND | 1.23 | ND | ND | 1.47 | 1.67 | ND | ND | 1.54 | ND | ND | ND |
| GAMMA-BHC | 7.56 | ND | 0.9 | 0.22 | 1.20 | ND | 0.3 | ND | 0.68 | 0.40 | 0.60 | 0.76 | ND | ND | ND | ND |
| HEPTACHLOR | ND | 0.70 | 0.95 | 9.91 | ND | 13.49 | ND | ND | ND | ND | ND | ND | 15.51 | ND | ND | ND |
| HEPTACHLOR EPOXIDE | 2.96 | 0.62 | 1.43 | 5.59 | ND | 9.64 | ND | ND | ND | ND | ND | ND | 6.71 | 6.72 | 3.72 | 4.29 |
| METHOXYCHLOR | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

TABLE 9-13 Nereis virens (SAND WORM): UPTAKE RATIOS FOR CHLORINATED PESTICIDES

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.





TABLE 9-14A Macoma nasuta (BLUNT-NOSE CLAM): MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| Anslyte (ug/kg) | Inside Site | Outside Site 104 | Ocean | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach | Craisbill | Craighill Apple Fast | Craighill | Craighli | Craighill | Cutoff | Suga Balat | Tolchester | Tolchester | Tolchester |
|--------------------|-------------|---------------------|---------|-----------------------------------|--------------------------------|-----------------|-----------|-------------------------|------------|----------|-------------|--------|------------|------------|------------|---------------|
| | 0.06 | 1.04 | ND | LACCISION 140 | ND | 1.0200 | LTAIght | Angle Last | Angre West | Entrance | Upper Range | Angie | Swan Point | North | South | Straigntening |
| 4,4-DDD | 0.90 | 1.04 | ND | 1.47 ND | ND | 4.44 | ND | 1.31 | 1.12 | 1.10 | 11/0-1 | ND ND | 1.47 | ND | ND | ND |
| 4,4-DDE | 0.73 | ND | ND | ND | 110 | ND | 0.70 | ND | ND | ND | ND | NU | NU | NU | ND | ND |
| 4,4-DD1 | 1.4/ | ND | ND | | 0.94 | | 0.78 | 1.88 | - 4.L.P | 1.48 | 344 | 3.04** | 5.08 | ND | ND | ND |
| ALDKIN | 2.69 | 1.04 | 0.71 | ND | 3.13 | 1.66 | 2.15 | 3.16 | 1.52 | 1.89 | 2.87 | 1.67 | 1.44 | 1.22 | 0.86 | 0.8 |
| ALPHA-BHC | UN | ND | ND | ND | ND | 1032-0 | ND | 0.55 | | 0.51 | 0.63 | ND | 1.05 | ND | ND | ND |
| BETA-BHC | 8.27 | 7.29 | 0.99 | 4.57 | 4.32 | 3.78 | 3.25 | 8.40 | 5.98 | 6.79 | 2.01 | 1.77 | ND | 2.48 | 1.00 | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.78 | ND | ND | ND |
| CHLORDANE | ND | ND | ND. | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | 3.54 | ND | ND | ND | ND | ND | ND | ND | ND | 2.71 | 3.70 | 4.13 | ND | 2.49 | ND | ND |
| DELTA-BHC | 1.45 | 2.57 | 2.84**B | ND | ND | ND | 0.62 | ND | ND | ND | ND | ND | ND | 1.07 | 1.85 | ND |
| DIELDRIN | 0.58 | 0.48 | 0.93* | ND | ND | 0.53 | 0.73 | 1.15 | 0.93 | 1.15 | ND | ND | ND | 0.81 | 1.20* | 38.35 * 1.20* |
| ENDOSULFAN 1 | 2.01 | 5.88** | ND | 1.56 | 0.88 | 3.59 | 0.50 | ND | ND | 1.18 | ND | ND | 0.90 | ND | ND | ND |
| ENDOSULFAN 11 | 0.81 | ND | ND | 4.44** | 1.44 | 2.13* | ND | ND | ND | ND | ND | ND | 3.15* | ND | ND | ND |
| ENDOSULFAN SULFATE | 0.61 | ND | 0.62 | 2.83 | 1.26 | 1.26 | 1.11 | ND | 1.36 | 1.58 | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | 0.61 | ND | ND | ND | ND | ND | ND | 0.63 | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | 0.73 | ND | 0.86 | 0.90 | 1.16 | ND | 0.74 | 1.20 | ND | 1.02 | 1.38 | 0.82 | ND | 0.80 | 0.76 | 0.8 |
| GAMMA-BHC | ND | ND | 0.72 | ND | ND | ND | 0.51 | ND | 0.57 | ND | 1.48** | 2.00** | ND | 0.55 | 0.59 | 0.55 |
| HEPTACHLOR | 2.58 | ND | 2.00 | 1.78 | 3.86 | 2.58 | ND | ND | 2.48 | ND | 3.88 | 4.20*B | 3.73 | ND | ND | 0.79 |
| HEPTACHLOR EPOXIDE | 1.29 | 3.00** | 0.82 | ND | ND | 0.49 | ND | ND | 2.31 | ND | 1.11 | ND | ND | 0.49 | 1.02 | 0.56 |
| METHOXYCHLOR | 2.45 | 7.20 | ND | ND | ND | ND | ND | ND | 2.06 | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ' ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

B = detected in laboratory blank in at least one of five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 ($p < 0.05^{+}$, $p < 0.01^{+*}$).

TABLE 9-14B Macoma nasuta (BLUNT-NOSE CLAM): MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

| Analyte (ug/kg) | Outside | Inside Site | Ocean | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach | Cralabill | Craighili Angle Fast | Craighill Angle | Cralghill | Craighill Linner Bange | Cutoff Angle | Swan Point | Tolchester | Tolchester | Thichester |
|--------------------|---------|-------------|----------|-----------------------------------|--------------------------------|-----------------|-----------|-------------------------|-----------------|-----------|---------------------------|--------------|------------|------------|------------|------------|
| | 1.04 | 0.96 | ND | 1 49 | ND | 2 0744 | ND | Augic Dast | 1 12 | L 10 | 1 76* | 1 00* | 147 | ND | ND | ND |
| 4.4'-DDF | ND | 0.75 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4 4'-DDT | ND | 147 | ND | 3.64** | 0.94 | 1.96** | 0.78 | 1 88 | 4.12** | 1 48 | 3.3244 | 3.02** | 3.084+ | ND | ND | ND |
| ALDRIN | 1.64 | 2 69 | 0.71 | ND | 3.13 | 1.66 | 2.15 | 3.16 | 1.52 | 1.89 | 2.87 | 1.67 | 44 | 1.22 | 0.86 | 0.8 |
| ALPHA-BHC | ND | ND | ND | ND | ND | 10.32** | ND | 0.55 | 1.94 | 0.51 | 0.63 | ND | 1.05 | ND | ND | ND |
| BETA-BHC | 7.29 | 8.27 | 0.99 | 4.57 | 4.32 | 3.78 | 3.25 | 8.40 | 5.98 | 6.79 | 2.01 | 1.77 | ND | 2.48 | 1.00 | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.78 | ND | ND | ND |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | 3.54 | ND | ND | ND | ND | ND | ND | ND | 2.71 | 3.70 | 4.13 | ND | 2.49 | ND | ND |
| DELTA-BHC | 2.57 | 1.45 | 2.84 | ND | ND | ND | 0.62 | ND | ND | ND | ND | ND | ND | 1.07 | 1.85 | ND |
| DIELDRIN | 0.48 | 0.58 | ·· 0.93* | ND | ND | 0.53 | 0.73 | 1.15 | 0.93 | 1.15 | ND | ND | ND | 0.81 | 1.20* | 1.20* |
| ENDOSULFAN I | 5.88 | 2.01 | ND | 1.56 | 0.88 | 3.59 | 0.50 | ND | ND | 1.18 | ND | ND | 0.90 | ND | ND | ND |
| ENDOSULFAN II | ND | 0.81 | ND | 4,44** | 1.44 | 2.13* | ND | ND | ND | ND | ND | ND | 3.25* | ND | ND | ND |
| ENDOSULFAN SULFATE | ND | 0.61 | 0.62 | 2.83 | 1.26 | 1.26 | 1.11 | ND | 1.36 | 1.58 | ND | ND | ND | ND | ND | ND |
| ENDRIN | 0.61 | ND | ND | ND | ND | ND | ND | ND | 0.63 | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | ND | 0.73 | 0.86 | 0.90 | 1.16 | ND | 0.74 | 1.20 | ND | 1.02 | 1.38 | 0.82 | ND | 0.80 | 0.76 | 0.8 |
| GAMMA-BHC | ND | ND | 0.72 | ND | ND | ND | 0.51 | ND | 0.57 | ND | 1.48* | 2.00** | ND | 0.55 | 0.59 | 0.55 |
| HEPTACHLOR | ND | 2.58*B | 2.00 | 1.78 | 3.86**B | 2.58 | ND | ND | 2.48 | ND | 3.88**B | 4.20**B | 3.73*B | ND | ND | 0.79 |
| HEPTACHLOR EPOXIDE | 3.00 | 1.29 | 0.82 | ND | ND | 0.49 | ND | ND | 2.31 | ND | 1.11 | ND | ND | 0.49 | 1.02 | 0.56 |
| METHOXYCHLOR | 7.20 | 2.45 | ND | ND | ND | ND | ND | ND | 2.06 | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).



TABLE 9-14C Macoma nasuta (BLUNT-NOSE CLAM): MEAN CHLORINATED PESTICIDE CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE COMPARED TO OCEAN REFERENCE

| Analyte (ug/kg) | Ocean Reference | Outside Site | Inside Site | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Cralghill Upper Range | Cutoff | Swan Point | Tolchester | Tolchester South | Tolchester |
|---------------------|--------------------|--------------|-------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|--------|------------|------------|---------------------|------------|
| 4,4'-DDD | ND | 1.04 | 0.96 | 1.49 | ND | 2.02** | ND | 1.31 | 1.12 | 1.10 | 1.76** | 1.99* | 1.47 | ND | ND | ND |
| 4.4'-DDE | ND | ND | 0.75 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4.4'-DDT | ND | ND | 1.47 | 3.64** | 0.94 | 1.96** | 0.78 | 1.88 | 4.12** | 1.48 | 3.32** | 3.02** | 3.08** | ND | ND | ND |
| ALDRIN | 0.71 | 1.64** | 2.69** | ND | 3.13* | 1.66 | 2.15* | 3.16** | 1.52 | 1.89* | 2.87 | 1.67 | 1.44 | 1.22 | 0.86 | 0.8 |
| ALPHA-BHC | ND | ND | ND | ND | ND | 10.32** | ND | 0.55 | 1.94 | 0.51 | 0.63 | ND | 1.05 | ND | ND | ND |
| BETA-BIIC | 0.99 | 7.29* | 8.27** | 4.57 | 4.32 | 3.78 | 3.25* | 8.40** | 5.98** | 6.79 | 2.01 | 1.77 | ND | 2.48 | 1.00 | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.78 | ND | ND | ND |
| CIILORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | ND | 3.54* | ND | ND | ND | ND | ND | ND | 2.71 | 3.70 | 4.13 | ND | 2.49 | ND | ND |
| DELTA-BHC | 2.84 | 2.57 | 1.45 | ND | ND | ND | 0.62 | ND | ND | ND | ND | ND | ND | 1.07 | 1.85 | ND |
| DIELDRIN | 0.93 | 0.48 | 0.58 | ND | ND | 0.53 | 0.73 | 1.15 | 0.93 | 1.15 | ND | ND | ND | 0.81 | 1.20 | 1.2 |
| ENDOSULFAN I | ND | 5.88** | 2.01** | 1.56 | 0.88 | 3.59* | 0.50 | ND | ND | 1.18 | ND | ND | 0.90 | ND | ND | ND |
| ENDOSULFAN 11 | ND | ND | 0.81 | 4,44** | 1.44 | 2.13* | ND | ND | ND | ND | ND | ND | 3.25* | ND | ND | ND |
| ENDOSULFAN SULFATE | 0.62 | ND | 0.61 | 2.83 | 1.26 | 1.26 | 1.11 | ND | 1.36 | 1.58 | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | 0.61 | ND | ND | ND | ND | ND | ND | 0.63 | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | 0.86 | ND | 0.73 | 0.90 | 1.16 | ND | 0.74 | 1.20 | ND | 1.02 | 1.38 | 0.82 | ND | 0.80 | 0.76 | 0.8 |
| GAMMA-BHC | 0.72 | ND | ND | ND | ND | ND | 0.51 | ND | 0.57 | ND | 1.48 | 2.00** | ND | 0.55 | 0.59 | 0.55 |
| HEPTACHLOR | 2.00 | ND | 2.58 | 1.78 | 3.86 | 2.58 | ND | ND | 2.48 | ND | 3.88 | 4.20 | 3.73 | ND | ND | 0.79 |
| IIEPTACHLOR EPOXIDE | 0.82 | 3.00** | 1.29 | ND | ND | 0.49 | ND | ND | 2.31 | ND | 1.11 | ND | ND | 0.49 | 1.02 | 0.56 |
| METHOXYCHLOR | ND | 7.20 | 2.45 | ND | ND | ND | ND | ND | 2.06 | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Cralghill | Craighill Angle East | Craighili Angle West | Cralghill Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Stralghtening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| 4,4'-DDD | ND | 1.39 | 0.73 | 0.34 | ND | 0.46 | ND | 0.29 | 0.25 | 0.25 | 0.40 | 0.45 | 0.33 | ND | ND | ND |
| 4,4'-DDE | ND | ND | 1.15 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4,4'-DDT | ND | ND | 0.75 | 1.48 | 0.38 | 0.80 | 0.32 | 0.76 | 1.67 | 0.6 | 1.35 | 2 3 1:23 | 1.25 | ND | ND | ND |
| ALDRIN | 0.33 | 365 AM 1:44 | 4.39 | ND | A 5.90 | 3.13 | 4.05 | Se | 2.86 | 3.56 | 5.42 | 3.15 | 2.71 | 2.3 | 1.63 | 1.52 |
| ALPHA-BHC | ND | ND | ND | ND | ND | 28.67 | ND | 1.52 | 5.40 | 1.41 | 1.74 | ND | 2.91 | ND | ND | ND |
| BETA-BHC | 0.18 | 5.04 | 223 4.43 | 0.6 | 0.57 | 0.5 | 0.43 | 111 | 0.79 | 0.9 | 0.27 | 0.23 | ND | 0.33 | 0.13 | ND |
| CHLORBENSIDE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.08 | ND | ND | ND |
| CHLORDANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DACTHAL | ND | ND | 1.13 | ND | ND | ND | ND | ND | ND | 1.64 | 2.24 | 2.5 | ND | 1.51 | ND | ND |
| DELTA-BHC | 8.23 | 7.45 | 4.21 | ND | ND | ND | 1.79 | ND | ND | ND | ND | ND | ND | 3.09 | 5.36 | ND |
| DIELDRIN | 2.42 | 0.12 | 0.83 | ND | ND | 1.37 | 1.89 | 3 | 2.42 | 3 | ND | ND | ND | 2.11 | 3.11 | 3.11 |
| ENDOSULFAN I | ND | 11.84 | 4.76 | 5.12 | 2.9 | 11.75 | 1.65 | ND | ND | 3.88 | ND | ND | 2.96 | ND | ND | ND |
| ENDOSULFAN II | ND | ND | 1.07 | 4.28 | 1.39 | 2.05 | ND | ND | ND | ND | ND | ND | 3.13 | ND | ND | ND |
| ENDOSULFAN SULFATE | 0.67 | ND | 0.73 | 3.92 | 1.76 | 1.76 | 1.53 | ND | 1.89 | 2.2 | ND | ND | ND | ND | ND | ND |
| ENDRIN | ND | 1.25 | ND | ND | ND | ND | ND | ND | 1.29 | ND | ND | ND | ND | ND | ND | ND |
| ENDRIN ALDEHYDE | 0.66 | ND | 0.82 | 0.32 | 0.41 | ND | 0.26 | 0.43 | ND | 0.36 | 0.49 | 0.29 | ND | 0.28 | 0.27 | 0.28 |
| GAMMA-BHC | 1.77 | ND | ND | ND | ND | ND | 0.14 | ND | 0.15 | ND | 0.40 | \$ 0.54 | ND | 0.15 | 0.16 | 0.15 |
| HEPTACHLOR | 0.96 | ND | 2.17 | 0.21 | 0.45 | 0.3 | ND | ND | 0.29 | ND | 0.45 | 216 0.48 | 0.43 | ND | ND | 0.09 |
| HEPTACHLOR EPOXIDE | 0.87 | 3.04 | 1.62 | ND | ND | 1.35 | ND | ND | 6.34 | ND | 3.04 | ND | ND | 1.34 | 2.79 | 1.54 |
| METHOXYCHLOR | ND | 4.8 | 1.63 | ND | ND | ND | ND | ND | 1.37 | ND | ND | ND | ND | ND | ND | ND |
| MIREX | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| TOXAPHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

TABLE 9-15 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR CHLORINATED PESTICIDES

ND = not detected

Shaded and bolded values = uptake ratios for tissue-residues that statistically exceeded atleast one of the placement site/reference tissue-residues



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| Analyte (ng/kg) | Inside Site | Outside Site | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Toichester South | Tolchester Straightening |
|------------------------|-------------|--------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| ACENAPHTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ACENAPHTHYLENE* | 121.13 | 184.00 | 117.00* | 5680.25 | 2515.75 | 2618.90 | 701.90 | ND | 62.60 | ND | ND | ND | 1216.80 | 3224.20 | 932.60 | ND |
| ANTIIRACENE | 0.72 | ND | ND | ND | 1.31 | ND | ND | ND | 1.16 | ND | ND | 1.18 | ND | ND | ND | ND |
| BENZ[A]ANTHRACENE | ND | ND | ND | ND | 1.29 | ND | ND | ND | 1.18 | ND | ND | 1.22 | ND | ND | ND | ND |
| BENZO(A)PYRENE | 1.52 | 2.50 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[B]FLUORANTHENE | 1.89 | ND | ND | ND | 2.39 | ND | ND | ND | 1.26 | ND | ND | 1.55 | 1.48 | 1.66 | ND | ND |
| BENZO[G,II,I]PERYLENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[K]FLUORANTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.65 | ND | ND | ND |
| CHRYSENE | 1.64 | 0.43 | ND | ND | 2.64 | ND | ND | ND | 1.42 | ND | ND | 0.85 | 1.58 | 1.42 | 0.81 | ND |
| DIBENZ(A.H)ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 4.56 | 2.00 | 1.97 | ND | 2.38 | 2.71 | ND | ND | 2.08 | ND | ND | 3.54 | ND | 5.80 | ND | ND |
| FLUORENE | 2.96 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.80 | ND | ND | ND | ND |
| INDENO[1,2,3-CD]PYRENE | ND | ND | ND. | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 11.63 | 26.33** | ND | ND | ND | ND | ND | ND | ND | ND | ND | 8.08 | 8.28 | ND | ND | ND |
| PHENANTHRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.26 | ND | ND | ND | ND |
| PYRENE | 7.19 | 8.13 | 1.63 | 0.98 | 1.01 | 1.36 | 1.58 | 1.68 | 1.01 | 1.32 | 0.93 | 1.27 | 0.94 | 1.03 | ND | 0.89 |
| TOTAL PAH (ND=0) | 27.9 | 39 | 3 | 0.98 | 9.21 | 3.36 | 1.58 | 1.68 | 5.65 | 1.32 | 0.744 | 14.9 | 8.57 | 8.95 | 0.62 | 0.184 |
| TOTAL PAH (ND=1/2DL) | 39.7 | 49.3 | 17.8 | 16 | 22.9 | 19.7 | 16.5 | 16.6 | 20 | 16.3 | 15.9 | 28 | 22.7 | 23 | 16 | 15.5 |
| TOTAL PAH (ND=DL) | 51.5 | 59.7 | 32.6 | 31 | 36.5 | 36 | 31.5 | 31.6 | 34.3 | 31.3 | 31.1 | 41.1 | 36.8 | 37 | 31.4 | 30.9 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

* not included in Total PAH calculations

| | | | | Brewerton | | C&D | | | | | | | | | | |
|------------------------|----------|-------------|-----------|-----------|--------------|----------|-----------|------------|-----------------|-----------|-----------------|--------|------------|------------|------------|---------------|
| | Outside | Inside Site | Ocean | Eastern | C&D Approach | Approach | | Craighill | Craighlll Angle | Craighill | Craighill Upper | Cutoff | | Tolchester | Tolchester | Tolchester |
| Analyte (ug/kg) | Site 104 | 104 | Reference | Extension | (Surficial) | (Cores) | Craighill | Angle East | West | Entrance | Range | Angle | Swan Polnt | North | South | Straightening |
| ACENAPHTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ACENAPHTHYLENE* | 184 | 121.13 | 117 | 5680.25 | 2515.75 | 2618.9 | 701.9 | ND | 62.6 | ND | ND | ND | 1216.8 | 3224.2 | 932.6 | ND |
| ANTHRACENE | ND | 0.72 | ND | ND | 1.31 | ND | ND | ND | 1.16 | ND | ND | 1.18 | ND | ND | ND | ND |
| BENZ[A]ANTHRACENE | ND | ND | ND | ND | 1.29 | ND | ND | ND | 1.18 | ND | ND | 1.22 | ND | ND | ND | ND |
| BENZO[A]PYRENE | 2.5 | 1.52 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[B]FLUORANTHENE | ND | 1.89 | ND | ND | 2.39 | ND | ND | ND | 1.26 | ND | ND | 1.55 | 1.48 | 1.66 | ND | ND |
| BENZO[G,H,I]PERYLENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[K]FLUORANTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.65 | ND | ND | ND |
| CHRYSENE | 0.43 | 1.64 | ND | ND | 2.64 | ND | ND | ND | 1.42 | ND | ND | 0.85 | 1.58 | 1.42 | 0.81 | ND |
| DIBENZ[A,H]ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 2 | 4.56 | 1.97 | ND | 2.38 | 2.71 | ND | ND | 2.08 | ND | ND | 3.54 | ND | 5.8 | ND | ND |
| FLUORENE | ND | 2.96 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.8 | ND | ND | ND | ND |
| INDENO[1,2,3-CD]PYRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 26.33 | 11.63 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 8.08 | 8.28 | ND | ND | ND |
| PHENANTHRENE | ND | ND | ND | NE | ND | ND | ND | ND | ND | ND | ND | 1.26 | ND | ND | ND | ND |
| PYRENE | 8.13 | 7.19 | 1.63 | 0.98 | 1.01 | 1.36 | 1.58 | 1.68 | 1.01 | 1.32 | 0.93 | 1.27 | 0.94 | 1.03 | ND | 0.89 |
| Total PAH (ND=0) | 39 | 27.9 | 3 | 0.98 | 9.21 | 3.36 | 1.58 | 1.68 | 5.65 | 1.32 | 0.744 | 14.9 | 8.57 | 8.95 | 0.62 | 0.184 |
| Total PAH (ND=1/2DL) | 49.3 | 39.7 | 17.8 | 16 | 22.9 | 19.7 | 16.5 | 16.6 | 20 | 16.3 | 15.9 | 28 | 22.7 | 23 | 16 | 15.5 |
| Total PAH (ND=DL) | 59.7 | 51.5 | 32.6 | 3 | 36.5 | 36 | 31.5 | 31.6 | 34.3 | 31.3 | 31.1 | 41.1 | 36.8 | 37 | 31.4 | 30.9 |

TABLE 9-16B Nereis virens (SAND WORM): MEAN PAH CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

.

ND = not detected in any of the five tested replicates.

Asterisks indicate sltes where mean tissue residues were statistically higher than Outside Site 104 ($p < 0.05^*$, $p < 0.01^*$).

* not included in total PAH calculations



TABLE 9-16C Nereis virens (SAND WORM): MEAN PAH CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (ng/kg) | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Stralghtening |
|------------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| ACENAPHTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ACENAPHTHYLENE* | 117.00 | 184.00 | 121.13 | 5680.25 | 2515.75 | 2618.90 | 701.90 | ND | 62.60 | ND | ND | ND | 1216.80 | 3224.20 | 9.32.60 | ND |
| ANTHRACENE | ND | ND | 0.72 | ND | 1.31 | ND | ND | ND | 1.16 | ND | ND | 1.18 | ND | ND | ND | ND |
| BENZ[A]ANTHRACENE | ND | ND | ND | ND | 1.29 | ND | ND | ND | 1.18 | ND | ND | 1.22 | ND | ND | ND | ND |
| BENZO[A]PYRENE | ND | 2.50** | 1.52 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[B]FLUORANTHENE | ND | ND | 1.89 | ND | 2.39 | ND | ND | ND | 1.26 | ND | ND | 1.55 | 1.48 | 1.66 | ND | ND |
| BENZO[G,H,I]PERYLENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[K]FLUORANTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.65 | ND | ND | ND |
| CHRYSENE | ND | 0.43 | 1.64* | ND | 2.64* | ND | ND | ND | 1.42 | ND | ND | 0.85 | 1.58 | 1.42 | 0.81 | ND |
| DIBENZ[A,H]ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 1.97 | 2.00 | 4.56* | ND | 2.38 | 2.71 | ND | ND | 2.08 | ND | ND | 3.54 | ND | 5.80* | ND | ND |
| FLUORENE | ND | ND | 2.96 | ND | ND | ND | ND | ND | ND | ND | ND | 2.80 | ND | ND | ND | ND |
| INDENO[1.2.3-CD]PYRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | ND | 26.33** | 11.63 | ND | ND | ND | ND | ND | ND | ND | ND | 8.08 | 8.28 | ND | ND | ND |
| PHENANTHRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.26 | ND | ND | ND | ND |
| PYRENE | 1.63 | 8.13**B | 7.19**B | 0.98 | 1.01 | 1.36 | 1.58 | 1.68 | 1.01 | 1.32 | 0.93 | 1.27 | 0.94 | 1.03 | ND | 0.89 |
| TOTAL PAH (ND=0) | 3 | 39** | 27.9** | 0.98 | 9.21 | 3.36 | 1.58 | 1.68 | 5.65 | 1.32 | 0.744 | 14.9 | 8.57 | 8.95** | 0.62 | 0.184 |
| TOTAL PAH (ND=1/2DL) | 17.8 | 49.3** | 39.7** | 16 | 22.9 | 19.7 | 16.5 | 16.6 | 20 | 16.3 | 15.9 | 28 | 22.7 | • 23* | 16 | 15.5 |
| TOTAL PAH (ND=DL) | 32.6 | 59.7** | 51.5** | 31 | 36.5 | 36 | 31.5 | 31.6 | 34.3 | 31.3 | 31.1 | 41.1 | 36.8 | . 37* | 31.4 | 30.9 |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statitically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

* not included in Total PAH calculation

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighili Angle West | Craighill Entrance | Craighili Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| ACENAPHTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| ACENAPHTHYLENE | 0.37 | 0.16 | 2.07 | 270.49 | 119.8 | 124.71 | 33.42 | ND | 2.98 | ND | ND | ND | 57.94 | 153.53 | 44.41 | ND |
| ANTHRACENE | ND | ND | 0.8 | ND | 2.39 | ND | ND | ND | 2.11 | ND | ND | 2.15 | ND | ND | ND | ND |
| BENZ[A]ANTHRACENE | ND | ND | ND | ND | 0.19 | ND | ND | ND | 0.18 | ND | ND | 0.18 | ND | ND | ND | ND |
| BENZO[A]PYRENE | ND | 0.95 | 1 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[B]FLUORANTHENE | ND | ND | 1.07 | ND | 0.92 | ND | ND | ND | 0.48 | ND | ND | 0.6 | 0.57 | 0.64 | ND | ND |
| BENZO[G,H,1]PERYLENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZO[K]FLUORANTHENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.49 | ND | ND | ND |
| CHRYSENE | ND | 0.47 | 0.62 | ND | 0.12 | ND | ND | ND | 0.06 | ND | ND | 0.04 | 0.07 | 0.06 | 0.04 | ND |
| DIBENZ[A,H]ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 0.12 | 0.58 | 0.80 | ND | 0.27 | 0.3 | ND | ND | 0.23 | ND | ND | 0.4 | ND | 0.65 | ND | ND |
| FLUORENE | ND | ND | 2.56 | ND | ND | ND | ND | ND | ND | ND | ND | 2.8 | ND | ND | ND | ND |
| INDENO[1,2,3-CD]PYRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | ND | 0.83 | 0.33 | ND | ND | ND | ND | ND | ND | ND | ND | 0.22 | 0.22 | ND | ND | ND |
| PHENANTHRENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.94 | ND | ND | ND | ND |
| PYRENE | 0.1 | et 1.51 | 2.49 | 0.59 | 0.61 | 0.82 | 0.95 | 1.01 | 0.6 | 0.79 | 0.56 | 0.76 | 0.56 | 0.62 | ND | 0.53 |
| TOTAL PAH (ND=0) | 0.01 | 0.79 | 0.43 | 0.01 | 0.12 | 0.04 | 0.02 | 0.02 | 0.07 | 0.02 | 0.01 | 0.19 | 0.11 | 0.11 | 0.01 | 0 |
| TOTAL PAH (ND=1/2DL) | 0.04 | 0.84 | 0.54 | 0.18 | 0.26 | 0.22 | 0.19 | 0.19 | 0.22 | 0.18 | 0.18 | 0.32 | 0.26 | 0.26 | 0.18 | 0.17 |
| TOTAL PAH (ND=DL) | 0.08 | .0.87 | 0.63 | 0.32 | 0.37 | 0.37 | 0.32 | 0.32 | 0.35 | 0.32 | 0.32 | 0.42 | 0.38 | 0.38 | 0.32 | 0.32 |

TABLE 9-17 Nereis virens (SAND WORM): UPTAKE RATIOS FOR PAHs

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



TABLE 9-18A Macoma nasuta (BLUNT-NOSE CLAM): MEAN PAH CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| | Inside | Outside | Ocean | Brewerton Eastern | C&D Approach | C&D Approach | | Craighill | Craighill | Craighill | Craighili | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
|------------------------|----------|----------|-----------|----------------------|-----------------|-----------------|-----------|------------|------------|-----------|-------------|--------|--------|------------|------------|---------------|
| Analyte (ug/kg) | Site 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | Cralghill | Angle East | Angle West | Entrance | Upper Range | Angle | Point | North | South | Stralghtening |
| ACENAPHTHENE | 25.49 | 39.75 | 5.44 | ND | ND | 3.40 | 4.26 | 6.92 | 3.14 | 5.98 | 3.96 | ND | 4.44 | ND | ND | ND |
| ACENAPHTHYLENE* | 495.20 | ND | ND | 548.40 | 208.00 | 526.20 | ND | ND | 400.00 | 52.80 | 494.00 | 630.20 | 602.40 | 784.00* | 512.40 | 546.00* |
| ANTHRACENE | 3.50 | 2.75 | ND | 2.44 | 2.44 | 1.83 | 1.18 | 1.85 | 3.21 | 1.87 | 2.72 | 3.86 | 3.11 | 1.81 | 2.46 | 4.1 |
| BENZ[A]ANTHRACENE | 1.15 | 3.38* | ND | 1.81 | 3.29* | 1.48 | 1.67 | 2.30 | 3.91* | 1.18 | 3.44** | 3.52** | 2.29 | A 3.24** | | · ··· 3.74** |
| BENZO[A]PYRENE | 4.45 | 6.85 | ND | 0.64 | 0.89 | 0.65 | ND | 0.57 | 1.04 | ND | 0.96 | 0.95 | 1.50 | 0.50 | ND | 1.1 |
| BENZO[B]FLUORANTHENE | 11.28 | 9.57 | 2.76 | 7.44 | 16.201 | 9.44 | 7.66 | 7.74 | 6.84 | 6.40 | 7.92 | 7.44 | 10.86 | 11.40 | 10.22 | 16.40* |
| BENZO[G,H,I]PERYLENE | 1.24 | 2.48** | ND | ND | ND | ND | ND | ND | 1.38 | ND | ND | ND | 1.74 | ND | ND | ND |
| BENZO[K]FLUORANTHENE | 2.10 | 2.01 | 0.96 | 0.54 | 0.65 | 0.57 | ND | 0.54 | 0.82 | ND | 0.82 | 0.79 | 0.91 | ND | ND | 0.63 |
| CHRYSENE | 8.43 | 1.90 | 3.22 | 4.52 | 4.90 | 3.78 | ND | 1.81 | 5.06 | 1.24 | 3.80 | 3.94 | 3.13 | 3.70 | 3.74 | 4.64 |
| DIBENZ[A,H]ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 16.09 | 7.58 | 7.48 | 9.32 | 9.60 | 9.34 | 10.18 | 9.20 | 9.78 | 11.12 | 16.52 | 12.80 | 10.10 | 8.66 | 16.86 | 9.74 |
| FLUORENE | 10.61 | 30.15 | ND | 7.88 | 10.64 | 9.80 | 10.02 | 9.94 | 11.00 | 10.18 | 9.40 | 9.80 | 8.22 | 10.50 | 10.10 | 18.00** |
| INDENO[1,2,3-CD]PYRENE | 1.14 | 2.70** | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 33.62 | 35.75 | 13.16 | 16.32 | ND | 10.76 | 36.00 | 37.80 | 32.60 | 44.00 | 32.80 | 35.40 | 23.48 | 39.60 | 41.40 | 48.60* |
| PHENANTHRENE | 3.58 | 3.13 | 2.28 | 2.70 | 2.94 | 3.16 | 2.18 | 2.24 | 3.08 | 2.52 | 3.62 | 2.80 | 4.04 | 2.18 | 2.28 | 4.64 |
| PYRENE | 43.40 | \$7.50 | 3.20 | 4.38 | 5.70 | 4.94 | 4.22 | 4.46 | 5.14 | 4.86 | 6.06 | 6.14 | 6.62 | 3.24 | 5.12 | 5.42 |
| TOTAL PAH (ND=0) | 162 | 204 | 30.5 | 51.6 | 56.1 | 49.4 | 74 | 81.8 | 83.5 | 87.1 | 89.5 | 86.8 | 72.5 | 83 | 93.5 | 117 |
| TOTAL PAH (ND=1/2DL) | 166 | 206 | 40.9 | 60.8 | 66.1 | 58.8 | 80.7 | 87.4 | 88.7 | 92.4 | 94.4 | 92.6 | 80.2 | 90.1 | 101 | 122 |
| TOTAL PAH (ND=DL) | 170 | 207 | 51.4 | 70.1 | 76.1 | 68.2 | 87.3 | 93 | 93.9 | 97.7 | 99.4 | 98.5 | 88 | 97.2 | 108 | 128 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05°, p < 0.01°°). * not included in Total PAH calculation

| Analyte (ug/kg) | Outside Site | Inside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighili Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------|--------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| ACENAPHTHENE | 39.75 | 25.49 | 5.44 | ND | ND | 3.40 | 4.26 | 6.92 | 3.14 | 5.98 | 3.96 | ND | 4.44 | ND | ND | ND |
| ACENAPHTHYLENE* | ND | 495.20 | ND | 548.40 | 208.00** | 526.20* | ND | ND | 400.00** | 52.80 | 494.00** | 630.20* | 602.40 | 784.00** | 512.40 | 546.00** |
| ANTHRACENE | 2.75 | 3.50 | ND | 2.44 | 2.44 | 1.83 | 1.18 | 1.85 | 3.21 | 1.87 | 2.72 | 3.86* | 3.11 | 1.81 | 2.46 | 4,10** |
| BENZ[A]ANTIIRACENE | 3.38 | 1.15 | ND | 1.81 | 3.29 | 1.48 | 1.67 | 2.30 | 3.91 | 1.18 | 3.44 | 3.52 | 2.29 | 3.24 | 3.14 | 3.74 |
| BENZO[A]PYRENE | 6.85 | 4.45 | ND | 0.64 | 0.89 | 0.65 | ND | 0.57 | 1.04 | ND | 0.96 | 0.95 | 1.50 | 0.50 | ND | 1.1 |
| BENZO[B]FLUORANTIIENE | 9.57 | 11.28 | 2.76 | 7.44 | 16.20** | 9.44 | 7.66 | 7.74 | 6.84 | 6.40 | 7.92 | 7.44 | 10.86 | 11.40 | 10.22 | 16.40** |
| BENZO[G,H,I]PERYLENE | 2.48 | 1.24 | ND | ND | ND | ND | ND | ND | 1.38 | ND | ND | ND | 1.74 | ND | ND | ND |
| BENZO[K]FLUORANTHENE | 2.01 | 2.10 | 0.96 | 0.54 | 0.65 | 0.57 | ND | 0.54 | 0.82 | ND | 0.82 | 0.79 | 0.91 | ND | ND | 0.63 |
| CHRYSENE | 1.90 | 8.43** | 3.22** | 4.52**B | 4.90** | 3.78 | ND | 1.81 | 5.06** | 1.24 | 3.80** | 3.94** | 3.13 | 3.70** | 3.74** | 4.64** |
| DIBENZ[A,H]ANTIIRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 7.58 | 16.09 | 7.48 | 9.32 | 9.60 | 9.34 | 10.18 | 9.20 | 9.78 | 11.12 | 16.52 | 12.80* | 10.10 | 8.66 | 16.86 | 9.74 |
| FLUORENE | 30.15 | 10.61 | ND | 7.88 | 10.64 | 9.80 | 10.02 | 9.94 | 11.00 | 10.18 | 9.40 | 9.80 | 8.22 | 10.50 | 10.10 | 18 |
| INDENO[1.2.3-CD]PYRENE | 2.70 | 1.14 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 35.75 | 33.62 | 13.16 | 16.32 | ND | 10.76 | 36.00 | 37.80 | 32.60 | 44.00 | 32.80 | 35.40 | 23.48 | 39.60 | 41.40 | 48.60* |
| PHENANTHRENE | 3.13 | 3.58 | 2.28 | 2.70 | 2.94 | 3.16 | 2.18 | 2.24 | 3.08 | 2.52 | 3.62 | 2.80 | 4.04 | 2.18 | 2.28 | 4.64 |
| PYRENE | 57.50 | 43.40 | 3.20 | 4.38 | 5.70 | 4.94 | 4.22 | 4.46 | 5.14 | 4.86 | 6.06 | 6.14 | 6.62 | 3.24 | 5.12 | 5.42 |
| Total PAH (ND=0) | 204 | 162 | 30.5 | 51.6 | 56.1 | 49.4 | 74 | 81.8 | 83.5 | 87.1 | 89.5 | 86.8 | 72.5 | 83 | 93.5 | 117 |
| Total PAH (ND=1/2DL) | 206 | 166 | 40.9 | 60.8 | 66.1 | 58.8 | 80.7 | 87.4 | 88.7 | 92.4 | 94.4 | 92.6 | 80.2 | 90.1 | 101 | 122 |
| Total PAH (ND=DL) | 207 | 170 | 51.4 | 70.1 | 76.1 | 68.2 | 87.3 | 93 | 93.9 | 97.7 | 99.4 | 98.5 | 88 | 97.2 | 108 | 128 |

TABLE 9-18B Macoma nasuta (BLUNT-NOSE CLAM): MEAN PAH CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).

* not included in total PAII calculation



TABLE 9-18C Macoma nasuta (BLUNT-NOSE CLAM): MEAN PAH CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (ug/kg) | Ocean Reference | Ontside Site 104 | Inside Site | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighili Angle East | Craighill Angle West | Craighill Entrance | Craighili Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------------|---------------------|-------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|------------|---------------------|---------------------|-----------------------------|
| ACENAPIITHENE | 5.44 | 39.75** | 25.49** | ND | ND | 3.40 | 4.26 | 6.92 | 3.14 | 5.98 | 3.96 | ND | 4.44 | ND | ND | ND |
| ACENAPIITHYLENE* | ND | ND | 495.20 | 548.40 | 208.00** | 526.20* | ND | ND | 400.00** | 52.80 | 494.00** | 630.20* | 602.40 | 784.00** | 512.40 | 546.00** |
| ANTHRACENE | ND | 2.75** | 3.50** | 2.44 | 2.44 | 1.83 | 1.18 | 1.85 | 3.21* | 1.87 | 2.72 | 3.86** | 3.11*B | 1.81 | 2.46 | 4.10** |
| BENZ[A]ANTHRACENE | ND | 3.38* | 1.15 | 1.81 | 3.29* | 1.48 | 1.67 | 2.30 | 3.91* | 1.18 | 3.44** | 3.52** | 2.29 | 3.24** | 3.14** | 3.74** |
| BENZO[A]PYRENE | ND | 6.85** | 4.45** | 0.64 | 0.89 | 0.65 | ND | 0.57 | 1.04 | ND | 0.96* | 0.95* | 1.50 | 0.50 | ND | 1.10* |
| BENZO[B]FLUORANTHENE | 2.76 | 9.57** | 11.28** | 7.44**B | 16.20** | 9.44**B | 7.66** | 7.74** | 6.84** | 6.40* | 7.92** | 7.44** | 10.86**B | 11.40** | 10.22** | 16.40** |
| BENZO[G.H.I]PERYLENE | ND | 2.48** | 1.24 | ND | ND | ND | ND | ND | 1.38 | ND | ND | ND | 1.74 | ND | ND | ND |
| BENZO[K]FLUORANTHENE | 0.96 | 2.01 | 2.10 | 0.54 | 0.65 | 0.57 | ND | 0.54 | 0.82 | ND | 0.82 | 0.79 | 0.91 | ND | ND | 0.63 |
| CHRYSENE | 3.22 | 1.90 | 8.43* | 4.52**B | 4.90** | 3.78 | ND | 1.81 | 5.06** | 1.24 | 3.80* | 3.94** | 3.13 | 3.70* | 3.74** | 4.64** |
| DIBENZ[A,H]ANTIIRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 7.48 | 7.58 | 16.09* | 9.32**B | 9.60** | 9.34*B | 10.18* | 9.20** | 9.78** | 11.12* | 16.52** | · 12.80** | 10.10*B | 8.66 | 16.86* | 9.74** |
| FLUORENE | ND | 30.15* | 10.61* | 7.88*B | 10.64** | 9.80**B | 10.02** | 9.94** | 11.00** | 10.18** | 9.40** | 9.80** | 8.22*B | 10.50** | 10.10** | 18.00** |
| INDENO[1,2,3-CD]PYRENE | ND | 2.70** | 1.14 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 13.16 | 35.75* | 33.62* | 16.32 | ND | 10.76 | 36.00* | 37.80* | 32.60 | 44.00** | 32.80* | 35.40* | 23.48 | 39.60** | 41.40** | 48.60** |
| PHENANTHRENE | 2.28 | 3.13 | 3.58 | 2.70 | 2.94 | 3.16 | 2.18 | 2.24 | 3.08* | 2.52 | 3.62 | 2.80 | 4.04 | 2.18 | 2.28 | 4.64** |
| PYRENE | 3.20 | 57.50** | 43.40** | 4.38*B | 5.70** | 4.94**B | 4.22* | 4.46 | 5.14* | 4.86 | 6.06** | 6.14** | 6.62*B | 3.24 | 5.12* | 5.42** |
| TOTAL PAH (ND=0) | 30.5 | 204** | 162** | 51.6 | 56.1* | 49 4 | 74** | 81.5** | \$3.5** | 87.1** | 89.5** | 86.8** | 72.5* | 83** | 93.5** | 117** |
| TOTAL PAH (ND=1/2DL) | 40.9 | 206** | 166** | 60.8* | 66.1** | 58.8* | 80.7** | 87.4** | 88.7** | 92.4** | 94.4** | 92.6** | 80.2* | 90.1** | 101** | 122** |
| TOTAL PAII (ND=DL) | 51.4 | 207** | 170** | 70.1* | 76.1** | 68.2* | 87.3** | . 93** | 93.9** | 97.7** | 99,4** | 98.5** | 88* | 97.2** | 108** | 128** |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site ($p < 0.05^{\circ}$, $p < 0.01^{\circ \circ}$). * not included in Total PAH calculation

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighlil | Craighlll Angle East | Craighlll Angle West | Craighill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|------------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| ACENAPHTHENE | 0.74 | 1.08 | 5.34 | ND | ND | 1.21 | 1.52 | 2.47 | 1.12 | 2.14 | 1.41 | ND | 1.59 | ND | ND | ND |
| ACENAPHTHYLENE | ND | ND | 2.04 | 1.97 | 0,75 | 1.89 | ND | ND | 1.43 | 0.19 | Sept. 51.77 | 2.26 | 2.16 | 2.81 | 1.84 | 1.96 |
| ANTHRACENE | ND | 5.00 | 3.94 | 1.97 | 1.97 | 1.48 | 0.95 | 1.49 | 2.59 | 1.51 | 2.19 | 8 3.11 | 2.51 | 1.46 | 1.98 | a a |
| BENZ[A]ANTHRACENE | ND | 4.50 | 1.53 | 2.41 | 4.39 | 1.97 | 2.23 | 3.07 | 5.21 | 1.57 | 4.59 | 4.69 | 3.05 | 1010103894.32 | 4.19 | 4.99 |
| BENZO[A]PYRENE | ND | 2.89 | 6.70 | 1.57 | 2.18 | 1.58 | ND | 1.39 | 2.55 | ND | 7 2.34 | 2.32 | 3.66 | 1.21 | ND | 2.69 |
| BENZO[B]FLUORANTHENE | 0.57 | 5.98 | 5.91 | 3.61 | 7.86 | 4.58 | 3.72 | 3.76 | My 331 | 3.11 | 3.84 | 3.61 | 5.27 | 5.53 | 4.96 | 7.96 |
| BENZO[G,H,I]PERYLENE | ND | 2.75 | 1.38 | ND | ND | ND | ND | ND | 1.53 | ND | ND | ND | 1.93 | ND | ND | ND |
| BENZO[K]FLUORANTHENE | 0.89 | 4.57 | 4.76 | 1.24 | 1.48 | 1.3 | ND | 1.23 | 1.87 | ND | 1.85 | 1.79 | 2.07 | ND | ND | 1.43 |
| CHRYSENE | 6.71 | 2.29 | 5.68 | 2 10.45 x 2.47 | 2.68 | 2.07 | ND | 0.99 | 9. 2.77 | 0.68 | 2.08 | 2.16 | 1.71 | 2.02 | 2.05 | 2.54 |
| DIBENZ[A,H]ANTHRACENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| FLUORANTHENE | 0.43 | 1.76 | 244 | | 1.10 | 1:07 | 1.16 | A 1.05 | 1.12 | # 388 8 1.1.27 | 1.89 | 1.46 | 1.15 | 0.99 | 1.92 | 1.11 Mar 1.11 |
| FLUORENE | ND | 2.56 | 1.05 | 0.87 | 1.17 | 1.08 | 1.10 | 1.09 | 1.1.21 | 1.12 | 1.04 | 1.08 | 0.91 | 1.16 | ILI CAL | 1.98 |
| INDENO[1,2,3-CD]PYRENE | ND | 3.37 | 1.43 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NAPHTHALENE | 0.09 | 1.07 | 1.24 | 0.77 | ND | 0.51 | 1.71 | 1.79 | 1.55 | 2.09 | 1.56 | 2.4.1.68 | 1.11 | 88.1 2 1.88 | 1.96 | AMA 43 43 1231 |
| PIIENANTHRENE | 0.38 | 1.56 | 1.54 | 1.07 | 1.17 | 1.25 | 0.87 | 0.89 | 1.22 | 1 | 1.44 | 1.11 | 1.6 | 0.87 | 0.9 | 1.84 |
| PYRENE | 0.29 | 14.26 | 14.05 | 1.58 | 2.06 | 1.78 | 1.52 | 1.61 | 1.86 | 1.76 | 2.19 | -2.22 | 2.39 | 1.17 | 1.85 | 1.96 |
| TOTAL PAH (ND=0) | 0.13 | 2.19 | 2.93 | 1.17 | 1.27 | 1.12 | 1.68 | 1.86 | 1.98 | 1.98 | 2.03 | 1.97 | 1.65 | 1.89 | 9 2.13 | 2.65 |
| TOTAL PAH (ND=1/2DL) | 0.17 | 2.07 | 2.53 | 1:13 | 1.23 | 1.09 | 1.50 | 1.62 | 1.65 | 1.72 | 1.75 | 1117 | 2.1.49 | 1.67 | 1.87 | 45 |
| TOTAL PAH (ND=DL) | 0.2 | 1.96 | 2.25 | 1.10 | 1.19 | 1.07 | 1.37 | 1.46 | 0.0147 | 1.53 | 1.56 | 1.54 | 1.38 | 1.52 | 383 1.69 | 2.01 |

TABLE 9-19 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR PAHs

ND = not detected

Shaded and bolded values = uptake ratios for tissue-residues that statistically exceeded atleast one of the placement site/reference tissue-residues



TABLE 9-20A Nereis virens (SAND WORM):MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED
TO INSIDE SITE 104

| • | | | | Brewerton | C&D | C&D | | | | | | | | | | l i i i i i i i i i i i i i i i i i i i |
|-----------------|----------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---|
| | Inside | Outside | Ocean | Eastern | Approach | Approach | | CraighIII | CraighIll | Craighill | Craighill | Cutoff | | Tolchester | Tolchester | Tolchester |
| Anaiyte (ug/kg) | Site 104 | Site 104 | Reference | Extension | (Surficiai) | (Cores) | Craighili | Angle East | Angle West | Entrance | Upper Range | Angie | Swan Point | North | South | Straightening |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

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ND = not detected in any of five tested replicates.

TABLE 9-20B Nereis virens (SAND WORM): MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED TOOUTSIDE SITE 104

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|-----------------|----------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Outside | Inside | Ocean | Eastern | Approach | Approach | | CraighIII | Craighill | CraighIII | Craighill | Cutoff | | Tolchester | Tolchester | Tolchester |
| Analyte (ug/kg) | Site 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | CraighIII | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Polnt | North | South | Stralghtening |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of the five tested replicates.

Asterisks indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p<0.05*, p<0.01**).



TABLE 9-20C Nereis virens (SAND WORM): MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| , | | | | Brewerton | C&D | C&D | | | | | | | | | | 1 |
|-----------------|-----------|----------|----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Ocean | Outside | Inside | Eastern | Approach | Approach | | Craighill | Craighlil | Craighill | Craighill | Cutoff | | Toichester | Tolchester | Tolchester |
| Analyte (ug/kg) | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Cores) | Cralghili | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Straightening |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NĐ | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of the five tested replicates.

TABLE 9-21 Nereis virens (SAND WORM): UPTAKE RATIOS FOR PCB AROCLORS

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| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Snrficial) | C&D Approach (Cores) | Craighili | Craighill Angle East | Craighill Angle West | Craighill Entrance | Craighiii Upper Range | Cntoff Angle | Swan Point | Toichester North | Toichester South | Tolchester Straightening |
|--------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



TABLE 9-22A Macoma nasuta (BLUNT-NOSE CLAM): MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| | | | 1 1 | Brewerton | C&D | C&D | | | | | | | | | 1 | |
|-----------------|----------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|---------------------|------------|------------|------------|---------------|
| | Inside | Outslde | Ocean | Eastern | Approach | Approach | 1 | Craighiil | Craighill | Cralghlil | Cralghill | | | Toichester | Tolchester | Toichester |
| Anaiyte (ug/kg) | Site 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | Craighlll | Angle East | Angie West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | South | Straightening |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of the five tested replicates.

Asterisks indicate sites where mean tissue residues were statistically higher than Inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-22B Macoma nasuta (BLUNT-NOSE CLAM): MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

| Analyte (ug/kg) | Outside Site 104 | Inside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Cralghill Angle East | Craighili Angle West | Craighili Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------|---------------------|--------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of the five tested replicates.

Asterisks indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p<0.05*, p<0.01**).



TABLE 9-22C Macoma nasuta (BLUNT-NOSE CLAM): MEAN PCB AROCLOR CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| | | · · · · · · · · · · · · · · · · · · · | | Brewerton | C&D | C&D | · · · · · | · · · · · · · · · · · · · · · · · · · | | | | · | (| F | | |
|-----------------|-----------|---------------------------------------|----------|-----------|-------------|----------|-----------|---------------------------------------|------------|-----------|-------------|--------|-------|------------|------------|---------------|
| | Ocean | Outside | Inside | Eastern | Approach | Approach | ! | CraighIII | Cralghill | CraighIII | Craighill | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| Analyte (ug/kg) | Reference | Site 104 | Site 104 | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Angle | Point | North | South | Straightening |
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

ND = not detected in any of five tested replicates.

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Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p<0.05*,p<0.01**).
| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| AROCLOR 1016 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1221 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1232 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1242 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NĎ | ND | ND | ND | ND | ND |
| AROCLOR 1248 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1254 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| AROCLOR 1260 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |

TABLE 9-23 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR PCB AROCLORS

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ND = not detected

Shaded and bolded values = uptake ratios for tissue-residues that statistically exceeded atleast one of the placement site/reference tissue-residues

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TABLE 9-24A Nereis virens (SAND WORM): MEAN PCB CONGENER^(a) AND TOTAL PCB CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| Analyte (11g/kg) | Inside Site 104 | Ontside Site 104 | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | CraighIII Entrance | Cralghill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|---------------------|--------------------|---------------------|--------------------|--------------------------------|-----------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| BZ# 8 | ND | ND | ND | ND | ND | ND | ND | ND | 17.7 | ND | 3.1 | 3.55 | ND | 15.1 | 1.58 | ND |
| BZ# 18 | 0.238 | 0.823 | ND | ND | ND | ND | ND | ND | 0.177 | ND | 0.164 | 0.153 | ND | ND | ND | ND |
| BZ# 28 · | 1.36 | 1.67 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.179 | ND | ND | ND | ND |
| BZ# 44 | 0.232 | ND | ND | ND | ND | ND | 0.146 | ND | ND | 0.164 | ND | 0.157 | 0.15 | ND | 0.128 | ND |
| BZ# 49 | 0.399 | 0.645 | ND | 0.364 | ND | ND | ND | ND | ND | 0.252 | 0.697 | 1.21 | ND | ND | 0.294 | ND |
| BZ# 52 | 1.27 | 0.835 | 3.15 | ND | ND | ND | ND | ND | ND | ND | ND | 0.183 | ND | ND | ND | ND |
| BZ# 66 | ND | ND | ND | ND | 1.59 | ND | ND | ND | ND | ND | ND | 0.449 | ND | ND | ND | ND |
| BZ# 77 | 1.88 | 0.404 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.622 | ND | ND | ND | 0.281 |
| BZ# 87 | 0.416 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.201 | 0.11 | ND | ND | ND |
| BZ# 101 | 1.14 | ND | ND | 0.484 | 0.802 | 0.454 | 0.864 | 0.258 | 0.307 | 0.354 | 0.443 | 0.806 | 0.246 | 0.586 | 0.266 | 0.308 |
| BZ# 105 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.173 | ND | ND | ND | ND |
| BZ# 118 | 0.273 | ND | ND | ND | ND | ND | 10 | 8.06 | 0.743 | 6.38 | ND | 2.48 | ND | 5.28 | ND | ND |
| BZ# 126 | 0.229 | ND | ND | 0.32 | 0.506 | ND | ND | ND | ND | 2.88 | ND | 5.88 | ND | ND | 0.44 | ND |
| BZ# 128 | 0.206 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.125 | ND | ND | ND | ND |
| BZ# 138 | 0.92 | 0.818 | ND | 0.728 | 0.54 | 0.896 | 0.824 | 0.282 | 0.436 | 0.752 | 0.886 | 0.834 | 0.31 | 0.288 | 0.886 | 0.58 |
| BZ# 153 | 2.11 | 1.92 | ND | 1.78 | 1.02 | 1.9 | 1.7 | 1.28 | 1.78 | 1.96 | 1.94 | 2.38 | 1.58 | 1.52 | 2.24 | 1.75 |
| BZ# 156 | ND | ND | ND | 0.103 | ND | ND | ND | 0.0992 | ND | ND | ND | 0.122 | ND | ND | 0.109 | ND |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.133 | ND | ND | ND | ND |
| BZ# 170 | 0.333 | 0.34 | ND | 0.195 | ND | 0.084 | ND | 0.088 | 0.0876 | 0.084 | 0.098 | 0.268 | 0.084 | 0.171 | 0.15 | 0.159 |
| BZ# 180 | 0.691 | 0.568 | ND | 0.65 | 0.391 | 0.642 | 0.612 | 0.411 | 0.478 | 0.622 | 0.64 | 0.75 | 0.472 | 0.418 | 0.768 | 0.525 |
| BZ# 183 | 0.107 | ND | ND | ND | ND | 0.127 | 0.159 | ND | 0.078 | 0.102 | 0.0788 | 0.0877 | 0.0844 | ND | 0.137 | ND |
| BZ# 184 | 0.856 | ND | ND | 0.172 | 0.32 | 0.108 | ND | 0.116 | 0.13 | 0.242 | 0.417 | 0.876 | 0.11 | 0.512 | 0.81 | 1.7 |
| BZ# 187 | 0.649 | 0.887 | ND | 0.518 | ND | 0.57 | 1.51 | 1.5 | 0.876 | 1.17 | 0.862 | 0.946 | 0.462 | 0.41 | 0.686 | 0.512 |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.251 | ND | ND | ND | ND |
| BZ# 206 | 0.158 | 0.525 | ND | 0.399 | ND | 0.526 | 0.177 | 0.116 | ND | 0.215 | 0.172 | 0.199 | 0.237 | 0.112 | 0.439 | 0.532 |
| BZ# 209 | 0.294 | 0.486 | ND | 0.279 | 0.495 | 0.53 | ND | 0.625 | ND | ND | ND | 0.159 | 0.408 | ND | 0.445 | 0.75 |
| TOTAL PCBs (ND=0) | 21.4 | 16 | 5.6 | 9.01 | 8.82 | 8.96 | 31.2 | 23 | 44.3 | 28.1 | 15.8 | 38.4 | 6 | 46.8 | 13.6 | 7.8 |
| TOTAL PCBs (ND=1/2) | 24.5 | 21.6 | 33.3 | 11.2 | 11.2 | 11.2 | 33.2 | 25.5 | 46.5 | 30.2 | 18 | 40.1 | 8.43 | 49 | 15.8 | 10.1 |
| TOTAL PCBs (ND=DL) | 27.5 | 27.2 | 61** | 13.4 | 13.6 | 13.4 | 35.3 | 27.9 | 48.7 | 32.4 | 20.2 | 41.8 | 10.9 | 51.3 | 18 | 12.3 |

ND = not detected in any of five tested replicates.

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for total PCB congeners only. Total PCBs determined by summing congeners as specified in iTM (Table 9-3) and multiplying total by a facor of 2 as per NOAA 1993 guidance. Note that the mean of total PCB for individual tissue replicates is not equivalent to the sum of mean congeners for ND=0 and ND=1/2DL. Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-24B Nereis virens (SAND WORM): MEAN PCB CONGENER^(a) AND TOTAL PCB CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|---------------------|----------|-------------|-----------|-----------|-------------|----------|-------------|------------|------------|-----------|-------------|--------------|------------|------------|------------------|---------------|
| A malute (wa/lum) | Outside | Inside Site | Ocean | Eastern | Approach | Approach | ~ · · · ··· | Craighill | Cralghill | Craighill | Craighill | | | Tolchester | | Tolchester |
| Analyte (lig/kg) | Site 104 | 104 | Reference | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | Tolchester Snuth | Stralghtening |
| BZ# 8 | ND | ND | ND | ND | ND | ND | ND | ND | 17.7 | ND | 3.1 | 3.55 | ND | 15.1 | 1.58 | ND |
| BZ# 18 | 0.823 | 0.238 | ND | ND | ND | ND | ND | ND | 0.177 | ND | 0.164 | 0.153 | ND | ND | ND | ND |
| BZ# 28 | 1.67 | 1.36 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.179 | ND | ND | ND | ND |
| BZ# 44 | ND | 0.232 | ND | ND | ND | ND | 0.146 | ND | ND | 0.164 | ND | 0.157 | 0.15 | ND | 0.128 | ND |
| BZ# 49 | 0.645 | 0.399 | ND | 0.364 | ND | ND | ND | ND | ND | 0.252 | 0.697 | 1.21 | ND | ND | 0.294 | ND |
| BZ# 52 | 0.835 | 1.27 | 3.15 | ND | ND | ND | ND | ND | ND | ND | ND | 0.183 | ND | ND | ND | ND |
| BZ# 66 | ND | ND | ND | ND | 1.59 | ND | ND | ND | ND | ND | ND | 0.449 | ND | ND | ND | ND |
| BZ# 77 | 0.404 | 1.88 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.622 | ND | ND | ND | 0.281 |
| BZ# 87 | ND | 0.416 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.201 | 0.11 | ND | ND | ND |
| BZ# 101 | ND | 1.14 | ND | 0.484 | 0.802 | 0.454 | 0.864 | 0.258 | 0.307 | 0.354 | 0.443 | 0.806 | 0.246 | 0.586 | 0.266 | 0.308 |
| BZ# 105 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.173 | ND | ND | ND | ND |
| BZ# 118 | * ND | 0.273 | ND | ND | ND | ND | 10 | 8.06 | 0.743 | 6.38 | ND | 2.48 | ND | 5.28 | ND | ND |
| BZ# 126 | ND | 0.229 | ND | 0.32 | 0.506 | ND | ND | ND | ND | 2.88 | ND | 5.88 | ND | ND | 0.44 | ND |
| BZ# 128 | ND | 0.206 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.125 | ND | ND | ND | ND |
| BZ# 138 | 0.818 | 0.92 | ND | 0.728 | 0.54 | 0.896 | 0.824 | 0.282 | 0.436 | 0.752 | 0.886 | 0.834 | 0.31 | 0.288 | 0.886 | 0.58 |
| BZ# 153 | 1.92 | 2.11 | ND | 1.78 | 1.02 | 1.9 | 1.7 | 1.28 | 1.78 | 1.96 | 1.94 | 2.38 | 1.58 | 1.52 | 2.24 | 1.75 |
| BZ# 156 | ND | ND | ND | 0.103 | ND | ND | ND | 0.0992 | ND | ND | ND | 0.122 | ND | ND | 0.109 | ND |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.133 | ND | ND | ND | ND |
| BZ# 170 | 0.34 | 0.333 | ND | 0.195 | ND | 0.084 | ND | 0.088 | 0.0876 | 0.084 | 0.098 | 0.268 | 0.084 | 0.171 | 0.15 | 0.159 |
| BZ# 180 | 0.568 | 0.691 | ND | 0.65 | 0.391 | 0.642 | 0.612 | 0.411 | 0.478 | 0.622 | 0.64 | 0.75 | 0.472 | 0.418 | 0.768 | 0.525 |
| BZ# 183 | ND | 0.107 | ND | ND | ND | 0.127 | 0.159 | ND | 0.078 | 0.102 | 0.0788 | 0.0877 | 0.0844 | ND | 0.137 | ND |
| BZ# 184 | ND | 0.856 | ND | 0.172 | 0.32 | 0.108 | ND | 0.116 | 0.13 | 0.242 | 0.417 | 0.876 | 0.11 | 0.512 | 0.81 | 1.7 |
| BZ# 187 | 0.887 | 0.649 | ND | 0.518 | ND | 0.57 | 1.51 | 1.5 | 0.876 | 1.17 | 0.862 | 0.946 | 0.462 | 0.41 | 0.686 | 0.512 |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.251 | ND | ND | ND | ND |
| BZ# 206 | 0.525 | 0.158 | ND | 0.399 | ND | 0.526 | 0.177 | 0.116 | ND | 0.215 | 0.172 | 0.199 | 0.237 | 0.112 | 0.439 | 0.532 |
| BZ# 209 | 0.486 | 0.294 | ND | 0.279 | 0.495 | 0.53 | ND | 0.625 | ND | ND | ND | 0.159 | 0.408 | ND | 0.445 | 0.75 |
| TOTAL PCBs (ND=0) | 16 | 21.4 | 5.6 | 9.01 | 8.82 | 8.96 | 31.2 | 23 | 44.3 | 28.1 | 15.8 | 38.4 | 6 | 46.8 | 13.6 | 7.8 |
| TOTAL PCBs (ND=1/2) | 21.6 | 24.5 | 33.3* | 11.2 | 11.2 | 11.2 | 33.2 | 25.5 | 46.5 | 30.2 | 18 | 40.1 | 8.43 | 49 | 15.8 | 10.1 |
| TOTAL PCBs (ND=DL) | 27.2 | 27.5 | 61** | 13.4 | 13.6 | 13.4 | 35.3 | 27.9 | 48.7 | 32.4 | 20.2 | 41.8 | 10.9 | 51.3 | 18 | 12.3 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for total PCB congeners only

Total PCBs determined by summing congeners as specified in 1TM (Table 9-3) and multiplying total by a facor of 2 as per NOAA 1993 guidance.

Note that the mean of total PCB for individual tissue replicates is not equivalent to the sum of mean congeners for ND=0 and ND=1/2DL.



TABLE 9-24C Nereis virens (SAND WORM): MEAN PCB CONGENER^(a) AND TOTAL PCB CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| Analyte (11g/kg) | Ocean Reference | Ontside Site | inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighili | Cralghiil Angie East | Craighili Angle West | Cralghili Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Toichester North | Toichester South | Tnlchester Straightening |
|---------------------|--------------------|--------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------------|------------|---------------------|---------------------|-----------------------------|
| BZ# 8 | ND | ND | ND | ND | ND | ND | ND | ND | 17.7 | ND | 3.1 | 3.55 | ND | 15.1 | 1.58 | ND |
| BZ# 18 | ND | 0.823 | 0.238 | ND | ND | ND | ND | ND | 0.177 | ND | 0.164 | 0.153 | ND | ND | ND | ND |
| BZ# 28 | ND | 1.67 | 1.36 | ND | ND | ND | ND | ND | ND | ND | ND | 0.179 | ND | ND | ND | ND |
| BZ# 44 | ND | ND | 0.232 | ND | ND | ND | 0.146 | ND | ND | 0.164 | ND | 0.157 | 0.15 | ND | 0.128 | ND |
| BZ# 49 | ND | 0.645 | 0.399 | 0.364 | ND | ND | ND | ND | ND | 0.252 | 0.697 | 1.21 | ND | ND | 0.294 | ND |
| BZ# 52 | 3.15 | 0.835 | 1.27 | ND | ND | ND | ND | ND | ND | ND | ND | 0.183 | ND | ND | ND | ND |
| BZ# 66 | ND | ND | ND | ND | 1.59 | ND | ND | ND | ND | ND | ND | 0.449 | ND | ND | ND | ND |
| BZ# 77 | ND | 0.404 | 1.88 | ND | ND | ND | ND | ND | ND | ND | ND | 0.622 | ND | ND | ND | 0.281 |
| BZ# 87 | ND | ND | 0.416 | ND | ND | ND | ND | ND | ND | ND | ND | 0.201 | 0.11 | ND | ND | ND |
| BZ# 101 | ND | ND | 1.14 | 0.484 | 0.802 | 0.454 | 0.864 | 0.258 | 0.307 | 0.354 | 0.443 | 0.806 | 0.246 | 0.586 | 0.266 | 0.308 |
| BZ# 105 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.173 | ND | ND | ND | ND |
| BZ# 118 | ND | ND | 0.273 | ND | ND | ND | 10 | 8.06 | 0.743 | 6.38 | ND | 2.48 | ND | 5.28 | ND | ND |
| BZ# 126 | ND | ND | 0.229 | 0.32 | 0.506 | ND | ND | ND | ND | 2.88 | ND | 5.88 | ND | ND | 0.44 | ND |
| BZ# 128 | ND | ND | 0.206 | ND | ND | ND | ND | ND | ND | ND | ND | 0.125 | ND | ND | ND | ND |
| BZ# 138 | NE | 0.818 | 0.92 | 0.728 | 0.54 | 0.896 | 0.824 | 0.282 | 0.436 | 0.752 | 0.886 | 0.834 | 0.31 | 0.288 | 0.886 | 0.58 |
| BZ# 153 | ND | 1.92 | 2.11 | 1.78 | 1.02 | 1.9 | 1.7 | 1.28 | 1.78 | 1.96 | 1.94 | 2.38 | 1.58 | 1.52 | 2.24 | 1.75 |
| BZ# 156 | NC | ND | ND | 0.103 | ND | ND | ND | 0.0992 | ND | ND | ND | 0.122 | ND | ND | 0.109 | ND |
| BZ# 169 | NE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.133 | ND | ND | ND | ND |
| BZ# 170 | ND | 0.34 | 0.333 | 0.195 | ND | 0.084 | ND | 0.088 | 0.0876 | 0.084 | 0.098 | 0.268 | 0.084 | 0.171 | 0.15 | 0.159 |
| BZ# 180 | ND | 0.568 | 0.691 | 0.65 | 0.391 | 0.642 | 0.612 | 0.411 | 0.478 | 0.622 | 0.64 | 0.75 | 0.472 | 0.418 | 0.768 | 0.525 |
| BZ# 183 | ND | ND | 0.107 | ND | ND | 0.127 | 0.159 | ND | 0.078 | 0.102 | 0.0788 | 0.0877 | 0.0844 | ND | 0.137 | ND |
| BZ# 184 | NE | ND ND | 0.856 | 0.172 | 0.32 | 0.108 | ND | 0.116 | 0.13 | 0.242 | 0.417 | 0.876 | 0.11 | 0.512 | 0.81 | 1.7 |
| BZ# 187 | NE | 0.887 | 0.649 | 0.518 | ND | 0.57 | 1.51 | 1.5 | 0.876 | 1.17 | 0.862 | 0.946 | 0.462 | 0.41 | 0.686 | 0.512 |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.251 | ND | ND | ND | ND |
| BZ# 206 | ND | 0.525 | 0.158 | 0.399 | ND | 0.526 | 0.177 | 0.116 | ND | 0.215 | 0.172 | 0.199 | 0.237 | 0.112 | 0.439 | 0.532 |
| BZ# 209 | ND | 0.486 | 0.294 | 0.279 | 0.495 | 0.53 | ND | 0.625 | ND | ND | ND | 0.159 | 0.408 | ND | 0.445 | 0.75 |
| TOTAL PCBs (ND=0) | 5.0 | 16* | 21.4 | 9.01 | 8.82 | 8.96 | * 31.2* | 23 | 44.3 | 28.1* | 15.8 | 38.4* | 6 | 46.8 | 13.6 | 7.8 |
| TOTAL PCBs (ND=1/2) | 33.3 | 21.6 | 24.5 | 11.2 | 11.2 | 11.2 | 33.2 | 25.5 | 46.5 | 30.2 | 18 | 40.1 | 8.43 | 49 | 15.8 | 10.1 |
| TOTAL PCBs (ND=DL) | 61 | 27.2 | 27.5 | 13.4 | 13.6 | 13.4 | 35.3 | 27.9 | 48.7 | 32.4 | 20.2 | 41.8 | 10.9 | 51.3 | 18 | 12.3 |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for total PCB congeners only

Total PCBs determined by summing congeners as specified in ITM (Table 9-3) and multiplying total by a facor of 2 as per NOAA 1993 guidance.

Note that the mean of total PCB for individual tissue replicates is not equivalent to the sum of mean congeners for ND=0 and ND=1/2DL.

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighili Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| BZ# 8 | ND | ND | ND | ND | ND | ND | ND | ND | 19.49 | ND | 3.42 | 3.92 | ND | 16.65 | 1.74 | ND |
| BZ# 18 | ND | 4.57 | 1.69 | ND | ND | ND | ND | ND | 2.95 | ND | 2.73 | 2.55 | ND | ND | ND | ND |
| BZ# 28 | ND | 1.74 | 1.93 | ND | ND | ND | ND | ND | ND | ND | ND | 2.98 | ND | ND | ND | ND |
| BZ# 44 . | ND | ND | 2.59 | ND | ND | ND | 2.65 | ND | ND | 2.98 | ND | 2.85 | 2.73 | ND | 2.33 | ND |
| BZ# 49 | ND | 0.33 | 0.4 | 1.19 | ND | ND | ND | ND | ND | 0.82 | 2.27 | 3.95 | ND | ND | 0.96 | ND |
| BZ# 52 | 0.93 | 1.24 | 1.87 | ND | ND | ND | ND | ND | ND | ND | ND | 0.26 | ND | ND | ND | ND |
| BZ# 66 | ND | ND | ND | ND | 1.82 | ND | ND | ND | ND | ND | ND | 0.51 | ND | ND | ND | ND |
| BZ# 77 | ND | 1.28 | 2.76 | ND | ND | ND | ND | ND | ND | ND | ND | 0.09 | ND | ND | ND | 0.04 |
| BZ# 87 | ND | ND | 1.16 | ND | ND | ND | ND | ND | ND | ND | ND | 0.44 | 0.24 | ND | ND | ND |
| BZ# 101 | ND | ND | 3.12 | 1.01 | 1.67 | 0.95 | 1.8 | 0.54 | 0.64 | 0.74 | 0.92 | 1.68 | 0.51 | 1.22 | 0.55 | 0.64 |
| BZ# 105 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.16 | ND | ND | ND | ND |
| BZ# 118 | ND | ND | 1.24 | ND | ND | ND | 95.64 | 76.79 | 7.08 | 60.78 | ND | 23.64 | ND | 50.32 | ND | ND |
| BZ# 126 | ND | ND | 1.13 | 3.2 | 5.06 | ND | ND | ND | ND | 28.8 | ND | 58.78 | ND | ND | 4.4 | ND |
| BZ# 128 | ND | ND | 1.35 | ND | ND | ND | ND | ND | ND | ND | ND | 1.79 | ND | ND | ND | ND |
| BZ# 138 | ND | 1.21 | 1.56 | 1.56 | 1.16 | 1.92 | 1.77 | 0.6 | 0.93 | 1.61 | 1.9 | 1.79 | 0.66 | 0.62 | 1.9 | 1.24 |
| BZ# 153 | ND | 1.38 | 1.89 | 2.39 | 1.38 | 2.56 | 2.29 | 1.73 | 2.39 | 2.64 | 2.61 | 3.20 | 2.13 | 2.04 | 3.01 | 2.35 |
| BZ# 156 | ND | ND | ND | 2.35 | ND | ND | ND | 2.25 | ND | ND | ND | 2.78 | ND | ND | 2.48 | ND |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.42 | ND | ND | ND | ND |
| BZ# 170 | ND | 2.17 | 2.55 | 5.2 | ND | 2.24 | ND | 2.35 | 2.34 | 2.24 | 2.61 | 7.15 | 2.24 | 4.56 | 4.01 | 4.23 |
| BZ# 180 | ND | 1.52 | 1.81 | 1.62 | 0.97 | 1.6 | 1.52 | 1.02 | 1.19 | 1.55 | 1.59 | 1.87 | 1.18 | 1.04 | 1.91 | 1.31 |
| BZ# 183 | ND | ND | 1.3 | ND | ND | 3.34 | 4.19 | ND | 2.05 | 2.69 | 2.07 | 2.31 | 2.22 | ND | 3.6 | ND |
| BZ# 184 | ND | ND | 5.08 | 0.69 | 1.28 | 0.43 | ND | 0.46 | 0.52 | 0.97 | 1.67 | 3.50 | 0.44 | 2.05 | 3.24 | 6.80 |
| BZ# 187 | ND | 1.69 | 1.69 | 5.89 | ND | 6.48 | 17.16 | 17.07 | 9.95 | 13.34 | 9.8 | 10.75 | 5.25 | 4.66 | 7.8 | 5.82 |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | 3.86 | ND | ND | ND | ND |
| BZ# 206 | ND | 3.18 | 1.33 | 7.25 | ND | 9.56 | 3.22 | 2.11 | ND | 3.91 | 3.13 | 3.62 | 4.31 | 2.04 | 7.98 | 9.68 |
| BZ# 209 | ND | 2.16 | 1.55 | 1.66 | 2.94 | 3.15 | ND | 3.71 | ND | ND | ND | 0.94 | 2.42 | ND | 2.64 | 4.46 |
| TOTAL PCB ND=0 | 0.06 | 1.51 | 1.73 | 0.39 | 0.39 | 0.39 | 1.37 | 1.01 | 1.94 | A 123 | 0.69 | 1.68 | 0.26 | 2.05 | 0.59 | 0.34 |
| TOTAL PCB ND=1/2DL | 0.36 | 1.32 | 1.35 | 0.45 | 0.45 | 0.45 | 1.34 | 1.03 | 1.87 | 1.22 | 0.72 | 1.62 | 0.34 | 1.98 | 0.64 | 0.41 |
| TOTAL PCB ND=DL | 0.64 | 1.22 | 1.18 | 0.5 | 0.51 | 0.5 | 1.32 | 1.04 | 1.82 | 1.21 | 0.75 | 1.56 | 0.41 | 1.91 | 0.67 | 0.46 |

TABLE 9-25 Nereis virens (SAND WORM): UPTAKE RATIOS FOR PCB CONGENERS AND TOTAL PCBs

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.





TABLE 9-26A Macoma nasuta (BLUNT-NOSE CLAM): MEAN PCB CONGENER^(a) AND TOTAL PCB CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| Analyte (ug/kg) | 1nslde Site 104 | Outside Site | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighili | Craighill Angle East | Cralghiii Angle West | Craighili Entrance | Cralghiii Upper Range | Cutoff | Swan Point | Toichester North | Tolchester | Toichester |
|---------------------|--------------------|--------------|--------------------|-----------------------------------|-----------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|--------|------------|---------------------|------------|------------|
| BZ# 8 | 0.282 | ND | ND | ND | ND | ND | ND | ND | 0.28 | 0.276 | ND | 0.2.54 | ND | 0.288 | 0.28 | ND |
| BZ# 18 | 0.817 | ND | ND | ND | 2.68 | 0.476 | 0.308 | 0.2 | ND | 0.42 | ND | ND | ND | ND | ND | ND |
| BZ# 28 | 0.31 | ND | ND | ND | 0.302 | ND | 0.49 | 0.142 | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 44 | 0.204 | ND | ND | 0.212 | 0.226 | 0.746 | 0.114 | 0.12 | ND | 0.15 | 0.182 | ND | 0.717 | ND | ND | ND |
| BZ# 49 | 0.695 | 0.608 | ND | 0.988 | 1.07 | ND | ND | ND | ND | ND | ND | ND | 1.67 | ND | ND | ND |
| BZ# 52 | 1.02 | 0.757 | ND | 3.62 | 0.68 | 3.5 | 1.87 | 1.77 | 0.25 | 2.76 B | 0.296 | ND | 3.9 | 0.242 | ND | 0.256 |
| BZ# 66 | 0.47 | ND | ND | 0.911 | ND | 0.751 | 0.984 | 0.539 | ND | 0.559 | ND | ND | 0.726 | ND | ND | 0.230 |
| BZ# 77 | 0.863 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 87 | ND | ND | ND | 0.695 | ND | 0.848 | 0.154 | 0.144 | 0.436 | 0.172 | ND | ND | 0.773 | 0.365 | ND | ND |
| BZ# 101 | 0.663 | ND | ND | ND | ND | 0.642 | ND | 0.404 | 0.26 | 0.48 | 0.452 | ND | 0.754 | 0.365 | ND | ND |
| BZ# 105 | 0.192 | 0.978 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 118 | 0.251 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 126 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 128 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 138 | 0.344 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ#153 | 0.469 | ND | ND | ND | ND | 0.186 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ#156 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | 0 112 | ND | ND | ND | ND | ND | 0.128 | ND |
| BZ# 170 | 0.0985 | ND | 1.01 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.128 | ND. |
| BZ# 180 | 0.15 | ND | ND | 0.18 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 183 | ND | ND | ND | ND | ND | ND | 0.114 | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 184 | 0.366 | ND | ND | 0.208 | ND | 0.158 | ND | 0.102 | ND | ND | 0 114 | ND | ND | ND | ND | ND |
| BZ# 187 | 0.158 | ND | ND | 0.105 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.972 | ND | ND | ND | ND | ND |
| BZ# 206 | ND | ND | ND | ND | ND | ND | ND | ND | 0 198 | ND | 0.264 | ND | ND | ND | ND | 0.114 |
| BZ# 209 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 0.114 |
| TOTAL PCBs (ND=0) | 9.25 | 3.47 | 0.96 | 9.76 | 7.49 | 12.3 | 7.28 | 5.74 | 1.27 | 8.67 | 1.52 | 0 316 | 110 | 1.14 | 0.536 | 0.1/8 |
| TOTAL PCBs (ND=1/2) | 13.5 | 11.1 | 46.2** | 12.1 | 10.1 | 14.4 | 9.56 | 8.2 | 3.97 | 11 | 4.27 | 3.11 | 14.3 | 3.87 | 3.3 | 3.21 |
| TOTAL PCBs (ND=DL) | 17.7 | 18.7 | 91,5** | 14.5 | 12.8 | 16.6 | 11.8 | 10.6 | 6.68 | 13.3 | 7.02 | 5.9 | 16.6 | 6.59 | 6.07 | 6.02 |

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for total PCB congeners only

Total PCBs determined by summing congeners as specified in 1TM (Table 9-3) and multiplying total by a facor of 2 as per NOAA 1993 guidance.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

Note that the mean of total PCB for individual tissue replicates is not equivalent to the sum of mean congeners for ND=0 and ND=1/2DL.

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Cralghill Entrance | Craighill Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|--------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| BZ# 8 | ND | ND | 1.12 | ND | ND | ND | ND | ND | 2.33 | 2.3 | ND | 2.12 | ND | 2.4 | 2.33 | ND |
| BZ#18 | ND | ND | 1.09 | ND | 0.63 | 0.11 | 0.07 | 0.05 | ND | 0.1 | ND | ND | ND | ND | ND | ND |
| BZ# 28 | ND | ND | 1.08 | ND | 0.36 | ND | 0.58 | 0.17 | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 44 | ND | ND | 0.61 | 0.19 | 0.2 | 0.67 | 0.1 | 0.11 | ND | 0.13 | 0.16 | ND | 0.64 | ND | ND | ND |
| BZ# 49 | ND | 1.08 | 0.84 | 0.72 | 0.78 | ND | ND | ND | ND | ND | ND | ND | 1.22 | ND | ND | ND |
| BZ# 52 | ND | 1.43 | 1.9 | 6.68 | 1.25 | 6.46 | 3.46 | 3.27 | 0.46 | 5.09 | 0.55 | ND | 7.20 | 0.45 | ND | 0.47 |
| BZ# 66 | ND | ND | 1.85 | 5.52 | ND | 4.55 | 5.96 | 3.27 | ND | 3.39 | ND | ND | 4.4 | ND | ND | ND |
| BZ# 77 | ND | ND | 4.8 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 87 | ND | ND | ND | 10.69 | ND | 13.05 | 2.37 | 2.22 | 6.71 | 2.65 | ND | ND | 11.89 | 5.62 | ND | ND |
| BZ# 101 | ND | ND | 2.3 | ND | ND | 2.01 | ND | 1.26 | 0.81 | 1.5 | 1.41 | ND | 2.36 | 0.83 | ND | ND |
| BZ# 105 | ND | 1.78 | 0.78 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 118 | ND | ND | 1.13 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 126 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 128 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 138 | ND | ND | 1.97 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 153 | ND | ND | 3.88 | ND | ND | 3.1 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 156 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 169 | ND | ND | ND | ND | ND | ND | ND | ND | 2.04 | ND | ND | ND | ND | ND | 2.33 | ND |
| BZ# 170 | 8.74 | ND | 1.59 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 180 | ND | ND | 2.02 | 4.24 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 183 | ND | ND | ND | ND | ND | ND | 3.01 | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 184 | ND | ND | 4.04 | 4.16 | ND | 3.16 | ND | 2.04 | ND | ND | 2.28 | ND | ND | ND | ND | ND |
| BZ#187 | ND | ND | 1.97 | 2.39 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BZ# 195 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 14.95 | ND | ND | ND | ND | ND |
| BZ# 206 | ND | ND | ND | ND | ND | ND | ND | ND | 3.6 | 5 ND | 4.8 | ND | ND | ND | ND | 2.07 |
| BZ# 209 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 2.37 |
| TOTAL PCB ND=0 | 0.51 | 1.87 | 2.79 | 0.70 | a a a o .54 | 0.88 | 0.52 | 0.41 | 0.09 | 0.62 | 0.11 | 0.02 | 0.85 | 0.08 | 0.04 | 0.03 |
| TOTAL PCB ND=1/2DL | 4.70 | 1.13 | 1.14 | 0.74 | 0.62 | 0.88 | 0.58 | 0.5 | 0.24 | 0.67 | 0.26 | 0.19 | 0.87 | 0.24 | 0.2 | 0.2 |
| TOTAL PCB ND=DL | 5.15 | 1.06 | 0.98 | 0.78 | 0.68 | 0.89 | 0.63 | 0.57 | 0.36 | 0.71 | 0.38 | 0.32 | 0.89 | 0.35 | 0.32 | 0.32 |

TABLE 9-27 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR PCB CONGENERS

ND = not detected

Shaded and bolded values = uptake ratios for tissue-residues that statistically exceeded atleast one of the placement site/reference tissue-residues





TABLE 9-28A Nereis virens (SAND WORM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|-------------------------------------|-------------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Inside Site | Outside | Ocean | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Craighill | Cutoff | | Tolchester | Tolchester | Toichester |
| Analyte (ug/kg) | 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | Craighili | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Straightening |
| 1,2,4-TRICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2-DIPHENYLHYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,4-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| I-METHYLNAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 7 |
| 2,2'-OXYBIS(1-CIILOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4,6-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4-DIMETHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.6-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLORONAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHY'L-4,6-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 6 | ND | ND | ND | ND |
| 2-METHYLPHENOL | 75 | 85 | 93 | 196** | 78 | - 254** | 73 | 79 | 80 | 57 | 73 | 106 | 214** | 65 | 94 | 64. |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,3'-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,4-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,5,5-TRIMETHYL-2-CYCLOHEXENE-I-ONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BROMOPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZOIC ACID | 284 | 223 | 272 | 268 | 220 | 141 | 101 | 152 | 198 | 150 | 143 | 95 | 207 | 176 | 178 | 162 |
| BENZYL ALCOHOL | ND | ND | ND | 59 | 56 | ND | ND | ND | 54 | ND | ND | ND | 65 | ND | 51 | ND |
| BENZYL BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHYL) ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-ETHYLHEXYL) PHTHALATE | 496 | 543 | 608 | 366 | 560 | 246 | 164 | 146 | 224 | ND | 176 | 152 | 302 | 128 | 232 | 128 |
| DI-N-BUTYL PHTHALATE | ND | ND | 54 | ND | ND | ND | ND | ND | 68* | ND | ND | ND | ND | ND | ND | ND |
| DI-N-OCTY'L PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 196** |
| DIBENZOFURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLORO-1,3-BUTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE, N-METHYL-N-NITROSO | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODI-N-PROPYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODIPHENYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NITROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PENTACHLOROPHENOL | ND | ND | ND | 124 | ND | ND | ND | ND | 142** | . 144** | ND | ND | ND | ND | ND | ND |
| PHENOL | ND | ND | ND | 54 | 45 | ND | ND | ND | ND | ND | ND | ND | ND | ND | 54** | ND |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-28B Nereis virens (SAND WORM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED TO OUTSIDE SITE 104

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|-------------------------------------|--------------|-------------|-----------|-----------|-------------|----------|-----------|------------|-----------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Outside Site | Inside Site | Ocean | Eastern | Approach | Approach | 1.1 | Cralghili | Craighill Angie | Craighiil | Craighili | Cutoff | | Toichester | Toiehester | Toiehester |
| Analyte (ug/kg) | 104 | 104 | Reference | Extension | (Surfieiai) | (Cores) | Craighili | Angle East | West | Entranee | Upper Range | Angle | Swan Point | North | South | Straightening |
| 1,2,4-TRICIILOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2-DIPHENYLHYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.4-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1-METHYLNAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 7 |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4,6-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DIMETHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,6-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLORONAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND. |
| 2-METHYL-4,6-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 6 | ND | ND | ND | ND |
| 2-METHYLPHENOL | 85 | 75 | 93 | 196** | 78 | 254** | 73 | 79 | 80 | 57 | 73 | 106 | 214** | 65 | 94 | 64 |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3.3'-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,4-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,5,5-TRIMETHYL-2-CYCLOHEXENE-1-ONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BROMOPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZOIC ACID | 223 | 284 | 272 | 268 | 220 | 141 | 101 | 152 | 198 | 150 | 143 | 95 | 207 | 176 | 178 | 162 |
| BENZYL ALCOHOL | ND | ND | ND | 59 | 56 | ND | ND | ND | 54 | ND | ND | ND | 65 | ND | 51 | ND |
| BENZYL BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHYL) ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-ETHYLHEXYL) PHTHALATE | 543 | 496 | 608 | 366 | 560 | 246 | 164 | 146 | 224 | ND | 176 | 152 | 302 | 128 | 232 | i 28 |
| DI-N-BUTYL PHTHALATE | ND | ND | 54 | ND | ND | ND | ND | ND | 68 | ND | ND | ND | ND | ND | ND | ND |
| DI-N-OCTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 196** |
| DIBENZOFURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLORO-1,3-BUTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE, N-METHYL-N-NITROSO | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODI-N-PROPYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODIPHENYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NITROBENZENE | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PENTACHLOROPHENOL | ND | ND | ND | 124 | ND | ND | ND | ND | 142* | 144* | ND | ND | ND | ND | ND | ND |
| PHENOL | ND | ND ND | ND | 54 | 45 | ND | ND | ND | ND ND | ND | ND | ND | ND ND | ND | 54* | ND |

ND = not detected in Asterisks, shaded a

five tested replicates.

cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 ($p < 0.05^*$, p





TABLE 9-28C Nereis virens (SAND WORM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| | | | | Brewerton | C&D | | | | | | | | | 1. And 1. | | |
|-------------------------------------|-----------|-----------|-------------|-----------|-------------|--------------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| Analyta (uplica) | Ocean | Outside | Inside Site | Eastern | Approach | C&D Approach | | Cralghill | Craighill | Craighill | Craightt | Cutoff | | Tolchester | Tolchester | Tolchester |
| | Reference | Sile 104 | 104 | Extension | (Surricial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Straightening |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2 DIRHENIXI HYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| LAFTIWI NA DUTUAL DUT | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1-METHT LNAFHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | 7 |
| 2.2-OATBIS(I-CHLOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4,0-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLORONAFHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 METHYL 4 6 DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND |
| 2-METHYL-4,0-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHTUNAPHTHALENE | ND | ND | ND | ND IO(11) | ND | ND | ND 72 | ND | ND | ND | ND | 6 | ND | ND | ND | ND |
| | 93 | 60 | /5 | 1980 | /8 | 434 | /3 | /9 | 80 | 5/ | /3 | 106 | 214** | 65 | 94 | 64 |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3.3-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.6.6 TRIMETHING & CVCLOUEVENE LONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1 DROMODULENVIL DUENVIL ETUER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BROMOPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BENZUIC ACID | 2/2 | | 284 | 268 | 220 | 141 | 101 | 152 | 198 | 150 | 143 | 95 | 207 | 176 | 1/8 | 162 |
| DENZYL ALLOHOL | ND | ND | ND | 39 | 30 | ND | ND | ND | 34 | ND | ND | ND | 00 | ND | 51 | ND |
| BENZYL BUTYL PHIMALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIS(2-CHLOROETHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIS(2-CHLOROETHTL) ETHER | 140 | ND 641 | 106 | 244 | ND | 246 | 164 | ND 146 | ND 224 | ND | ND | UN LCO | 102 | 108 | 222 | ND |
| DIN(2-EINTLNEATL) FRITALATE | 600 | 343 MD | 490 | 300 | 300 | 240 | 104 | 140 | 224 | ND | 1/0 | 152 | 302 | 120 | 232 | 128 |
| DI-N-BUTTL PHTHALATE | 34 | ND | ND | ND | ND | ND | ND | ND | 68. | ND | ND | ND | ND | ND | ND | ND |
| DIPENZOEURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NU | ND | ND | 190** |
| DIBENZOFUKAN | ND | ND | ND | ND | ND | ND | ND | UN ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | DN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEAACHLOROPENZENE | ND | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND |
| HEXACHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N NITROSODI N BRODYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PUENOL | ND | ND | ND | 124 | ND | ND | ND | ND | 142** | 144** | ND | ND | ND | ND | ND | ND |
| ITTENUL | L ND | ND | ND | 54 | 45 | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | 54** | ND |

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Occan Reference site (p < 0.05*, p < 0.01**).

TABLE 9-29 Nereis virens (SAND WORM): UPTAKE RATIOS FOR SEMIVOLATILE ORGANIC COMPOUNDS

| Analyte Reference Sile 104 Extension (Surficial) (Cores) Angle East Angle West Upper Range Angle Volter Forth South | ND |
|---|------|
| I 74 IVE VIOL VIOL VIOL VIOL VIOL VIOL | ND |
| ND N | 110 |
| 1,2-DICHLOROBENZENE ND | ND |
| 1,2-DIPHENYLHYDRAZINE ND | ND |
| 1.4-DICHLOROBENZENE ND | ND |
| I-METHYLNAPHTHALENE ND | 1.51 |
| 2.2-OXYBIS(I-CHLOROPROPANE) ND | ND |
| 2,4,6-TRICHLOROPHENOL ND | ND |
| 2.4-DICHLOROPHENOL ND | ND |
| 2.4-DIMETHYLPHENOL ND | ND |
| 2.4-DINITROPHENOL ND | ND |
| 2,4-DINITROTOLUENE ND | ND |
| 2.6-DINITROTOLUENE ND | ND |
| 2-CHLORONAPHTHALENE ND | ND |
| 2-CHLOROPHENOL ND | ND |
| 2-METHYL-4,6-DINITROPHENOL ND | ND |
| 2-METHYLNAPHTHALENE ND | ND |
| 2-METHYLPHENOL 1.3 0.96 0.84 2.17 0.87 2.81 0.81 0.87 0.88 0.63 0.81 1.18 2.37 0.72 1.04 | 0.71 |
| 2-NITROPHENOL ND | ND |
| 3,3'-DICHLOROBENZIDINE ND | ND |
| 3.4-METHYLPHENOL ND | ND |
| 3.5.5-TRIMETHYL-2-CYCLOHEXENE-I-ONE ND | ND |
| 4-BROMOPHENYL PHENYL ETHER ND | ND |
| 4-CHLORO-3-METHYLPHENOL ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER ND | ND |
| 4-NITROPHENOL ND | ND |
| BENZOIC ACID 0.56 0.49 1.61 2.44 2 1.28 0.92 1.38 1.8 1.37 1.3 0.86 1.88 1.6 1.62 | 147 |
| BENZYLALCOHOL ND ND ND 1.22 1.15 ND ND ND 1.1 ND ND 1.34 ND 1.05 | ND |
| BENZYL BUTYL PHTHALATE ND | ND |
| BIS(2-CHLOROETHOXY)METHANE ND | ND |
| BIS(2-CHLOROETHYL) ETHER ND | ND |
| BIS(2-ETHYLHEXYL) PHTHALATE 1.2 0.79 1.77 2 3.05 1.34 0.89 0.8 1.22 ND 0.96 0.83 1.65 0.7 1.27 | 0.7 |
| DI-N-BUTYL PHTHALATE 0.98 ND | ND |
| DI-N-OCTYL PHTHALATE ND | 1.02 |
| DIBENZOFURAN ND | ND |
| DIETHYL PHTHALATE ND | ND |
| DIMETHYLPHTHALATE ND | ND |
| HEXACHLORO-1.3-BUTADIENE ND | ND |
| HEXACHLOROBENZENE ND | ND |
| HEXACHLOROCYCLOPENTADIENE ND | ND |
| HEXACHLOROFTHANE ND | ND |
| M-DICHLOROBENZENE ND | ND |
| METHANAMINE NMETHYL-N-NITROSO ND | ND |
| N-NITROSDEN-PROPYLAMINE ND | ND |
| NUTROSODIPHENYLAMINE ND | ND |
| NITROBENZENE ND | ND |
| PENTACHLOROPHENOL ND ND ND 114 ND ND ND ND 110 ND | ND |
| PHENOL ND ND ND 108 091 ND | ND |

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.









TABLE 9-30A Macoma nasuta (BLUNT-NOSE CLAM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED TO INSIDE SITE 104

| Analyte (ug/kg) | Inside Site 104 | Outside Site | Ocean Reference | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | Craighili Entrance | Craighill Upper Range | Cutoff | Swan | Tolchester North | Tolchester Snuth | Tolchester Stralghtening |
|-----------------------------------|--------------------|--------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|----------|-----------|---------------------|---------------------|-----------------------------|
| 1.2.4-TRICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1 2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2-DIPHENYLHYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 14-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| I-METHYI NAPHTHALENE | 12 | ND | ND | 20 | ND | ND | 8 | ND | 11 | 14 | ND | ND | ND | 11 | 16 | 16 |
| 2 2'-OXYBIS(I-CHLOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 246-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DIMETHYL PHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 24-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 26-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLORONAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOI | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYL-4 6-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | 14 | 17 | ND | ND | 6 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLPHENOL | 54 | ND | 42 | ND | 45 | ND | ND | ND | ND | NU ARE | 62 | ND 61 | ND | ND | NU | 9 |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | J2 | JI | ND | ND | NID. | 10 - + |
| 3 3'-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | 120 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3.4. METHYL PHENOL | 66 | 40 | ND | 100 | 10 | 127 | 60 | 20 | ND | NU | ND | NU | ND PT AND | ND | NU | NU |
| 3.5 STRIMETHVL 2 CVCLOHEVENE LONE | ND | ND | ND | ND | J31- | ND | bitto | 39 | 00 | | ND | 45 | ND | 30 | 45 | 52 |
| 4.BROMOPHENYI PHENYI ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NU |
| 4-CHLORO.3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENVI PHENVI ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOI | ND | ND | ND | ND | ND | ND | ND | ND | ND | 12440 | ND | ND | ND | ND | ND | ND |
| BENZOIC ACID | 4411 | 3050 | 5170 | 2560 | 4500 | 1740 | 3440 | 1500 | 3090 | 4590 | 3100 | 2600 | 2000 | 100 MD | ND SOART | ND 6040 |
| BENZYL ALCOHOL | 170 | 3030 | 63 | 2300 | 4300 | 59 | 3440 | 1300 | 3000 ND | 4380 | SIU | 3000 | 2000 | 129 | 05 | 3040 |
| RENZVI DUTVI PHTHALATE | ND | ND | ND | ND | J4 | ND | ND | ND | ND | ND | ND | ND | NID | 120 ND | 9J | 09 |
| BIS(2-CHLOROFTHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHVI) ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CTILOROETHTE) ETHER | 203 | Salar Adcorp | Rest BEACOR | 278 | 170 | 376 | 304 | 206 | 202 | 322 | 300 | 224 | 136 | 107 | 115 | 160 |
| DIS(2-EINTEREATE) FRINALATE | ND | ND | ND | 2/0 ND | ND | 370 | 374 ND | 290 | 202 | | ND | ALC NID | 430 ND | ND | ND | ND |
| DLN OCTVI PUTUALATE | ND | ND | ND | ND | 102 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIBENZOFURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIFTHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEY ACHLOPO 1 3 PUTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROPENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACILOROCTCLOFENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M.DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE N.METHYL.N.NETBOSO | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N.NITROSODI NI DRODVI AMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | UN | ND | ND | NU | ND |
| N NITROSODIRUENVI AMDIE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NITROBENZENE | UN ND | ND | ND | ND | ND | ND | ND | ND | ND | NU | ND | ND | ND | ND | ND | ND |
| PENTACHLOPOPUENOL | ND | ND | UN ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PHENOI | 124 | ND | ND EL | ND | ND | ND | ND | ND | ND | ND | ND | UN | NU | ND | NU | NU |
| TIMITUM | 1 24 | 09 | 20 | 100 | ND | 23 | 23 | 40 | ND | 00 | ND | NU | 07 | 44 | 22 | 48 |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-30B Macoma nasuta (BLUNT-NOSE CLAM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED **TO OUTSIDE SITE 104**

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|-------------------------------------|----------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|-------|------------|------------|---------------|
| | 0 | Inside | Ocean | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Craighill | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| Analyte (ug/kg) | Site 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Angle | Point | North | South | Straightening |
| 1.2.4-TRICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND. | ND | ND | ND | ND | ND |
| 1,2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2-DIPHENYLHYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| I,4-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| I-METHYLNAPHTHALENE | ND | 12 | ND | 20 | ND | ND | 8 | ND | 11 | 14 | ND | ND | ND | 11 | 16* | 16* |
| 2,2'-OXY BIS(1-CHLOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | . ND | ND | ND | ND | ND | ND | ND |
| 2.4.6-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DIMETHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2,4-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.6-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLORONAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYL-4,6-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | 17 | 14 | ND | ND | 6 | ND | ND | ND | ND | ND | ND | ND | 6 | ND | ND. | 9 |
| 2-METHYLPHENOL | ND | 54 | 42 | ND | 45 | ND | ND | ND | ND | 5 79* | 52 | 51 | ND | 81** | 77* | 77** |
| 2-NITROPHENOL | ND | ND | ND. | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,3'-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | 129 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,4-METHYLPHENOL | 49 | 66 | ND | 109 | 351* | 189* | 69 | 39 | 80 | 91 | ND | 45 | 734** | 36 | 45 | 52 |
| 3.5,5-TRIMETHYL-2-CYCLOHEXENE-I-ONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BROMOPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | 134** | ND | ND | ND | ND | ND | ND |
| BENZOIC ACID | 3050 | 4411* | 5170 | 2560 | 4500* | 1740 | 3440 | 1500 | 3080 | 4580** | 3100 | 3600 | 2000 | 5980** | 6940** | 5040** |
| BENZYL ALCOHOL | 373 | 170 | 63 | 97 | 94 | 58 | 48 | ND | ND | 64 | ND | ND | 64 | 128 | 95 | 89 |
| BENZYL BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHYL) ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-ETHYLHEXYL) PHTHALATE | 445 | 303 | 884##B | 278 | 170 | 376 | 394 | 296 | 202 | 322 | 300 | 224 | 436 | 107 | 115 | 160 |
| DI-N-BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DI-N-OCTYL PHTHALATE | ND | ND | ND | ND | 192 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIBENZOFURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL PHTHALATE | . ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLORO-1,3-BUTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROBENZENE | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M-DICHLOROBENZENE | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE, N-METHYL-N-NITROSO | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODI-N-PROPY LAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODIPHENYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NITROBENZENE | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND | ND |
| PENIACHLOROPHENOL | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| IPHENOL | 69 | 9 124 | 56 | 100 | ND | 53 | 53 | -18 | ND | 60 | ND | ND | 69 | 44 | 55 | 48 |

B =detected in laboratory,

ND = not detected in an

tested replicates.

at least one of the five test replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).



TABLE 9-30C Macoma nasuta (BLUNT-NOSE CLAM): MEAN SEMIVOLATILE ORGANIC COMPOUND CONCENTRATIONS (UG/KG) COMPARED TO OCEAN REFERENCE

| Ansiste (na/ka) | Ocean Reference | Outside Site | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Cralabili | Craighill Angle | Craighill Angle Wast | Craighill | Craighill | Cutoff | Su an Point | Tolchester | Tokhester | Tolchester |
|-------------------------------------|--------------------|--------------|-----------------|--------------------------------|--------------------------------|----------------------------|-----------|-----------------|-------------------------|-----------|-----------|--------|-------------|------------|-----------|---------------|
| 1.2.4.TRICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | NOTIN | ND | Straightening |
| L 2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2 DIPHENVI HYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| LADICIII OROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | 17+ | 20 | ND | ND | 8 | ND | 11 | 14 | ND | ND | ND | 11 | 16* | 160 |
| 2.2'-OXYBIS(LCHLOROPROPANE) | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4.6-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4-DIMETHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4. DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.4 DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2.6-DINITROTOLUENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND. | ND | ND | ND | ND |
| 2-CHLORONAPHTHALENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYL-4.6-DINITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | ND | 17** | 14* | ND | 6 | ND | ND | ND | ND | ND | ND | ND | 6 | ND | ND | 9 |
| 2-METHYLPHENOL | 42 | ND | 54 | ND | 45 | ND | ND | ND | ND | 79** | 52* | 51* | ND | 81** | 77** | 77** |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,3'-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | 129 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,4-METHYLPHENOL | ND | 49** | 56** | 109** | 351 ** | 189** | 69** | 39** | 80** | 91== | ND | 45** | 734** | 3600 | 45** | 52** |
| 3.5.5-TRIMETHYL-2-CYCLOHEXENE-1-ONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BROMOPHENYL PIIENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | 134** | ND | ND | ND | ND | ND | ND |
| BENZOIC ACID | 5170 | 30.50 | 4411 | 2560 | 4500 | 1740 | 3440 | 1500 | 3080 | 4580 | 3100 | 3600 | 2000 | 5980 | 6940 | .5040 |
| BENZYL ALCOHOL | 63 | 373** | 170 | 97 | 94 | 58 | 48 | ND | ND | 64 | ND | ND | 64 | 128* | 9.5 | 89 |
| BENZYL BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHOXYIMETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHYL) ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-ETHYLHEXYL) PHTHALATE | 884 | 44.5 | 303 | 278 | 170 | 376 | 394 | 296 | 202 | 322 | 300 | 224 | 436 | 107 | 115 | 160 |
| DI-N-BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DI-N-OCTYL PHTHALATE | ND | ND | ND | ND | 192 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIBENZOFURAN | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIETHYL PHTHALATE | ND | ND | ND | ND. | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLORO-1,3-BUTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROETHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| M-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE, N-METHYL-N-NITROSO | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODI-N-PROPYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N-NITROSODIPHENYLAMINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| NITROBENZENE | ND | NE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PENTACHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PHENOL | 56 | 69* | 1 124* | 100* | ND | 53 | 53 | 48 | ND | 60 | ND | ND | 69 | 44 | 55 | 48 |

ND = not detected in any of five tested replicates.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

TABLE 9-31 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR SEMIVOLATILE ORGANIC COMPOUNDS

| | | 0.11 | | Brewerton | C&D | C&D | Contraction | | | | | | | | | |
|--------------------------------------|-----------|----------|-------------|----------------|------------|----------|-------------|------------|------------|--------------|-------------|--------|-------|---------------|------------|---------------|
| Analyte | Reference | Site 104 | Site 104 | Eastern | Approach | Approach | Craigniii | Craighill | Craighill | Craighili | Craighill | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| 1.2.4-TRICHI OROBENZENE | ND | ND | SILE IO4 | ND | (Surneial) | (Cores) | NID | Angle East | Angle west | Entrance | Upper Range | Angle | Point | North | South | Straightening |
| 1 2-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1 2-DIPHENYL HYDRAZINE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| L4-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| I-METHYI NAPHTHAI ENE | ND | ND | 1 1 20 | 4.15 | ND | ND | 1.62 | ND | 221 | 204 | ND | ND | ND | 2.20 | ND | NU |
| 2 2'-OXYBIS(I-CHI OROPROPANE) | ND | ND | ND | IJ | ND | ND | 1.03 | ND | 2.31 | 2.94 | ND | ND | ND | 2.39 | 2.41 J.41 | 3.33 |
| 2.4.6-TRICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DICHLOROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 4-DIMETHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 24-DINITROPHENOI | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 CHI OPONADUTHAI ENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2 CHLORONAFHTHALENE | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-CHLOROPHENOL | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYL-4,0-DINTIKUPHENUL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 2-METHYLNAPHTHALENE | ND | 2.14 | 2.09 | ND | 1.35 | ND | ND | ND | ND | ND | ND | ND | 1.31 | ND | ND | 1.88 |
| | 0.34 | ND | 1.22 | ND | 1.02 | ND | ND | ND | ND | 2 3 2 4 1.79 | 1.19 | E.15 | ND | 1.83 | 1.74 | 1.75 |
| 2-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,3-DICHLOROBENZIDINE | ND | ND | ND | ND | ND | 0.92 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 3,4-METHYLPHENOL | ND ND | 0.70 | 0.54 | 8.58 | 1.83 | 1.00 | 0.36 | 0.70 | 0.A1 | 0.43 | ND | 0.24 | 3.88 | 0.19 | 0.24 | 0.27 |
| 3,5,5-1 KIMETHYL-2-CYCLOHEXENE-1-ONE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-BRUMUPHEN YL PHEN YL ETHER | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLORO-3-METHYLPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-CHLOROPHENYL PHENYL ETHER | ND | ND | ND ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 4-NITROPHENOL | ND | ND | ND | ND | ND | ND | ND | ND | ND | 1.91 | ND | ND | ND | ND | ND | ND |
| DENZUIC ACID | 1.04 | 0.00 | 2095 | 0.53 | 0.93 | 0.36 | 0.71 | 0.31 | 0.64 | 0.95 | 0.64 | 0.74 | 0.41 | - and 51 1.24 | 143 | 1.04 |
| DENZYL ALCOHOL | 0.31 | 145 | 3.05 | 0.73 | 0.7 | 0.43 | 0.36 | ND | ND | 0.48 | ND | ND | 0.48 | 0.96 | 0.71 | 0.66 |
| BENZYL BUTYL PHIHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHOXY)METHANE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND | ND |
| BIS(2-CHLOROETHYL) ETHEK | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| BIS(2-ETHYLHEXYL) PHIHALATE | 1611 | . 0.8 | 1.14 | 1.95 | 1.19 | 2.64 | 2.76 | 2.08 | 1.42 | 2.26 | 2.1 | 1.57 | 3.06 | 0.75 | 0.81 | 1.12 |
| DI-N-BUTYL PHTHALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DIPENZOEURAN | ND | NU | ND ND | ND | 1.04 | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND | ND |
| DIETUVI DUTUALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND | ND | ND | ND | ND | ND |
| DIMETHYL DUTUALATE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROPENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCYCLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| HEXACHLOROCT CLOPENTADIENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND ND | ND |
| M-DICHLOROBENZENE | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| METHANAMINE N.METHVI N.NITROSO | ND | ND | IND NID | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| N.NITROSODI N PROPYLAMDIE | ND | NU | ND | ND | ND | ND | ND | ND | ND | NU | ND | ND | ND | ND | ND | ND |
| NUTROSODIDUENVI AVDIE | ND | NU | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| | ND | NU | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| DENTA CHI ODOBLENOI | ND | NU | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| PLENOL | ND | ND | ND | ND | ND | ND | UN 0.72 | ND | I ND | NU | ND | ND | ND | ND | ND | ND |
| THENOL | 0.04 | A STATE | A STATE AND | 1 2 7 1 1 Last | ND ND | 0.72 | 0.75 | 0.66 | ND ND | 0.83 | I ND | I ND | 0.94 | 0.61 | 0.75 | 0.65 |

ND = not detected

Shaded and bolded values = uptake ratios for tissue-residues that statistically exceeded atleast one of the placement site/reference tissue-residues



TABLE 9-32ANereis virens (SAND WORM):MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED TOINSIDE SITE 104^(a)

| | Inside Site | Outside | Осеал | Brewerton Fastern | C&D Approach | C&D Approach | | Cralabill | Croichill | Craiabill | Crolabill | Cutoff | | Talahasta | T-1-1 | |
|------------------------|-------------|----------|-----------|----------------------|-----------------|-----------------|-----------|------------|------------|-----------|-------------|--------|------------|-----------|--------|-------------------------------|
| Analyte (ng/kg) | 104 | Site 104 | Reference | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | · loichester Straightening |
| 2,3.7,8-TCDD | 0.0783 | 0.0712 | 0.102 | ND | 0.0602 | ND | ND | 0.132 | 0.0819 | ND | 0.118 | ND | 0.0861 | ND | ND | ND |
| 1,2,3,7,8-PECDD | 0.139 | 0.272 | 0.364B | 0.148 | 0.279 | 0.132 | 0.187 | 0.154 | 0.131 | 0.318 | 0.27 | 0.573B | 0.224 | 0.177 | 0.302B | 0.0831 |
| 1,2,3,4,7.8-HXCDD | 0.0707 | 0.0479 | 0.111 | 0.167 | 0.0608 | ND | ND | 0.174 | 0.0742 | 0.0843 | 0.108 | ND | 0.0691 | ND | ND | 0.125 |
| 1,2,3,6,7,8-IIXCDD | 0.266 | 0.258 | 0.399 | 0.417 | 0.267 | 0.295 | 0.39 | 0.318 | 0.245 | 0.304 | 0.307 | 0.255 | 0.264 | 0.26 | 0.206 | 0.223 |
| 1,2,3.7,8,9-HXCDD | 0.147 | 0.139 | 0.196B | 0.228 | 0.139 | 0.105 | 0.121 | 0.133 | 0.138 | 0.202 | 0.163 | ND | 0.114 | 0.101 | ND | 0.107 |
| 1,2,3,4,6,7,8-HPCDD | 1.48 | 1.48 | 2.83 | 1.95 | 1.31 | 1.24 | 1.76 | 1.47 | 1.57 | 1.35 | 1.51 | 1.46 | 1.3 | 1.31 | 1.07 | 1.04 |
| OCDD | 7.31 | 7.05 | 7.44 | 7.78 | 6.56 | 6.95 | 7.51 | 8.42 | 10.4B | 6.45 | 7.54 | 7.44 | 7.91 | 8.03 | 7.56 | 5.5 |
| 2,3,7,8-TCDF | 0.742 | 0.712 | 0.844 | 0.828 | 0.54 | 0.598 | 0.729 | 0.814 | 0.665 | 0.692 | 0.954 | 0.894 | 0.724 | 0.703 | 0.716 | 0.609 |
| 1,2,3,7,8-PECDF | 0.101 | 0.0944 | 0.152B | 0.0814 | 0.104 | 0.0844 | ND | 0.105 | 0.0962 | 0.115 | 0.178B | 0.115 | 0.088 | 0.1 | 0.0535 | 0.0717 |
| 2,3,4,7,8-PECDF | 0.489 | 0.192 | 0.249 | 0.231 | 0.167 | 0.199 | 0.212 | 0.223 | 0.194 | 0.202 | 0.249 | 0.2 | 0.194 | 0.207 | 0.16 | 0.141 |
| 1,2,3,4,7,8-HXCDF | 0.112 | 0.0749 | 0.104 | 0.0931 | 0.12 | ND | 0.0857 | 0.0841 | 0.115 | 0.0829 | 0.169 | 0.109 | 0.0978 | 0.121 | 0.0881 | 0.0942 |
| 1,2,3,6,7,8-HXCDF | 0.0829 | 0.0694 | 0.103 | ND | 0.0853 | ND | 0.084 | 0.0612 | 0.0799 | 0.0992 | 0.108 | ND | 0.0626 | 0.0872 | ND | 0.061 |
| 2,3,4,6,7,8-HXCDF | 0.0772 | 0.0918 | 0.119B | ND | 0.0889 | ND | 0.121 | 0.0667 | 0.0838 | 0.0775 | 0.117 | 0.0629 | 0.0773 | 0.0783 | ND | 0.0687 |
| 1,2,3,7,8,9-HXCDF | ND | ND | ND | ND | 0.0639 | ND | ND | ND | ND | 0.0968 | 0.0775 | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 0.245 | 0.229 | 0.432 | 0.357 | 0.224 | 0.251 | 0.318 | 0.273 | 0.248 | 0.243 | 0.304 | 0.245 | 0.231 | 0.247 | 0.199 | 0.225 |
| 1,2,3,4,7,8,9-HPCDF | ND | ND | ND | ND | 0.0452 | ND | ND | 0.0917 | 0.0472 | ND | 0.0817 | ND | ND | ND | ND | ND |
| OCDF | 0.251 | 0.216 | 0.333B | ND | 0.22 | ND | ND | 0.207 | 0.205 | 0.201 | 0.363 | 0.285 | 0.213 | 0.246 | ND | ND |
| DIOXINS TEQ (ND=0) | 0.37 | 0.153 | 0.251 | 0.273 | 0.159 | 0.151 | 0.211 | 0.211 | 0.218 | 0.187 | 0.205 | 0.121 | 0.22 | 0.103 | 0.1 | 0.147 |
| DIOXINS TEQ (ND=1/2DL) | 0.436 | 0.176 | 0.256 | 0.427 | 0.167 | 0.33 | 0.437 | 0.358 | 0.263 | 0.302 | 0.225 | 0.258 | 0.301 | 0.229 | 0.204 | 0.26 |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for Individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-32B Nereis virens (SAND WORM): MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED TO OUTSIDE SITE 104^(a)

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|------------------------|----------|-------------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Outside | Inside Site | Ocean | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Craighill | Cutoff | | Tolchester | Tolchester | Tolchester |
| Analyte (ng/kg) | Site 104 | 104 | Reference | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Straightening |
| 2,3.7.8-TCDD | 0.0712 | 0.0783 | 0.102 | ND | 0.0602 | ND | ND | 0.132 | 0.0819 | ND | 0.118 | ND | 0.0861 | ND | ND | ND |
| 1.2.3.7.8-PECDD | 0.272 | 0.139 | 0.364 | 0.148 | 0.279 | 0.132 | 0.187 | 0.154 | 0.131 | 0.318 | 0.27 | 0.573 | 0.224 | 0.177 | 0.302 | 0.0831 |
| 1,2,3,4,7,8-11XCDD | 0.0479 | 0.0707 | 0.111 | 0.167 | 0.0608 | ND | ND | 0.174 | 0.0742 | 0.0843 | 0.108 | ND | 0.0691 | ND | ND | 0.125 |
| 1,2,3,6,7,8-HXCDD | 0.258 | 0.266 | 0.399 | 0.417 | 0.267 | 0.295 | 0.39 | 0.318 | 0.245 | 0.304 | 0.307 | 0.255 | 0.264 | 0.26 | 0.206 | 0.223 |
| 1,2,3,7,8,9-HXCDD | 0.139 | 0.147 | 0.196 | 0.228 | 0.139 | 0.105 | 0.121 | 0.133 | 0.138 | 0.202 | 0.163 | ND | 0.114 | 0.101 | ND | 0.107 |
| 1,2,3,4.6,7,8-HPCDD | 1.48 | 1.48 | 2.83 | 1.95 | 1.31 | 1.24 | 1.76 | 1.47 | 1.57 | 1.35 | 1.51 | 1.46 | 1.3 | 1.31 | 1.07 | 1.04 |
| OCDD | 7.05 | 7.31 | 7.44 | 7.78 | 6.56 | 6.95 | 7.51 | 8.42 | 10.4B | 6.45 | 7.54 | 7.44 | 7.91 | 8.03 | 7.56 | 5.5 |
| 2.3.7,8-TCDF | 0.712 | 0.742 | 0.844 | 0.828 | 0.54 | 0.598 | 0.729 | 0.814 | 0.665 | 0.692 | 0.954 | 0.894 | 0.724 | 0.703 | 0.716 | 0.609 |
| 1,2,3,7,8-PECDF | 0.0944 | 0.101 | 0.152 | 0.0814 | 0.104 | 0.0844 | ND | 0.105 | 0.0962 | 0.115 | 0.178 | 0.115 | 0.088 | 0.1 | 0.0535 | 0.0717 |
| 2,3.4,7.8-PECDF | 0,192 | 0.489B | 0.249B | 0.231 | 0.167 | 0.199 | 0.212 | 0.223 | 0.194 | 0.202 | 0.249B | 0.2 | 0.194 | 0.207 | 0.16 | 0.141 |
| 1,2,3,4,7,8-HXCDF | 0.0749 | 0.112 | 0.104B | 0.0931 | 0.12 | ND | 0.0857 | 0.0841 | 0.115 | 0.0829 | 0.169B | 0.109 | 0.0978 | 0.121 | 0.0881 | 0.0942 |
| 1,2,3,6,7,8-HXCDF | 0.0694 | 0.0829 | 0.103B | ND | 0.0853 | ND | 0.084 | 0.0612 | 0.0799 | 0.0992 | 0.108B | ND | 0.0626 | 0.0872 | ND | 0.061 |
| 2,3,4,6,7,8-HXCDF | 0.0918 | 0.0772 | 0.119 | ND | 0.0889 | ND | 0.121 | 0.0667 | 0.0838 | 0.0775 | 0.117 | 0.0629 | 0.0773 | 0.0783 | ND | 0.0687 |
| 1,2,3,7.8,9-HXCDF | ND | ND | ND | ND | 0.0639 | ND | ND | ND | ND | 0.0968 | 0.0775 | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 0.229 | 0.245 | 0.432 | 0.357 | 0.224 | 0.251 | 0.318 | 0.273 | 0.248 | 0.243 | 0.304 | 0.245 | 0.231 | 0.247 | 0.199 | 0.225 |
| 1,2,3,4,7,8,9-HPCDF | ND | ND | ND | ND | 0.0452 | ND | ND | 0.0917 | 0.0472 | ND | 0.0817 | ND | ND | ND | ND | ND |
| OCDF | 0.216 | 0.251 | 0.333 | ND | 0.22 | ND | ND | 0.207 | 0.205 | 0.201 | 0.363 | 0.285 | 0.213 | 0.246 | ND | ND |
| DIOXINS TEQ (ND=0) | 0.153 | 0.37** | 0.251* | 0.273** | 0.159 | 0.151 | 0.211 | 0.211 | 0.218 | 0.187* | 0.205 | 0.121 | 0.22 | 0.103 | 0.1 | 0.147 |
| DIOXINS TEQ (ND=1/2DL) | 0.176 | 0.436** | 0.256* | 0.427** | 0.167 | 0.33** | 0.437** | 0.358** | 0.263* | 0.302** | 0.225 | 0.258* | 0.301** | 0.229 | 0.204 | 0.26* |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than Outside Site 104 (p < 0.05*, p < 0.01**).



TABLE 9-32C Nereis virens (SAND WORM): MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED TO OCEAN REFERENCE^(a)

| 1 1 2 6 1 1 1 | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|------------------------|-----------|--------------|-------------|-----------|-------------|----------|------------|------------|------------|-----------|-------------|--------------|------------|------------|------------|---------------|
| Applute (no/lin) | Ocean | Outside Site | Inside Site | Eastern | Approach | Approach | Contabilit | Craighill | Craighill | Craighill | Cralghill | | | Tolchester | Tolchester | Tolchester |
| Analyte (ng/kg) | Reference | 104 | 104 | Extension | (Surnicial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Cutoff Angle | Swan Point | North | South | Straightening |
| 2.3.7.8-TCDD | 0.102 | 0.0712 | 0.0783 | ND | 0.0602 | ND | ND | 0.132 | 0.0819 | ND | 0.118 | ND | 0.0861 | ND | ND | ND |
| 1,2,3.7,8-PECDD | 0.364 | 0.272 | 0.139 | 0.148 | 0.279 | 0.132 | 0.187 | 0.154 | 0.131 | 0.318 | 0.27 | 0.573 | 0.224 | 0.177 | 0.302 | 0.0831 |
| 1.2.3,4.7.8-HXCDD | 0.111 | 0.0479 | 0.0707 | 0.167 | 0.0608 | ND | ND | 0.174 | 0.0742 | 0.0843 | 0.108 | ND | 0.0691 | ND | ND | 0.125 |
| 1.2.3.6,7,8-HXCDD | 0.399 | 0.258 | 0.266 | 0.417 | 0.267 | 0.295 | 0.39 | 0.318 | 0.245 | 0.304 | 0.307 | 0.255 | 0.264 | 0.26 | 0.206 | 0.223 |
| 1.2,3,7,8.9-HXCDD | 0.196 | 0.139 | 0.147 | 0.228 | 0.139 | 0.105 | 0.121 | 0.133 | 0.138 | 0.202 | 0.163 | ND | 0.114 | 0.101 | ND | 0.107 |
| 1,2.3,4,6.7,8-11PCDD | 2.83 | 1.48 | 1.48 | 1.95 | 1.31 | 1.24 | 1.76 | 1.47 | 1.57 | 1.35 | 1.51 | 1.46 | 1.3 | 1.31 | 1.07 | 1.04 |
| OCDD | 7.44 | 7.05 | 7.31 | 7.78 | 6.56 | 6.95 | 7.51 | 8.42 | 10.4B | 6.45 | 7.54 | 7.44 | 7.91 | 8.03 | 7.56 | 5.5 |
| 2.3.7.8-TCDF | 0.844 | 0.712 | 0.742 | 0.828 | 0.54 | 0.598 | 0.729 | 0.814 | 0.665 | 0.692 | 0.954 | 0.894 | 0.724 | 0.703 | 0.716 | 0.609 |
| 1,2,3,7,8-PECDF | 0.152 | 0.0944 | 0.101 | 0.0814 | 0.104 | 0.0844 | ND | 0.105 | 0.0962 | 0.115 | 0.178 | 0.115 | 0.088 | 0.1 | 0.0535 | 0.0717 |
| 2,3,4,7,8-PECDF | 0.249 | 0.192 | 0.489B | 0.231 | 0.167 | 0.199 | 0.212 | 0.223 | 0.194 | 0.202 | 0.249 | 0.2 | 0.194 | 0.207 | 0.16 | 0.141 |
| 1,2,3,4,7,8-HXCDF | 0.104 | 0.0749 | 0.112 | 0.0931 | 0.12 | ND | 0.0857 | 0.0841 | 0.115 | 0.0829 | 0.169*B | 0.109 | 0.0978 | 0.121 | 0.0881 | 0.0942 |
| 1,2,3,6.7,8-IIXCDF | 0.103 | 0.0694 | 0.0829 | ND | 0.0853 | ND | 0.084 | 0.0612 | 0.0799 | 0.0992 | 0.108 | ND | 0.0626 | 0.0872 | ND | 0.061 |
| 2.3.4.6.7.8-HXCDF | 0.119 | 0.0918 | 0.0772 | ND | 0.0889 | ND | 0.121 | 0.0667 | 0.0838 | 0.0775 | 0.117 | 0.0629 | 0.0773 | 0.0783 | ND | 0.0687 |
| 1,2,3,7,8,9-HXCDF | ND | ND | ND | ND | 0.0639 | ND | ND | ND | ND | 0.0968 | 0.0775 | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 0.432 | 0.229 | 0.245 | 0.357 | 0.224 | 0.251 | 0.318 | 0.273 | 0.248 | 0.243 | 0.304 | 0.245 | 0.231 | 0.247 | 0.199 | 0.225 |
| 1,2,3,4,7.8,9-HPCDF | ND | ND | ND | ND | 0.0452 | ND | ND | 0.0917 | 0.0472 | ND | 0.0817 | ND | ND | ND | ND | ND |
| OCDF | 0.333 | 0.216 | 0.251 | ND | 0.22 | ND | ND | 0.207 | 0.205 | 0.201 | 0.363 | 0.285 | 0.213 | 0.246 | ND | ND |
| DIOXINS TEQ (ND=0) | 0.251 | 0.153 | 0.37 | 0.273 | 0.159 | 0.151 | 0.211 | 0.211 | 0.218 | 0.187 | 0.205 | 0.121 | 0.22 | 0.103 | 0.1 | 0.147 |
| DIOXINS TEQ (ND=1/2DL) | 0.256 | 0.176 | 0.436** | 0.427** | 0.167 | 0.33 | 0.437** | 0.358* | 0.263 | 0.302 | 0.225 | 0.258 | 0.301 | 0.229 | 0.204 | 0.26 |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and bolded cells Indicate sltes where mean tlssue residues were statistically higher than the Ocean Reference slte (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighiil | Craighiil Angle East | Craighill Angle West | Craighiil Entrance | Craighill Upper Range | Cutoff Angie | Swan Point | Tolchester North | Toichester South | Tolchester Straightening |
|-----------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| 2,3,7,8-TCDD | 1.43 | 1.03 | 1.14 | ND | 0.87 | ND | ND | 1.92 | 1.19 | ND | 1.72 | ND | 1.25 | ND | ND | ND |
| 1,2,3,7,8-PECDD | 2.48 | 1.35 | 0.69 | 0.73 | 1.38 | 0.65 | 0.93 | 0.76 | 0.65 | 1.58 | 1.34 | 2.84 | 1.11 | 0.88 | 1.50 | 0.41 |
| 1,2,3,4,7,8-HXCDD | 0.96 | 0.68 | 1.01 | 2.38 | 0.86 | ND | ND | 2.47 | 1.05 | 1.2 | 1.53 | ND | 0.98 | ND | ND | 1.77 |
| 1,2,3,6,7,8-HXCDD | 1.60 | 2.1 | 2.16 | 3.39 | 2.17 | 2.4 | 3.17 | 2.59 | 1.99 | 2.47 | 2.5 | 2.07 | 2.14 | 2.12 | 1.68 | 1.81 |
| 1,2,3,7,8,9-HXCDD | 2.00 | 1.53 | 1.61 | 2.51 | 1.52 | 1.15 | 1.32 | 1.46 | 1.51 | 2.22 | 1.78 | ND | 1.25 | 1.11 | ND | 1.18 |
| 1,2,3,4,6,7,8-HPCDD | 0.91 | 1.01 | 1 | 1.33 | 0.89 | 0.84 | 1.2 | 1 | 1.07 | 0.92 | 1.03 | 0.99 | 0.88 | 0.89 | 0.73 | 0.71 |
| OCDD | 0.32 | 0.71 | 0.73 | 0.78 | 0.66 | 0.69 | 0.75 | 0.84 | 1.04 | 0.64 | 0.75 | 0.74 | 0.79 | 0.8 | 0.76 | 0.55 |
| 2,3,7,8-TCDF | 2.11 | 1.5 | 1.56 | 1.74 | 1.14 | 1.26 | 1.54 | 1.71 | 1.4 | 1.46 | 2.01 | 1.88 | 1.53 | 1.48 | 1.51 | 1.28 |
| 1,2,3,7,8-PECDF | 3.14 | 1.45 | 1.54 | 1.25 | 1.59 | 1.29 | ND | 1.61 | 1.48 | 1.77 | 2.73 | 1.76 | 1.35 | 1.54 | 0.82 | 1.1 |
| 2,3,4,7,8-PECDF | 1.56 | 1.75 | 4.46 | 2.1 | 1.52 | 1.81 | 1.93 | 2.03 | 1.77 | 1.85 | 2.27 | 1.83 | 1.77 | 1.88 | 1.46 | 1.29 |
| 1,2,3,4,7,8-HXCDF | 1.36 | 1.08 | 1.63 | 1.35 | 1.74 | ND | 1.24 | 1.22 | 1.67 | 1.2 | 2.45 | 1.57 | 1.42 | 1.76 | 1.27 | 1.36 |
| 1,2,3,6,7,8-HXCDF | 1.44 | 1.38 | 1.65 | ND | 1.69 | ND | 1.67 | 1.21 | 1.59 | 1.97 | 2.15 | ND | 1.24 | 1.73 | ND | 1.21 |
| 2,3,4,6,7,8-HXCDF | 1.32 | 1.43 | 1.2 | ND | 1.39 | ND | 1.88 | 1.04 | 1.31 | 1.21 | 1.83 | 0.98 | 1.2 | 1.22 | ND | 1.07 |
| 1,2,3,7,8,9-HXCDF | ND | ND | ND | ND | 3.4 | ND | ND | ND | ND | 5.15 | 4.12 | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 0.59 | 0.5 | 0.54 | 0.78 | 0.49 | 0.55 | 0.69 | 0.6 | 0.54 | 0.53 | 0.66 | 0.53 | 0.5 | 0.54 | 0.43 | 0.49 |
| 1,2,3,4,7,8,9-HPCDF | ND | ND | ND | ND | 1.63 | ND | ND | 3.31 | 1.7 | ND | 2.95 | ND | ND | ND | ND | ND |
| OCDF | 0.27 | 0.26 | 0.3 | ND | 0.26 | ND | ND | 0.24 | 0.24 | 0.24 | 0.43 | 0.34 | 0.25 | 0.29 | ND | ND |
| Dioxin TEQ (ND=0) | 1.66 | 1.1 | 2.67 | 1.97 | 1.15 | 1.09 | 1.52 | 1.52 | 1.58 | 135 | 1.48 | 0.87 | 1.59 | 0.75 | 0.72 | 1.06 |
| Dioxin TEQ (ND=1/2DL) | 0.78 | 1.25 | 3:10 | 3.03 | 1.19 | 2.35 | 3:11 | 2:55 | 1.87 | 2414 | 1.6 | 1.84 | 2.14 | 1.62 | 1.45 | 1.85 |

TABLE 9-33 Nereis virens (SAND WORM): UPTAKE RATIOS FOR DIOXIN AND FURAN CONGENERS

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



TABLE 9-34A Macoma nasuta (BLUNT-NOSE CLAM): MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED

TO INSIDE SITE 104^(a)

| | | | | Brewerton | C&D | C&D | | | | | | | | | | |
|------------------------|----------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------------|--------|------------|------------|---------------|
| | Inside | Outside | Ocean | Eastern | Approach | Approach | | Craighill | Craighlll | Craighill | Craighill | | Swan | Tolchester | Tolchester | Tolchester |
| Analyte (ng/kg) | Site 104 | Site 104 | Reference | Extension | (Surficlai) | (Cores) | Craighill | Angle East | Angle West | Entrance | Upper Range | Cutoff Angle | Point | North | South | Straightening |
| 2,3,7,8-TCDD | ND | ND | 0.0956 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,7,8-PECDD | ND | ND | 0.157B | 0.0521 | ND | ND | ND | 0.0633 | 0.0756 | ND | ND | ND | ND | ND | ND | 0.0562 |
| 1,2,3,4,7,8-HXCDD | ND | ND | 0.0903 | 0.11 | ND | ND | ND | 0.105 | 0.117 | 0.0768 | ND | ND | ND | ND | ND | ND |
| 1,2,3,6,7,8-HXCDD | 0.226 | 0.234 | 0.185 | 0.258 | ND | 0.192 | 0.208 | 0.272 | 0.297 | 0.202 | 0.269 | 0.186 | 0.26 | 0.237 | 0.202 | 0.227 |
| 1,2,3,7,8,9-HXCDD | 0.112 | 0.231 | 0.142 | 0.141 | ND | 0.0851 | 0.108 | 0.191 | 0.171 | 0.0941 | 0.12 | ND | 0.137 | 0.0902 | 0.101 | 0.0773 |
| 1,2,3,4,6,7,8-HPCDD | 2.6 | 2.44 | 1.77 | 2.33 | 2.16 | 2.36 | 1.86 | 2.66 | 2.2 | 1.66 | 3.4 | 2.13 | 2.45 | 2.94 | 2.02 | 1.75 |
| OCDD | 22.8 | 21.9 | 4.58 | 21.4 | 21.9 | 22.4 | 16.5 | 20 | 18 | 13.2 | 21.2 | 19.6 | 21.1 | 18.3 | 14.8 | 19.4 |
| 2,3,7,8-TCDF | ND | ND | 0.107 | ND | ND | NĐ | ND | ND | ND | ND | ND | ND | ND | NĎ | ND | ND |
| 1,2,3.7,8-PECDF | ND | ND | 0.126 | 0.0703 | ND | ND | ND | 0.0582 | 0.0607 | ND | ND | ND | ND | ND | 0.0627 | ND |
| 2,3,4,7,8-PECDF | 0.179 | ND | 0.104 | 0.0644 | ND | ND | ND | 0.065 | 0.0605 | ND | ND | 0.042 | ND | 0.0355 | 0.0481 | ND |
| 1,2,3,4,7,8-HXCDF | 0.136 | ND | 0.11 | 0.106 | ND | 0.0693 | 0.057 | 0.152 | 0.127 | 0.0663 | 0.0908 | 0.105 | 0.072 | ND | 0.0951 | 0.124 |
| 1,2,3,6,7,8-HXCDF | 0.108 | ND | 0.0933 | 0.0727 | ND | ND | ND | 0.0994 | 0.084 | ND | ND | ND | ND | ND | ND | ND |
| 2,3,4,6,7,8-HXCDF | 0.173 | ND | 0.0962 | 0.0838 | ND | ND | ND | 0.122 | 0.0783 | ND | ND | ND | ND | ND | ND | 0.0608 |
| 1,2,3,7,8,9-HXCDF | ND | ND | 0.0848 | ND | ND | ND | ND | ND | 0.0741 | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 0.675 | 0.254 | 0.238 | 0.336 | 0.308 | 0.281 | 0.221 | 0.542 | 0.335 | 0.211 | 0.397 | 0.277 | 0.283 | 0.298 | 0.277 | 0.237 |
| 1,2,3,4,7,8,9-HPCDF | ND | ND | 0.0651 | ND | ND | ND | ND | ND | 0.105 | ND | ND | ND | ND | ND | ND | ND |
| OCDF | 0.485 | ND | 0.239 | 0.34 | ND | 0.251 | 0.168 | 0.403 | 0.406 | 0.2 53 | 0.399 | 0.347 | 0.227 | 0.213 | 0.228 | 0.204 |
| DIOXINS TEQ (ND=0) | 0.114 | 0.0609 | 0.121 | 0.1 | 0.0383 | 0.0488 | 0.0219 | 0.0938 | 0.0908 | 0.0188 | 0.0786 | 0.0533 | 0.0485 | 0.0596 | 0.0461 | 0.0533 |
| DIOXINS TEQ (ND=1/2DL) | 0.362 | 0.254 | 0.125 | 0.239 | 0.395 | 0.197 | 0.155 | 0.242 | 0.201 | 0.15 | 0.266 | 0.208 | 0.197 | 0.218 | 0.2 | 0.208 |

B = detected in laboratory blank in at least one of the five test replicates

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for Individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and holded cells indicate sites where mean tissue residnes were statistically higher than inside Site 104 (p < 0.05*, p < 0.01**).

TABLE 9-34B Macoma nasuta (BLUNT-NOSE CLAM): MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED TOOUTSIDE SITE 104^(a)

| | | | | Brewerton | C&D | C&D | | | | | | | | | | _ |
|------------------------|--------------|-------------|-----------|-----------|-------------|----------|-----------|-----------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| | Ontside Site | Inside Site | Ocean | Eastern | Approach | Approach | | Craighiii Angie | Craighiii | Craighiii | Craighiii | Cutoff | | Toichester | Toichester | Toichester |
| Anaiyte (ng/kg) | i 04 | i 04 | Reference | Extension | (Surficiai) | (Cores) | Craighiii | East | Angie West | Entrance | Upper Range | Angie | Swan Point | North | South | Straightening |
| 2,3,7.8-TCDD | ND | ND | 0.0956 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2,3,7,8-PECDD | ND | ND | 0.157B | 0.0521 | ND | ND | ND | 0.0633 | 0.0756 | ND | ND | ND | ND | ND | ND | 0.0562 |
| 1,2,3,4,7,8-HXCDD | ND | ND | 0.0903 | 0.11 | ND | ND | ND | 0.105 | 0.117 | 0.0768 | ND | ND | ND | ND | ND | ND |
| i.2,3,6.7,8-HXCDD | 0.234 | 0.226 | 0.185 | 0.258 | ND | 0. i 92 | 0.208 | 0.272 | 0.297 | 0.202 | 0.269 | 0.186 | 0.26 | 0.237 | 0.202 | 0.227 |
| 1,2,3,7,8,9-HXCDD | 0.231 | 0.112 | 0.i42 | 0.14i | ND | 0.0851 | 0.i08 | 0.191 | 0.171 | 0.0941 | 0.12 | ND | 0.137 | 0.0902 | 0.10i | 0.0773 |
| 1,2,3.4,6,7,8-HPCDD | 2.44 | 2.6 | 1.77 | 2.33 | 2.16 | 2.36 | 1.86 | 2.66 | 2.2 | 1.66 | 3.4 | 2.13 | 2.45 | 2.94 | 2.02 | 1.75 |
| OCDD | 21.9 | 22.8 | 4.58 | 21.4 | 21.9 | 22.4 | 16.5 | 20 | 18 | i3.2 | 21.2 | i9.6 | 21.i | 18.3 | i4.8 | 19.4 |
| 2,3,7.8-TCDF | ND | ND | 0.i07 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3.7,8-PECDF | ND | ND | 0.126B | 0.0703 | ND | ND | ND | 0.0582 | 0.0607 | ND | ND | ND | ND | ND | 0.0627 | ND |
| 2,3,4.7,8-PECDF | ND | 0.179 | 0.104B | 0.0644 | ND | ND | ND | 0.065 | 0.0605 | ND | ND | 0.042 | ND | 0.0355 | 0.0481 | ND |
| 1,2,3,4,7,8-HXCDF | ND | 0.136 | 0.ilB | 0.106 | ND | 0.0693 | 0.057 | 0.152B | 0.i27B | 0.0663 | 0.0908 | 0.105 | 0.072 | ND | 0.095 i | 0.12413 |
| 1,2.3.6,7,8-HXCDF | ND | 0.108 | 0.0933B | 0.0727 | ND | ND | ND | 0.0994 | 0.084 | ND | ND | ND | ND | ND | ND | ND |
| 2.3.4.6.7.8-HXCDF | ND | 0.173 | 0.0962B | 0.0838 | ND | ND | ND | 0.122 | 0.0783 | ND | ND | ND | ND | ND | ND | 0.0608 |
| 1,2,3,7,8,9-HXCDF | ND | ND | 0.0848 | ND | ND | ND | ND | ND | 0.074 i | ND | ND | ND | ND | ND | ND | ND |
| 1.2.3.4.6.7.8-11PCDF | 0.254 | 0.675 | 0.238 | 0.336 | 0.308 | 0.281 | 0.22 i | 0.542B | 0.335 | 0.211 | 0.397 | 0.277 | 0.283 | 0.298 | 0.277 | 0.237 |
| i,2,3,4,7,8,9-HPCDF | ND | ND | 0.0651 | ND | ND | ND | ND | ND | 0.105 | ND | ND | ND | ND | ND | ND | ND |
| OCDF | ND | 0.485 | 0.239B | 0.34 | ND | 0.251 | 0.168 | 0.403 | 0.406 | 0.253 | 0.399 | 0.347 | 0.227 | 0.213 | 0.228 | 0.204 |
| DIOXINS TEQ (ND=0) | 0.0609 | 0.114 | 0.121 | 0.1 | 0.0383 | 0.0488 | 0.0219 | 0.0938 | 0.0908 | 0.0188 | 0.0786 | 0.0533 | 0.0485 | 0.0596 | 0.0461 | 0.0533 |
| DIOXINS TEQ (ND=1/2DL) | 0.254 | 0.362 | 0.i25 | 0.239 | 0.395* | 0.197 | 0.155 | 0.242 | 0.201 | 0.15 | 0.266 | 0.208 | 0.197 | 0.218 | 0.2 | 0.208 |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

.

(a) Statistical comparisons not conducted for Individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and boided celis indicate sites where mean tissue residues were statistically higher than Outside Site i 04 (p < 0.05*, p < 0.0i**).



TABLE 9-34C Macoma nasuta (BLUNT-NOSE CLAM): MEAN DIOXIN AND FURAN CONCENTRATIONS (NG/KG) COMPARED TO

OCEAN REFERENCE^(a)

| | | Outelde Site | Include Site | Brewerton | C&D | C&D | 1.5.1.3 | Craighill Angle | Crolebill | Crolabill | Crolobill | Cutoff | | Talabastas | Talahastar | Tolebastan |
|------------------------|-----------|--------------|--------------|-----------|-------------|---------|-----------|-----------------|------------|-----------|-------------|--------|------------|------------|------------|---------------|
| Analyte (ng/kg) | Reference | 104 | 104 | Extension | (Surficial) | (Cores) | Craighill | East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Stralghtening |
| 2.3.7.8-TCDD | 0.0956 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2.3.7.8-PECDD | 0.157 | ND | ND | 0.0521 | ND | ND | ND | 0.0633 | 0.0756 | ND | ND | ND | ND | ND | ND | 0.0562 |
| 1.2.3.4.7.8-HXCDD | 0.0903 | ND | ND | 0.11 | ND | ND | ND | 0.105 | 0.117 | 0.0768 | ND, | ND | ND | ND | ND | ND |
| 1.2.3.6.7.8-IIXCDD | 0.185 | 0.234 | 0.226 | 0.258 | ND | 0.192 | 0.208 | 0.272 | 0.297 | 0.202 | 0.269 | 0.186 | 0.26 | 0.237 | 0.202 | 0.227 |
| 1.2,3,7.8.9-HXCDD | 0.142 | 0.231 | 0.112 | 0.141 | ND | 0.0851 | 0.108 | 0.191 | 0.171 | 0.0941 | 0.12 | ND | 0.137 | 0.0902 | 0.101 | 0.0773 |
| 1.2,3.4.6.7.8-HPCDD | 1.77 | 2.44 | 2.6 | 2.33 | 2.16 | 2.36 | 1.86 | 2.66 | 2.2B | 1.66 | 3.4 | 2.13 | 2.45 | 2.94 | 2.02 | 1.75 |
| OCDD | 4.58 | 21.9 | 22.8 | 21.4 | 21.9 | 22.4 | 16.5B | 20B | 18B | 13.2B | 21.2 | 19.6 | 21.1 | 18.3 | 14.8 | 19.4 |
| 2.3.7.8-TCDF | 0.107 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1.2,3.7,8-PECDF | 0.126 | ND | ND | 0.0703 | ND | ND | ND | 0.0582 | 0.0607 | ND | ND | ND | ND | ND | 0.0627 | ND |
| 2.3.4.7.8-PECDF | 0.104 | ND | 0.179 | 0.0644 | ND | ND | ND | 0.065 | 0.0605 | ND | ND | 0.042 | ND | 0.0355 | 0.0481 | ND |
| 1.2,3,4.7.8-HXCDF | 0.11 | ND | 0.136 | 0.106 | ND | 0.0693 | 0.057 | 0.152 | 0.127 | 0.0663 | 0.0908 | 0.105 | 0.072 | ND | 0.0951 | 0.124 |
| 1.2.3.6.7.8-HXCDF | 0.0933 | ND | 0.108 | 0.0727 | ND | ND | ND | 0.0994 | 0.084 | ND | ND | ND | ND | ND | ND | ND |
| 2,3,4.6.7,8-HXCDF | 0.0962 | ND | 0.173 | 0.0838 | ND | ND | ND | 0.122 | 0.0783 | ND | ND | ND | ND | ND | ND | 0.0608 |
| 1.2.3.7.8,9-HXCDF | 0.0848 | ND | ND | ND | ND | ND | ND | ND | 0.0741 | ND | ND | ND | ND | ND | ND | ND |
| 1.2,3,4,6,7,8-HPCDF | 0.238 | 0.254 | 0.675 | 0.336 | 0.308 | 0.281 | 0.221 | 0.542B | 0.335 | 0.211 | 0.397 | 0.277* | 0.283 | 0.298 | 0.277 | 0.237 |
| 1,2,3,4.7,8,9-HPCDF | 0.0651 | ND | ND | ND | ND | ND | ND | ND | 0.105 | ND | ND | ND | ND | ND | ND | ND |
| OCDF | 0.239 | ND | 0.485 | 0.34 | ND | 0.251 | 0.168 | 0.403 | 0.406 | 0.253 | 0.399 | 0.347 | 0.227 | 0.213 | 0.228 | 0.204 |
| DIOXINS TEQ (ND=0) | 0.121 | 0.0609 | 0.114 | 01 | 0.0383 | 0.0488 | 0.0219 | 0.0938 | 0.0908 | 0.0188 | 0.0786 | 0.0533 | 0.0485 | 0.0596 | 0.0461 | 0.0533 |
| DIOXINS TEQ (ND=1/2DL) | 0.125 | 0.254* | 0.362** | 0.239* | 0.395** | 0.197 | 0.155 | 0.242* | 0.201 | 0.15 | 0.266** | 0.208* | 0.197 | 0.218* | 0.2 | 0.208* |

B = detected in laboratory blank in at least one of the five test replicates.

ND = not detected in any of the five tested replicates.

(a) Statistical comparisons not conducted for individual congeners; statistical comparisons conducted for TEQs only.

Asterisks, shaded and bolded cells indicate sites where mean tissue residues were statistically higher than the Ocean Reference site (p < 0.05*, p < 0.01**).

| Analyte | Ocean Reference | Outside Site 104 | Inside Site 104 | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Cores) | Craighill | Craighill Angle East | Craighill Angle West | CraighIII Entrance | Craighlll Upper Range | Cutoff Angle | Swan Point | Tolchester North | Tolchester South | Tolchester Straightening |
|-----------------------|--------------------|---------------------|--------------------|-----------------------------------|--------------------------------|----------------------------|-----------|-------------------------|-------------------------|-----------------------|--------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| 2,3,7,8-TCDD | 1.06 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,7,8-PECDD | 1.96 | ND | ND | 0.64 | ND | ND | ND | 0.78 | 0.94 | ND | ND | ND | ND | ND | ND | 0.7 |
| 1,2,3,4,7,8-HXCDD | 0.9 | ND | ND | 1.78 | ND | ND | ND | 1.7 | 1.88 | 1.24 | ND | ND | ND | ND | ND | ND |
| 1,2,3,6,7,8-HXCDD | 1.03 | 1.84 | 1.78 | 2.03 | ND | 1.51 | 1.63 | 2.14 | 2.34 | 1.59 | 2.12 | 1.46 | 2.04 | 1.86 | 1.59 | 1.78 |
| 1,2,3,7,8,9-HXCDD | 1.68 | 3.53 | 1.71 | 2.16 | ND | 1.3 | 1.66 | 2.92 | 2.61 | 1.44 | 1.84 | ND | 2.09 | 1.38 | 1.54 | 1.18 |
| 1,2,3,4,6,7,8-HPCDD | 1.25 | 2.61 | 2.78 | 2.49 | 2.32 | 2.52 | 1.99 | 2.85 | 2.36 | 1.78 | 3.63 | 2.28 | 2.62 | 3.14 | 2.16 | 1.87 |
| OCDD | 0.75 | 7.35 | 7.64 | 7.18 | 7.34 | 7.51 | 5.55 | 6.72 | 6.04 | 4.42 | 7.12 | 6.56 | 7.07 | 6.15 | 4.97 | 6.51 |
| 2,3,7,8-TCDF | 1.43 | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,7,8-PECDF | 2.52 | ND | ND | 1.45 | ND | ND | ND | 1.2 | 1.25 | ND | ND | ND | ND | ND | 1.29 | ND |
| 2,3,4,7,8-PECDF | 2.98 | ND | 3.78 | 1.36 | ND | ND | ND | 1.37 | 1.27 | ND | ND | 0.88 | ND | 0.75 | 1.01 | ND |
| 1,2,3,4,7,8-HXCDF | 1.99 | ND | 2.56 | 2 | ND | 1.31 | 1.07 | 2.87 | 2.39 | 1.25 | 1.71 | 1.99 | 1.36 | ND | 1.79 | 2.34 |
| 1,2,3,6,7,8-HXCDF | 1.70 | ND | 2.21 | 1.5 | ND | ND | ND | 2.04 | 1.73 | ND | ND | ND | ND | ND | ND | ND |
| 2,3,4,6,7,8-HXCDF | 1.37 | ND | 2.98 | 1.44 | ND | ND | ND | 2.1 | 1.35 | ND | ND | ND | ND | ND | ND | 1.05 |
| 1,2,3,7,8,9-HXCDF | 1 | ND | ND | ND | ND | ND | ND | ND | 1.21 | ND | ND | ND | ND | ND | ND | ND |
| 1,2,3,4,6,7,8-HPCDF | 1.4 | 1.5 | 3.97 | 1.98 | 1.81 | 1.66 | 1.3 | 3.19 | 1.97 | 1.24 | 2.33 | 1.63 | 1.67 | 1.75 | 1.63 | 1.39 |
| 1,2,3,4,7,8,9-HPCDF | 0.77 | ND | ND | ND | ND | ND | ND | ND | 1.63 | ND | ND | ND | ND | ND | ND | ND |
| OCDF | 1.33 | ND | 2.28 | 1.60 | ND | 1.18 | 0.79 | 1.90 | 1.91 | 1.19 | 1.88 | 1.63 | 1.07 | 1 | 1.07 | 0.96 |
| DIOXIN TEQ (ND=0) | 5.47 | 1.18 | 2.21 | 1.94 | 0.74 | 0.95 | 0.43 | 1.82 | 1.76 | 0.36 | 1.52 | 1.03 | 0.94 | 1.16 | 0.89 | 1.03 |
| DIOXIN TEQ (ND=1/2DL) | 0.53 | 1.45 | 2.06 | 1.36 | 2.25 | 1.12 | 0.88 | 138 | 1.15 | 0.85 | 1.52 | 1.19 | 1.12 | 1.24 | 1.14 | 1.19 |

TABLE 9-35 Macoma nasuta (BLUNT-NOSE CLAM): UPTAKE RATIOS FOR DIOXIN AND FURAN CONGENERS

ND = not detected

Shaded and bolded values = uptake ratios for tissue residues that statistically exceeded atleast one of the placement site/ reference tissue-residues.



TABLE 9-36 Nereis virens (SAND WORM): MEAN LIPID (%) AND MOISTURE CONCENTRATIONS

| | | | | Brewerton | C&D | C&D | | | | | Cralghlii | | | | | |
|----------------------|-------------|----------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------------|-------|------------|------------|---------------|
| | Inside Site | Outside | Осеап | Eastern | Approach | Approach | | Craighili | Craighiil | Craighiii | Upper | | Swan | Toichester | Toichester | Toichester |
| Analyte | 104 | Site 104 | Reference | Extension | (Surficiai) | (Cores) | Craighiil | Angle East | Angle West | Entrance | Range | Cutoff Angie | Point | North | South | Straightening |
| LIPIDS | 0.27 | 0.28 | 0.25 | 0.35 | 0.33 | 0.35 | 0.55 | 0.40 | 0.63 | 0.58 | 0.60 | 0.67 | 0.28 | 0.54 | 1.39 | 0.34 |
| MOISTURE CONTENT (%) | 84.5 | 85.0 | NT | 85.0 | 85.4 | 85.8 | 84.8 | 85.4 | 85.6 | 84.5 | 85.2 | 85.5 | 87.1 | 86.2 | 84.9 | 85.3 |

NT = not tested (not measured)

TABLE 9-37 Macoma nasuta (BLUNT-NOSE CLAM): MEAN LIPID (%) AND MOISTURE CONCENTRATIONS

.

| | | | | Brewerton | C&D | C&D | | | | | | | | · · · · · · · · · · · · · · · · · · · | | |
|----------------------|-------------|--------------|-----------|-----------|-------------|----------|-----------|------------|------------|-----------|-------------|--------|------------|---------------------------------------|------------|---------------|
| | Inside Site | Outside Site | Ocean | Eastern | Approach | Approach | | Craighiil | Craighlll | Craighill | Craighill | Cutoff | | Tolchester | Tolchester | Tolchester |
| Analyte | 104 | 104 | Reference | Extension | (Surficlai) | (Cores) | Craighlll | Angle East | Angle West | Entrance | Upper Range | Angle | Swan Point | North | South | Straightening |
| LIPIDS | ND | ND | 0.24 | 0.08 | ND | 0.07 | 0.06 | 0.07 | 0.17 | 0.10 | 0.07 | 0.10 | 0.05 | ND | 0.05 | ND |
| MOISTURE CONTENT (%) | 86.5 | 86.2 | NT | 86.3 | 86.7 | 87.0 | 86.0 | 86.6 | 86.5 | 85.8 | 85.7 | 86.6 | 87.0 | 86.6 | 87.5 | 86.7 |

ND = not detected in any of the five test replicates.

NT = not tested

TABLE 9-38A:Nereis virens (SAND WORM): NUMBER OF CHEMICAL ANALYTES IN
CHANNEL TISSUES THAT STATISTICALLY EXCEED THE PLACEMENT SITE OR
REFERENCE TISSUE RESIDUES

| | Inside Site | 104 | Outside Site | 104 | Ocean Refer | ence |
|---------------------|-----------------|---------|-----------------|---------|-----------------|---------|
| ANALYTE | # Exceed/Total* | Percent | # Exceed/Total* | Percent | # Exceed/Total* | Percent |
| METALS | 43/208 | 21 | 67/208 | 32 | 14/208 | 7 |
| PESTICIDES | 34/286 | 12 | 43/286 | 15 | 21/286 | 7 |
| PAHs | 0/208 | 0 | 0/208 | 0 | 2/208 | <1 |
| PCB AROCLORS | 0/91 | 0 | 0/91 | 0 | 0/91 | 0 |
| TOTAL PCB CONGENERS | 0/13 | 0 | 0/13 | 0 | 3/13 | 23 |
| SVOCs | 8/611 | 1 | 7/611 | 1 | 8/611 | 1 |
| DIOXIN/FURAN TEQ | 0/13 | 0 | 9/13 | 69 | 3/13 | 23 |
| TOTAL | 85/1430 | 6 | 126/1430 | 9 | 51/1430 | 4 |

*Total number of analytes tested times 13 channel reaches

TABLE 9-38B:Macoma nasuta (BLUNT NOSE CLAM): NUMBER OF CHEMICAL ANALYTES
IN CHANNEL TISSUES THAT STATISTICALLY EXCEED THE PLACEMENT SITE OR
REFERENCE TISSUE RESIDUES

| | Inside Site 1 | 04 | Outside Site | 104 | Ocean Refer | ence |
|--------------------|-----------------|---------|-----------------|---------|-----------------|---------|
| ANALYTE | # Exceed/Total* | Percent | # Exceed/Total* | Percent | # Exceed/Total* | Percent |
| METALS | 61/208 | 29 | 104/208 | 50 | 49/208 | 24 |
| PESTICIDES | 19/286 | 7 | 21/286 | 7 | 22/286 | 8 |
| PAHs | 14/208 | 7 | 21/208 | 10 | 87/208 | 42 |
| PCB AROCLORS | 0/91 | 0 | 0/91 | 0 | 0/91 | 0 |
| TOTAL PCB CONGENER | 0/13 | 0 | 4/13 | 31 | 7/13 | 54 |
| SVOCs | 11/611 | 2 | 15/611 | 2 | 23/611 | 4 |
| DIOXIN/FURAN TEQ | 0/13 | 0 | 1/13 | 8 | 7/13 | 54 |
| TOTAL | 105/1430 | 7 | 166/1430 | 12 | 195/1430 | 14 |

*Total number of analytes tested times 13 channel reaches



TABLE 9-39A Nereis virens (SAND WORM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTEFRACTIONS VS. INSIDE SITE 104

| | # Tested | Brewerton Eastern | C&D Approach | C&D Approach | Grelshill | Craighll Angle | Cralghill Angle | Craighill | Cralghill Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
|-------------------------------|----------|----------------------|-----------------|-----------------|-----------|-------------------|--------------------|-----------|--------------------|--------|-------|------------|------------|---------------|
| ANALTIE | Analytes | Extension | (Surnelal) | (Core) | Craigain | Lasi | west | Entrance | Range | Angle | Foint | North | South | Straigntening |
| METALS | 16 | 4 | 3 | 5 | 1 | 0 | 4 | 0 | 4 | 5 | 3 | 5 | 5 | 4 |
| PAHs | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| PESTICIDES | 22 | 4 | 6 | 5 | 0 | 0 | I | i | 2 | 2 | 4 | 5 | 2 | 2 |
| TOTAL PCBs (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | Ō | 0 | 0 | 0 | 0 |
| PCB AROCLORS | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SVOCs | 47 | 1 | 0 | 1 | 0 | 0 | 2 | 1 | 0 | 0 | I | 0 | I | l |
| DIOXIN/FURAN CONGENERS (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 110 | 9 | 9 | 11 | 1 | 0 | 7 | 2 | 6 | 7 | 8 | 10 | 8 | 7 |

TABLE 9-39B Macoma nasuta (BLUNT NOSE CLAM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTEFRACTIONS VS. INSIDE SITE 104

| | # Tested | Brewerton Eastern | C&D Approach | C&D Approach | | Craighlll Angle | Cralghlll Angle | Cralghill | Cralghlll Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
|-------------------------------|----------|----------------------|-----------------|-----------------|-----------|--------------------|--------------------|-----------|--------------------|--------|-------|------------|------------|---------------|
| ANALYTE | Analytes | Extension | (Surficial) | (Core) | Cralghill | East | West | Entrance | Range | Angle | Point | North | South | Stralghtening |
| METALS | 16 | 8 | 4 | 3 | 9 | 7 | 5 | 8 | 5 | 2 | 3 | 3 | 2 | 2 |
| PAHs | 16 | 0 | 2 | 0 | 0 | 0 | 1 | ō | 1 | 1 | 0 | 2 | 1 | 6 |
| PESTICIDES | 22 | 2 | 0 | 4 | 0 | 0 | 2 | 0 | 3 | 4 | 2 | 0 | l | 1 |
| TOTAL PCBs (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| PCB AROCLORS | 7 | 0 | Ö | 0 | Ö | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SVOCs | 47 | 0 | 1 | 1 | 0 | 0 | 0 | 3 | 0 | 0 | 1 | 2 | 2 | 1 |
| DIOXIN/FURAN CONGENERS (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 110 | 10 | 7 | 8 | 9 | 7 | 8 | 11 | 9 | 7 | 6 | 7 | 6 | 10 |

TABLE 9-40A Nereis virens (SAND WORM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTEFRACTIONS VS. OUTSIDE SITE 104

| | # Tested | Brewerton Eastern | C&D Approach | C&D Approach | | Craighill | Craighill Angle | CraighIII | Cralghill Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
|-------------------------------|----------|----------------------|-----------------|-----------------|-----------|------------|--------------------|-----------|--------------------|--------|-------|------------|------------|---------------|
| ANALYTE | Analytes | Extension | (Surficial) | (Core) | Craighill | Angle East | West | Entrance | Range | Angle | Point | North | South | Straightening |
| METALS | 16 | 4 | 5 | 6 | 4 | 4 | 5 | 4 | 5 | 6 | 6 | 6 | 7 | 5 |
| PAHs | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| PESTICIDES | 22 | 3 | 6 | 3 | 1 | 0 | 4 | 5 | 5 | 4 | 4 | 3 | 2 | 1 |
| TOTAL PCBs (ND=0) | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| PCB AROCLORS | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SVOCs | 47 | 1 | 0 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | | <u> </u> | | |
| DIOXIN/FURAN CONGENERS (ND=0) | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | i i | 0 |
| TOTAL | 110 | 9 | 11 | 10 | 6 | 4 | 10 | 12 | 10 | 11 | 11 | 9 | 10 | 9 |

TABLE 9-40B Macoma nasuta (BLUNT NOSE CLAM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTE FRACTIONS VS. OUTSIDE SITE 104

| | # Tested | Brewerton Eastern | C&D Approach | C&D Approach | | Craighill | Craighill Angle | Craighill | Craighill | Cutoff | Swan | Tolchester | Talabestar | Tolchestor |
|-------------------------------|----------|----------------------|-----------------|-----------------|-----------|------------|--------------------|-----------|-----------|--------|-------|------------|------------|---------------|
| ANALYTE | Analytes | Extension | (Surficial) | (Core) | Craighill | Angle East | West | Entrance | Range | Angle | Point | North | South | Straightening |
| METALS | 16 | 10 | 9 | 8 | 9 | 9 | 8 | 7 | 8 | 8 | 8 | 7 | 5 | 8 |
| PAHs | 16 | 1 | 3 | I | 0 | Ó | 2 | 0 | 2 | 4 | 0 | 2 | I | 5 |
| PESTICIDES | 22 | 2 | 1 | 4 | 0 | 0 | i | 0 | 4 | 4 | 3 | 0 | 1 | 1 |
| TOTAL PCBs (ND=0) | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| PCB AROCLORS | - 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | Ö | 0 | 0 |
| SVOCs | 47 | 0 | 2 | 1 | 0 | 0 | 0 | 3 | 0 | 0 | I | 2 | 3 | 3 |
| DIOXIN/FURAN CONGENERS (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 110 | 14 | 15 | 15 | 9 | 9 | 11 | 11 | 14 | 16 | 13 | 11 | 10 | 17 |



TABLE 9-41A Nereis virens (SAND WORM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTEFRACTIONS VS. OCEAN REFERENCE

| | # Tested | Brewerton Eastern | C&D Approach | C&D Approach | | Craighili | Cralghill Angie | Craighill | Craighill Upper | Cutoff | Swan | Tolchester | Toichester | Tolchester |
|-------------------------------|----------|----------------------|-----------------|-----------------|-----------|------------|--------------------|-----------|--------------------|--------|-------|------------|------------|---------------|
| ANALYTE | Analytes | Extension | (Surficial) | (Core) | Craighiii | Angle East | West | Entrance | Kange | Angie | Point | North | South | Straightening |
| METALS | 16 | 2 | | II | | 0 | 1 | 0 | 1 | | 2 | 2 | | <u> </u> |
| PAHs | 16 | 0 | i | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | | 0 | 0 |
| PESTICIDES | 22 | | 5 | 2 | 0 | 0 | 1 | | 0 | 2 | 2 | 2 | 2 | 3 |
| TOTAL PCBs (ND=0) | | 0 | 0 | 0 | | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| PCB AROCLORS | 7 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SVOCs | 47 | | 0 | | 0 | 0 | 2 | 1 | 0 | 0 | | 0 | <u> </u> | <u> </u> |
| DIOXIN/FURAN CONGENERS (ND=0) | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 110 | 4 | 7 | 4 | 2 | 0 | 4 | 3 | I | 4 | 5 | 5 | 4 | 5 |

TABLE 9-41B Macoma nasuta (BLUNT NOSE CLAM) NUMBER OF STATISTICAL EXCEEDANCES IN TARGET ANALYTE FRACTIONS VS. OCEAN REFERENCE

| ANALYTE | # Tested Analytes | Brewerton Eastern Extension | C&D Approach (Surficial) | C&D Approach (Core) | Craighiii | Craighiil Angie East | Craighill Angle West | Craighill Entrance | Craighiil Upper Range | Cutoff Angle | Swan Point | Toichester North | Tolchester South | Tolchester Straightening |
|-------------------------------|----------------------|-----------------------------------|--------------------------------|---------------------------|-----------|-------------------------|----------------------------|-----------------------|-----------------------------|-----------------|---------------|---------------------|---------------------|-----------------------------|
| METALS | 16 | 5 | 5 | 5 | 4 | 5 | 3 | 2 | 3 | 3 | 4 | 3 | 3 | 4 |
| РАНе | 16 | 5 | 7 | 5 | 5 | 4 | 9 | 4 | 9 | 10 | 5 | 6 | 7 | 11 |
| PESTICIDES | 22 | 2 | 1 | 5 | 2 | 2 | 2 | i | 2 | 3 | 2 | 0 | 0 | 0 |
| TOTAL PCBs (ND=0) | 1 | 1 | 1 | i | I | I | 0 | I | 0 | 0 | 1 | 0 | 0 | 0 |
| PCB AROCLORS | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| SVOCs | 47 | 2 | 1 | 1 | I | 1 | I | 3 | | 2 | <u> </u> | 3 | 3 | 3 |
| DIOXIN/FURAN CONGENERS (ND=0) | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL | 110 | 15 | 15 | 17 | 13 | 13 | 15 | 11 | 15 | 18 | 13 | 12 | 13 | 18 |

TABLE 9-42 COPCs* IDENTIFIED IN TISSUES OF WORMS AND CLAMS EXPOSED TO CHANNEL SEDIMENTS PROPOSED FOR OPEN-WATER AND **OCEAN PLACEMENT**

| ······ | Inside Site | Outside Site | Ocean |
|-----------------------|-------------|----------------|-----------|
| | 104 | 104 | Reference |
| COPCs (in tissue) | 104 | c.w | c,w |
| DIOXIN TEQ | | C W | c |
| ALUMINUM | <u>c,w</u> | C,W | |
| ANTIMONY | <u> </u> | <u> </u> | |
| ARSENIC | C | | C W |
| BERYLLIUM | c,w | <u> </u> | |
| CADMIUM | | <u> </u> | |
| CHROMIUM | c,w | c,w | |
| COPPER | w | | |
| IRON | C | <u> </u> | C |
| MANGANESE | c,w | c,w | <u> </u> |
| MERCURY | c | c | C |
| NICKEL | c,w | c,w | c,w |
| SELENIUM | с | c | c |
| SILVER | | L | c,w |
| ZINC | c | c,w | w |
| ACENAPHTHYLENE | c | c | с |
| ANTHRACENE | | c | |
| BENZ[A]ANTHRACENE | c | | c |
| BENZO[A]PYRENE | | | |
| BENZO[B]FLUORANTHENE | с | с | с |
| CHRYSENE | | c,w | с |
| FLUORANTHENE | | с | c,w |
| FLUORENE | с | | с |
| NAPHTHALENE | с | с | с |
| PHENANTHRENE | с | | c |
| PYRENE | | | с |
| TOTAL PAHs | c,w | с | с |
| TOTAL PCBs | | с | c,w |
| 4 4'-DDD | c,w | c,w | c,w |
| 4 4'-DDT | c,w | С | с |
| | w | w | c,w |
| | c,w | c,w | c,w |
| BETA-BHC | w | w | c,w |
| CHLORBENSIDE | w | w | w |
| DACTHAL | w | w | w |
| | w | w | |
| | c | с | |
| | | | С |
| ENDOSUL FAN II | c,w | с | с |
| ENDRIN | w | w | |
| GAMMA-BHC | c.w | c,w | с |
| HEPTACHLOR | c.w | c,w | w |
| HERTACHLOR FPOXIDE | w | w | |
| I-METHYI NAPHTHAI FNF | | c | c |
| 2 METHYL PHENOL | c.w | c.w | c,w |
| | | c | c |
| A NITROPHENOI | | - c | c |
| 4-NITKOPHENOL | | - <u> </u> | + |
| BENZUIC ACID | | - | c |
| BENZYL ALCOHOL | | | w |
| DI-N-BUTYL PHTHALATE | | | |
| DI-N-OCTYL PHTHALATE | | -+ | |
| PENTACHLOROPHENOL | | | |
| PHENOL | | W | ,w |

c = clain tissue; w = worm tissue * Any constituent that statistically exceeded a placement site/reference area tissue-residue was retained as a COPC.

| Analyte | Log ₁₀ K _{ow} | Reference |
|--|-----------------------------------|-----------|
| DIOXINS TEO (ND=0) | 6.64 | 1 |
| DIOXING TEQ ($ND=1/2$) | 6.64 | 1 |
| ACENAPHTHENE | 3.9 | 3 |
| | 4.1 | 3 |
| | 4.3 | 3 |
| BENZIAIANTHRACENE | 5.6 | 3 |
| BENZOLAIPVRENE | 6 | 3 |
| BENZOBIEI LIORANTHENE | 6.6 | 3 |
| BENZOIG H IPER YI ENE | 7 | 3 |
| BENZOR IT UORANTHENE | 6.8 | 3 |
| CHRVSENE | 5.6 | 3 |
| DIPENZIA HIANTHRACENE | 6.69 | 1 |
| FILIOPANTHENE | 5.5 | 3 |
| FLUORANTILINE | 4.4 | 3 |
| NIDENOU 2 3-CDIPYRENE | 7.7 | 3 |
| | 3.6 | 3 |
| DUENANTHRENE | 4.5 | 3 |
| PVRENE | 4.9 | 3 |
| 2 2' 3 3' 4 4' 5 5' 6 6'-DECACHLOROBIPHENYL (BZ#209) | 8.18 | 4 |
| 2 2' 3 3' 4 4' 5 5' 6-NONACHLOROBIPHENYL (BZ# 206) | 8.09 | 4 |
| 2 2' 3 3' 4 4' 5 6-OCTACHLOROBIPHENYL (BZ# 195) | 7.56 | 4 |
| 2 2' 3 3' 4 4' 5-HEPTACHLOROBIPHENYL (BZ# 170) | 7.27 | 4 |
| 2 2' 3 3' 4 4'-HEXACHI OROBIPHENYL (BZ# 128) | 6.74 | 4 |
| 2 2' 3 4' 5 5' 6-HEPTACHLOROBIPHENYL (BZ# 187) | 7.17 | 4 |
| 2 2' 3 4 4' 5' 6-HEPTACHLOROBIPHENYL (BZ# 183) | 7.2 | 4 |
| 2 2' 3 4 4' 5'-HEXACHLOROBIPHENYL (BZ# 138) | 6.83 | 4 |
| 2 2' 3 4 4' 5 5'-HEPTACHLOROBIPHENYL (BZ# 180) | 7.36 | 4 |
| 2 2' 3 4 4' 6 6'-HEPTACHLOROBIPHENYL (BZ# 184) | 6.85 | 4 |
| 2 2' 3 4 5'-PENTACHLOROBIPHENYL (BZ# 87) | 6.29 | 4 |
| 2 2' 3 5'-TETRACHLOROBIPHENYL (BZ# 44) | 5.75 | 4 |
| 2 2' 4 4' 5 5'-HEXACHLOROBIPHENYL (BZ# 153) | 6.92 | 4 |
| 2 2' 4 5'-TETRACHLOROBIPHENYL (BZ# 49) | 5.85 | 4 |
| 2 2' 4 5 5'-PENTACHLOROBIPHENYL (BZ# 101) | 6.38 | 8 4 |
| 2 2' 5 5'-TETRACHLOROBIPHENYL (BZ# 52) | 5.84 | 4 |
| 2 2' 5-TRICHLOROBIPHENYL (BZ# 18) | 5.24 | 4 |
| 2. 3' 4.4', 5-PENTACHLOROBIPHENYL (BZ# 118) | 6.74 | 4 |
| 2 3'44'-TETRACHLOROBIPHENYL (BZ# 66) | 6.2 | 2 4 |
| 2.3.3'4.4'.5HEXACHLOROBIPHENYL (BZ# 156) | 7.18 | 3 4 |
| 2.3.3' 4.4'-PENTACHLOROBIPHENYL (BZ# 105) | 6.6 | 5 4 |
| 2.4'-DICHLOROBIPHENYL (BZ# 8) | 5.0 | 7 4 |
| 2 4 4'-TRICHLOROBIPHENYL (BZ# 28) | 5.6 | 7 4 |
| 3.3'.4.4'.5.5'-HEXACHLOROBIPHENYL (BZ# 169) | 7.42 | 2 4 |
| 3,3',4,4',5-PENTACHLOROBIPHENYL (BZ# 126) | 6.8 | 9 4 |
| 3.3'.4.4'-TETRACHLOROBIPHENYL (BZ# 77) | 6.3 | 6 4 |
| AROCLOR 1016 | 5. | 9 3 |
| AROCLOR 1221 | | 4 3 |
| AROCLOR 1232 | 4. | 5 3 |
| AROCLOR 1242 | | 6 3 |
| AROCLOR 1248 | 6. | 1 3 |
| AROCLOR 1260 | 6. | 1 3 |

TABLE 9-43 OCTANOL-WATER PARTITION COEFFICIENTS (K_{ow}) FOR NEUTRAL ORGANIC COMPOUNDS

| Analyte | Log ₁₀ K _{ow} | Referen |
|--------------------------------------|-----------------------------------|-------------------------------|
| 4,4'-DDD | 6.1 | 3 |
| 4,4'-DDE | 6.76 | 3 |
| 4,4'-DDT | 6.83 | 3 |
| ALDRIN | 3 | 3 |
| ALPHA-BHC | 3.8 | 3 |
| BETA-BHC | 3.8 | 3 |
| CHLORDANE | 6.32 | 1 |
| DACTHAL | 4.4 | 1 |
| DELTA-BHC | 3.8 | 3 |
| DIFLORIN | 5.5 | 3 |
| ENDOSUL FAN 1 | 3.6 | 3 |
| ENDOSUI FAN II | 3.6 | 3 |
| ENDOSULTAN SULFATE | 3.6 | 3 |
| | 4.6 | 3 |
| | 5.6 | 3 |
| | 3.8 | 3 |
| GAMMA-BHC | 5.0 | 3 |
| | 5.4 | 3 |
| HEPTACHLOR EPOXIDE | | 3 |
| METHOXYCHLOR | 4.3 | 3 |
| MIREX | | 3 |
| TOXAPHENE | 3.3 | 3 |
| 1,2,4-TRICHLOROBENZENE | 4.2 | 3 |
| 1,2-DICHLOROBENZENE | 3.4 | 3 |
| 1,2-DIPHENYLHYDRAZINE | 2.9 | 3 |
| 1,4-DICHLOROBENZENE | 3.5 | 3 |
| 1-METHYLNAPHTHALENE | 3.87 | |
| 2,2'-OXYBIS(1-CHLOROPROPANE) | 2.6 | 3 |
| 2,4,6-TRICHLOROPHENOL | 3.7 | 3 |
| 2,4-DICHLOROPHENOL | 3.1 | 3 |
| 2,4-DIMETHYLPHENOL | 2.4 | 3 |
| 2,4-DINITROPHENOL | 1.5 | 3 |
| 2,4-DINITROTOLUENE | 2.1 | 3 |
| 2.6-DINITROTOLUENE | 2 | 3 |
| 2-CHLORONAPHTHALENE | 4.7 | 3 |
| 2-CHLOROPHENOL | 2.2 | 3 |
| 2-METHYL-4 6-DINITROPHENOL | 2.564 | 1 |
| 2-METHYLNAPHTHALENE | 3.36 | 1 |
| 2-METHYLPHENOL | 1.99 | 1 |
| 2-NTTROPHENOL | 1.8 | 3 |
| 3 3'-DICHLOROBENZIDINE | 3 | 3 |
| 3.4 METHVI PHENOI | 2.36 | 5 1 |
| 2.5.5 TRIMETHYL 2 CVCI OHEVENE-1-ONE | 2.22 | 1 |
| 4 DROMODUENVI DUENVI ETHER | 5.1 | 3 |
| 4-BROMOPHENTL PHENTL ETTER | 3 | 1 |
| 4-CHLORO-3-METHTLFHENOL | 40 | 3 |
| 4-CHLOROPHENYL PHENYL EINER | | |
| A-NITKOPHENUL | 1 94 | <u> </u> |
| BENZUIC ACID | 1.00 | 1 |
| BENZYL ALCOHOL | | |
| BIS(2-CHLOROETHOXY)METHANE | | |
| BIS(2-CHLOROETHYL) ETHER | 2.0 | $\frac{1}{3}$ |
| BIS(2-ETHYLHEXYL) PHTHALATE | 4.2 | $\frac{2}{1}$ - $\frac{3}{2}$ |
| DI-N-BUTYL PHTHALATE | 5. | |
| DI-N-OCTYL PHTHALATE | 9.2 | $\frac{2}{3}$ |
| DIETHYL PHTHALATE | 1.4 | 4 3 |
| DIMETHYL PHTHALATE | 1.0 | 5 3 |
| HEXACHI OBO-1 3-BUTADIENE | 4. | 3 3 |

TABLE 9-43 CONTINUED

TABLE 9-43 CONTINUED

| Analyte | Log ₁₀ K _{ow} | Reference |
|---------------------------------|-----------------------------------|-----------|
| HEXACHI OROBENZENE | 5.2 | 3 |
| HEXACHLOROCYCLOPENTADIENE | 5.5 | 3 |
| HEXACHI OROFTHANE | 3.8 | 3 |
| M-DICHLOROBENZENE | 3.42 | 1 |
| METHANAMINE, N-METHYL-N-NITROSO | -0.57 | 1 |
| N-NITROSODI-N-PROPYLAMINE | 1.3 | 3 |
| N-NITROSODIPHENYLAMINE | 1.3 | 3 |
| NTROBENZENE | 1.9 | 3 |
| PENTACHLOROPHENOL | 5 | 3 |
| PHENOL | 1.5 | 3 |

References:

- 1) Syracuse Research Corporation (SRC 2000)
- 2) Mabey et al. (1982)
- 3) USEPA/USACE (1998): ITM
- 4) Hawker, D.W. and D.W. Connell (1988)











| Chemical Constituent (s) ^(b) | Action/Tolerance / Guidance Level | UNITS | Fish Commodity | Reference |
|---|--------------------------------------|-------|--------------------|--------------------------------------|
| Aldrin/Dieldrin ^(c) | 300 | UG/KG | All Fish | Compliance Policy Guide sec. 575.100 |
| Chlordane | 300 | UG/KG | All Fish | Compliance Policy Guide sec. 575.100 |
| DDT,DDD, DDE ^(c) | 5000 | UG/KG | All Fish | Compliance Policy Guide sec. 575.100 |
| Arsenic | 86 | MG/KG | Molluscan bivalves | FDA Guidance Document |
| Cadmium | 4 | MG/KG | Molluscan bivalves | FDA Guidance Document |
| Chromium | 13 | MG/KG | Molluscan bivalves | FDA Guidance Document |
| Lead | 1.7 | MG/KG | Molluscan bivalves | FDA Guidance Document |
| Nickel | 80 | MG/KG | Molluscan bivalves | FDA Guidance Document |
| Methyl Mercury | 1 | MG/KG | Molluscan bivalves | Compliance Policy Guide sec. 540.600 |
| Heptachlor/Heptachlor Epoxic | 300 | UG/KG | All Fish | Compliance Policy Guide sec. 575.100 |
| Mirex | 100 | UG/KG | All Fish | Compliance Policy Guide sec. 575.100 |
| Total PCBs ^(d) | 2000 | UG/KG | All Fish | 21 CFR 109.30 |

TABLE 9-44 USFDA ACTION LEVELS AND USEPA TOLERANCE/GUIDANCE LEVELS^(a)

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(a) primary reference for all values: (USFDA 1998) Fish and Fishery Products Hazards and Control Guide. U.S. Food

and Drug Administration, Center for Food Safety and Applied Nutrition. January.

(b) Values provided only for chemcial constituents tested and relevant to this project.

(c) Action Level applies to residues for the individual pesticides or in combination.

(d) Tolerance value, rather than guidance level or action level.



TABLE 9-45A Nereis virens (SAND WORM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUERESIDUE TO THE USFDA ACTION LEVELS

| | FDA | | Brewerton | C&D | C&D | | | ······ | | Craighill | | | | | |
|----------------------|--------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | ACTION | | Eastern | Approach | Approach | | Craighill | Cralghill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | LEVEL | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| METHYL MERCURY | 1000 | UG/KG | 46 | 49 | 49 | 49 | 50 | 48 | 49 | 50 | 51 | 45 | 49 | 49 | 45 |
| TOTAL PCB (ND=0) | 2000 | UG/KG | 16 | 18 | 15 | 49 | 39 | 64 | 42 | 26 | 56 | 12 | 52 | 23 | 13 |
| TOTAL PCB (ND=1/2DL) | 2000 | UG/KG | 36 | 36 | 34 | 126 | 99 | 92 | 109 | 46 | 124 | 26 | 92 | 49 | 28 |
| TOTAL PCB (ND=DL) | 2000 | UG/KG | 41 | 42 | 39 | 130 | 105 | 98 | 113 | 50 | 130 | 32 | 102 | 53 | 33 |
| ALDRIN+DIELDRIN | 300 | UG/KG | 1 | 39 | 1 | 1 | 1 | 6 | 1 | 6 | 1 | 1 | 1 | 1 | 1 |
| CHLORDANE | 300 | UG/KG | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 |
| DDD+DDT+DDE | 5000 | UG/KG | 5 | 36 | 5 | 10 | 8 | 8 | 10 | 8 | 11 | 5 | 47 | 38 | 29 |
| MIREX | 100 | UG/KG | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| TOTAL HEPTACHLOR | 300 | UG/KG | 10 | 1 | 14 | 1 | 1 | | 1 | 1 | 1 | 11 | 5 | 3 | 3 |
TABLE 9-45B Macoma nasuta (BLUNT-NOSE CLAM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUERESIDUE TO THE USFDA ACTION LEVELS

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| | FDA | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|-----------------------|--------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | ACTION | | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | LEVEL | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| METHYL MERCURY | 1000 | UG/KG | 121 | 122 | 128 | 49 | 45 | 51 | 49 | 48 | 52 | 120 | 49 | 50 | 50 |
| TOTAL PCBs (ND=0) | 2000 | UG/KG | 17 | 16 | 19 | 14 | 13 | 8 | 15 | 9 | 6 | 20 | 7 | 6 | 7 |
| TOTAL PCBs (ND=1/2DL) | 2000 | UG/KG | 29 | 19 | 33 | 22 | 19 | 11 | 26 | 11 | 9 | 33 | 11 | 9 | 9 |
| TOTAL PCBs (ND=DL) | 2000 | UG/KG | 34 | 26 | 38 | 28 | 26 | 18 | 31 | 18 | 16 | 38 | 18 | 17 | 17 |
| ALDRIN+DIELDRIN | 300 | UG/KG | 1 | 5 | 3 | 4 | 7 | 4 | 5 | 5 | 3 | 2 | 3 | 3 | 3 |
| CHLORDANE | 300 | UG/KG | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 |
| DDD+DDT+DDE | 5000 | UG/KG | 18 | 7 | 13 | 6 | 13 | 22 | ii îi | 16 | 18 | 18 | 5 | 5 | 5 |
| MIREX | 100 | UG/KG | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| TOTAL HEPTACHLOR | 300 | UG/KG | 3 | 6 | 6 | 1 | 1 | 9 | 1 | 6 | 6 | 8 | 1 | 2 | 2 |



TABLE 9-46A Nereis virens (SAND WORM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUERESIDUE TO USEPA TOLERANCE VALUES

| | EPA | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|----------|-----------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | TOLERANCE | | Eastern | Approach | Approach | | Craighill | Cralghill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | VALUE | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| ARSENIC | 76000 | UG/KG | 1810 | 1790 | 2020 | 2430 | 1730 | 2310 | 1870 | 1950 | 1810 | 1790 | 1600 | 1940 | 1980 |
| CADMIUM | 3000 | UG/KG | 20 | 20 | 38.4 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 |
| CHROMIUM | 12000 | UG/KG | 630 | 188 | 543 | 283 | 238 | 405 | 149 | 636 | 662 | 531 | 549 | 1020 | 632 |
| LEAD | 2000 | UG/KG | 146 | 378 | 209 | 145 | 145 | 145 | 145 | 245 | 258 | 235 | 145 | 145 | 145 |
| NICKEL | 70000 | UG/KG | 1370 | 854 | 1480 | 1180 | 811 | 696 | 683 | 1200 | 1220 | 1000 | 1140 | 1070 | 1050 |

TABLE 9-46B Macoma nasuta (BLUNT-NOSE CLAM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATETISSUE RESIDUE TO USEPA TOLERANCE VALUES

| | EPA | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|-----------|-----------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | TOLERANCE | | Eastern | Approach | Approach | | Craighill | Cralghill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | VALUE | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| ARSENIC . | 76000 | UG/KG | 2720 | 2910 | 2390 | 2940 | 2680 | 2760 | 3040 | 2700 | 3280 | 2790 | 2690 | 2640 | 2600 |
| CADMIUM | 3000 | UG/KG | 20 | 50.7 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 | 20 |
| CHROMIUM | 12000 | UG/KG | 887 | 678 | 837 | 3630 | 2660 | 993 | 2350 | 962 | 622 | 728 | 543 | 426 | 469 |
| LEAD | 2000 | UG/KG | 480 | 513 | 211 | 145 | 220 | 365 | 234 | 388 | 245 | 144 | 146 | 145 | 328 |
| NICKEL | 70000 | UG/KG | 1170 | 1220 | 1100 | 3420 | 2720 | 1190 | 2220 | 1290 | 964 | 1030 | 924 | 789 | 995 |

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| | Screening Value | | |
|---|---------------------|-------|--|
| Chemical Constituent (s) ^(b) | (SV) ^(c) | UNITS | Notes |
| Arsenic (Inorganic) | 3 | MG/KG | non-carcinogen |
| Cadmium | 10 | MG/KG | non-carcinogen |
| Mercury | 0.6 | MG/KG | non-carcinogen; assumes methyl mercury |
| Selenium | 50 | MG/KG | non-carcinogen; selenious acid or selenium sulfate |
| Total DDT | 300 | UG/KG | carcinogen; DDD+DDE+DDT |
| Dieldrin | 7 | UG/KG | carcinogen |
| Endodulfan (I and II) | 60,000 | UG/KG | non-carcinogen |
| Endrin | 3,000 | UG/KG | non-carcinogen |
| Heptachlor Epoxide | 10 | UG/KG | carcinogen |
| Gamma-BHC (Lindane) | 80 | UG/KG | carcinogen |
| Dioxins/Furans | 0.7 | NG/KG | carcinogen |

TABLE 9-47 USEPA RECOMMENDED FISH TISSUE SCREENING VALUES^(a)

(a) primary reference for screening values: USEPA. 1995. Guidance for Assessing Chemical Contaminant Data for Use in Fish Advisories. Volume I: Fish Sampling and Analysis. Second Edition. EPA 823-R-95-007. September
(b) Values provided only for chemical constituents tested and relevant to this project.

(c) SV are target analyte concentrations in fish tissue that are equal to the oral reference dose (RfD) (mg/kg/d or (1 + 1)) for (1 + 1) for (1 + 1) (DL) for

ug/kg/d) for non-carcinogens or the oral slope factor (SF) $[(mg/kg/d)^{-1} \text{ or } (ug/kg/d)^{-1}]$ and a risk level (RL) of 10⁻⁵ for carcinogens. Assumes an average consumption rate (CR) of 6.5 g/day for a 70 kg adult. Values represent the recommended SVs for use in State fish/shellfish comumption advisory programs for the general adult population.

TABLE 9-48A Nereis virens (SAND WORM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUE RESIDUE TO USEPA SCREENING VALUES

| | EPA | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|----------------------|-----------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | SCREENING | | Eastern | Approach | Approach | | Cralghill | Cralghill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | VALUE | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| DIOXINS TEQ (ND=0) | 0.7 | NG/KG | 0.882 | | | | | | 0.577 | | | | | | |
| DIOXINS TEQ (ND=1/2) | 0.7 | NG/KG | 1.36 | | 1.08 | 1.41 | 1.15 | 0.838 | 0.972 | | 0.826 | 0.946 | | | 0.832 |
| 4,4'-DDT | 300 | UG/KG | | | | | | | | | | | 9 | | |
| ENDOSULFAN II | 60000 | UG/KG | 26 | | 22 | | | | 4 | | | 31 | 13 | | |
| ENDRIN | 3000 | UG/KG | | | | | | | | 2 | | | | | |
| GAMMA-BHC | 80 | UG/KG | | 4 | | | | 3 | | 2 | 2 | | | | |
| TOTAL HEPTACHLOR | 10 | UG/KG | 9.68 | | 14 | | | | | | | 11 | 5 | | |



TABLE 9-48B Macoma nasuta (BLUNT-NOSE CLAM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUE RESIDUE TO USEPA SCREENING VALUES

| | EPA | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|----------------------|-----------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | SCREENING | | Eastern | Approach | Approach | | Cralghill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | VALUE | UNITS | Extension | (Surficial) | (Cores) | Cralghill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| DIOXINS TEQ (ND=1/2) | 0.7 | NG/KG | 0.755 | 1.32 | | | 0.753 | | | 0.829 | 0.627 | | 0.659 | | 0.61 |
| MERCURY | 1000 | UG/KG | 121 | 122 | 128 | | | | | | | 120 | | | |
| 4,4'-DDT | 300 | UG/KG | 13 | | 7 | | | 17 | | 10 | 10 | 12 | | | |
| DIELDRIN | 7 | UG/KG | | | | | | | | | | | | 2 | 2 |
| ENDOSULFAN I | 60000 | UG/KG | | | 5 | | | | | | | | | | |
| ENDOSULFAN II | 60000 | UG/KG | 5 | | 3 | | | | | | | 5 | | | |
| GAMMA-BHC | 80 | UG/KG | | | | | | | | 2 | 2 | | | | |
| SELENIUM | 50000 | UG/KG | 910 | | 765 | 938 | 885 | 716 | 863 | 697 | 697 | 711 | 777 | 782 | 917 |
| TOTAL HEPTACHLOR | 10 | UG/KG | | 6 | | | | | | 6 | 6 | 8 | | | |

| Chemical Constitue | Test Species | Exposure Concentratio n | UNITS | Effect | Endpoint | Exposure Route | Body Part | Life Stage | Reference |
|--------------------|--|-------------------------------|-------|----------------------------|--------------|--------------------------|--------------------------|------------|--|
| Silver | <i>Mya arenaria</i> (soft-shell clam) | 10.4 | MG/KG | Physiological | LOED | Absorption | Whole Body | Adult | Thurberg et al. 1974; ERED database |
| Zinc | <i>Mytilus edulis</i> (blue mussel) | 130 | MG/KG | Mortality | ED100 | - | Whole Body | - | Burbridge et al. 1994; USEPA 2000d |
| Benzo(a)pyrene | Mercenaria mercenaria (Quahog) | 2.21 | UG/KG | Physiological Mortality | LOED NOED | Absorption Absorption | Whole Body Whole Body | NA NA | Anderson et al. 1981; ERED database Anderson et al. 1981; ERED database |
| Fluoranthene | <i>Mytilus edulis</i> (blue mussel) | 112 | UG/KG | Physiological | LOED | - | Whole Body | - | Malins et al. 1985; USEPA 2000d |
| Phenanthrene | <i>Mytilus edulis</i> (blue mussel) | 30,700 | UG/KG | Physiological | ED50 | - | Whole Body | - | Donkin et al. 1989; USEPA 2000d |
| Pyrene | <i>Mytilus edulis</i> (blue mussel) | 189,000 | UG/KG | Physiological | ED50 | - | Whole Body | - | Donkin et al.; USEPA 2000d |
| Pentachlorophenol | Nereis virens (sand worm) | 28,000 | UG/KG | Physiological | LOED | Absorption | Whole Body | Adult | Carr and Neff 1981; ERED database |

TABLE 9-49 RESIDUE-EFFECT DATA* FROM PUBLISHED LITERATURE

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* If exposure concentrations from multiple residue-effect studies were available, the lowest exposure concentration that produced an effect was used in the project comparisons.



TABLE 9-50A Nereis virens (SAND WORM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUE RESIDUE TO
RESIDUE EFFECT DATA

| | | | Brewerton | C&D | C&D | | | | | Craighill | | 1 | | | |
|-------------------|----------------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | EFFECT | | Eastern | Approach | Approach | | Craighill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | CONCENTRATIONS | UNITS | Extension | (Surficial) | (Cores) | Craighill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| PENTACHLOROPHENOL | 28000 | UG/KG | | | | | | 188 | 192 | | Γ | | | | |

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TABLE 9-50B Macoma nasuta (BLUNT-NOSE CLAM): COMPARISON OF THE UPPER 95% CONFIDENCE LEVEL OF THE MEAN STEADY-STATE TISSUE RESIDUE TO RESIDUE-EFFECT DATA

| | | | Brewerton | C&D | C&D | | | | | Craighill | | | | | |
|----------------|---------------|-------|-----------|-------------|----------|-----------|------------|------------|-----------|-----------|--------|-------|------------|------------|---------------|
| | EFFECT | | Eastern | Approach | Approach | | Cralghill | Craighill | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| ANALYTE | CONCENTRATION | UNITS | Extension | (Surficial) | (Cores) | Cralghill | Angle East | Angle West | Entrance | Range | Angle | Point | North | South | Straightening |
| SILVER | 10000 | UG/KG | | 315 | 147 | 187 | | | | | | | | 263 | |
| ZINC | 130000 | UG/KG | 18600 | 16500 | 15200 | 18400 | 15900 | 20600 | 18400 | 18300 | 23900 | | 16300 | | 14800 |
| BENZO[A]PYRENE | 2 | UG/KG | | | | | | | | 2.49 | 2.47 | | | | 2.96 |
| FLUORANTHENE | 112 | UG/KG | 13.7 | 14.6 | 14.5 | 16.6 | 13.7 | 14.6 | 19.3 | 30.4 | 21.3 | 16.7 | | 31 | 14.5 |
| PHENANTHRENE | 30700 | UG/KG | | | | | | 3.46 | | | | | | | 5.48 |
| PYRENE | 189000 | UG/KG | 5.16 | 6.79 | 5.77 | 4.67 | | 6.26 | | 7.41 | 7.13 | 8.82 | | 6.47 | 6.32 |



TABLE 9-51A Nereis virens (SAND WORM): ACUTE AND CHRONIC CRITICAL BODY RESIDUE FOR PAHS AND PESTICIDES

| | EFFECT | | Brewerton | C&D | C&D | | Craighill | Craighill | | CraighIll | | | | | |
|-----------------------------|---------------------|--------------|-----------|-------------|----------|-----------|-----------|-----------|-----------|-----------|--------|-------|------------|------------|---------------|
| | CONCENTRATION | UNITS | Eastern | Approach | Approach | | Angle | Angle | Cralghill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| NEUTRAL ORGANICS | (RANGE) | (wet welght) | Extension | (Surficial) | (Cores) | Cralghill | East | West | Entrance | Range | Angle | Point | North | South | Straightening |
| PAHs (acute) | 2 - 8 (acute) | mmol/KG | 0.4 | 0.2 | 0.2 | 0.05 | 0.003 | 0.006 | 0.003 | 0.003 | 0.003 | 0.08 | 0.2 | 0.06 | 0.003 |
| PAHs (chronic) | 0.2 - 0.8 (chronic) | mmol/KG | 0.04 | 0.02 | 0.02 | 0.005 | 0.0003 | 0.0006 | 0.0003 | 0.0003 | 0.0003 | 0.008 | 0.02 | 0.006 | 0.0003 |
| PAHs + pesticides (acute) | 2 - 8 (acute) | mmol/KG | 0.4 | 0.2 | 0.2 | 0.05 | 0.006 | 0.01 | 0.006 | 0.007 | 0.009 | 0.09 | 0.2 | 0.07 | 0.008 |
| PAHs + pesticides (chronic) | 0.2 - 0.8 (chronic) | mmol/KG | 0.04 | 0.02 | 0.02 | 0.005 | 0.0006 | 0.001 | 0.0006 | 0.0007 | 0.0009 | 0.009 | 0.02 | 0.007 | 0.0008 |

TABLE 9-51B Macoma nasuta (BLUNT-NOSE CLAM): ACUTE AND CHRONIC CRITICAL BODY RESIDUE FOR PAHs AND PESTICIDES

| | EFFECT | | Brewerton | C&D | C&D | | Craighill | Craighill | | Craighill | | | | | |
|-----------------------------|---------------------|--------------|-----------|-------------|----------|-----------|-----------|-----------|-----------|-----------|--------|-------|------------|------------|---------------|
| | CONCENTRATION | UNITS | Eastern | Approach | Approach | | Angle | Angle | Craighill | Upper | Cutoff | Swan | Tolchester | Tolchester | Tolchester |
| NEUTRAL ORGANICS | (RANGE) | (wet weight) | Extension | (Surficial) | (Cores) | Craighill | East | West | Entrance | Range | Angle | Point | North | South | Straightening |
| PAHs (acute) | 2 - 8 (acute) | mmol/KG | 0.04 | 0.02 | 0.04 | 0.007 | 0.008 | 0.03 | 0.01 | 0.04 | 0.05 | 0.05 | 0.06 | 0.04 | 0.04 |
| PAHs (chronic) | 0.2 - 0.8 (chronic) | mmol/KG | 0.004 | 0.002 | 0.004 | 0.0007 | 0.0008 | 0.003 | 0.001 | 0.004 | 0.005 | 0.005 | 0.006 | 0.004 | 0.004 |
| PAHs + pesticides (acute) | 2 - 8 (acute) | mmol/KG | 0.04 | 0.02 | 0.04 | 0.01 | 0.01 | 0.04 | 0.01 | 0.04 | 0.05 | 0.05 | 0.06 | 0.04 | 0.05 |
| PAHs + pesticides (chronic) | 0.2 - 0.8 (chronic) | mmol/KG | 0.004 | 0.002 | 0.004 | 0.001 | 0.001 | 0.004 | 0.001 | 0.004 | 0.005 | 0.005 | 0.006 | 0.004 | 0.005 |

TABLE 9-54 COPCs IDENTIFIED IN TISSUE-RESIDUES FOR EACH CHANNEL/PLACEMENT OPTION

| | E | rewe | rton | | | | T | | | T | | | 1 | | | T | | | T | | | T | | | | - | | r= | | - | 1 | _ | | T | | - | T | _ | |
|-----------------------|----------------|-----------------|-------|------------|--------|----------------|------|----------------|--------|----------|--|------|------|--------|-------|----------|--------|----------|------|----------|----------|----------|------------------|-------|------|-------|------|------|------------|------|------|------------|----------|--|------------|----------|------------|----------|----------|
| CHANNE | 4 | Easte | ern | C& | D Ap | proac | h C& | D App | oroach | | | | Cra | ighill | Angle | Cra | ighill | Angle | e (| Craigl | hill | Сга | ighill U | Jpper | | | | | | | Т | olche | ster | T | olches | ster | T/ | olche | ster |
| | | Exten | sion | 1 | Surfic | cial) | - | (Core | s) | <u> </u> | Craigl | hill | | Eas | t | | Wes | t | 1 | Entra | nce | | Range | 2 | Cut | off A | ngle | Sv | van Pe | oint | | Nort | h | | Sout | h | Str | aigth | ening |
| | × I | 104 | | T | 104 | | 4 | 2 | 1 | 4 | 8 | | 4 | 5 | | 7 | 8 | | - | 8 | | - | 3 | | | Z | | | 8 | | | 7 | | | Z | | | 4 | |
| | e e | ie l | | e 10 | ite | | 10 | ite | | 9 | ite | | 2 | E | | 2 | te] | | | te 1 | | 2 | te 1 | | 104 | le 10 | | 10 | te 1 | | 5 | e 10 | | 101 | ie 10 | 1 | 104 | 10 | |
| | e Si | de S | - | Sit | de S | | Sit | le S | | Sit | de S | | Sit | c S | | Site | e Si | | Site | c Si | | Site | e Si | | Site | c Si | | Site | c Si | | Site | Sil | | Site | e Sil | | Site | e Si | |
| COPCs | lsid | utsi | ceal | side | utsi | cear | side | utsic | cear | side | utsic | cear | side | Itsic | can | side | Itsid | can | side | Itsid | can | side | Itsid | can | ide | tsid | can | e. | Itsid | can | e e | tsid | can | ide | tsid | ean | ide | tsid | can |
| DIOXIN TEO | ╉╧ | 0 | -0 | - <u>-</u> | - 0 | 10 | - 5 | 0 | -0 | 5 | <u> </u> | Ó | | ļō. | ŏ | Ins | Ő | ŏ | Ĕ | ō | ŏ | Ins | õ | ð | Ins | Ou | õ | Ins | õ | ő | Ins | õ | ő | Ins | on | ő | Ins | õ | ő |
| ALLIMINUM | | - W | w,c | - | c | c | | W | | | l w | w,c | - | W | w,c | - | W | - | | W | | <u> </u> | | С | | w | с | w | ÷ | | 1 | | с | | | | | w | с |
| ANTIMONY | | + | | W | - | - | - | | | c | C | | c | L C | | W | | | С | | | w | W | _ | w | W | | | - | | w | W | | w | W | | w | w,c | |
| ARSENIC | ι - · | | + | + | + | + | w,c | W | | c | c | L C | C C | C | | w | W | ┥── | С | | | I | | | | _ | | w,c | W | | w,c | w | | w,c | w | | с | | + |
| BERYLLIUM | 111/ | L W C | | | C | + | | - | | c | c | ─ | ┨─── | C C | | | C | _ | С | с | | | С | | | с | | | c | | | | | | | | | с | - |
| CADMILIM | " | + ", | . w,c | - C | C | C | W | w,c | w,c | | ┨──── | ┨─── | | ┥─── | | w | w,c | w,c | - | - | | w | w,c | w,c | w | w,c | w,c | W | w,c | w,c | w | w,c | W,C | w | w,c | w,c | w | w,c | w,c |
| CHROMIUM | wc | 1 W C | + | + | | - | - | | | | | | | | | | | - | - | - | | | | | | | | _ | | | | | | | | | | | |
| COPPER | w,c | W,C | | | w,c | + | W | w,c | | c | w,c | | c | w,c | c | w,c | w,c | | С | w,c | | w,c | w,c | | w | w,c | | w | w,c | | w | w,c | | w | w,c | | W | w,c | |
| IRON | " | + <u>"</u> | | W | w | | - | W | | | W | | | W | | | W | - | - | w | - | | W | _ | - | w | _ | _ | - W' | | | w | | | w | | | w | |
| MANGANESE | | | L C | | C | C | | C | C | C | c | c | c | c | c | | С | c | - | c | c | - | С | с | | С | с | | с | с | | С | с | | | | | С | C |
| MERCURY | | | + c | w,c | . w,c | w,c | w,c | w,c | C | C | - C | c | c | w,c | С | С | C | С | c | w,c | c | с | c | С | w,c | w,c | с | с | w,c | с | с | С | c | | C | с | | | c |
| NICKEL | wc | wc | 1 | | W.C | | | C | C | | | - | | | - | | | <u> </u> | | | - | | | | | | | | С | с | | - | <u> </u> | | | | | | |
| SELENIUM | <u> </u> | ,c | + | <u>-</u> | w,c | + | w | w,c | | w,c | w,c | | c | w,c | c | с | w,c | | С | w,c | | w,c | w,c | - | w | w,c | | _ | w,c | | w | w,c | | W | w,c | | w | w,c | - |
| SILVER | Ť | <u> </u> | WC | + | + | | C C | C | | C | C | - | c | C | С | c | С | | С | c | | с | С | | с | с | | С | с | | С | с | | С | С | | С | c | c |
| ZINC | C | | w,c | + | we | 0 | | | C | | | | | | | | | | | | <u> </u> | | | | | | | | | w | | <u> </u> | w | <u> </u> | | c | | | |
| ACENAPHTHYLENE | Ť | <u>ب</u> | | - | w,c | C | | | | C | w,c | l w | | C | | c | c | | c | c | | c | с | - | | с | - | | | | | С | _ | ' | W | <u> </u> | | c | + |
| ANTHRACENE | | ┼── | + | | | | 1 | - ^c | L L | | | | | | | | c | c | - | | | | C | с | | С | с | | | | с | C | С | ──' | ' | ┣── | С | c | c |
| BENZIAIANTHRACENE | 1 | † | | C | | C | | + | | | | | | | | | | c | - | | | | + | | | | с | | | с | | ' | ┣── | ' | ' | | | c | c |
| BENZOLAIPYRENE | 1 | 1 | | - · | - | - ^c | - | | | | | | | | | c | | c | - | - | | c | $\left \right $ | С | с | | с | | | | с | ' | С | с | | с | с | \vdash | c |
| BENZO[B]FLUORANTHENE | 1 | | C | C | C | C | t | <u> </u> | 6 | | | 6 | | | 0 | <u> </u> | ┨─── | | | | | | + | C | | | c | | | | - | ' | + | ' | | <u> </u> | <u> </u> ' | <u> </u> | C |
| CHRYSENE | | C | C | Ť | C | wc | 1 | t | | | | L. | | | C | | | C C | ┣── | | C | - | | c | | | с | | | с | - | | C | | - | C | С | C C | c |
| FLUORANTHENE | | <u> </u> | c | | | C | | | C | | | - C | | | 6 | | | | | | | | | c | | c | c | - | | | | с | С | ┝──′ | c | C | | c | c |
| FLUORENE | | | c | | | c | | | Ċ | | | C | | | 6 | | | | | - | C | - | | 0 | | C | C O | | _ | C | | | W | | | C | | <u> </u> | C |
| NAPHTHALENE | | | | | 1 | - | | | | | | c | | | C | | | | | | C | | + | 0 | | | - | | - | c | | \vdash | C C | \vdash | | C | C | | |
| PHENANTHRENE | | | | | | | | | | | | | | _ | - | | | C | | | - C | | | C | | - | - | | _ | - | - | | L | \vdash | <u> </u> ' | | C | - | |
| PYRENE | | | с | | | с | | | с | | | с | | | | 1.1 | | c | | | | - | | c | | - 1 | c | | | 6 | | | | | \vdash | C | | | C |
| TOTAL PAHs | | | с | | | с | | | с | | | с | | | с | _ | | c | | | с | | c | c | | с | c | | | c | wc | C | | | | C | | C | - C |
| TOTAL PCBs | | с | С | | | с | | с | с | | | w,c | | | с | | | | | c | w.c | | | - | | | w | | с | c | ,. | | Ť | | \square | | | <u> </u> | <u> </u> |
| 4,4'-DDD | | | | w | w | w | с | С | с | | | | | | | | | | | w | | с | с | с | c | w.c | c | | | - | w | w | w | w | w | w | w | w | w |
| 4,4'-DDT | с | с | с | | | | с | с | с | | | | | | | с | с | с | | | | с | с | с | c | c | с | с | с | с | w | | | | | | | | + |
| ALDRIN | | | | w | w | w,c | | | | | | с | | | с | | w | | | w | с | | w | | | | | | | | | | | | | | | | |
| ALPHA-BHC | | | | w | w | w | с | с | с | | | | | | | с | | | | | | | | | | | | | | | | | | | | | | | |
| BETA-BHC | | | | | | | | | | | | с | | | С | w | w | w,c | | | | | | | | | | | | w | | | | | | | | | |
| CHLORBENSIDE | | | | w | w | w | | | | | w | | | | | | w | | | w | w | | w | | | w | w | | | | | | | | | | | w | W |
| DACTHAL | | w | | w | w | w | w | w | w | | | | | | | | | | | | | w | w | | w | w | w | | w | | w | w | w | w | w | w | w | w | w |
| DELTA-BHC | w | w | | | | | w | w | | | | | | | | | | | w | w | | | | | | | | w | w | | | | | | | | | | |
| DIELDRIN | | | | | | | | | | | | | | | | | | | | | | | | | | | | | - | | | | | с | с | | С | с | |
| ENDOSULFAN I | | | | | | | | | с | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| ENDOSULFAN II | w,c | с | с | | | | w,c | с | с | | | | | | _ | | | | | | | | | | | | | w,c | с | с | w | | | | | | | | |
| ENDRIN | | | | | | | | | | | | | | | | | | | | | | w | w | _ | | | | | | | | | | | | | | | |
| GAMMA-BHC | | | | W | w | | | _ | | | | | | | | | w | | | w | | с | w,c | | w,c | w,c | с | | | | | | | | | | | | |
| HEPTACHLOR | w | | w | | С | | w | | w | | | | | | | | | | | | | | с | | с | с | | w | W,C | w | | | | | | | | | |
| HEPTACHLOR EPOXIDE | w | W | | | | _ | w | w | _ | | | | | | | | | | | | | | | | | | | w | w | | w | w | | | | | | | |
| I-METHYLNAPHTHALENE | | | | | | _ | | | | | | | | | | _ | - | | | | | | | | | | | | | | | | | | С | С | | с | с |
| 2-METHYLPHENOL | w | W | w | | | | w | w | w | | | | | | | | | | с | с | с | | | с | | | с | w | w | w | с | с | с | с | с | с | с | с | с |
| 3,4-METHYLPHENOL | | | С | с | С | С | С | С | с | | | с | | | с | | | С | с | | с | | | | | | с | с | с | с | | | с | | | с | | | с |
| 4-NITROPHENOL | | | | | | | | | | | | | | | | | | | с | с | с | | | | | | | | | | | | | | | | | | |
| BENZUIC ACID | | | | | С | | | | | | | | | | | | | | | с | | | | | | | | | | | с | С | | с | с | | | С | |
| BENZYL ALCOHOL | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | с | | | | | | |
| DI-N-BUTYL PHTHALATE | | 1.1 | | | | - | | - | - | | | | - 17 | 1 | 1.1 | w | | w | | | 1.5 | | | | _ | | 1.1 | | | 1 | | - | 4 17 | | | | | | |
| DI-IN-OCTYL PHTHALATE | - 1 | | _ | | | | | | | | | | | | | 1 | | - | | | | 1 | 1 | | | | | | t Estina a | | | | | | | | w | w | W |
| PUENOL | | | | | ļ | | | | | | | | | | | W | W | W | W | w | w | | | | | | | | | | | | | | | | | | |
| PHENOL | | | с | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | W | w | w | | | |

c = clam tissue; w = worm tissue

* Any constituent that exceeded a placement site/reference area tissue-residue was retained as a COPC.

| Chemical Constituent | Units (2) | Detected in Bulk | Detected in | Mean Tissue | | Mean Tissue Residue | Frequency of Detection | Uptake | nside 104 W | agnitu eedano | de of ce (%) | Propensity to | ITM Ev Exceeds USFDA Action | Exceeds USEPA | USH Supple Evid Exceeds USEPA Screening | CPA Region I mental Weigh ence Screenin Exceeds Residue-Effect | II nt-of- ng Exceeds | Propos Keep as | ed Action | |
|-----------------------|-----------|---------------------|----------------|-------------|------|---------------------------|---------------------------|--------|-------------|------------------|-----------------|---------------|--------------------------------------|------------------|--|--|-------------------------------|-------------------|-----------|--------------------|
| DIOXIN TEO (ND=0) | NG/KG | Yes | Yes | 0 273 | | < TDL? | in Tissue | Ratio | - | 0 | | Biomagnify? | Level? | Tolerance? | Value? | Data? | RBC? | COPC | as COPC | Comments |
| | | | | 0.275 | IVA | IN/A | 5/5 | 1.97 | | 19.2 | | Yes | * | * | Yes | * | Yes | ~ | | |
| DIOXIN TEQ (ND=1/2DL) | NG/KG | Yes | Yes | 0.427 | N/A | N/A | 5/5 | 3.03 | | 143 | 66.9 | Yes | * | * | Yes | * | Yes | ~ | | |
| BERYLLIUM | MG/KG | Yes | No | 0.07 | 0.1 | Yes | 4/4 | 6.75 | 575 | 575 | 575 | No | * | * | * | * | No | | - | |
| CHROMIUM | MG/KG | Yes | Yes | 0.53 | 0.1 | No | 4/4 | 1.84 | 249 | 428 | _ | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.68 | 0.1 | No | 4/4 | 1.28 | 234 | 90.8 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 1.00 | 0.1 | No | 4/5 | 8.85 | 104 | 255 | | No | * | No | * | * | No | | ~ | |
| SILVER | MG/KG | Yes | NO | 0.13 | 0.1 | No | 3/4 | 1.00 | | - | 225 | No | * | * | * | * | No | | ~ | |
| DACTHAL | UG/KG | No | No | 50.60 | 2 | No | 5/5 | 1.64 | | 406 | | Yes | * | * | * | * | No | | ~ | |
| DELTA-BHC | UG/KG | No | No | 2.02 | NONE | N/A | 5/5 | 0.36 | 486 | 486 | | Yes | * | * | * | * | Yes | | ~ | UR <i< td=""></i<> |
| ENDOSULFAN II | UG/KG | No | No | 17.68 | 10 | No | 5/5 | 3.17 | 124 | | | Yes | * | * | No | * | No | | ~ | |
| HEPTACHLOR | UG/KG | No | Yes | 3.37 | 10 | Yes | 4/5 | 9.91 | 188 | | 891 | Yes | No | * | * | * | Yes | | ~ | |
| HEPTACHLOR EPOXIDE | UG/KG | No | Yes | 2.04 | 10 | Yes | 5/5 | 5.59 | 181 | 326 | | Yes | No | * | * | * | Yes | | ~ | |
| 2-METHYLPHENOL | UG/K·G | No | No | 196 | 20 | No | 5/5 | 2.17 | 160 | 131 | 111 | No | * | * | * | * | No | | ~ | |

TABLE 9-55A BREWERTON EASTERN EXTENSION: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

.

N/A=not applicable

*no criterion

(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

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TABLE 9-55B: BREWERTON EASTERN EXTENSION: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | | | | | | | | M | agnitu eedano | de of ce (%) | | ITM E | aluation | USE Suppler Evid | CPA Region mental We ence Scree | n III eight-of- ening | Ргоро | oscd Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|--------------------------------|---------------|--------------------------------------|--|-----------------|------------|------------------|-----------------|---------------------------------|--------------------------------------|--------------------------------|--|--|-----------------------------|--------------------|----------------------|--|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissu Concentrat on | e i TDL | Mean Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screenin g Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL) | NG/KG | Yes | Yes | 0.239 | N/A | N/A | 5/5 | 1.36 | | | 90.5 | Yes | * | * | Yes | + | Yes | ~ | | |
| ALUMINUM | MG/KG | Yes | Yes | 67.44 | 1.0 | No | 5/5 | 7.59 | 108.0 | 122.0 | 105.0 |) No | * | * | * | * | No | | | |
| ANTIMONY | MG/KG | Yes | Yes | 0.11 | 0.1 | No | 5/5 | 3.03 | 155.0 | | | No | * | * | * | * | Yes | | - | UCLM exceeds 1/10 of RBC, but does not exceed RBC |
| ARSENIC | MG/KG | Yes | Yes | 2.60 | 0.1 | No | 5/5 | 1.07 | | 22.4 | - | No | * | No | No | * | Yes | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 0.47 | | 70.0 | 70.0 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.76 | 0.1 | No | 5/5 | 1.92 | 69.1 | 656.0 | | No | * | No | * | * | Yes | - | ~ | |
| IRON | MG/KG | Yes | Yes | 172.6 | 10 | No | 5/5 | 2.63 | | 44.7 | 42.9 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 of RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 22.46 | 0.5 | No | 5/5 | 8.20 | 204.0 | 261.0 | 806.0 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 of RBC, but does not exceed RBC |
| MERCURY | MG/KG | Yes | No | 0.12 | 0.01 | No | 5/5 | 0.97 | 43.9 | 88.8 | 151.0 | Yes | No | * | No | * | Yes | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 1.08 | 0.1 | No | 5/5 | 1.69 | 48.1 | 259.0 | | No | * | No | + | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | Yes | 0.80 | 0.2 | No | 5/5 | 1.93 | 91.4 | 122.0 | | No | * | * | No | * | Yes | | ~ | |
| ZINC | MG/KG | Yes | No | 17.14 | 2.0 | No | 5/5 | 1.05 | 19.6 | 37.9 | | No | * | * | * | No | No | | ~ | |
| BENZO[B]FLUORANTHENE | UG/KG | Yes | No | 7.44 | 20 | Yes | 5/5 | 3.61 | | | 170.0 | No | * | * | * | * | Yes | | ~ | Detected in laboratory method blank |
| CHRYSENE | UG/KG | Yes · | No | 4.52 | 20 | Yes | 5/5 | 2.47 | - | 138.0 | 40.4 | No | * | * | * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 9.32 | 20 | Yes | 5/5 | 1.06 | 1 | | 24.6 | No | * | * | * | No | No | | ~ | |
| FLUORENE | UG/KG | Yes | No | 7.88 | 20 | Yes | 5/5 | 0.87 | | | 294.0 | No | * | * | + | * | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 4.38 | 20 | Yes | 5/5 | 1.58 | | | 36.9 | No | * | + | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 51.6-70.1 | NONE | N/A | 5/5 | 1.85 | - | | 36.4, 48.6 | No | * | + | * | + | * | | ~ | CBR <acute chronic="" or="" td="" threshold<=""></acute> |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 9.76 | NONE | N/A | 5/5 | 0.70 | - | 181.0 | 917.0 | Yes | No | * | * | * | Yes | | ~ | UR<1 |
| 4,4'-DDT | UG/KG | No | No | 3.64 | 10 | Yes | 2/5 | 0.34 | 148.0 | 507.0 | 507.0 | Yes | No | * | No | * | Yes | | ~ | UR<1 |
| ENDOSULFAN II | UG/KG | No | No | 4.44 | 10 | Yes | 5/5 | 4.28 | 450.0 | 933.0 | 933.0 | Yes | * | * | No | * | No | | ~ | |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 109 | NONE | N/A | 4/5 | 0.58 | _ | | 227.0 | No | * | * | * | * | No | | ~ | |
| PHENOL | UG/KG | No | No | 100 | 20 | No | 5/5 | 1.37 | | | 77.3 | No | * | * | * | * | No | -+ | ~ | |

N/A=not applicable

*no criterion

• .

| | | | | | | | | | M Exc | agnitud eedanco | le of e (%) | | ITM E | aluation | USE Supplen Evide | PA Region lental Wei | III ght-of- ning | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|------|-------------------------------------|--|-----------------|------------|--------------------|----------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Delected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uplake Ralio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effec1 Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Commenis |
| ALUMINUM | MG/KG | Yes | Yes | 14.6 | 1.0 | No | 5/5 | 0.30 | 69.5 | | | NO | * | * | * | * | No | <u> </u> | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.15 | 0.1 | No | 1/5 | 0.53 | - | 52 | | No | * | No | * | * | No | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.86 | 0.1 | No | 5/5 | 1.42 | 195 | 112 | | Yes | * | No | * | * | No | | ~ | |
| MANGANESE | MG/KG | Yes | Yes | 3.78 | 0.5 | No | 5/5 | 1.43 | 278 | 473 | 215 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.65 | 0.1 | No | 5/5 | 5.70 | | 129 | | No | * | * | * | * | No | | ~ | |
| ZINC | MG/KG | Yes | Yes | 26.76 | 2.0 | No | 5/5 | 2.11 | | 104 | | No | * | * | | * | No | | ~ | |
| CHRYSENE | UG/KG | Yes | No | 2.64 | 20 | Yes | 3/4 | 0.12 | | | 998 | Yes | * | * | * | * | No | | ~ | |
| 4,4'-DDD | UG/KG | No | No | 5.88 | 10 | Yes | 5/5 | 7.84 | 135 | 684 | 187 | Yes | No | * | * | * | Yes | | ~ | <tdl; detected="" in="" not="" sediment<="" td=""></tdl;> |
| ALDRIN | UG/KG | No | No | 28.60 | 10 | No | 5/5 | 2.25 | 107 | 10700 | 484 | Yes | No | * | * | * | Yes | | ~ | <tdl; detected="" in="" not="" sediment<="" td=""></tdl;> |
| ALPHA-BHC | UG/KG | No | No | 3.54 | NONE | N/A | 5/5 | 9.83 | 75.8 | 641 | 287 | Yes | * | * | * | * | Yes | ~ | | not detected in sediment |
| CHLORBENSIDE | UG/KG | No | No | 35.20 | 2 | No | 5/5 | 3.60 | 136 | 2030 | 628 | Yes | * | * | * | * | * | ~ | | not detected in sediment; no criterion |
| DACTHAL | UG/KG | No | No | 116.80 | 2 | No | 5/5 | 3.80 | 363 | 1070 | 363 | Yes | * | * | * | * | Nọ | | ~ | |
| GAMMA-BHC | UG/KG | No | Yes | 3.16 | 10 | Yes | 5/5 | 1.20 | 130 | 671 | | Yes | * | * | No | * | Yes | | ~ | <tdl< td=""></tdl<> |

TABLE 9-56A C&D APPROACHES (SURFICIAL): INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

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*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

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TABLE 9-56B C&D APPROACH (SURFICIAL): INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | | | | | | | | Magni | tude of E (%) | xceedance | | ITM E | valuation | USEI Supplem Evide | PA Region ental Wei nce Screer | III ght-of- ning | Propos | sed Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|------------------|-----------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|--|
| Chemical Constituent | Units ⁽²⁾ | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL | NG/KG | Yes | No | 0.395 | N/A | N/A | 5/5 | 2.25 | | 55.4 | 215.0 | Yes | * | * | Yes | * | Yes | ~ | | |
| ARSENIC | MG/KG | Yes | Yes | 2.58 | 0.1 | No | 5/5 | 1.07 | | 21.4 | - | No | * | No | No | * | Yes | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.08 | 0.1 | Yes | 5/5 | 1.06 | 68.9 | 280.0 | 280.0 | No | * | * | * | * | No | | ~ | |
| CADMIUM | MG/KG | Yes | No | 0.04 | 0.1 | Yes | 2/5 | 0.76 | | 204.0 | | No | * | No | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.58 | 0.1 | No | 5/5 | 1.48 | | 482.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | No | 200.20 | 10 | No | 5/5 | 3.05 | | 67.9 | 65.7 | No | * | * | * | * | No | | ~ | |
| MANGANESE | MG/KG | Yes | Yes | 12.30 | 0.5 | No | 5/5 | 8.20 | 66.4 | 97.6 | 396.0 | No | * | * | * | * | No | | ~ | |
| MERCURY | MG/KG | Yes | Yes | 0.12 | 0.01 | No | 5/5 | 0.95 | 41.5 | 85.6 | 147.0 | Yes | No | * | No | * | Yes | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 1.14 | 0.1 | No | 5/5 | 1.79 | 56.6 | 280.0 | | No | * | No | * | * | No | | ~ | |
| SILVER | MG/KG | Yes | No | 0.23 | 0.1 | No | 4/5 | 4.88 | | | 333.0 | No | * | * | * | No | No | | ~ | |
| ZINC | MG/KG | Yes | Yes | 15.42 | 2 | No | 5/5 | 0.94 | | 24.1 | | No | * | * | * | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | Yes | No | 208 | 20 | No | 5/5 | 0.75 | | 890.0 | 890.0 | No | * | * | * | * | No | | ~ | |
| BENZO[A]ANTHRACEN | UG/KG | Yes | No | 3.29 | 20 | Yes | 4/5 | 4.39 | 186.0 | - | 339.0 | No | * | * | * | * | Yes | ~ | | |
| BENZO[B]FLUORANTHE | UG/KG | Yes | No | 16.20 | 20 | Yes | 5/5 | 7.86 | 43.6 | 69.2 | 487.0 | No | * | * | * | * | Yes | ~ | | |
| CHRYSENE | UG/KG | Yes | No | 4.90 | 20 | Yes | 5/5 | 2.68 | | 158.0 | 52.2 | No | * | * | * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 9.60 | 20 | Yes | 5/5 | 1.10 | | | 28.3 | No | * | * | * | No | No | | ~ | |
| FLUORENE | UG/KG | Yes | No | 10.64 | 20 | Yes | 5/5 | 1.17 | | | 432.0 | No | * | * | * | * | No | | ~ | · · · · · · · · · · · · · · · · · · · |
| PYRENE | UG/KG | Yes | No | 5.70 | 20 | Yes | 5/5 | 2.06 | | | 78.1 | No | • * | * | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 56.1-76.1 | NONE | N/A | 5/5 | 0.82 | | | 48.1-83.8 | No | * | * | * | * | * | | ٢ | CBR <acute chronic="" or="" td="" threshold<=""></acute> |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 7.49 | NONE | N/A | 5/5 | 0.54 | | | 680.0 | Yes | No | * | * | * | Yes | | < | UR<1 |
| ALDRIN | UG/KG | No | No | 3.13 | 10 | Yes | 4/5 | 5.90 | | | 339.0 | No | No | * | + | * | Yes | | ~ | |
| HEPTACHLOR | UG/KG | No | Yes | 3.86 | 10 | Yes | 5/5 | 0.45 | | 1040.0 | · | Yes | No | * | * | * | Yes | | ~ | detected in laboratory method blank |
| 3,4-METHYLPHENOL | UG/KG | Yes | Noo | 351 | NONE | N/A | 4/5 | 1.85 | 430.0 | 623.0 | 947.0 | No | * | * | * | * | Noo | | | |
| BENZOIC ACID | UG/KG | No | No | 4500 | 100 | No | 5/5 | 0.93 | | 47.5 | | No | * | * | * | * | No | | ~ | |

N/A=not applicable

*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

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| | | | Detected | | | Mean | | | Magnitu | de of Ex (%) | ceedance | Description | ITM Ev | valuation | USE Suppler Evid | PA Regio nental We ence Scree | n III eight-of- ening | Propos | sed Action | |
|----------------------|-----------|----------------------------------|----------------------|----------------------------------|------|-----------------------------|--|-----------------|-------------|-----------------|----------|-----------------------|---------------------------|--------------------------------|-------------------------------|--|-----------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | in Elutriate ? | Mean Tissue Concentrati on | TDL | Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 10 | Outside 10 | Ocean | to Biomagnify ? | USFDA Action Level? | Exceeds USEPA Tolerance? | USEPA Screenin g Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2I | NG/KG | Yes | | 0.33 | N/A | N/A | 5/5 | 2.35 | | 87.8 | | Yes | * | * | Yes | * | Yes | ~ | | |
| ANTIMONY | MG/KG | Yes | Yes | 0.13 | 0.1 | No | 5/5 | 2.22 | 170 | 163 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.07 | 0.1 | Yes | 5/5 | 7.00 | 600 | 600 | 600 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.45 | 0.1 | No | 5/5 | 1.56 | 195 | 346 | | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.58 | 0.1 | No | 5/5 | 1.21 | | 80 | | Yes | * | * | * | * | No | | ~ | |
| MANGANESE | MG/KG | Yes | Yes | 2.90 | 0.5 | No | 4/5 | 1.09 | 190 | 339 | | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 1.11 | 0.1 | No | 5/5 | 9.79 | 126 | 294 | | No | * | No | * | * | No | | ~ | |
| DACTHAL | UG/KG | No | No | 41.00 | 2 | No | 5/5 | 1.33 | 62.6 | 310 | 624 | Yes | * | * | * | * | No | | ~ | |
| DELTA-BHC | UG/KG | No | No | 2.70 | NONE | N/A | 5/5 | 0.49 | 683 | 683 | | Yes | * | * | * | * | Yes | | ~ | UR<1 |
| ENDOSULFAN II | UG/KG | No | No | 16.78 | 10 | No | 5/5 | 3.01 | 112 | | | Yes | * | * | No | * | No | | ~ | |
| HEPTACHLOR | UG/KG | No | Yes | 4.59 | 10 | Yes | 4/5 | 13.49 | 292 | | 1250 | Yes | No | * | * | * | Yes | | ~ | <tdl action="" and="" fda="" level<="" td=""></tdl> |
| HEPTACHLOR EPOXI | UG/KG | No | No | 3.52 | 10 | Yes | 5/5 | 9.64 | 384 | 635 | - | Yes | No | * | * | * | Yes | | ~ | |
| 2-METHYLPHENOL | UG/KG | No | No | 254 | 20 | No | 5/5 | 2.81 | 237 | 200 | | No | * | * | * | * | No | | ~ | |

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TABLE 9-57A C&D APPROACHES (CORES): INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

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*no criterion

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| | 1 | | 1 | 1 | | 1 | 1 | T | 1 | | | | ······· | | | | | | | T |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|----------|------------------------------|--|-----------------|------------|---------------------|----------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|---------|-------------|---|
| | | | | | | | | | N Ex | Aagnitud ceedanc | ie of e (%) | | ITM E | valuation | USE Supplen Evide | PA Region nental Wei ence Scree | III ght-of- ning | Prop | osed Action | |
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissu Concentration n | e TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds | Keep as | Eliminate | Comments |
| ANTIMONY | MG/KG | Ves | Ves | 0.14 | | N | | | | | 1 | · · · · · | | | , and c. | | KDC. | Core | ascore | Comments |
| BERYLLIUM | MG/KG | Vec | No | 0.14 | 0.1 | | 5/5 | 4.00 | 237.0 | | | No | * | | · · | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| CHROMUM | MG/KG | Vee | Ver | 0.03 | 0.1 | Yes | 5/5 | 0.44 | | 60.0 | 60.0 | No | * | * | * | * | No | | · · | |
| | MORO | 1 CS | Tes | 0.66 | 0.1 | No | 5/5 | 1.69 | | 564.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | No | 196.00 | 10 | No | 5/5 | 2.98 | | 64.4 | 62.3 | No | * | + | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 10.86 | 0.5 | No | 5/5 | 7.24 | 47.0 | 74.5 | 338.0 | No | * | * | * | + | No | | ~ | |
| MERCURY | MG/KG | Yes | Yes | 0.12 | 0.01 | No | 5/5 | 0.95 | | 85.6 | 147.0 | Yes | No | * | No | * | Yes | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.97 | 0.1 | No | 5/5 | 1.52 | | 223.0 | | No | * | No | * | * | No | | > | |
| SELENIUM | MG/KG | Yes | No | 0.65 | 0.2 | No | 5/5 | 1.56 | 54.3 | 78.8 | | No | + | * | No | * | Yes | | ~ | |
| SILVER | MG/KG | Yes | No | 11 | 0.1 | No | 4/5 | 2.25 | | | 100.0 | No | * | * | * | No | No | | ~ | |
| ZINC | MG/KG | Yes | No | 14.30 | 2.0 | No | 5/5 | 0.87 | | 15.1 | - | No | * | * | * | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | No | No | 526.20 | 20 | No | 4/5 | 1.89 | | 2410.0 | 2410.0 | Yes | + | * | * | * | No | | ~ | |
| BENZO[B]FLUORANTHEN | UG/KG | Yes | No | 9.44 | 20 | Yes | 5/5 | 4.58 | - | | 242.0 | Yes | * | * | * | * | Yes | | ~ | detected in the laboratory method blank |
| FLUORANTHENE | UG/KG | Yes | No | 9.34 | 20 | Yes | 5/5 | 1.07 | | | 24.9 | Yes | * | + | * | No | No | | ~ | detected in the laboratory method blank |
| FLUORENE | UG/KG | Yes | No | 9.80 | 20 | Yes | 5/5 | 1.08 | | | 390.0 | Yes | * | * | * | * | No | | ~ | detected in the laboratory method blank |
| PYRENE | UG/KG | Yes | No | 4.94 | 20 | Yes | 5/5 | 1.78 | | | 54.4 | Yes | * | • | + | No | No | | | detected in the laboratory method blank |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 49.4-68.2 | NONE | N/A | 5/5 | 1.77 | | | 43.7, 32.8 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| TOTAL PCBs (ND=0) | UG/KG | Yes | No | 12.3 | NONE | N/A | 5/5 | 0.88 | | 253.0 | 1180.0 | Yes | No | • | • | * | Yes | | ~ | UR<1 |
| 4,4'-DDD | UG/KG | Yes | No | 2.02 | 10 | Yes | 5/5 | 0.46 | 110.0 | 94.2 | 169.0 | Yes | No | * | * | + | No | | ~ | · · · · · · · · · · · · · · · · · · · |
| 4,4'-DDT | UG/KG | No | No | 1.96 | 10 | Yes | 5/5 | 0.80 | 33.3 | 227.0 | 227.0 | Yes | No | * | No | * | No | | ~ | |
| ALPHA-BHC | UG/KG | No | No | 10.32 | NONE | N/A | 5/5 | 28.67 | 2770.0 | 2770.0 | 2770.0 | Yes | * | * | + | * | Yes | ~ | | |
| ENDOSULFAN I | UG/KG | No | No | 3.59 | 10 | Yes | 4/5 | 11.75 | - | | 1080.0 | Yes | + | * | No | + | No | | ~ | · · · · · · · · · · · · · · · · · · · |
| ENDOSULFAN II | UG/KG | No | No | 2.13 | 10 | Yes | 4/5 | 2.05 | 163.0 | 394.0 | 394.0 | Yes | + | * | No | * | No | | ~ | |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 189 | NONE | N/A | 5/5 | 1.00 | 185.0 | 289.0 | 463.0 | No | * | * | * | * | No | | ~ | |

 TABLE 9-57B
 C&D APPROACHES (CORES): INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

N/A=not applicable

*no criterion

| | | | | | | | | | N Ex | Magnitud | e of e (%) | | ITM E | valuation | USE Suppler Evid | PA Regio nental We ence Scree | n III eight-of- ening | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|------|-------------------------------------|--|-----------------|------------|-------------|---------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|-----------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/21 | NG/KG | Yes | No | 0.437 | N/A | N/A | 5/5 | 3.11 | | 148 | 70.8 | Yes | * | * | Yes | * | Yes | ~ | | |
| CHROMIUM | MG/KG | Yes | No | 0.22 | 0.1 | No | 2/5 | 0.75 | | 116 | | No | * | No | * | * | No | | ~ | |
| COPPER | MG/KG | Yes | No | 1.54 | 0.1 | No | 5/5 | 1.18 | | 754 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.92 | 0.1 | No | 5/5 | 8.1 | 86.9 | 226 | | No | * | No | * | * | No | | ~ | |
| ZINC | MG/KG | Yes | No | 33.00 | 2.0 | No | 5/5 | 2.61 | | 152 | 80. I | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| TOTAL PCBs (ND=0) | UG/KG | Yes | No | 31.2 | NONE | N/A | 5/5 | 1.37 | | | 457 | Yes | No | * | * | * | Yes | | ~ | |
| CHLORBENSIDE | UG/KG | No | No | 9.07 | 2 | No | 4/5 | 0.93 | | 450 | | Yes | * | * | * | + | * | | ~ | UR <i< td=""></i<> |

TABLE 9-58A CRAIGHILL: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

*no criterion

TABLE 9-58B CRAIGHILL: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

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| | | | | | | | | | | | | | | | USE | PA Region | III abt of | | | |
|---------------------------------------|----------------------|-------------------|---------------------------|-------------|------|-----------|------------|--------|--------|--------------------|-----------------|---------------|---------|------------|-----------|------------|---------------|---------|-----------|---|
| | | | | | | | | | E, | agnitu (ceedan) | de of ce (%) | | ITM E | aluation | Evide | ence Scree | ning | Propos | ed Action | |
| | | Detected in | | Mean Tissue | | Tissue | Frequency | | 104 | c 104 | u a | | Exceeds | Encode | Exceeds | Exceeds | | | | |
| Chemical Constituent | Units ⁽²⁾ | Bulk Sediment? | Detected in Elutriate? | Concentrati | TO | Residue < | Dectection | Uptake | Inside | Dutsid | 066 | Propensity to | Action | USEPA | Screening | Effect | Exceeds | Keep as | Eliminate | |
| ALUMINUM | MG/KG | Yes | Yes | 48.42 | | No. | in Tissue | Katio | 40 (| | <u> </u> | Biomagnity? | Level? | Tolerance? | Value? | Data? | RBC? | COPC | as COPC | Comments |
| | | | | 40.42 | 1.0 | 140 | 3/3 | 3.43 | 49.0 | 59.3 | | | | | * | * | No | | | UCLM exceeds 1/10 RBC |
| ANTIMONY | MG/KG | Yes | Yes | 0.19 | 0.1 | No | 5/5 | 5.49 | 363.0 | 96.9 | 15.7 | No | * | * | * | * | Yes | | ~ | but does not exceed RBC |
| ARSENIC | MG/KG | Yes | Yes | 2.78 | 0.1 | No | 5/5 | 1.15 | 16.3 | 30.8 | | No | * | * | No | No | Yes | | ~ | |
| CHROMIUM | MG/KG | Yes | No | 2.51 | 0.1 | No | 5/5 | 6.38 | 462.0 | 2410.0 | | No | * | * | No | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 232.00 | 10 | No | 5/5 | 3.53 | 33.0 | 94.5 | 62.3 | No | * | * | * | * | Yes | | - | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 14.90 | 0.5 | No | 5/5 | 9.93 | 102.0 | 139.0 | 501.0 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 2.54 | 0.1 | No | 5/5 | 3.99 | 249.0 | 747.0 | | No | * | * | No | * | Yes | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.83 | 0.2 | No | 5/5 | 1.99 | 96.7 | 128.0 | - | No | * | * | No | * | Yes | | ~ | |
| SILVER | MG/KG | No | No | 0.13 | 0.1 | No | 4/5 | 2.75 | 18.2 | | 144.0 | No | * | * | * | No | No | | ~ | |
| ZINC | MG/KG | Yes | No | 16.94 | 2.0 | No | 5/5 | 1.03 | | 36.3 | | No | * | * | * | No | No | | ~ | |
| BENZO[B]FLUORANTH | UG/KG | Yes | No | 7.66 | 20 | Yes | 5/5 | 3.72 | - | | 178.0 | No | * | * | * | * | Yes | ~ | | |
| FLUORANTHENE | UG/KG | Yes | No | 10.18 | 20 | Yes | 5/5 | 1.16 | - | _ | 36.1 | No | * | * | * | No | No | | ~ | , , , , , , , , , , , , , , , , , , , |
| FLUORENE | UG/KG | Yes | No | 10.02 | 20 | Yes | 5/5 | 1.10 | - | | 401.0 | No | * | * | * | * | No | | ~ | |
| NAPHTHALENE | UG/KG | Yes | No | 36.00 | 20 | No | 5/5 | 1.71 | | - | 174.0 | No | * | * | * | * | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 4.22 | · 20 | Yes | 5/5 | 1.52 | | | 31.9 | No | * | * | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 74-87.3 | NONE | N/A | 5/5 | 0.23 | - | 1 | 143, 97, 70 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| TOTAL PCBs (ND=0) | UG/KG | Yes | No | 7.28 | NONE | N/A | 5/5 | 0.52 | | | 658.0 | Yes | No | * | * | * | Yes | | > | UR<1 |
| ALDRIN | UG/KG | No | No | 2.15 | 10 | Yes | 4/5 | 4.05 | | | 201.0 | Yes | No | * | * | * | Yes | | ~ | |
| BETA-BHC | UG/KG | No | Yes | 3.25 | NONE | N/A | 4/5 | 0.43 | | | 230.0 | Yes | * | * | * | * | Yes | | ~ | UR<1; <tdl< td=""></tdl<> |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 69 | NONE | N/A | 4/5 | 0.36 | | | 105.0 | No | * | * | * | * | No | | ~ | |

N/A=not applicable

*no criterion



TABLE 9-59A CRAIGHILL ANGLE EAST: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

| | | | | | | | | | M | agnitude :eedance | of (%) | | ITM Ev | aluation | USEF Supplem Evide | PA Region ental Weignce Screen | III ght-of- iing | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|-----|-------------------------------------|--|-----------------|------------|----------------------|-----------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|----------|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2 | NG/KG | Yes | No | 0.358 | N/A | N/A | 5/5 | 2.55 | | 104 | 40 | Yes | * | * | Yes | * | Yes | ~ | | |
| CHROMIUM | MG/KG | Yes | Yes | 0.18 | 0.1 | No | 2/5 | 0.64 | | 84 | | No | . * | No | • | * | No | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.48 | 0.1 | No | 5/5 | 1.13 | | 68.6 | | Yes | * | * | * | * | No | | ~ | |
| MANGANESE | MG/KG | No | Yes | 1.46 | 0.5 | No | 2/5 | 0.55 | | 121 | · | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.62 | 0.1 | No | 5/5 | 5.51 | | 121 | | No | * | No | * | * | No | | ~ | |

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N/A=not applicable

*no criterion

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^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

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| | | | | | | | | | r Es | Magnitu kceedan | ide of ce (%) | | ITM E | aluation | USEI Supplem Evide | PA Region ental Weignce Screer | III ght-of- ning | Propo | sed Action | |
|---------------------------------------|---------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|--------------------|-------------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units ^{(a} | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL | NG/KG | Yes | No | 0.242 | N/A | N/A | 5/5 | 1.38 | | | 93.4 | Yes | * | * | Yes | * | Yes | ~ | | |
| ALUMINUM | MG/KG | Yes | Yes | 45.82 | 1.0 | No | 5/5 | 5.16 | 41.6 | 50.7 | | No | * | * | * | * | No | | ~ | |
| ANTIMONY | MG/KG | Yes | Yes | 0.18 | 0.1 | No | 5/5 | 5.09 | 329.0 | 82.6 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| ARSENIC | MG/KG | Yes | Yes | 2.54 | 0.1 | No | 5/5 | 1.05 | | 19.5 | | No | * | No | No | * | Yes | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 2.24 | 0.1 | No | 5/5 | 5.69 | 401.0 | 2140.0 | 31.8 | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 22.40 | 10 | No | 5/5 | 3.39 | 27.5 | 86.5 | 84.1 | No | * | * | * | * | Yes | | - | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | No | Yes | 21.28 | 0.5 | No | 5/5 | 14.19 | 188.0 | 242.0 | 758.0 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| NICKEL | MG/KG | Yes | Yes | 2.36 | 0.1 | No | 5/5 | 3.71 | 224.0 | 687.0 | 63.9 | No | * | No | * | * | Yes | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.82 | 0.2 | No | 5/5 | 1.97 | 94.8 | 126.0 | 15.9 | No | * | * | No | * | No | | ~ | |
| ZINC | MG/KG | Yes | No | 15.04 | 2.0 | No | 5/5 | 0.92 | | 21.0 | | No | * | * | * | No | No | | ~ | |
| BENZO[B]FLUORANTHE | UG/KG | Yes | No | 7.74 | 20 | Yes | 5/5 | 3.76 | | | 180.0 | No | * | * | * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 9.20 | 20 | Yes | 5/5 | 1.05 | | | 23.0 | No | * | * | * | No | No | | ~ | |
| FLUORENE | UG/KG | Yes | Yes | 9.94 | 20 | Yes | 5/5 | 1.09 | | | 397.0 | No | * | * | * | * | No | | ~ | |
| NAPHTHALENE | UG/KG | Yes | No | 37.80 | 20 | No | 5/5 | 1.79 | | | 187.0 | No | * | * | * | * | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | Yes | 81.8-93 | NONE | N/A | 5/5 | 0.25 | | | 168, 114, 81.1 | No | * | * | * | * | * | | > | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 5.74 | NONE | N/A | 5/5 | 0.41 | | | 498.0 | Yes | No | * | * | * | Yes | | ~ | UR<1 |
| ALDRIN | UG/KG | No | No | 3.16 | 10 | Yes | 5/5 | 5.96 | - | | 344.0 | Yes | No | * | * | * | Yes | | ~ | |
| BETA-BHC | UG/KG | No | Yes | 8.40 | NONE | N/A | 5/5 | 1.11 | - | | 752.0 | Yes | * | * | * | * | Yes | ~ | | not detected in sediment |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 39 | NONE | N/A | 2/5 | 0.20 |] | | 15.2 | No | * | * | * | * | No | | . 🗸 | |

TABLE 9-59B CRAIGHILL ANGLE EAST: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

N/A=not applicable

*no criterion

| | | | | | | | | | Magnitu | de of Exa (%) | ceedance | | ITM Ev | aluation | Supplen Evide | iental Wei ince Scree | ight-of- ning | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|------|-------------------------------------|--|-----------------|------------|------------------|----------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------|-----------------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnif y? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2D | NG/KG | Yes | No | 0.263 | N/A | N/A | 5/5 | 1.87 | | 49.4 | | Yes | * | * | Yes | * | Yes | ~ | | |
| ALUMINUM | MG/KG | Yes | Yes | 20.88 | 1.0 | No | 5/5 | 0.43 | 142 | | | No | * | * | * | * | No | | 4 | |
| ANTIMONY | MG/KG | Yes | Yes | 0.11 | 0.1 | No | 4/5 | 1.89 | 129 | 123 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 3.00 | 200 | 200 | 200 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.32 | 0.1 | No | 5/5 | 1.12 | 113 | 222 | | No | * | No | * | * | No | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.52 | 0.1 | No | 5/5 | 1.16 | | 73.1 | | Yes | * | * | * | * | No | | * | |
| NICKEL | MG/KG | Yes | Yes | 0.61 | 0.1 | No | 5/5 | 5.36 | | 116 | | No | * | No | * | * | No | | · 🗸 | |
| ALDRIN | UG/KG | No | No | 3.35 | 10 | Yes | 4/5 | 0.26 | | 1170 | | Yes | No | * | * | * | Yes | | ~ | UR <i< td=""></i<> |
| BETA-BHC | UG/KG | No | Yes | 4.86 | NONE | N/A | 5/5 | 12.62 | 461 | 558 | 1160 | Yes | * | * | * | * | Yes | ~ | | |
| CHLORBENSIDE | UG/KG | No | No | 8.60 | 2 | No | 5/5 · | 0.88 | | 421 | | Yes | * | * | * | * | * | | ~ | UR <i< td=""></i<> |
| GAMMA-BHC | UG/KG | No | Yes | 1.80 | 10 | Yes | 4/5 | 0.68 | | 340 | | Yes | * | * | No | * | Yes | | • | |
| DI-N-BUTYL PHTHAL | UG/KG | Yes | No | 68 | 20 | No | 3/5 | 1.24 | 24.4 | | 26.7 | No | * | * | * | * | No | | ~ | |
| PENTACHLOROPHENO | UG/KG | No | No | 142 | 100 | No | 1/5 | 1.31 | 35.2 | 35.2 | 35.2 | No | * | * | * | No | Yes | | ~ | |

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TABLE 9-60A CRAIGHILL ANGLE WEST: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

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N/A=not applicable

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*no criterion

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^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

TABLE 9-60B CRAIGHILL ANGLE WEST: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | Detected in | | Mean Ticsue | | Tissue | Frequency | | 401 Ex | Magnitu sceedance 2 | de of ce (%) ≡ | Propensity | ITM Exceeds | valuation | USE Supplen Evide Exceeds | PA Region nental We ence Scree Exceeds | n III ight-of- ning | Ргоро | sed Action | |
|---------------------------------------|----------------------|-------------------|---------------------------|--------------|------|-------------------|-------------------------|-----------------|-----------|---------------------------|----------------------|-----------------------|---------------------------|--------------------------------|------------------------------------|---|---------------------------|-----------------|----------------------|--|
| Chemical Constituent | Units ^(a) | Bulk Sediment? | Detected in Elutriate? | Concentratio | TDL | Residue < TDL? | Dectection in Tissue | Uptake Ratio | Inside | Outside | Occa | to Biomagnify ? | USFDA Action Level? | Exceeds USEPA Tolerance? | USEPA Screenin g Value? | Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ARSENIC | MG/KG | Yes | Yes | 2.58 | 0.1 | No | 5/5 | 1.07 | | 21.4 | | No | * | No | No | * | Yes | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 0.47 | | 70.0 | 70.0 | No | * | * | * | * | No | | | |
| CHROMIUM | MG/KG | Yes | Yes | 0.84 | 0.1 | No | 5/5 | 2.13 | 87.5 | 738.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 219.80 | 10 | No | 5/5 | 3.35 | | 84.3 | 82.0 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 21.10 | 0.5 | No | 5/5 | 14.07 | 125.0 | 239.0 | 751.0 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| NICKEL | MG/KG | Yes | Yes | 1.06 | 0.1 | No | 5/5 | 1.67 | 136.0 | 254.0 | | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.62 | 0.2 | No | 5/5 | 1.50 | 75.2 | 72.1 | | No | * | * | No | * | Yes | | ~ | |
| ZINC | MG/KG | Yes | No | 18.54 | 2.0 | No | 5/5 | 1.13 | 19.1 | 49.2 | | No | * | * | * | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | No | No | 400 | 20 | No | 5/5 | 1.43 | | 1800.0 | 1800.0 | No | * | * | * | * | No | | ~ | |
| ANTHRACENE | UG/KG | Yes | No | 3.21 | 20 | Yes | 4/5 | 2.59 | | | 484.0 | No | * | * | * | * | No | | ~ | ······································ |
| BENZO[A]ANTHRACENE | UG/KG | Yes | No | 3.91 | 20 | Yes | 4/5 | 5.21 | 240.0 | | 421.0 | No | * | * | * | * | Yes | > | | |
| BENZO[B]FLUORANTHE | UG/KG | Yes | No | 6.84 | 20 | Yes | 5/5 | 3.32 | | - | 148.0 | No | * | * | * | * | Yes | ~ | | <tdl< td=""></tdl<> |
| CHRYSENE | UG/KG | Yes | No | 5.06 | 20 | Yes | 5/5 | 2.77 | - | 166.0 | | No | * | * | · * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 9.78 | 20 | Yes | 5/5 | 1.12 | | | 30.7 | No | * | * | * | No | No | | ~ | |
| FLUORENE | UG/KG | Yes | No | 11.00 | 20 | Yes | 5/5 | 1.21 | | | 450.0 | No | * | * | * | * | No | | ~ | |
| PHENANTHRENE | UG/KG | Yes | No | 3.08 | 20 | Yes | 5/5 | 1.22 | | | 35.1 | No | * | * | * | No | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 5.14 | 20 | Yes | 5/5 | 1.86 | | | 60.6 | No | * | * | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 83.5-93.9 | NONE | N/A | 5/5 | 1.50 | | | 174, 117, 82.8 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic<br="">threshold values</acute> |
| 4,4'-DDT | UG/KG | No | No | 4.12 | 10 | Yes | 5/5 | 1.67 | 180.0 | 587.0 | 587.0 | Yes | No | * | No | * | Yes | | ~ | |
| ALPHA-BHC | UG/KG | No | No | 1.94 | NONE | N/A | 3/5 | 5.40 | 440.0 | | | Yes | * | * | * | * | Yes | ~ | | not detected in sediment |
| BETA-BHC | UG/KG | No | Yes | 5.98 | NONE | N/A | 5/5 | 0.79 | | | 506.0 | Yes | * | * | * | * | Yes | | | 1/R<1 |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 80 | NONE | N/A | 2/5 | 0.42 | | | 140.0 | | * | * | * | * | No | | ~ | |

N/A=not applicable

*no criterion

| | | | | | | | | | M Ex | lagnitud ceedance | e of e (%) | | ITM E | valuation | USI Supple Evid | EPA Regio mental We lence Scree | n IIJ ight-of- ening | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|------------------------------|------|-------------------------------------|--|-----------------|------------|----------------------|---------------|-------------------------------------|--------------------------------------|--------------------------------|--|--|----------------------------|---------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentration | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screenin g Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as | Eliminate as COPC | Commonis |
| DIOXIN TEQ (ND=O) | NG/KG | Yes | No | 0.187 | N/A | N/A | 5/5 | 1.35 | - | 22.5 | | Yes | * | * | No | * | Yes | | | Comments |
| DIOXIN TEQ (ND=1/2D | NG/KG | Yes | No | 0.302 | N/A | N/A | 5/5 | 2.14 | | 71.5 | | Yes | * | * | Yes | * | Yes | ~ | | |
| CHROMIUM | MG/KG | Yes | Yes | 0.13 | 0.1 | No | 1/5 | 0.46 | | 32 | | No | * | No | * | * | Yes | | - | |
| COPPER | MG/KG | Yes | No | 1.44 | 0.1 | No | 5/5 | 1.10 | | 64 | | Yes | * | * | * | * | No | | ~ | |
| MANGANESE | MG/KG | Yes | Yes | 1.40 | 0.5 | No | 2/5 | 0.53 | | 112 | | No | * | * | * | * | No | | | |
| NICKEL | MG/KG | Yes | Yes | 0.52 | 0.1 | No | 5/5 | 4.62 | | 85.8 | | No | * | No | * | * | No | | | |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 28.10 | NONE | N/A | 5/5 | 1.23 | | | 402 | Yes | No | * | * | * | Vec | | | |
| 4,4'-DDD | UG/KG | No | No | 2.39 | 10 | Yes | 4/5 | 3.19 | | 219 | | Yes | No | * | * | * | No | | | |
| ALDRIN | UG/KG | No | No | 0.57 | 10 | Yes | 4/5 | 0.04 | | 115 | | Yes | No | * | * | * | Var | | | |
| CHLORBENSIDE | UG/KG | No | No | 15.40 | 2 | No | 5/5 | 1.57 | | 833 | 218 | Yes | * | * | * | * | * | ~ | • | no criterion; not detected in sediment |
| DELTA-BHC | UG/KG | No | No | 1.80 | NONE | N/A | 5/5 | 0.32 | 422 | 422 | | Yes | * | * | * | * | Yes | | ~ | UR<1; detected in the the laboratory method blank |
| GAMMA-BHC | UG/KG | No | Yes | 1.04 | 10 | Yes | 4/5 | 0.40 | | 154 | _ | Yes | * | * | No | * | No | | ~ | |
| PENTACHLOROPHENO | UG/KG | No | No | 144 | 100 | No | 1/5 | 1.33 | 37.1 | 37.1 | 37.1 | No | * | * | * | No | Yes | | ~ | |

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TABLE 9-61A CRAIGHILL ENTRANCE: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

*no criterion

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TABLE 9-61B CRAIGHILL ENTRANCE: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

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| | | | | | | | | | M Exc | agnitud eedance | e of : (%) | | ITM Ex | aluation | USE Supplem Evide | PA Region lental Wei ence Screer | lll ght-of- ling | Propose | d Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|--------------------|----------------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ALUMINUM | MG/KG | Yes | Yes | 38.34 | 1.0 | No | 5/5 | 4.32 | 18.5 | | | No | * | * | * | * | No | | ~ | |
| ANTIMONY | MG/KG | Yes | Yes | 0.13 | 0.1 | No | 4/5 | 3.74 | 216.0 | | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| ARSENIC | MG/KG | Yes | Yes | 2.8 | 0.1 | No | 5/5 | 1.16 | 17.2 | 31.8 | | No | * | No | Yes | * | Yes | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 1.61 | 0.1 | No | 5/5 | 4.08 | 260.0 | 1510.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 188.2 | 10 | No | 5/5 | 2.87 | | 57.8 | 55.8 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 16.66 | 0.5 | No | 5/5 | 11.11 | 125.0 | 168.0 | 572.0 | No | * | * | * | * | No | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| NICKEL | MG/KG | Yes | Yes | 1.72 | 0.1 | No | 5/5 | 2.70 | 136.0 | 473.0 | | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.74 | 0.2 | No | 5/5 | 1.77 | 75.2 | 103.0 | | No | * | + | No | * | Yes | | | |
| ZINC | MG/KG | Yes | No | 17.06 | 2.0 | No | 5/5 | 1.04 | 19.1 | 37.3 | | No | * | * | * | No | No | | | |
| BENZO[B]FLUORANTH | UG/KG | Yes | No | 6.4 | 20 | Yes | 5/5 | 3.11 | | | 132.0 | No | * | * | * | * | Yes | ~ | | <tdi< td=""></tdi<> |
| FLUORANTHENE | UG/KG | Yes | No | 11.12 | 20 | Yes | 5/5 | 1.27 | | | 48.7 | No | * | * | * | No | No | | | |
| FLUORENE | UG/KG | Yes | No | 10.18 | 20 | Yes | 5/5 | 1.12 | | | 409.0 | No | * | * | * | * | No | | | |
| NAPHTHALENE | UG/KG | Yes | No | 44 | 20 | No | 5/5 | 2.09 | | | 234.0 | No | * | * | * | * | No | | | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 87.1-97.7 | NONE | N/A | 5/5 | 0.38 | | | 185, 126, 90.3 | No | * | * | * | * | * | | | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| TOTAL PCBs (ND=0) | UĠ/KG | Yes | Yes | 8.67 | NONE | N/A | 5/5 | 0.62 | | 150.0 | 803.0 | Yes | No | + | * | * | Yes | | | UR<1 |
| ALDRIN | UG/KG | No | No | 1.89 | 10 | Yes | 4/5 | 3.56 | | | 165.0 | No | No | * | * | * | Ves | | | not detected in sediment |
| 2-METHYLPHENOL | UG/KG | No | No | 79 | 20 | No | 2/5 | 1.79 | 47.0 | 78.9 | 86.9 | No | * | * | * | * | No | | | |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 91 | NONE | N/A | 5/5 | 0.48 | 37.4 | | 172.0 | No | * | * | * | * | No | | | |
| 4-NITROPHENOL | UG/KG | No | No | 134 | NONE | N/A | 3/5 | 1.91 | 91.4 | 91.4 | 91.4 | No | * | * | * | * | No | | | |
| BENZOIC ACID | UG/KG | No | No | 4580 | 100 | No | 5/5 | 0.95 | | 50.2 | | No | * | * | * | * | No | | | |

.

N/A=not applicable

*no criterion

| | | | | | | | | | M Exc | agnitude ceedance (| of (%) | | ITM Ev | aluation | USE Supplen Evide | PA Region tental Wei | III ght-of- ning | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|-----|-------------------------------------|--|-----------------|------------|------------------------|-----------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|----------|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ALUMINUM | MG/KG | Yes | Yes | 12.2 | 1.0 | No | 4/4 | 0.25 | 41.7 | 42.9 | | No | * | * | * | * | No | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 4/4 | 2.75 | 175 | 175 | 175 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.49 | 0.1 | No | 4/4 | 1.70 | | 387 | | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | No | 1.58 | 0.1 | No | 4/4 | 1.21 | | 79.4 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.87 | 0.1 | No | 4/4 | 7.70 | 77.7 | 209 | | No | * | No | * | * | No | | ~ | |
| ALDRIN | UG/KG | No | No | 2.87 | 10 | Yes | 3/4 | 0.23 | | 482 | | Yes | No | * | * | * | Yes | | ~ | |
| CHLORBENSIDE | UG/KG | No | No | 8.35 | 2 | No | 4/4 | 0.85 | | 406 | | Yes | * | * | * | * | * | | ~ | UR<1 |
| DACTHAL | UG/KG | No | No | 40.00 | 2 | No | 4/4 | 1.30 | 58.7 | 300 | | Yes | * | * | * | * | No | | ~ | |
| ENDRIN | UG/KG | No | No | 1.27 | 10 | Yes | 3/4 | 0.17 | 160 | 160 | | Yes | * | * | No | * | * | | ~ | |
| GAMMA-BHC | UG/KG | No | Yes | 1.58 | 10 | Yes | 4/4 | 0.60 | | 287 | | Yes | * | + | No | * | No | | ~ | |

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TABLE 9-62A CRAIGHILL UPPER RANGE: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

.

N/A=not applicable

.

*no criterion

.

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

TABLE 9-62B CRAIGHILL UPPER RANGE: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | | | | | | | | M Ex | lagnitud ceedanc | le of e (%) | | ITM Ev | aluation | USE Suppler Evid | CPA Regio nental We ence Scree | n III eight-of- ening | Prop | osed Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|---------------------|----------------------|-------------------------------------|--------------------------------------|------------------------------------|---|--|-----------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance ? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL) | NG/KG | Yes | No | 0.266 | N/A | N/A | 5/5 | 1.52 | | | 112.0 | Yes | * | * | Yes | * | Yes | ~ | | |
| ARSENIC | MG/KG | Yes | Yes | 2.48 | 0.1 | No | 5/5 | 1.02 | | 16.7 | | No | * | No | No | * | Yes | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 0.47 | | 70.0 | 70.0 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.82 | 0.1 | No | 5/5 | 2.08 | 83.0 | 738.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | No | 173.00 | 10 | No | 5/5 | 2.63 | | 45.1 | 43.2 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 11.36 | 0.5 | No | 5/5 | 7.57 | 53.7 | 82.5 | 358.0 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 1.14 | 0.1 | No | 5/5 | 1.79 | 56.0 | 279.0 | | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.62 | 0.2 | No | 5/5 | 1.50 | 48.6 | 72.1 | | No | * | * | No | * | Yes | | ~ | |
| ZINC | MG/KG | Yes | No | 16.98 | 2.0 | No | 5/5 | 1.04 | 18.5 | 36.7 | | No | * | * | * | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | No | No | 494.00 | 20 | No | 5/5 | 1.77 | | 2250.0 | 2250.0 | Yes | * | * | * | * | Yes | | ~ | not detected in sediment; results may be biased high |
| BENZ[A]ANTHRACENE | UG/KG | Yes | No | 3.44 | 20 | Yes | 5/5 | 4.59 | 199.0 | | 359.0 | Yes | * | * | * | * | No | | * | |
| BENZO[A]PYRENE | UG/KG | Yes | No | 0.96 | 20 | Yes | 4/5 | 2.34 | | | 134.0 | Yes | * | * | * | Yes | Yes | ~ | | |
| BENZO[B]FLUORANTHEN | UG/KG | Yes | No | 7.92 | 20 | Yes | 5/5 | 3.84 | | | 187.0 | Yes | * | * | * | * | Yes | ~ | | |
| CHRYSENE | UG/KG | Yes | No | 3.80 | 20 | Yes | 5/5 | 2.08 | | 100.0 | 18.0 | Yes | * | * | * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 16.52 | 20 | Yes | 5/5 | 1.89 | | | 121.0 | Yes | * · | * | * | No | No | | ~ | · · · · · · · · · · · · · · · · · · · |
| FLUORENE | UG/KG | Yes | No | 9.40 | 20 | Yes | 5/5 | 1.04 | | | 370.0 | Yes | * | * | * | * | No | | ~ | |
| NAPHTHALENE | UG/KG | Yes | No | 32.80 | 20 | No | 5/5 | 1.56 | | | 149.0 | Yes | * | * | * | * | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 6.06 | 20 | Yes | 5/5 | 2.19 | | | 89.4 | Yes | * | * | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 89.5-99.4 | NONE | N/A | 5/5 | 1.81 | - | | 193, 131, 93.5 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| 4,4'-DDD | UG/KG | No | No | 1.76 | 10 | Yes | 5/5 | 0.40 | 83.3 | 69.2 | 135.0 | Yes | No | * | * | * | No | | × . | |
| 4,4'-DDT | UG/KG | No | No | 3.32 | 10 | Yes | 5/5 | 1.35 | 126.0 | 453.0 | 453.0 | Yes | No | * | No | * | Yes | | ~ | |
| GAMMA-BHC | UG/KG | No | Yes | 1.48 | 10 | Yes | 4/5 | 0.40 | 261.0 | 261.0 | | Yes | * | * | No | No | No | | ~ | |
| HEPTACHLOR | UG/KG | No | No | 3.88 | 10 | Yes | 5/5 | 0.45 | | 1040.0 | | Yes | No | * | * | * | Yes | | · · | detected in laboratory method blank |
| 2-METHYLPHENOL | UG/KG | No | No | 52 | 20 | No | 2/5 | 1.19 | | | 24.0 | No | * | * | * | * | No | | ~ | |

N/A=not applicable

*no criterion

| TABLE 9-63A CU' | TOFF A | NGLE: | INTEG | RATED EV | ⁷ ALU. | ATIC | ON OF I | 310A(| CCUN | 1ULA | ATION le of | N IN Nereis | s virens | (SAND V | VORM USE Suppl | I) CPA Regio emental V | on 111 Weight- | | | |
|----------------------|-----------|----------------------------------|------------------------------|------------------------------|-------------------|--|--|-----------------|------------|-------------|----------------|------------------------------|--------------------------------------|--------------------------------|---|---|-------------------|-------------------|-----------|--|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentration | TDL | Mean Tissue Residu e <tdl ?</tdl | Frequenc y of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | of-Evi Exceeds USEPA Screenin g Value? | dence Sci Exceeds Residue- Effect Data? | Exceeds | Propos Keep as | Eliminate | |
| DIOXIN TEQ (ND=1/2DL | NG/KG | Yes | Yes | 0.258 | N/A | N/A | 5/5 | 1.84 | | 46.7 | | Yes | * | * | Yes | * | Yes | <u> </u> | | |
| ALUMINUM | MG/KG | Yes | Yes | 32.92 | 1.0 | No | 5/5 | 0.68 | 282 | 285 | | No | * | * | * | * | No | · | | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 3.40 | 240 | 240 | 240 | No | * | * | * | * | No | | | |
| CHROMIUM | MG/KG | Yes | No | 0.50 | 0.1 | No | 5/5 | 1.74 | 230 | 398 | | No | * | No | * | * | Yes | | | |
| COPPER | MG/KG | Yes | Yes | 1.50 | 0.1 | No | 5/5 | 1.15 | | 70.8 | | Yes | * | * | * | * | No | | ~ | 1 |
| MANGANESE | MG/KG | Yes | Yes | 8.92 | 0.5 | No | 2/5 | 3.37 | 792 | 1250 | | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.94 | 0.1 | No | 5/5 | 8.28 | 91 | 233 | | No | * | No | * | * | No | | ~ | |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 38.4 | NONE | N/A | 5/5 | 1.68 | | | 586 | Yes | No | * | * | * | Yes | | ~ | · · · · · · · · · · · · · · · · · · · |
| 4,4'-DDD | UG/KG | No | No | 2.75 | 10 | Yes | 4/5 | 3.67 | | 267 | | Yes | No | * | * | * | No | | ~ | |
| CHLORBENSIDE | UG/KG | No . | No | 14.80 | 2 | No | 5/5 | 1.51 | | 797 | 206 | Yes | * | * | * | * | * | ~ | | no criterion; not detected in sediment |
| DACTHAL | UG/KG | No | No | 76.00 | 2 | No | 5/5 | 2.47 | 201 | 660 | 201 | Yes | * | * | * | * | No | | ~ | |
| GAMMA-BHC | UG/KG | No | Yes | 2.00 | 10 | Yes | 5/5 | 0.76 | 45.7 | 388 | | Yes | * | * | No | * | No | | ~ | |

N/A=not applicable

*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

TABLE 9-63B CUTOFF ANGLE: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | | | | | | | | E | Magnitu xceedanc | de of e (%) | | ITM E | valuation | USI Supple Evid | EPA Regio mental We ence Scree | n 111 eight-of- ening | Propos | sed Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|---------------------|----------------------|-------------------------------------|--------------------------------------|--------------------------------|--|--|-----------------------------|-----------------|----------------------|--|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screenin g Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ARSENIC | MG/KG | Yes | Yes | 2.78 | 0.1 | No | 5/5 | 1.15 | | 30.8 | | No | * | No | Yes | * | Yes | | ~ | |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 0.47 | | 70.0 | 70.0 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | No | 0.57 | 0.1 | No | 5/5 | 1.44 | | 466.0 | | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 187.40 | 10 | No | 5/5 | 2.85 | | 57.1 | 55.1 | No | * | * | * | * | Yes | | > | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 16.32 | 0.5 | No | 5/5 | 10.88 | 121.0 | 162.0 | 558.0 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.87 | 0.1 | No | 5/5 | 1.36 | | 189.0 | | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | Yes | 0.57 | 0.2 | No | 5/5 | 1.38 | 36.2 | 57.8 | | No | * | No | * | * | Yes | | ~ | |
| ZINC | MG/KG | Yes | No | 19.62 | 2.0 | No | 5/5 | 1.20 | | 57.9 | | No | * | * | + | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | No | No | 630.20 | 20 | No | 4/5 | 2.26 | | 2900.0 | 2900.0 | No | * | * | * | * | No | | ~ | |
| ANTHRACENE | UG/KG | Yes | No | 3.86 | 20 | Yes | 5/5 | 3.11 | | 40.4 | 602.0 | No | * | * | * | * | No | | ~ | |
| BENZ[A]ANTHRACEN | UG/KG | Yes | No | 3.52 | 20 | Yes | 5/5 | 4.69 | 206.0 | | 369.0 | No | * | * | * | * | Yes | ~ | | |
| BENZO[A]PYRENE | UG/KG | Yes | No | 0.95 | 20 | Yes | 4/5 | 2.32 | | | 132.0 | No | * | * | * | Yes | Yes | ~ | | |
| BENZO[B]FLUORANTI | UG/KG | Yes | No | 7.44 | 20 | Yes | 5/5 | 3.61 | | | 170.0 | No | * | * | * | * | Yes | ~ | | |
| CHRYSENE | UG/KG | Yes | No | 3.94 | 20 | Yes | 5/5 | 2.16 | | 107.0 | 22.4 | No | * | * | * | * | No | | ~ | |
| FLUORANTHENE | UG/KG | Yes | No | 12.80 | 20 | Yes | 5/5 | 1.46 | | 69.0 | 71.1 | No | * | * | * | No . | No | | ~ | · · · · · · · · · · · · · · · · · · · |
| FLUORENE | UG/KG | Yes | No | 9.80 | 20 | Yes | 5/5 | 1.08 | | | 390.0 | No | * | * | * | * | No | | * | |
| NAPHTHALENE | UG/KG | Yes | No | 35.40 | 20 | No | 5/5 | 1.68 | | | 169.0 | No | * | * | * | * | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 6.14 | 20 | Yes | 5/5 | 2.22 | - | | 91.9 | No | * | * | | No | No | | ~ | ····· |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 86.8-98.5 | NONE | N/A | 5/5 | 2.21 | | | 184, 126, 91.8 | No | * | * | * | .* | * | | ~ | CBR <acute and="" chronic<br="">threshold values</acute> |
| 4,4'-DDD | UG/KG | No | No | 1.99 | 10 | Yes | 4/5 | 0.45 | 107.0 | 91.3 | 165.0 | Yes | No | * | * | * | No | | ~ | |
| 4,4'-DDT . | UG/KG | No | No | 3.02 | 10 | Yes | 5/5 | 1.23 | 105.0 | 403.0 | 403.0 | Yes | No | * | No | * | Yes | | ~ | |
| GAMMA-BHC | UG/KG | No | Yes | 2.00 | 10 | Yes | 5/5 | 0.54 | 388.0 | 388.0 | 176.0 | Yes | * | * | No | No | No | | ~ | |
| HEPTACHLOR | UG/KG | No | Yes | 4.20 | 10 | Yes | 5/5 | 0.48 | 63.0 | 1140.0 | | Yes | No | * | * | * | Yes | | ~ | UR<1 |
| 2-METHYLPHENOL | UG/KG | No | No | 51 | 20 | No | 2/5 | 1.15 | | | 20.2 | No | * | * | * | * | No | | ~ | |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 45 | NONE | N/A | 1/5 | 0.24 | [| [| 35.5 | No | * | * | * | * | No | | ~ | |

.

N/A=not applicable

*no criterion

.

| | | | | | | | | | Magni | tude of Ex (%) | ceedance | | ITM E | aluation | USE Supplen Evide | PA Region lental Wei ence Scree | III ght-of- ning | Propos | ed Action | |
|----------------------|-----------|----------------------------------|---------------------------|----------------------------------|------|-------------------------------------|--|-----------------|------------|-------------------|----------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|--|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2I | NG/KG | Yes | No | 0.301 | N/A | N/A | 5/5 | 2.14 | | 70.8 | | Yes | * | * | Yes | * | Yes | ~ | | |
| ANTIMONY | MG/KG | Yes | Yes | 0.16 | 0.1 | No | 4/4 | 2.78 | 238 | 228 | | No | * | * | * | * | Yes | | - | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.07 | 0.1 | Yes | 4/4 | 7.00 | 600 | 600 | 600 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.47 | 0.1 | No | 4/4 | 1.62 | 208 | 365 | | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | No | 1.55 | 0.1 | No | 4/4 | 1.19 | | 76.8 | | Yes | * | * | * | * | No | | ~ | |
| MANGANESE | MG/KG | | | 1.25 | 0.5 | No | | 1.17 | | 89.4 | | No | * | * * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.73 | 0.1 | No | 4/4 | 6.44 | | 159 | | No | * | No | * | * | No | | ~ | |
| SILVER | MG/KG | | | 0.15 | 0.1 | No | | 0.47 | | | 281 | No | * | * | * | * | No | | ~ | |
| ВЕТА-ВНС | UG/KG | No | Yes | 1.22 | NONE | N/A | 2/4 | 3.17 | | | 217 | Yes | * | * | * | * | Yes | > | | detected in 2 of 4 replicates; not detected in sediment |
| DACTHAL | UG/KG | No | No | 37.50 | 2 | No | 4/4 | 1.22 | | 275 | | Yes | * | · * | * | * | No | | > | |
| DELTA-BHC | UG/KG | No | No | 2.63 | NONE | N/A | 4/4 | 0.47 | 66 1 | 661 | | Yes | * | * | * | * | Yes | | > | UR <i< td=""></i<> |
| ENDOSULFAN II | UG/KG | No | No | 24.25 | 10 | No | 4/4 | 4.35 | 207 | | | Yes | * | * | No | * | No | | > | ······································ |
| HEPTACHLOR | UG/KG | No | Yes | 5.27 | 10 | Yes | 4/4 | 15.51 | 350 | 205 | 1450 | Yes | No | * | * | * | Yes | | ~ | |
| HEPTACHLOR EPOXII | UG/KG | No | Yes | 2.45 | 10 | Yes | 4/4 | 6.71 | 237 | 412 | | Yes | No | * | * | * | Yes | | ~ | |
| 2-METHYLPHENOL | UG/KG | No | No | 214 | 20 | No | 5/5 | 2.37 | 184 | 153 | 130 | No | * | * | * | * | No | | ~ | |

TABLE 9-64A SWAN POINT: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

| | 1 | | <u> </u> | | | γ | 1 | T | 1 | | | | | | <u></u> | <u> </u> | | | | · · · · · · · · · · · · · · · · · · · |
|-------------------------|----------------------|-------------------|---------------------------|------------------|----------|----------------|-------------------------|-----------------|--------------|--------------|---------------|-------------|-------------|---------------------|-----------|------------|------------------|---------|------------|--|
| | | | | | | | | | | Ananitu | la af | | | | USE | PA Region | n III iaht of | | | |
| | | | | 1 | | | | | Ex | ceedanc | e (%) | | ITM E | valuation | Evide | ence Scree | ning | Propo | sed Action | |
| | | Detected in | | Mean Tissu | e | Tissue- | Frequency of | | - 11 - | t de | - | Propensity | Exceeds | Excoods | Exceeds | Exceeds | | 11000 | | - |
| Chemical Constituent | Units ^(a) | Bulk Sediment? | Detected in Elutriate? | Concentrati n | 0 TDL | Residue < TDL? | Dectection in Tissue | Uptake Ratio | Insie 104 | Outsi 104 | Ocea | Biomagnify | Action | USEPA Tolerance? | Screening | Effect | Exceeds | Keep as | Eliminate | |
| ANTIMONY | MG/KG | Yes | Yes | 0.10 | 0.1 | EOUAL | 5/5 | 2.91 | 146.0 | | | No | | tokrance. | Value: | Data: | RBC? | COPC | ascore | UCLM exceeds 1/10 RBC, |
| ARSENIC | MG/KG | Yes | Yes | 2.54 | 0.1 | No | 5/5 | 1.05 | | 19.5 | | No | * | No | | | Yes | | | but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 0.42 | | 50.0 | 50.0 | No | * | * | * | | Yes | | | |
| CHROMIUM | MG/KG | Yes | Yes | 0.60 | 0.1 | No | 5/5 | 1.52 | | 498.0 | | No | * | No | * | * | N0 Vac | | | |
| IRON | MG/KG | Yes | Yes | 161.60 | 10 | No | 5/5 | 2.46 | | 25.5 | 1 22.0 | No | | | | | 105 | | | UCLM exceeds 1/10 RBC, |
| MANGANESE | MG/KG | Yes | Yes | 11.34 | 0.5 | No | 5/5 | 7.56 | 53.5 | 82.2 | 357.0 | No | | * | * | * | Yes | | ~ | but does not exceed RBC |
| MERCURY | MG/KG | Yes | Yes | 0.11 | 0.01 | No | 5/5 | 0.93 | | 82.2 | 143.0 | Yes | No | * | * | * | No | | | |
| NICKEL | MG/KG | Yes | Yes | 0.87 | 0.1 | No | 5/5 | 1.36 | | 189.0 | | No | * | No | N0 * | | Yes | | | |
| SELENIUM | MG/KG | Yes | No | 0.67 | 0.2 | No | 5/5 | 1.61 | 59.0 | 84.3 | | No | * | * | No | * | N0 Var | | | |
| ANTHRACENE | UG/KG | Yes | No | 3.11 | 20 | Yes | 4/5 | 2.51 | | | 465.0 | No | * | | | | 163 | | | detected in laboratory |
| BENZO[B]FLUORANTHE | UG/KG | Yes | No | 10.86 | 20 | Vaa | | | | | 405.0 | | | | | * | No | | ~ | detected in laboratory |
| ELLIOR ANTHENE | | | | 10.00 | | res | 5/5 | 5.27 | | | 293.0 | No | * | * | * | * | Yes | | ~ | method blank detected in laboratory |
| LOOKANTIENE | 00/KG | Yes | No | 10.10 | 20 | Yes | 5/5 | 1.15 | | | 35.0 | No | * | * | * | No | No | | | method blank |
| FLUORENE | UG/KG | Yes | No | 8.22 | 20 | Yes | 4/5 | 0.91 | | | 311.0 | No | * | * | * | * | No | | ~ | detected in laboratory method blank |
| PYRENE | UG/KG | Yes | No | 6.62 | 20 | Yes | 5/5 | 2.39 | | | 107:0 | No | * | * | * | No | No | | ~ | detected in laboratory method blank |
| TOTAL PAHs | | | | | | | | | | | 138, 95.9, | | | | | | | | | CPP coouto and share is |
| (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 72.5-88 | NONE | N/A | 5/5 | 2.06 | | | 71.3 | No | * | * | * | * | * | | ~ | threshold values |
| TOTAL PCBs (ND=0) | UG/KG | Yes | Yes | 11.9 | NONE | N/A | 5/5 | 0.85 | | 244.0 | 1140.0 | Yes | No | * | * | * | Yes | | ~ | UR <i< td=""></i<> |
| 4,4'-DDT | UG/KG | No | No | 3.08 | 10 | Yes | 5/5 | 1.25 | 110.0 | 413.0 | 413.0 | Yes | No | * | No | * | Yes | _ | < | |
| ENDOSULFAN II | UG/KG | No | No | 3.25 | 10 | Yes | 4/5 | 3.13 | 302.0 | 655.0 | 655.0 | Yes | * | * | No | * | No | | ~ | |
| HEPTACHLOR | UG/KG | No | Yes | 3.73 | 10 | Yes | 4/5 | 0.43 | | 996.0 | | Yes | No | * | * | * | Yes | | | detected in laboratory method blank |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 734 | NONE | N/A | 5/5 | 3.88 | 1010.0 | 1410.0 | 2090.0 | No | * | * | * | * | No | | | |

TABLE 9-64B SWAN POINT: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

.

N/A=not applicable

*no criterion

.

| | | | | | | | | | Magnitu | de of Exc (%) | eedance | | ITM E | aluation | USE Suppler Evid | PA Regio nental We ence Scree | n III eight-of- ening | Propos | ed Action | |
|---------------------------------------|-----------|----------------------------------|---------------------------|----------------------------------|------|-------------------------------------|--|-----------------|------------|------------------|---------|-------------------------------------|--------------------------------------|--------------------------------|--|--|-----------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units (a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Mean Tissue Residue < TDL? | Frequency of Detection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screenin g Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ALUMINUM | MG/KG | Yes | Yes | 10.96 | 1.0 | No | 5/5 | 0.23 | 27.3 | 28.3 | | No | * | * | * | * | No | | ~ | |
| ANTIMONY | MG/KG | Yes | Yes | 0.15 | 0.1 | No | 5/5 | 2.65 · | 221 | 213 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 3.00 | 200 | 200 | 200 | NO | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | No | 0.44 | 0.1 | No | 5/5 | 1.53 | 190 | 338 | | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.64 | 0.1 | No | 5/5 | 1.26 | | 86.8 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.85 | 0.1 | No | 5/5 | 7.54 | 73.9 | 203 | | No | * | No | * | * | No | | | |
| SILVER | MG/KG | Yes | No | 0.10 | 0.1 | EQUAL | 4/5 | 0.75 | | | 145 | No | * | * | * | * | Yes | | | 11/R<1 |
| FLUORANTHENE | UG/KG | Yes | No | 5.80 | 20 | Yes | 5/5 | 0.75 | | | 194 | No | * | * | * | * | No | | | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 8.95-37 | NONE | N/A | 5/5 | 0.65 | 13.6-198 | | | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" effect<br="">threshold</acute> |
| 4,4'-DDD | UG/KG | No | No | 13.74 | 10 | No | 5/5 | 2.05 | 450 | 1730 | 570 | Yes | No | * | * | * | Yes | | ~ | |
| 4,4'-DDT | UG/KG | No | No | 2.20 | 10 | Yes | 3/5 | 18.32 | 267 | | - | Yes | No | * | * | * | Yes | | > | |
| DACTHAL | UG/KG | No | No | 44.00 | 2 | No | 5/5 | 1.06 | 74.5 | 340 | 74.3 | Yes | * | * | * | * | No | | > | |
| ENDOSULFAN II | UG/KG | No | No | 9.88 | 20 | Yes | 5/5 | 1.77 | 25 | | | Yes | * | * | No | * | No | | ~ | |
| HEPTACHLOR EPOXIC | UG/KG | No | Yes | 2.45 | 20 | Yes | 4/5 | 6.72 | 238 | 412 | | Yes | No | * | * | * | Yes | | ~ | |

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TABLE 9-65A TOLCHESTER NORTH: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

.

N/A=not applicable

*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

| | | | | | | | | | N Ex | Aagnitu ceedan | de of ce (%) | | ITM E | aluation | USEI Supplem Evide | PA Region lental Wei nce Screer | III ght-of- ning | Propo | osed Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|--|-----------------|------------|-------------------|----------------------|-------------------------------------|--------------------------------------|--------------------------------|---|--|------------------------|-----------------|----------------------|---|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detected in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Occan | Propensity to Biomagnify ? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL | NG/KG | Yes | No | 0.218 | N/A | N/A | 5/5 | 1.24 | | | 74.1 | Yes | * | * | No | * | Yes | | ~ | <epa screening="" td="" value<=""></epa> |
| ANTIMONY | MG/KG | Yes | Yes | 0.13 | 0.1 | No | 5/5 | 3.66 | 208.0 |) | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.04 | 0.1 | Yes | 5/5 | 0.50 | - | 90.0 | 80.0 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | No | 0.46 | 0.1 | No | 5/5 | 1.16 | | 300.0 | ' | No | * | No | * | * | Yes | | ~ | |
| IRON | MG/KG | Yes | Yes | 161.20 | 10 | No | 5/5 | 2.45 | | 35.2 | 33.4 | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| MANGANESE | MG/KG | Yes | Yes | 10.04 | 0.5 | No | 5/5 | 6.69 | 35.9 | 61.3 | 305.0 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.86 | 0.1 | No | 5/5 | 1.35 | | 187.0 | | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | No . | 0.73 | 0.2 | No | 5/5 | 1.76 | 74.8 | 102.0 | - | No | * | * | No | * | Yes | | ~ | |
| ZINC | MG/KG | Yes | No | 15.22 | 2.0 | No | 5/5 | 0.93 | + | 22.5 | ' | No | * | * | * | No | No | | ~ | |
| ACENAPHTHYLENE | UG/KG | Yes | No | 784.00 | 20 | No | 5/5 | 2.81 | 58.3 | 3630.0 | 3630.0 | No | * | * | * | * | No | | ~ | |
| BENZO[A]ANTHRACEN | UG/KG | Yes | No | 3.24 | 20 | Yes | 5/5 | 4.32 | 182.0 | | 332.0 | No | * | * | * | * | Yes | ~ | | |
| BENZO[B]FLUORANTH | UG/KG | Yes | No | 11.40 | 20 | Yes | 5/5 | 5.53 | | | 313.0 | No | * | * | * | * | Yes | ~ | | <tdl< td=""></tdl<> |
| CHRYSENE | UG/KG | Yes | No | 3.70 | 20 | Yes | 5/5 | 2.02 | | 94.7 | 14.9 | No | * | * | * | * | No | | ~ | |
| FLUORENE | UG/KG | Yes | No | 10.50 | 20 | Yes | 5/5 | 1.16 | | | 425.0 | No | * | * | * | * | No | | ~ | |
| NAPHTHALENE | UG/KG | Yes | No | 39.60 | 20 | No | 5/5 | 1.88 | | | 201.0 | No | * | * | * | * | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 83-97.2 | NONE | N/A | 5/5 | 2.69 | | · | 172, 120, 89.2 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| 2-METHYLPHENOL | UG/KG | No | No | 81 | 20 | No | 5/5 | 1.83 | 50.5 | 83.2 | 91.4 | No | * | * | * | * | No | | ~ | |
| 3,4-METHYLPHENOL | UG/KG | Yes | No | 36 | NONE | N/A | 1/5 | 0.19 | | | 7.5 | No | * | * | * | * | No | | ~ | |
| BENZOIC ACID | UG/KG | No | No | 5980 | 100 | No | 5/5 | 1.24 | 35.6 | 96.1 | | No | * | * | * | * | No | | ~ | |
| BENZYL ALCOHOL | UG/KG | No | No | 128 | 100 | No | 5/5 | 0.96 | | | 103.0 | No | * | * | * | * | No | | ~ | |

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TABLE 9-65B TOLCHESTER NORTH: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

N/A=not applicable

*no criterion

| Chemical Constituent | Units (a) | Detected ir Bulk Sediment? | Detected in Flutriate? | Mean Tissue Concentratio | TDI | Mean Tissue Residue | Frequency of Detection | Uptake | nside 104 Wa | agnitude eedance 104 | of (%) Occan | Propensity to | ITM E Exceeds USFDA Action | Exceeds USEPA | USE Suppler Evid Exceeds USEPA Screenin | CPA Regio mental We ence Scree Exceeds Residue- Effect | n IIJ ight-of- ening Exceeds | Propos Keep as | ed Action Eliminate | |
|----------------------|-----------|----------------------------------|------------------------------|-----------------------------|-----|---------------------------|---------------------------|--------|--------------------|----------------------------|--------------------|---------------|-------------------------------------|------------------|--|---|---------------------------------------|-------------------|------------------------|---|
| | MG/KG | Ves | | | | < IDL: | in rissue | Ratio | | | | Biomagnify? | Level? | Tolerance? | g Value? | Data? | RBC? | COPC | as COPC | Comments |
| | MO/KO | res | Yes | 12.34 | 1.0 | No | 5/5 | 0.25 | 43.3 | 44.5 | | No | * | * | * | * | No | | ~ | |
| ANTIMONY | MG/KG | Yes | Yes | 0.16 | 0.1 | No | 5/5 | 2.75 | 234 | 225 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.03 | 0.1 | Yes | 5/5 | 3.00 | 200 | 200 | 200 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.79 | 0.1 | No | 5/5 | 2.76 | 424 | 692 | - | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.54 | 0.1 | No | 5/5 | 1.18 | | 75.4 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.82 | 0.1 | No | 5/5 | 7.25 | 67.4 | 191 | | No | * | No | * | * | No | | ~ | |
| ZINC | MG/KG | Yes | No | 41.66 | 2.0 | No | 4/5 | 3.29 | | 218 | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| 4,4'-DDD | UG/KG | Yes | No | 10.32 | 10 | No | 5/5 | 13.76 | 313 | 1280 | 403 | Yes | No | * | * | * | Yes | | ~ | |
| DACTHAL | UG/KG | No | No | 41.20 | 2 | No | 5/5 | 1.34 | 63.4 | 312 | 63.2 | Yes | * | * | * | * | No | | ~ | |
| PHENOL | UG/KG | Yes | No | 54 | 20 | No | 1/5 | I.09 | 9.72 | 9.09 | 10.2 | No | * | * | * | * | No | | ~ | |

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TABLE 9-66A TOLCHESTER SOUTH: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

.

N/A=not applicable

*no criterion

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^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)

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TABLE 9-66B TOLCHESTER SOUTH: INTEGRATED EVALUATION OF BIOACCUMULATION IN Macoma nasuta (BLUNT-NOSE CLAM)

| | | | | | | | | | N Ex | 1agnitud ceedanc | de of e (%) | | ITM | Evaluation | USE Suppler Evide | PA Region I nental Weig ence Screeni | II ht-of- ng | Prop | osed Action | |
|---------------------------------------|----------------------|----------------------------------|---------------------------|----------------------------------|------|------------------------------|---|-----------------|------------|---------------------|---------------------|------------------------------|--------------------------------------|--------------------------------|---|--|--------------------|-----------------|----------------------|---|
| Chemical Constituent | Units ^(a) | Detected in Bulk Sediment? | Detccted in Elutriate? | Mean Tissue Concentratio n | TDL | Tissue- Residue < TDL? | Frequency of Dectection in Tissue | Uptake Ratio | Inside 104 | Outside 104 | Ocean | Propensity to Biomagnify? | Exceeds USFDA Action Level? | Exceeds USEPA Tolerance? | Exceeds USEPA Screening Value? | Exceeds Residue- Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| ANTIMONY | MG/KG | Yes | Yes | 0.16 | 0.1 | No | 5/5 | 4.57 | 286.0 | | | No | * | * | * | * | Yes | | ~ | UCLM exceeds 1/10 RBC, but does not exceed RBC |
| BERYLLIUM | MG/KG | Yes | No | 0.04 | 0.1 | Yes | 5/5 | 0.53 | | 90.0 | 90.0 | No | * | * | * | * | No | | | |
| CHROMIUM | MG/KG | Yes | Yes | 0.40 | 0.1 | No | 5/5 | 1.02 | | 300.0 | | No | * | No | * | * | Yes | | - | |
| MANGANESE | MG/KG | Yes | Yes | 8.98 | 0.5 | No | 5/5 | 5.99 | | 44.3 | 262.0 | No | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.72 | 0.1 | No | 5/5 | 1.13 | | 140.0 | - | No | * | No | * | * | No | | ~ | |
| SELENIUM | MG/KG | Yes | No | 0.69 | 0.2 | No | 5/5 | 1.65 | 63.3 | 89.2 | | No | + | * | No | * | Yes | | ~ | |
| SILVER | MG/KG | Yes | No | 0.18 | 0.1 | No | 4/5 | 3.71 | | | 230.0 | No | * | * | * | No | No | | ~ | |
| BENZO[A]ANTHRACENE | UG/KG | Yes | No | 3.14 | 20 | Yes | 5/5 | 4.19 | 173.0 | - | 319.0 | No | * | * | * | * | Yes | ~ | | <tdl< td=""></tdl<> |
| BENZO[B]FLUORANTHENE | UG/KG | Yes | No | 10.22 | 20 | Yes | 5/5 | 4.96 | | | 270.0 | No | * | * | * | * | Yes | ~ | | <tdl< td=""></tdl<> |
| CHRYSENE | UG/KG | Yes | No | 3.74 | 20 | Yes | 5/5 | 2.05 | | 96.8 | 16.1 | No | * | * | * | * | No | | ~ | + |
| FLUORANTHENE | UG/KG | Yes | No | 16.86 | 20 | Yes | 5/5 | 1.92 | | - | 125.0 | No | * | * | * | No | No | | ~ | |
| FLUORENE | UG/KG | Yes | No | 10.10 | 20 | Yes | 5/5 | 1.11 | | | 405.0 | No | * | * | * | * | No | | ~ | |
| NAPHTHALENE | UG/KG | Yes | No | 41.40 | 20 | No | 5/5 | 1.96 | | | 215.0 | No | * | + | * | * | No | | ~ | |
| PYRENE | UG/KG | Yes | No | 5.12 | 20 | Yes | 5/5 | 1.85 | | | 60.0 | No | * | + | * | No | No | | ~ | |
| TOTAL PAHs (ND=0, ND=1/2DL, ND=DL) | UG/KG | Yes | No | 93.5-108 | NONE | N/A | 5/5 | 1.85 | | | 207, 146, 110 | No | * | * | * | * | * | | ~ | CBR <acute and="" chronic="" td="" threshold="" values<=""></acute> |
| DIELDRIN | UG/KG | No | No | 1.20 | 10 | Yes | 4/5 | 3.11 | 107.0 | 149.0 | | No | No | * | No | * | Yes | | ~ | |
| 1-METHYLNAPHTHALENE | UG/KG | Yes | No | 16 | NONE | N/A | 4/5 | 3.41 | - | 241.0 | 241.0 | No | * | * | * | * | No | | ~ | |
| 2-METHYLPHENOL | UG/KG | No | No | 77 | 20 | No | 4/5 | 1.74 | 43.0 | 74.1 | 81.9 | No | * | * | * | * | No | | ~ | |
| 3,4-METHYLPHENOL | UG/KG | No | No | 45 | NONE | N/A | 2/5 | 0.24 | | | 34.9 | No | * | * | * | * | No | | ~ | |
| BENZOIC ACID | UG/KG | No | No | 6940 | 100 | No | 5/5 | 1.43 | 57.3 | 128.0 | | No | * | * | * | * | No | | ~ | |

N/A=not applicable

*no criterion
| | | Detected in | | Mean Tissue | | Mean Tissue | Frequency | | 101 N | Aagnitud ceedance 70 0 0 | e of (%) | Propensity | ITM EN Exceeds | aluation | USEP Suppleme Evider Exceeds | A Regior ental We nce Scree Exceeds Residue | n III ight-of- ning | Propos | ed Action | |
|----------------------|-----------|-------------------|---------------------------|-------------------|-----|-------------------|---------------------------|-----------------|----------|--------------------------------------|-------------|-----------------|-------------------|---------------------|---------------------------------------|---|---------------------------|-----------------|----------------------|----------|
| Chemical Constituent | Units (a) | Bulk Sediment? | Detected in Elutriate? | Concentratio n | TDL | Residue < TDL? | of Detection in Tissue | Uptake Ratio | Inside | Outsid | Oce | Biomagnify ? | Action Level? | USEPA Tolerance? | Screening Value? | Effect Data? | Exceeds RBC? | Keep as COPC | Eliminate as COPC | Comments |
| DIOXIN TEQ (ND=1/2DL | NG/KG | Yes | No | 0.26 | N/A | N/A | 5/5 | 1.85 | | 47.7 | | Yes | * | * | Yes | * | Yes | ~ | | |
| ALUMINUM | MG/KG | Yes | Yes | 11.28 | 1.0 | No | 5/5 | 0.23 | 31 | 32.1 | | No | * | * | * | * | No | | > | |
| BERYLLIUM | MG/KG | Yes | Yes | 0.03 | 0.1 | Yes | 5/5 | 3.00 | 200 | 200 | 200 | No | * | * | * | * | No | | ~ | |
| CHROMIUM | MG/KG | Yes | Yes | 0.44 | 0.1 | No | 4/5 | 1.54 | 192 | 342 | | No | * | No | * | * | Yes | | ~ | |
| COPPER | MG/KG | Yes | Yes | 1.38 | 0.1 | No | 5/5 | 1.06 | - | 57.2 | | Yes | * | * | * | * | No | | ~ | |
| NICKEL | MG/KG | Yes | Yes | 0.90 | 0.1 | No | 5/5 | 7.92 | 82.9 | 218 | | No | * | No | * | * | No | | ~ | |
| 4,4'-DDD | UG/KG | Yes | No | 10.73 | 10 | No | 4/4 | 14.30 | 329 | 1330 | 423 | Yes | No | * | * | * | Yes | | ~ | |
| CHLORBENSIDE | UG/KG | No | No | 13.40 | 2 | No | 4/4 | 1.37 | | 712 | 177 | Yes | * | * | * | * | * | ~ | | |
| DACTHAL | UG/KG | No | No | 50.50 | 2 | No | 4/4 | 1.64 | 100 | 405 | 100 | Yes | * | * | * | * | No | | ~ | |
| DI-N-OCTYL PHTHALA | UG/KG | No | No | 196 | 20 | No | 1/5 | 1.02 | 5.95 | 5.95 | 5.95 | No | * | * | * | * | No | | ~ | |

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TABLE 9-67A TOLCHESTER STRAIGHTENING: INTEGRATED EVALUATION OF BIOACCUMULATION IN Nereis virens (SAND WORM)

N/A=not applicable

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*no criterion

^(a) all values reported to specified units (i.e., mean tissue concentration, TDL)



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CHAPTER 10 -SUMMARY OF FINDINGS



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CHAPTER 11 -REFERENCES

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October 2000

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15 Loveton Circle Sparks, MD 21152 Telephone: 410-771-4950 Fax: 410-771-4204

27 October 2000



Christina E. Correale Chief - Operations Division USACE Baltimore District P.O. Box 1715 10 North Howard Street - Room 8620 Baltimore, MD 21203-1715

RE: Draft Sediment Report - Site 104, Chapter 8

Dear Ms. Correale:

Accompanying this letter please find the draft of Chapter 8 of the above referenced report. This has been carefully reviewed by senior technical scientists and by our senior editor. We are currently in the process of completing Chapters 9 and 10 and will forward them to you as soon as they are completed. Please feel free to call either Peggy Derrick or me with any questions.

Sincerely,

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Peggy Derrick, Project Manager/Scientist

Frank W. Pine, Ph.D. Senior Scientist/Project Director

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8. TOXICITY TESTING

EA's Ecotoxicology Laboratory performed water column and whole sediment toxicity testing on sediment composites collected from the approach channels, from Placement Site 104, from an Outside Placement Site 104 reference area, and from the Norfolk Ocean Disposal Site reference area. The testing was in compliance with ITM requirements. The toxicity testing program consisted of acute water column bioassays with *Mysidopsis bahia* (opossum shrimp), *Cyprinodon variegatus* (sheepshead minnow), *Mytilus* sp. (blue mussel), and *Menidia beryllina* (inland silverside), and 10-day whole sediment toxicity tests with *Neanthes arenaceodentata* (estuarine polychaete) and *Leptocheirus plumulosus* (estuarine amphipod). Acute water column bioassays with *Arbacia punetulata* (purple sea urchin) were initially conducted. However, USEPA Region III did not recommend this species, and USEPA representatives indicated that this species did not accurately represent aquatic organisms that could potentially be impacted in Attachment V. The acute water column bioassays and the whole sediment toxicity tests evaluated the effects of exposure to the sediment elutriates and whole-sediment, respectively, on survival of the test organisms.

8.1 METHODS

The toxicity testing program consisted of three separate sequential rounds of testing:

- *Round 1*: Initial water column and solid phase testing with sediment from the approach channels and from Inside Placement Site 104 (September-November 1999).
- *Round 2:* Additional water column and solid phase testing conducted in response to recommendations from USEPA Region III–Philadelphia. Additional testing with sediment from an Outside Placement Site 104 reference area and additional testing of sediment from approach channels and Inside Placement Site 104 with additional test species (December 1999-February 2000) was conducted.
- *Round 3:* Testing of Ocean Reference sediment (in conjunction with Woodrow Wilson Bridge sediment testing program) (February-March 2000).

A summary of the toxicity testing schedule is provided in Table 8-1. The water column and whole sediment toxicity testing was conducted in accordance with USEPA/USACE guidance (1998) and EA (1996). The testing procedures, acceptability criteria, and quality assurance protocols are fully documented in the Quality Assurance Project Plan (QAPP) for the ecotoxicological testing program (EA 2000c) (Appendix B). To take advantages in efficiencies from testing on two concurrent programs using the same ocean placement reference site, testing procedures for the Ocean Reference sediment followed the same methodology as the Site 104 testing and are documented in the ecotoxicology QAPP for the Woodrow Wilson Bridge project (EA 2000b).

Original data sheets, records, memoranda, notes, and computer printouts for toxicity testing components are archived at EA's Baltimore Office in Sparks, Maryland. These data will be retained for a period of 5 years unless a longer period of time is requested by USACE–Baltimore District.

8.1.1 Sample Receipt and Preparation

Approximately 20 gallons of composited sediment from each sampling reach were required for the ecotoxicological (including bioaccumulation) testing. Sediment composites for the ecotoxicological testing are described in Table 8-2. Processing and homogenization of sediment followed procedures described in Chapter 4.

After completion of processing, compositing, and homogenization, reach composites and site water samples were logged in and assigned EA laboratory accession numbers. The sediment and water samples were stored in the dark in a secured walk-in cooler at 4°C until used for testing. Prior to use in toxicity testing, large rocks and debris were manually removed and discarded from each sample. Table 8-3A summarizes the sample identifications, accession numbers, and collection and receipt information for sediment and water from the approach channels, inside Placement Site 104, and outside Placement Site 104. Table 8-3B summarizes the sample identifications, accession numbers, and collected from the Ocean Reference site. Copies of chain-of-custody records are provided in Attachments V-A and V-B for approach channels/Site 104 and Ocean Reference, respectively.

8.1.2 Water Column Testing

For the water column toxicity testing, elutriates were prepared from each composited sediment sample. Prior to elutriate preparation, the site water for each sampling reach was salinity adjusted to 30 ppt, as per USEPA/USACE (1998) guidance for the selected test species, using Forty Fathoms synthetic sea salts. Following USEPA/USACE (1998) guidance, a subsample of each homogenized sediment was combined with its respective site water in a 1:4 sediment to water ratio, on a volume/volume basis. The sediment/water combination was thoroughly mixed by vigorous aeration and manually stirring for 30 minutes at 20°C, and was then allowed to settle for one hour. After settling, the supernatant was decanted off and used for the water column acute toxicity testing. All elutriates were used for testing within 24 hours of preparation.

Static, non-renewal bioassays were conducted on the prepared elutriates using *Mysidopsis bahia* (opossum shrimp), *Cyprinodon variegatus* (sheepshead minnow), *Menidia beryllina* (inland silverside), and *Mytilus* sp. (blue mussel). The test organisms were acquired from scientific organism vendors and, when appropriate, gradually acclimated to test temperature and salinity prior to use in testing. The *Mytilus* sp. testing was performed by Northwestern Aquatic Sciences (NAS) located in Newport, Oregon. Elutriates for the *Mytilus* sp. testing were prepared by EA's Ecotoxicology Laboratory, shipped via overnight delivery on the day of preparation, and were used by NAS for the *Mytilus* sp. test within 24 hours of preparation.

8.1.2.1 Mysidopsis bahia and Cyprinodon variegatus Water Column Toxicity Testing

Approach Channels and Inside Placement Site 104 (Round 1)

The 96-hour toxicity tests with *M. bahia* and *C. variegatus* were conducted in two batches (25-29 October 1999, 26-30 October 1999). Two lots of test organisms per species were acquired from Cosper Environmental Services, Inc. (Bohemia, New York). For use in the 25 October testing, stocks of *M. bahia* (lot number MB-402) and *C. variegatus* (CV-287) were received at EA on 23 October from Cosper Environmental Services. Additional lots from Cosper Environmental Services, *M. bahia* (lot MB-403) and *C. variegatus* (lot CV-288), were received on 26 October for use in toxicity testing the same day. The opossum shrimp and sheepshead minnows were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding.

Outside Placement Site 104 (Round 2)

The 96-hour toxicity tests with *M. bahia* and *C. variegatus* were conducted 4-8 January 2000. An additional test with the Inside Site 104 sediments was conducted concurrently with the Outside Site 104 sediment to compare the consistency of results for the Inside Site 104 testing in Round 1 and Round 2. Test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The lots of organisms MB-413 (*M. bahia*) and CV-304 (*C. variegatus*) were received on 4 January 2000 from Aquatic BioSystems for use in toxicity testing the same day. The opossum shrimp and sheepshead minnows were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding.

Ocean Reference Site (Round 3)

The 96-hour toxicity test with *M. bahia* was conducted on 16-20 February 2000. The test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The organisms. *M. bahia* (lot MB-422), were received at EA on 15 February 2000. The opossum shrimp were fed freshly hatched *Artemia* sp. nauplii (<24 hours old) during holding. *C. variegatus* tests were not conducted under the Woodrow Wilson Bridge testing program.

Test concentrations of 100, 50, and 10 percent elutriate were prepared by measuring aliquots of elutriate in a graduated cylinder and bringing to final volume with artificial seawater. A dilution water control of artificial seawater was also prepared. The artificial seawater was prepared by mixing Forty Fathoms synthetic sea salts with laboratory water to a final salinity of 30 ppt. The source of the laboratory water was the City of Baltimore municipal tap water which was passed through a high-capacity, activated carbon filtration system. This synthetic seawater formulation has proven acceptable for aquatic toxicological studies, and has been used successfully at EA for maintaining multigeneration cultures of *M. bahia*, and for holding healthy populations of estuarine and marine species. Batches of artificial seawater were aerated and aged at least 24 hours prior to use in testing.

8-3

The opossum shrimp and sheepshead minnow testing utilized 1-L beakers as test chambers. Each beaker contained 250 ml of test solution, and there wcrc five replicate beakers per test concentration. Ten organisms were randomly introduced into each replicate for a total of 50 organisms per concentration. The test chambers were maintained at 20±1°C with a 16-hour light/8-hour dark photoperiod. Temperature, pH, and dissolved oxygen were measured in one replicate of each concentration daily for the 96-hour exposure period. Due to the size of the water quality probe and the danger of injury to the test organisms, salinity was measured in each concentration only at test initiation (prior to introduction of the test organisms) and at termination. A summary of water quality measurements is presented in Table 8-4A and 8-4B (M. bahia-approach channels and Site 104), Table 8-4C (M. bahia- ocean reference), and Tables 8-5A and 8-5B (C. variegatus-approach channels and Inside Site 104). The number of live organisms were counted daily and recorded on the test data sheets. The opossum shrimp were fed freshly hatched Artemia sp. daily to avoid cannibalism during testing. Copies of the opossum shrimp and sheepshead minnow acute toxicity test data sheets are included in Attachments V-A and V-B, for the approach channel/Site 104 and the Ocean Reference, respectively.

8.2.1.2 Menidia beryllina Water Column Toxicity Testing

Approach Channels, Inside Placement Site 104, and Outside Placement Site 104 (Round 2)

The 96-hour *M. beryllina* acute toxicity tests were conducted in five batches (20-24, 21-25, and 25-29 January 2000, and 4-8 and 7-11 February 2000). The test methodologics followed those of the *M. bahia* and *C. variegatus* water column testing with the exception of the test volume per chamber. Each test chamber for the *M. beryllina* testing contained 200 ml of test solution.

The *M. beryllina* toxicity testing was performed with 9-14 day old fish (hatched within a 24-hour period). During the testing program, five lots of *M. beryllina* were acquired from Aquatic BioSystems (Fort Collins, Colorado), and were gradually acclimated to test conditions in EA's Culture Facility prior to use in testing. The *M. beryllina* were fed *Artemia* nauplii at the 48-hour intermediate observation period during testing.

Ocean Reference (Round 3)

The 96-hour toxicity test with *M. beryllina* was conducted on 16-20 February 2000. The test organisms were acquired from Aquatic BioSystems (Fort Collins, Colorado). The organisms, *M. menidia* (lot MS-080), were received at EA on 15 February 2000, and were gradually acclimated to test conditions in EA's Culture Facility prior to use in testing. The *M. beryllina* were fed *Artemia* sp, nauplii (<24 hours old) during the holding period.

A summary of water quality parameters measured during the *M. beryllina* testing is presented in Tables 8-6A (approach channels, Inside Site 104, and Outside Site 104) and 8-6B (Ocean Reference). Attachments V-A and V-B contain copies of the data sheets from the *M. beryllina* toxicity testing for the approach channels/Site 104 and ocean testing, respectively.

8.1.2.3.1 Mytilus sp. Water Column Toxicity Testing

The *Mytilus* sp. bivalve embryo-larval toxicity testing was conducted by Northwestern Aquatic Sciences (NAS) located in Newport, Oregon. The 30 ppt salinity elutriates were prepared by EA on 20 December 1999 (approach channels and Site 104) and on 14 February 2000 (Ocean Reference), packed on wet ice the same day as the elutriate preparation, and shipped by overnight courier to NAS. The bivalve toxicity tests were initiated by Northwestern upon receipt of the elutriate samples on 21 December 1999 and on 15 February 2000.

The adult mussels were acquired from Carlsbad Aquafarm (Carlsbad, California) on 21 December 1999 and 15 February 2000. The mussels were induced to spawn by gradually cycling the temperature of the holding water several times through the range of 15-23 °C. Spawning animals were rinsed and isolated in small dishes containing clean filtered seawater for collection of gametes. Four females and three males were isolated for gamete collection.

Unfertilized eggs were rinsed and suspended in clean seawater at a concentration of approximately 5,000 eggs/ml. Sufficient sperm was added to the egg suspension to achieve an approximate sperm to eggs ratio of 5:1. Ten minutes after sperm addition, the suspension was filtered through a 25 μ m Nitex® screen to remove remaining sperm, and the embryos were resuspended and adjusted to achieve a stock concentration of about 2,500 embryos/ml.

Test chambers were 30 ml borosilicate glass vials containing 10 ml of test solution. Test concentrations of 100, 50, 10, and 0 percent elutriate were prepared using 30 ppt site water (Yaquina Bay, Oregon) for the dilution. A laboratory control of natural clean seawater was also utilized. At test initiation, $100 \ \mu$ l of well-mixed embryo suspension was added to each test chamber. The tests were maintained at 15 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Temperature, pH, dissolved oxygen, and salinity were measured daily on surrogate test chambers (without test organisms). A summary of water quality measurements from the *Mytilus* sp. testing is provided in Table 8-7A (approach channels, Inside Site 104, Outside Site 104) and Table 8-7B (Ocean Reference). The toxicity tests were terminated on 23 December 1999 and 17 February 2000 by adding 1 ml of 37 percent buffered formalin to each test chamber. The preserved embryos were observed microscopically, to determine the percentage of normally developed larvae. The complete reports for the *Mytilus* sp. elutriate testing are presented in Attachments V-A and V-B.

8.1.3 Whole Sediment Testing

8.1.3.1 Neanthes arenaceodentata and Leptocheirus plumulosus Whole Sediment Testing

Approach Channels and Inside Site 104 (Round 1)

Whole sediment toxicity testing was conducted with the estuarine polychaete *Neanthes* arenaceodentata and the estuarine amphipod *Leptocheirus plumulosus*. The approach channel and Inside Site 104 sediments were evaluated on 22 October-1 November with

N. arenaceodentata and on 23 October-2 November 1999 with *L. plumulosus*. The polychaete worms (organism lot numbers NA-005) were acquired from Dr. Donald Riesh, California State University (Long Beach, California) on 21 October 1999, and the amphipods (LP-010) were acquired from University of Maryland, Wye Research and Education Center (Queenstown, Maryland) on 21 October 1999. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C and to the appropriate test salinity [20 ($\pm 10\%$) ppt for *Leptocheirus plumulosus* and 30 ($\pm 10\%$) ppt for *Neanthes arenaceodentata*, as per USEPA/USACE (1998)].

Outside Site 104 (Round 2)

Additional lots of *N. arenaceodentata* and *L. plumulosus* were acquired for the later testing of the Outside Site 104 reference sediment. The Inside Site 104 sediment was re-run concurrently with the Outside Site 104 tests to compare the consistency of results for Inside Site 104 testing in Round 1 and Round 2. The *N. arenaceodentata* lot (NA-006) was received on 28 December from Dr. Donald Reish for use in toxicity testing on 30 December 1999. The *L. plumulosus* lot (LP-012) was acquired from Aquatic BioSystems on 27 January for use in toxicity testing on 28 January 2000.

Ocean Reference (Round 3)

L. plumulosus (organism lot # NA-005) were acquired from Aquatic BioSystems (Fort Collins, CO) on 23 February 2000. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C. Based on recommendations from USEPA Region III, a lower Chesapeake Bay control sediment (collected from a USEPA Region III approved location) was tested in place of the Ocean Reference Site sediment. L. plumulosus are sensitive to sediment grain size characteristics (USEPA 1993b). This species of estuarine amphipod prefers fine sediment characteristic of the Chesapeake Bay and does not typically survive well in coarse-grained sandy sediment. The lower Chesapeake Bay control sediment was tested in place of the Ocean Reference the potential for adverse grain-size effects on survival.

The whole sediment toxicity tests were conducted as static, non-renewal tests with 10 days of exposure to the whole sediments and overlying water. Artificial seawater (Forty Fathoms sea salts) at 30 ppt salinity for *N. arenaceodentata* and 20 ppt salinity for *L. plumulosus* was used as the overlying water. The sediments and overlying water were added to the test chambers, and the suspended sediments were allowed to settle 1-3 days. During the settling period, ammonia was monitored in the overlying water of each sediment test chamber as directed by USEPA/USACE guidance to ensure that ammonia was not an artifact of the sediment test. No replacement of the overlying water was required due to the low measured levels of ammonia (<2 mg/L NH₃-N). The addition of the test organisms to the exposure chambers marked the initiation of the toxicity tests. A natural Chesapeake Bay sediment was used as a laboratory control for the whole sediment toxicity tests.

The *N. arenaceodentata* and *L. plumulosus* tests utilized 1-L beakers as the exposure chambers, with each beaker containing 200 ml of sediment and 700 ml of overlying water. There were five

replicate chambers for each sediment sample and control. Test organisms were randomly assigned to the test chambers. The *N. arenaceodentata* test had 5 organisms per replicate for a total of 25 organisms exposed per sample, while the *L. plumulosus* test had either 20 or 25 organisms per replicate chamber for a total of 100 organisms per sample (23 October 1999 testing) or 125 organisms per sample (28 January 2000 testing).

The tests were maintained at 20 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Water quality measurements of temperature, pH, dissolved oxygen, and salinity were recorded daily on one replicate of each sample and control. Water quality parameters measured during the *N. arenaceodentata* testing are summarized in Tables 8-8A (approach channels and Inside Site 104) and 8-8B (Outside and Inside Site 104). Water quality parameters measured during the *L. plumulosus* testing are summarized in Tables 8-9A (approach channels and Inside Site 104), 8-9B (Outside and Inside Site 104), and 8-9C (Ocean Reference). Additionally, test chambers were visually inspected daily for abnormal organism behavior/lack of burrowing. The test organisms were not fed during the 10-day exposure period.

After 10 days of exposure, the test organisms were retrieved from the samples and the number of live organisms per replicate was recorded. Copies of the original data sheets for the *N. arenaceodentata* and the *L. plumulosus* testing are included in Attachments V-A (approach channels/Site 104) and V-B (Ocean Reference), respectively.

8.1.3.2 Mysidopsis bahia Whole Sediment Testing

For the Ocean Reference sediment, *Mysidposis bahia* was tested as the second species for the whole sediment bioassays. The Ocean Reference sediment was tested in conjunction with another testing program (Woodrow Wilson Bridge) that requested *M. bahia* testing, rather than *Neanthes arenaceodentata*. Either species is acceptable to USEPA Region III for evaluating open water and ocean placement.

The opossum shrimp (lot MB-423) were acquired from Aquatic BioSystems on 18 February 2000. During the holding period, the organisms were gradually acclimated to laboratory water at 20°C and at the required test salinity (30 ppt).

The whole sediment toxicity tests were conducted as static, non-renewal tests with 10 days of exposure to the whole sediments and overlying water. Artificial seawater (Forty Fathoms sea salts) at 30 ppt salinity for *M. bahia* was used as the overlying water. The sediments and overlying water were added to the test chambers on 17 February 2000, and the suspended sediments were allowed to settle 1-5 days. During the settling period, ammonia was monitored in the overlying water of each sediment chamber. Due to the low measured levels of ammonia (<2 mg/L NH₃-N), no replacement of the overlying water was required. The addition of the test organisms to the exposure chambers marked the initiation of the toxicity tests. The *M. bahia* whole sediment toxicity tests were initiated on 18 February 2000. A natural Chesapeake Bay sediment was used as a laboratory control for the *M. bahia* tests.

The *M. bahia* tests utilized 1-L beakers as the exposure chambers, with each beaker containing 300 ml of sediment and 650 ml of overlying water. There were five replicate chambers for each

sediment sample and control. Test organisms were randomly assigned to the test chambers. The *M. bahia* tests had 10 organisms per replicate for a total of 50 organisms exposed per sample.

The tests were maintained at 20 ± 1 °C with a 16-hour light/8-hour dark photoperiod. Water quality measurements of temperature, pH, dissolved oxygen, and salinity were recorded daily on one replicate of each sample and control. Water quality parameters measured during the Ocean Reference *M. bahia* testing are summarized in Table 8-10. The test organisms were not fed during the 10-day exposure period. After 10 days of exposure, the test organisms were retrieved from the samples and the number of live organisms per replicate was recorded. Copies of the original data sheets for the *M. bahia* testing are included in Attachment V-B.

8.1.4 Data Analysis / Statistics

Statistical analyses were performed on the water column and whole sediment test data according to USEPA/USACE (1998) guidance. Survival (or larval development) of the organisms exposed to the test material for the prescribed time period was statistically compared (p=0.05) to either the laboratory control (elutriate tests) or the reference sediment (whole-sediment tests) as appropriate for each test using the student t-test.

For the elutriate testing, a 96-hour LC50 (median lethal concentration), or EC50 (median effective concentration for *Mytlius* sp.), was calculated for each test species using either the linear interpolation, trimmed Spearman-Karber method (Hamilton et al., 1977), or probit method (as described by Stephan 1977). The LC50 is an estimate of the elutriate concentration that is lethal to 50 percent of the test organisms, or that creates a sub-lethal effect on the development of 50 percent (EC50) of the test organisms, in the time period prescribed by the test. If survival in the 100 percent elutriate concentration was at least 10 percent lower than the dilution water control, then a statistical comparison (t-Test) was performed between the 100 percent elutriate concentrations were independent and normally distributed, and that the variances of the observations were equal between the two groups. The F-Test was used to test for homogeneity of variance. The test for normality was the Shapiro-Wilk's Test. When the data did not meet the normality assumption, the nonparametric test, Wilcoxon's Rank-Sum Test, was used to analyze the data. An arc sine (square root [Y]) transformation was performed on the survival percentages, where appropriate.

For data sets in which the 100 percent concentration was statistically different from the control, an addition analysis was performed to determine the No Observed Adverse Effect Concentration (NOAEC). Based on USEPA (1993) guidance for standard multi-concentration effluent toxicity tests, a multiple mean comparison was conducted to statistically compare the 100, 50, and 10 percent concentrations to the control. This comparison utilized the dose response data from the three test concentrations and control. A concentration which had no surviving organisms was excluded from the analysis. A parametric or nonparametric statistical test was utilized based on the assumptions of normality and homogeneity of variance. The test for normality was the Shapiro-Wilk's Test, and the test for homogeneity of variance was the Bartlett's Test. For parametric data, an analysis of variance (ANOVA) and either Dunnett's Mean Comparison test or Bonferroni's T-test was used (depending on equal or unequal replicate numbers). Steel's Many-One Rank Test or the Wilcoxon Rank Sum Test were the alternative nonparametric tests.

For the whole sediment toxicity test data, statistical analyses were performed to determine if exposure to any of the sediment samples resulted in significantly lower survival of the test organisms as compared to the reference site. If survival in a test sample was at least 10 percent lower than the reference, then a t-test or Wilcoxon's Rank-Sum Test (depending on data characteristics) was performed to compare the single test sample to the reference (Inside Site 104).

For the whole sediment bioassays, only results generated within a single round of testing should be statistically compared to each other because of potential differences associated with survival in the control organisms. In this testing program, data from different rounds of whole-sediment testing could not be statistically compared against each other, but were compared qualitatively (i.e., toxic or non-toxic). For the water column bioassays, statistical comparisons are conducted against a laboratory or test control (not a reference sample). In these cases, each sample had an independent statistical comparison against a control, and the results generated by different rounds of testing stand as independent measures of toxicity.

8.1.5 Reference Toxicant Testing

In conformance with EA's QA/QC program requirements, reference toxicant testing was performed on the acquired lots of organisms utilized in the testing program or reference toxicant data were obtained from the test organism supplier. The reference toxicant tests consisted of a graded concentration series of a specific toxicant in water only tests, with no sediment present in the test chambers.

The reference toxicant for *M. bahia*, *C. variegatus*, and *M. beryllina* was potassium chloride (KCl); the reference toxicant for *Mytilus* sp. was copper sulfate (CuSO₄·5H₂0); and the reference toxicant for *N. arenaceodentata* and *L. plumulosus* was cadmium chloride (CdCl₂). Reference toxicant testing was also conducted for the species utilized in the bioaccumulation studies (Chapter 9). The results of the reference toxicant tests were compared to established control chart limits.

8.2 **RESULTS**

8.2.1 Water Column Testing

8.2.1.1 Mysidopsis bahia

The results of the *M. bahia* toxicity testing are presented in Tables 8-11A (approach channels and Inside Site 104), 8-11B (Inside and Outside Site 104), and 8-11C (Ocean Reference).

Approach Channels and Inside Placement Site 104 (Round 1)

The results of the initial 14 elutriate tests with *M. bahia* are provided in Table 8-11A. Five of the prepared elutriates exhibited some inhibition of survival to *M. bahia*. Samples from the Craighill Entrance (CRE), Craighill Angle-West (CRA-W), Tolchester-South (TLC-S),

Brewerton Eastern Extension (BE), and C& D Approaches (surficial CD) had 96-hour LC50s of >100 percent elutriate; however, the survival in the 100 percent elutriate concentration for each sample was statistically lower than the control. Calculation of NOAECs for these samples indicated that no effect would be expected at elutriate concentrations of >50%. Samples from the Craighill (CR), Craighill Angle-East (CRA-E), Craighill Upper Range (CRU), Cutoff Angle (CUT), Tolchester North (TLC-N), Tolchester Straightening (TLS), Swan Point (SWP), C&D Approaches (cores CD-VC), and Inside Site 104 (KI-Reference) were not acutely toxic to M. *bahia*, with 96-hour LC50s of >100 percent elutriate and no statistically difference in survival between the control and the 100 percent elutriate concentration.

Outside Placement Site 104 (Round 2)

The results of the *M. bahia* testing on the Outside Site 104 (KI-OUT) reference elutriate are presented in Table 8-11B. The Outside Site 104 sample was not acutely toxic to *M. bahia* with 94 percent survival in the 100 percent elutriate concentration (48-hour LC50 >100 percent elutriate). The Inside Site 104 reference elutriate was re-analyzed concurrently with the Outside Site 104 elutriate using a fresh sample of Inside Site 104 sediment and site water. The Inside Site 104 elutriate was again not acutely toxic to *M. bahia* (48-hour LC50 >100 percent elutriate), with 100 percent survival in the 100 percent elutriate concentration.

Ocean Reference (Round 3)

The results of the Ocean Reference elutriate testing with *M. bahia* are provided in Table 8-11C. The Ocean Reference sample was not acutely toxic to *M. bahia*, with a 96-hour LC50 of >100 percent elutriate and no statistical difference in survival between the control and the 100 percent elutriate concentration.

8.2.1.2 Cyprinodon variegatus

The results of the sheepshead minnow acute toxicity tests are summarized in Tables 8-12A (approach channels and Inside Site 104) and 8-12B (Inside and Outside Site 104).

Approach Channels and Inside Placement Site 104 (Round 1)

Results of the *C. variegatus* acute toxicity tests conducted on the 13 channel elutriates and Inside Site 104 are provided in Table 8-12A. None of the elutriates was acutely toxic to *C. variegatus*. Survival in the 100 percent elutriate concentrations was at least 92 percent, while survival in the controls was a minimum of 98 percent. The 96-hour *C. variegatus* LC50s for all of the tested elutriates, including the Inside Site 104 reference, were >100 percent elutriate.

Outside Placement Site 104 (Round 2)

The results of the Outside Site 104 reference elutriate testing with *C. variegatus* are summarized in Table 8-12B. The Outside Site 104 elutriate was not acutely toxic to *C. variegatus*, with 96 percent survival in the 100 percent clutriate concentration. The re-analyzed Inside Site 104 reference elutriate was also not acutely toxic (100 percent survival in the 100 percent elutriate

concentration). The 48-hour LC50s for the Inside Site 104 and Outside Site 104 elutriates were both >100 percent elutriate.

Ocean Reference (Round 3)

C. variegatus was not tested with the Ocean Reference sediment.

8.2.1.3 Menidia beryllina

Results of the inland silverside testing are summarized in Table 8-13A (approach channels, Inside Site 104, and Outside Site 104) and Table 8-13B (Ocean reference). These tests were performed as an additionally requested species during the testing program.

Approach Channels, Inside Placement Site 104 and Outside Placement Site 104 (Round 2)

The *M. beryllina* acute toxicity test results are summarized in Table 8-13A. Four of the elutriates (Inside Site 104, Outside Site 104, Swan Point, and Tolchester Straightening) were not acutely toxic to *M. beryllina*, with 48-hour LC50s of >100 percent elutriate and no statistical difference in survival between the 100 percent elutriate concentration and the control. Samples for the Craighill Upper Range and the C&D Approaches (cores) had 48-hour LC50s of >100 percent elutriate; however, there was a statistically significant decrease in survival in the 100 percent elutriate concentration sof NOAECs indicated that no effects to survival would be expected in the 10% elutriate and 50% elutriate for the Craighill Upper Range and C&D Approach (cores), respectively.

The remaining nine elutriates [Craighill, Craighill Entrance, Craighill Angle-East, Craighill Angle West, Cutoff Angle, Brewerton Eastern Extension, C&D Approaches (surficial), Tolchester-South, and Tolchester-North] were all acutely toxic to *M. beryllina* with 48-hour LC50s ranging from 23.8 percent elutriate (Tolchester-South) to 70.7 percent elutriate (Craighill Entrance). Calculation of NOAECs indicated that no effects to survival would be expected at a concentration of 10% elutriate in 8 of the 9 channel clutriates. The NOAEC was <10% elutriate for Brewerton Eastern Extension.

Ocean Reference (Round 3)

The *M. beryllina* acute toxicity test results for the Ocean Reference site are presented in Table 8-13B. The Ocean Reference elutriate was not acutely toxic to *M. beryllina*, with 96-hour LC50 of >100 percent elutriate and no statistical difference in survival between the control and the 100 percent elutriate concentration.

8.2.1.4 Mytilus sp.

Results of the blue mussel testing are summarized in Table 8-14A (approach channels, Inside Site 104, Outside Site 104) and Table 8-14B (Ocean Reference).

Approach Channels, Inside Placement Site 104 and Outside Placement Site 104 (Round 2)

| Baltimore Harbor Approach Channel | s and Placement Site 104 |
|-----------------------------------|--------------------------|
| Dredged Material Evaluation | Draft Report |
| 5 | 8-11 |

Results of the *Mytilus* sp. embryo larval toxicity tests are presented in Table 8-14A. Two samples (Craighill and Craighill Upper Range) were not acutely toxic to *Mytilus* sp., with 87 percent normal development in the 100 percent elutriate concentration and 48-hour EC50s of >100 percent elutriate. The Inside Site 104 reference elutriate (48-hour EC50 >100 percent elutriate) was marginally toxic with 82 percent normal development in the 100 percent elutriate concentration which was statistically lower than the control (97 percent normal development). The Outside Site 104 reference elutriate was more toxic than the Inside Site 104 elutriate, with a 48-hour EC50 of 63.8 percent elutriate.

Elutriates for the Cutoff Angle, Tolchester-South, Tolchester-North, Tolchester Straightening, Brewerton Eastern Extension, and the C&D Approaches (surficial) had 48-hour EC50s ranging from 56.4 to 79.7 percent elutriate. Elutriates for the Craighill Angle-East, Craighill Angle-West, and Swan Point had 48-hour EC50s of 43.5, 46.8, and 47.8 percent elutriate, respectively. Elutriates for the Craighill Entrance and C&D Approaches (cores) were the most toxic with 48-hour EC50s of 22.0 and 21.2 percent elutriate, respectively. Overall, calculation of NOECs indicated that only one elutriate (C&D Approach cores) would be expected to effect larval development at concentrations of <10% elutriate.

Ocean Reference (Round 3)

The results of the elutriate testing with *Mytilus* sp. for the Ocean Reference are presented in Table 8-14B. The Ocean Reference sediment elutriate was not acutely toxic to *Mytilus* sp. (EC50 > 100 percent).

8.2.2 Whole Sediment Testing

8.2.2.1 Neanthes arenaceodentata

The results of the estuarine polychaete toxicity tests are summarized in Tables 8-15A (approach channels and Inside Site 104) and 8-15B (Outside and Inside Site 104).

Approach Channels and Inside Placement Site 104 (Round 1)

The results of the *N. arenaceodentata* toxicity testing conducted on the approach channel and Inside Site 104 sediments are summarized in Table 8-15A. After ten days of exposure, the lowest survival (92 percent) was recorded in the sediment from Craighill Angle-West. Sediments from Craighill, Cutoff Angle, Tolchester-South, Tolchester Straightening, and C&D Approaches (cores) had 96 percent survival, while sediments from Craighill Entrance, Craighill Angle-East, Craighill Upper Range, Tolchester North, Brewerton Eastern Extension, Swan Point, and C&D Approaches (surficial) had 100 percent survival. Survival in all of the sediments was within 10 percent of the Inside Side 104 sediment, which had 96 percent survival, indicating that the sediments are not toxic and are not statistically different from those at the proposed placement site. The laboratory control sediment for the *N. arenaceodentata* testing had 100 percent survival.

Outside Placement Site 104 (Round 2)

As summarized in Table 8-15B, the Outside Site 104 reference sediment had 80 percent survival compared to 88 percent survival in the Inside Site 104 reference and 96 percent survival in the control. When tested statistically, survival in the Outside Site 104 sediment was not significantly different from the Inside Site 104 survival or the control survival.

Ocean Reference (Round 3)

N. arenaceodentata was not tested for the Ocean Reference site.

8.2.2.2 Leptocheirus plumulosus

The results of the estuarine amphipod toxicity tests are summarized in Tables 8-16A (approach channels and Inside Site 104), 8-16B (Outside and Inside Site 104), and 8-16C (Ocean Reference).

Approach Channels and Inside Placement Site 104 (Round 1)

The *L. plumulosus* whole sediment toxicity test results are presented in Table 8-16A. Survival in the channel sediments ranged from 86 to 96 percent, which was within the allowable 20 percent difference from the Inside Site 104 reference survival of 93 percent. Based on the survival results, the whole sediments were not acutely toxic to *L. plumulosus*. The control sediment for the *L. plumulosus* toxicity testing had 96 percent survival.

Outside Placement Site 104 (Round 2)

The Outside Site 104 reference sediment had 96 percent survival indicating that this sample was not acutely toxic to *L. plumulosus* (Table 8-16B). Survival for Inside Site 104 and the control were 87 and 98 percent, respectively.

Ocean Reference (Round 3)

The *L. plumulosus* whole sediment toxicity test results are presented in Table 8-16C. The lower Chesapeake Bay control sediment had 94 percent survival to estuarine amphipods.

8.2.2.3 Mysidopsis bahia

Ocean Reference (Round 3 only)

The results of the opossum shrimp whole sediment toxicity testing for the Ocean Reference are summarized in Table 8-17. The Ocean Reference sediment had 94 percent survival, indicating that the sediment was not acutely toxic to opossum shrimp.

8.2.3 Reference Toxicant Tests

The results of the reference toxicant tests are summarized in Table 8-18A (Round 1 and Round 2) and Table 8-18B (Round 3). The LC50s from the *M. bahia*, *C. variegatus*, *Mytilus* sp., *M. beryllina*, and *L. plumulosus* (96-hours; Round 1 and Round 2) reference toxicant tests all fell within the established laboratory control chart limits.

For *N. arenaceodentata*, control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 6.4 mg/L Cd was similar to the LC50 of 5.7 mg/L Cd for a previous lot of *N. arenaceodentata* from September 1999.

For *L. plumulosus* (reference toxicant tests for Ocean Reference-Round 3), control chart limits have not yet been established (48-hour LC50), because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 5.9 mg/L Cd (48-hour) falls within the range of LC50's from previous *L. plumulosus* reference toxicant tests (2.2-9.0 mg/L Cd).

8.3 DISCUSSION AND TIER III TOXICITY EVALUATION

The Tier III toxicity evaluation requires an analysis of water column and benthic toxicity test data. The following sections discuss the results of the water column and whole-sediment bioassays and the potential for impacts to the aquatic environment.

8.3.1 Water Column Bioassays

According to the ITM (USEPA/USACE 1998), after considering water column test results and expected mixing at the placement site, one of the following conclusions is reached:

- 1) The 100% dredged material elutriate toxicity is not statistically higher than the dilution water (laboratory control). Therefore, the dredged material is not predicted to be acutely toxic to water column organisms. However, benthic impact must also be evaluated.
- 2) The concentration of dissolved plus suspended contaminants, after allowance for mixing, does not exceed 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material is not predicted to be acutely toxic to water column organisms. However, benthic impact must also be evaluated.
- 3) The concentration of dissolved plus suspended contaminants, after allowance for mixing, exceeds 0.01 (1%) of the toxic LC50 or EC50 concentration beyond the boundaries of the mixing zone. Therefore, the dredged material may have the potential to be acutely toxic to water column organisms.

The evaluation guidelines in the ITM assume that state regulatory agencies will issue or permit an allowable mixing zone for open water placement events. It is important to recognize that the evaluation protocols in the ITM are guidelines, not regulations. More specifically, 404 CFR Part 230.10 (c) states that dredged material placement may not result in unacceptable adverse impact; however, the guidance in the ITM *suggests* the 0.01 (1%) LC50/EC50 concentrations as a quantitative method for assessing whether unacceptable adverse impacts might occur in the water column.

The ITM explicitly states that regional modifications of the national guidelines may be required or may be appropriate based on project-specific requirements or circumstances (ITM, p.1-1). In a situation where a mixing zone is not issued, the evaluation of the water column impacts requires alternative methods to demonstrate whether an unacceptable adverse impact is expected during placement or as a result of placement. Therefore, the evaluation of water column impacts for open-water placement at Site 104 is based on an assessment of elutriate concentration and duration of exposure to aquatic organisms. Evaluation of dredged material placement at the NODS is based on the Ocean Testing Manual (USEPA/USACE 1991) guidance which specifies that the 0.01 (1%) of LC50/EC50 must occur within a 4-hour time period inside the boundaries of the placement site (NODS).

Results of the water column toxicity testing are summarized in Table 8-19A and 8-19B. LC50/EC50 and NOAEC/NOEC values are provided in Table 8-19A, and mixing factors that would be required to determine compliance for ocean placement are provided in Table 8-19B.

In the water column tests, survival was the endpoint for the opposum shrimp, sheepshead minnow, and inland silverside tests. The endpoint of the blue mussel test was normal hinge development. All water column tests were conducted with larval or juvenile tests organisms within the age range specified by the USEPA/USACE (1998) testing guidelines: opposum shrimp (1-5 days old); sheepshead minnow (1-14 days old); inland silverside (9-14 days old); and blue mussels (< 4 hours old). In water column tests, results for 100% test elutriates are statistically compared (single-point comparison) to results of the laboratory controls as per ITM evaluation protocols, not to the results for the placement site or reference area.

Results of the water column tests indicated that the blue mussel and inland silverside were the most sensitive water column species to the project elutriates. None of the tested 100% elutriates was acutely toxic to sheepshead minnow. Although the LC50 values were >100% elutriate for all of the opposum shrimp tests, mean survival in 5 of the 13 elutriates was statistically lower than the mean laboratory control survival.

Normal development in the blue mussel is defined as transformation to the fully shelled, straight hinged, D-Shaped prodissoconch I stage. In the blue mussel tests, the 48-hr EC50 (median effect concentration) for the channel sediments ranged from 21.2 to >100% elutriate. Eleven of the 13 reaches had EC50 values for 100% elutriate that were statistically lower than the laboratory controls, indicating that eleven of the 100% elutriates affected normal development in the larval (<4 hrs old) organisms. Craighill Channel and Craighill Upper Range were the only channels where development was not statistically lower than the laboratory control. The lowest EC50 values were reported for the C&D Approach Channel-cores (21.2% elutriate) and the Craighill Entrance (22% elutriate).
Results of the inland silverside bioassays indicated that 11 of the 13 channel elutriates elicited some level of acute toxicity to juvenile (9-14 day old) inland silversides when exposed to undiluted elutriate. Tolchester Straightening and Swan Point were the only channel elutriates (100%) that were not acutely toxic to the juvenile inland silverside (LC50 >100% and not significantly different than the laboratory control). LC50 values for inland silverside ranged from 23.8% elutriate to >100% elutriate. The lowest LC50 value was reported for Tolchester-South (23.8% elutriate).

Calculation of NOAECs for oppossum shrimp, sheepshead minnow, and inland silverside and NOECs for blue mussel water column tests indicated that 11 of the 13 test reaches had NOAECs or NOECs of >10% elutriate for all of the test species data. Brewerton Eastern Extension (inland silverside test) and C&D Approaches-cores (blue mussel test) yielded an NOAEC of <10% and an NOEC of <10%, respectively. These results indicate that no effect to survival or larval development would be expected to occur at elutriate concentrations below 10% based on a continuous 96-hour exposure period (oppossum shrimp, sheepshead minnow, and inland silverside) or based on a 48-hour continuous exposure period (blue mussel), with the exception of the two specified channel/species combinations with an NOAEC or NOEC of <10% elutriate. Based on the results of the STFATE modeling, an elutriate concentration of 10% would be expected at the placement site within approximately one-hour after placement occurs. The NOAECs indicate that no effect is expected for the majority of test organisms/channels at a concentration of 10% over either a 96-hour (oppossum shrimp, sheepshead minnow, and inland silverside) or a 48-hour (blue mussel) exposure period. For each placement event at Site 104, the duration of organism exposure to elutriate constituents in the water column would be expected to be short (acute), not a long-term continuous chronic exposure. Therefore, assuming that material would not be placed successively or consecutively at the same location within Site 104, no unacceptable adverse acute impact to water column organisms would be expected for a single placement event.

Evaluation of the elutriate data using methodology for whole effluent toxicity (USEPA 1991) yields similar results. The most restrictive acute toxicity test value presented in Table 8-19A is a 48-hour EC50 value for the blue mussel (21.2 percent elutriate). Per USEPA guidance, converting this value to acute toxic units (TUa) yields a value of 4.7 TUa (100/21.2). USEPA (1991) guidance in the Technical Support Document for Water Quality-Based Toxics Control requires compliance with the Agency's 0.3 TUa acute toxicity criterion at the edge of an acute mixing zone. Further, the 0.3 TUa criterion is generally interpreted as a 1-hour average concentration (p. 35), which "is expected to be fully protective for the fast-acting toxicants [e.g., chlorine, ammonia], and even more protective for slower-acting toxicants" (p. 35). Based on this guidance, the most restrictive of the elutriate test results would require 15.7 fold dilution [4.7/0.3] within one-hour to comply with USEPA's Technical Support Document guidance. STFATE modeling results presented in Chapter 7 indicates that a 16:1 dilution factor would be achieved within one hour under conservative modeling conditions. Using this approach, the other acute toxicity results presented in Table 8-19 would also not be expected to result in acute toxicity in the water column. Furthermore, this evaluation approach is believed to be, conservative because it compares 48- and 96-hour continuous exposure acute toxicity test results to a one-hour exposure duration criterion, and "fast-acting toxicants" are not expected to be present in the elutriate samples in meaningful concentrations (e.g., free chlorine, ammonia).

Ocean Placement

Based on the results of the water column toxicity testing, a maximum mixing factor of 472-fold would be required for all reaches to comply with the 0.01 LC50/EC50 requirement at the edge of the allowable mixing zone for ocean placement (Table 8-19B). This value is based on the lowest EC50 of 21.2% (C&D approach –core elutriate for blue mussel) in combination with a very conservative acute to chronic conversion factor of 0.01. Modeling of conditions at the ocean placement site indicated that a 1000-fold dilution would occur within the disposal site boundary during the allowable 4-hour ocean placement mixing period (see Chapter 7). Therefore, none of the channel elutriates is expected to be acutely toxic to aquatic organisms during ocean placement.

8.3.2 Whole-Sediment Bioassays

According to Tier III of the ITM (USEPA/USACE 1998), benthic toxicity testing of contaminants in the dredged material in Tier III will result in one of the following possible conclusions:

- Mean test organism mortality in the dredged material is not statistically greater than in the reference sediment, or does not exceed mean mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). Therefore, the dredged material is predicted not to be acutely toxic to benthic organisms. However, bioaccumulation of contaminants must also be considered.
- 2) Mean test organism mortality in the dredged material is statistically greater than in the reference sediment *and* exceeds mortality in the reference sediment by at least 10 percentage points (or 20 percentage points for amphipods). In this case, the dredged material has the potential to be acutely toxic to benthic organisms.

Results of the whole-sediment bioassays are summarized in Table 8-20. Results of the first round of the whole-sediment bioassays indicated that none of the mean survival values in the channel sediments was statistically lower than survival in Inside Site 104 sediments. None of the channel sediments was acutely toxic to either the estuarine polychaete (*Neanthes arenaceodentata*) or estuarine amphipod (*Leptocheirus plumulosus*). Therefore, the dredged material is not predicted to be acutely toxic to benthic organisms after placement occurs.

The whole sediment test results demonstrate that the sediment proposed for dredging is not predicted to be toxic to benthic organisms post-placement. The evaluation of benthic-effects for whole sediment bioassays is based on the Limiting Permissible Concentration (LPC). The LPC is defined as "...that concentration which will not cause unreasonable acute or chronic toxicity or sublethal adverse effects based on bioassay results using...appropriate sensitive marine organisms..." (USEPA/USACE 1991 and USEPA/USACE 1998). Based on the results of the whole sediment bioassays, the proposed dredged material from the channels is not significantly toxic to the tested benthic organisms. The statistical comparisons of the channel sediments to Inside Site 104 indicate that all of the channel sediments comply with Conclusion 1, above.

Because the whole-sediment test results indicate that the channel sediments are not acutely toxic to aquatic tests organisms, the channel sediments are expected to be suitable for occan placement

8.3.3 Conclusions

In summary, conditions that would be expected to produce adverse effects in the water column at a placement site exist for a short-duration (minutes to a few hours following placement). Laboratory elutriate tests represent continuous exposure periods (48-96 hours) that greatly exceed the exposure periods that would be expected in the field. The laboratory elutriate tests provide conservative estimates of the potential for adverse water column effects. The conservatism is compounded by multiplying the LC50 concentration by a factor 0.01 for use in the mixing model. Overall, results of the whole-sediment toxicity tests are considered to be much more significant measures of the potential for adverse effects as a result of dredged material placement. Post-placement, benthic organisms and communities will be exposed to dredged material for weeks, months, or years, in comparison to the minutes or few hours of exposure experienced by organisms in the water column during the placement event. No toxicity was observed in the whole-sediment tests for the channel sediments, indicating that there would be little potential for long-term, adverse effects following open-water placement. The water column toxicity that could occur during the placement event will be short-term and localized. In addition, placement would occur during a time period when larval organisms (such as those testes in the laboratory tests) would not be expected to occur in the water column.

TABLE 8-1 SUMMARY OF TOXICITY TESTING SCHEDULE

| | | | WA | WATER COLUMN TESTING | | | WHOLE SEDIMENT TESTING | | |
|---------------|--------------------------|----------------------|---------------------|--------------------------|-----------------------|----------------------|-----------------------------|----------------------------|------------------------------------|
| | | | opossum shrimp | sheepshead minnow | blue mussel | inland silverside | estuarine polychaete | estuarine amphipod | opossum shrimp |
| TEST ROUND | DATES | TEST SEDIMENT | Mysidopsis bahia | Cyprinodon variegatus | <i>Mytilus</i> sp. | Menidia beryllina | Neanthes arenaceodentata | Leptocheirus plumulosus | Mysidopsis bahia ^(a) |
| 1 | September – | Inside Site 104 | х | x | | | x | X | |
| | November 1999 | Approach Channels | X | X | | | X | Х | |
| | December 1999– | Inside Site 104 | х | x | X | x | x | X | |
| 2 | February 2000 | Outside Site 104 | X | x | X | X | X | Х | |
| | | Approach Channels | | | X | x | | | |
| 3 | February – March 2000 | Ocean Reference | x | | x | X | | X ^(b) | X |

(a) Mysidopsis bahia tested in conjunction with Woodrow Wilson Bridge testing, Neanthes arenaceodentata not tested for Ocean Reference.

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(b) Lower Chesapeake Bay control sediment substituted for Ocean Reference Site sediment (as requested by USEPA Region III Philadelphia) in Leptocheirus tests to minimize potential grain-size effects

TABLE 8-2 SAMPLE COMPOSITES FOR ECOTOXICOLOGICAL TESTING

| Sampling Reach | Station | Sediment Volume | Composite |
|----------------------------|-----------------------------------|-----------------|---------------------------------------|
| | | (gallons) | Sample |
| Inside Placement Site 104 | KI-3 | 4 | KI-TOX |
| | KI-5 | 4 | |
| | KI-7-Ref | 4 | |
| | KI-S-1 | 4 | |
| · | KI-S-2 | 4 | |
| Outside Placement Site 104 | KI-11 | 5 | KI-OUT-TOX |
| | KI-14 | 5 | |
| | KI-15 | 5 | |
| | KI-16 | 5 | · · · · · · · · · · · · · · · · · · · |
| Brewerton Eastern | BE-1 | 5. | BE-TOX |
| Extension | BE-2 | 5 | |
| | BE-3 | 5 | |
| | BE-4 | 5 | |
| C&D Approach Channel | CD-001VC | 10 | CD-TOX |
| | CD-002VC | 10 | |
| · · · | CD003 | 5 | |
| | CD004 | 5 | |
| | CD005 | 5 | |
| | CD006 | <u> </u> | |
| Craighill | CRI | 7 | CR-TOX |
| | <u> </u> | 1 | |
| | | | |
| Craighill Angle–East | CRA-E-001VC | / | CRA-E-IUX |
| | | / | |
| | | - <u> </u> | CDA W TOV |
| Craighill Angle-west | | / / | UKA-W-IUA |
| | $\frac{CRA-W-002VC}{CRA-W-003VC}$ | 7 | |
| | | <u> </u> | CPE TOY |
| Craignill Entrance | CRE-001 VC | 5 | UKE-IUA |
| | CRF-003VC | 5 | |
| i | CRE-003 VC | 5 | |
| Cutoff Angle | | 7 | CUT-TOX |
| | CUT2 | 7 | 001 1011 |
| · · · · | CUT3 | 7 | |
| Swan Point | SWP-001VC | 3.5 | SWP-TOX |
| | SWP-002VC | 3.5 | |
| l f | SWP-003VC | 3.5 | |
| | SWP-004VC | 3.5 | |
| - E | SWP-005VC | 3.5 | |
| [| SWP-006VC | 3.5 | |
| Tolchester Channel – | TLC-005VC | 3.5 | TLC-N-TOX |
| North | TLC-006VC | 3.5 | |
| | TLC-007VC | 3.5 | |
| | TLC-008VC | 3.5 | |
| | TLC-009VC | 3.5 | |
| | TLC-010VC | 3.5 | |
| Tolchester Channel – South | TLC-001VC | 5 | TLC-S-TOX |
| | TLC-002VC | 5 | ١ |
| L | TLC-003VC | 5 | |
| | TLC-004VC | 5 | · · |
| Tolchester Straightening | TLS-001VC | 10 | TLS-TOX |
| | TLS-002VC | 10 | |

TABLE 8-3A SUMMARY OF COLLECTION AND RECEIPT INFORMATION FOR SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104

| Sample | Sample | Sample | EA Accession | Collection | Receipt |
|-----------------------|----------------------------------|-------------|---------------------|-------------------------|------------------------|
| Identification | Description | <u>Type</u> | <u>Number</u> | <u>Time and Date</u> | Time and Date |
| CRE | Craighill Entrance | Sediment | AT9-1430 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CR | Craighill | Sediment | AT9-1431 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRA-E | Craighill Angle East | Sediment | AT9-1432 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| · CRA-W | Craighill Angle West | Sediment | AT9-1433 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRU | Craighill Upper Range | Sediment | AT9-1434 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CUT | Cutoff Angle | Sediment | AT9-1435 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLC-S | Tolchester Channel South | Sediment | AT9-1436 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLC-N | Tolchester Channel North | Sediment | AT9-1437 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| TLS | Tolchester Straightening | Sediment | AT9-1438 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| BE | Brewerton Eastern Extension | Sediment | AT9-1439 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| SWP | Swan Point | Sediment | AT9-1440 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CD-VC | C&D Approaches - Cores | Sediment | AT9-1441 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CD | C&D Approaches - Surficial Grabs | Sediment | AT9-1442 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| KI - Reference | Inside Site 104 | Sediment | AT9-1443 | 1600, 14 October 1999 | 1600, 14 October 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1444 | 1052, 19 September 1999 | 1500, 13 October 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1611 | 1118, 22 November 1999 | 1515, 22 November 1999 |
| CR | Craighill | Site Water | AT9-1445 | 1205, 28 September 1999 | 1500, 13 October 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1446 | 1315, 18 September 1999 | 1500, 13 October 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1609 | 1029, 22 November 1999 | 1515, 22 November 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1447 | 1645, 18 September 1999 | 1500, 13 October 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1610 | 1041, 22 November 1999 | 1515, 22 November 1999 |
| CRU | Craighill Upper Range | Site Water | AT9-1448 | 1020, 28 September 1999 | 1500, 13 October 1999 |

TABLE 8-3A (CONTINUED)

| Sample | Sample | Sample | EA Accession | Collection | Receipt |
|-----------------------|----------------------------------|-------------|--------------|-------------------------|------------------------|
| Identification | Description | <u>Type</u> | Number | Time and Date | <u>Time and Date</u> |
| CUT | Cutoff Angle | Site Water | AT9-1449 | 0835, 28 September 1999 | 1500, 13 October 1999 |
| CUT | Cutoff Angle | Site Water | AT9-1608 | 0958, 22 November 1999 | 1515, 22 November 1999 |
| TLC-S | Tolchester Channel South | Site Water | AT9-1450 | 0850, 20 September 1999 | 1500, 13 October 1999 |
| TLC-N | Tolchester Channel North | Site Water | AT9-1451 | 1130, 20 September 1999 | 1500, 13 October 1999 |
| TLS | Tolchester Straightening | Site Water | AT9-1452 | 1520, 27 September 1999 | 1500, 13 October 1999 |
| BE | Brewerton Eastern Extension | Site Water | AT9-1453 | 1645, 27 September 1999 | 1500, 13 October 1999 |
| SWP | Swan Point | Site Water | AT9-1454 | 1530, 19 September 1999 | 1500, 13 October 1999 |
| CD-VC | C&D Approaches - Cores | Site Water | AT9-1455 | 1230, 21 September 1999 | 1500, 13 October 1999 |
| CD-SW | C&D Approaches - Surficial Grabs | Site Water | AT9-1456 | 1420, 27 September 1999 | 1500, 13 October 1999 |
| KI – Reference | Inside Site 104 | Site Water | AT9-1457 | 1540, 28 September 1999 | 1500, 13 October 1999 |
| KI – Reference | Inside Site 104 | Site Water | AT9-1612 | 1147, 22 November 1999 | 1515, 22 November 1999 |
| | | | | | |
| CRE | Craighill Entrance | Sediment | AT9-1702 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRE | Craighill Entrance | Site Water | AT9-1703 | 1300, 15 December 1999 | 0830, 20 December 1999 |
| ĊR | Craighill | Sediment | AT9-1704 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CR | Craighill | Site Water | AT9-1705 | 1600, 13 December 1999 | 0830, 20 December 1999 |
| CRA-E | Craighill Angle East | Sediment | AT9-1706 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRA-E | Craighill Angle East | Site Water | AT9-1707 | 0930, 15 December 1999 | 0830, 20 December 1999 |
| CRA-W | Craighill Angle West | Sediment | AT9-1708 | 1430, 16 December 1999 | 0830, 20 December 1999 |
| CRA-W | Craighill Angle West | Site Water | AT9-1709 | 0950, 14 December 1999 | 0830, 20 December 1999 |
| CRU | Craighill Upper Range | Sediment | AT9-1710 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CRU | Craighill Upper Range | Site Water | AT9-1711 | 0915, 15 December 1999 | 0830, 20 December 1999 |
| CUT | Cutoff Angle | Sediment | AT9-1712 | (a), 17 December 1999 | 0830, 20 December 1999 |
| CUT | Cutoff Angle | Site Water | AT9-1713 | 1035, 15 December 1999 | 0830, 20 December 1999 |

(a) Time of collection not provided by sampler.

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| TABLE 8-3A (CONTINUED) | | | | | | |
|--------------------------|----------------------------------|-----------------------|-------------------------------|------------------------------------|---------------------------------|--|
| Sample Identification | Sample Description | Sample <u>Type</u> | EA Accession <u>Number</u> | Collection <u>Time and Date</u> | Receipt <u>Time and Date</u> | |
| TLC-S | Tolchester Channel South | Sediment | AT9-1714 | 1310, 17 December 1999 | 0830, 20 December 1999 | |
| TLC-S | Tolchester Channel South | Site Water | AT9-1715 | 0950, 9 December 1999 | 0830, 20 December 1999 | |
| TLC-N | Tolchester Channel North | Sediment | AT9-1716 | 1040, 17 December 1999 | 0830, 20 December 1999 | |
| TLC-N | Tolchester Channel North | Site Water | AT9-1717 | 1315, 8 December 1999 | 0830, 20 December 1999 | |
| TLS | Tolchester Straightening | Sediment | AT9-1718 | 1310, 17 December 1999 | 0830, 20 December 1999 | |
| TLS | Tolchester Straightening | Site Water | AT9-1719 | 0830, 9 December 1999 | 0830, 20 December 1999 | |
| BE | Brewerton Eastern Extension | Sediment | AT9-1720 | (a), 17 December 1999 | 0830, 20 December 1999 | |
| BE | Brewerton Eastern Extension | Site Water | AT9-1721 | 1420, 14 December 1999 | 0830, 20 December 1999 | |
| SWP | Swan Point | Sediment | AT9-1722 | 1420, 17 December 1999 | 0830, 20 December 1999 | |
| SWP | Swan Point | Site Water | AT9-1723 | 1145, 9 December 1999 | 0830, 20 December 1999 | |
| CD-VC | C&D Approaches - Cores | Sediment | AT9-1724 | 1040, 17 December 1999 | 0830, 20 December 1999 | |
| CD-VC | C&D Approaches - Cores | Site Water | AT9-1725 | 0845, 8 December 1999 | 0830, 20 December 1999 | |
| CD | C&D Approaches - Surficial Grabs | Sediment | - AT9-1726 | (a), 17 December 1999 | 0830, 20 December 1999 | |
| CD | C&D Approaches - Surficial Grabs | Site Water | AT9-1727 | 1210, 14 December 1999 | 0830, 20 December 1999 | |
| KI – Reference | Inside Site 104 | Sediment | AT9-1728 | (a), 17 December 1999 | 0830, 20 December 1999 | |
| KI – Reference | Inside Site 104 | Site Water | AT9-1729 | 1430, 13 December 1999 | 0830, 20 December 1999 | |
| KI-OUT - Reference | Outside Site 104 | Sediment | AT9-1730 | (a), 17 December 1999 | 0830, 20 December 1999 | |
| KI-OUT - Reference | Outside Site 104 | Site Water | AT9-1731 | 1235, 13 December 1999 | 0830, 20 December 1999 | |

(a) Time of collection not provided by sampler.

TABLE 8-3BSUMMARY OF COLLECTION AND RECEIPT INFORMATION FOR SAMPLESFROM THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| Sample <u>Identification</u> | Sample <u>Type</u> | EA Accession <u>Number</u> | Collection <u>Time and Date</u> | Receipt <u>Time and Date</u> |
|---------------------------------|-----------------------|-------------------------------|------------------------------------|---------------------------------|
| Lower Chesapeake Bay Control | Sediment | AT0-050 | 1200, 12 January 2000 | 0910, 19 January 2000 |
| Ocean Reference | Sediment | AT0-097 | 1130, 1 February 2000 | 0903, 3 February 2000 |
| Ocean Reference | Water | AT0-098 | 1130, 1 February 2000 | 0903, 3 February 2000 |

TABLE 8-4AWATER QUALITY PARAMETERS MEASURED DURING Mysidopsis bahia
(OPOSSUM SHRIMP) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|------------------|-------------|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | |
| | (°C) | рН | (mg/L) | (ppt) | | |
| Inside Site 104 | 19.8 (±0.6) | 8.2 (±0.1) | 6.9 (±0.5) | 29.0 (±0.7) | | |
| Brewerton Eastern Extension | 20.1 (±0.4) | 8.2 (±0.1) | 6.8 (±0.4) | 29.0 (±0.6) | | |
| C&D Approaches - Cores | 20.4 (±0.5) | 8.1 (±0.2) | 6.7 (±0.5) | 29.6 (±1.2) | | |
| C&D Appr Surf. Grabs | 20.6 (±0.7) | 8.2 (±0.1) | 7.0 (±0.4) | 29.1 (±0.7) | | |
| Craighill | 20.0 (±0.6) | 8.1 (±0.1) | 6.6 (±0.4) | 29.2 (±0.6) | | |
| Craighill Angle East | 20.2 (±0.6) | 8.1 (±0.2) | 6.9 (±0.3) | 29.3 (±0.7) | | |
| Craighill Angle West | 19.9 (±0.3) | 8.2 (±0.1) | 6.7 (±0.4) | 29.4 (±0.7) | | |
| Craighill Entrance | 20.2 (±0.4) | 8.1 (±0.1) | 6.7 (±0.6) | 29.1 (±0.8) | | |
| Craighill Upper Range | 19.6 (±0.4) | 8.1 (±0.1) | 6.8 (±0.4) | 29.3 (±0.7) | | |
| Cutoff Angle | 20.4 (±0.6) | 8.1 (±0.1) | 6.8 (±0.4) | 29.2 (±0.5) | | |
| Swan Point | 20.0 (±0.5) | 8.2 (±0.1) | 6.9 (±0.8) | 29.4 (±0.5) | | |
| Tolchester North | 19.4 (±0.5) | 8.1 (±0.2) | 6.9 (±0.5) | 28.9 (±0.3) | | |
| Tolchester South | 20.0 (±0.6) | 8.1 (±0.1) | 6.9 (±0.4) | 29.2 (±0.8) | | |
| Tolchester Straightening | 20.0 (±0.7) | 8.1 (±0.1) | 6.8 (±0.4) | 29.0 (±0.6) | | |

TABLE 8-4BWATER QUALITY PARAMETERS MEASURED DURING Mysidopsis bahia
(OPOSSUM SHRIMP) ELETRIATE TOXICITY TESTING ON SAMPLES
FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT
SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (±Standard Deviation) | | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|---------------------------|--|--|--|
| Test Élutriate | Temperature (°C) | рП | Dissolved Oxygen (mg/L) | Salinity <u>(</u> ppt) | | | |
| Inside Site 104 | 20.2 (±0.6) | 8.3 (±0.1) | 7.0 (±0.3) | 30.5 (±1.7) | | | |
| Outside Site 104 | 19.8 (±1.0) | 8.3 (±0.1) | 7.0 (±0.3) | 30.8 (±1.6) | | | |
| Laboratory Control Sediment | 19.5 (±1.0) | 8.3 (±0.1) | 7.1 (±0.4) | 30.4 (±1.8) | | | |

TABLE 8-4C SUMMARY OF WATER QUALITY PARAMETERS MEASURED
DURING Mysidopsis bahia (OPOSSUM SHRIMP) ELUTRIATE
TOXICITY TESTING ON SAMPLES FROM THE NORFOLK OCEAN
DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000)

| | Water Quality Parameters – Mean (±Standard Deviation) | | | | | | | |
|--------------------|---|------------|----------------------------|-------------------|--|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | | |
| Ocean Reference | 19.3 (±0.5) | 8.0 (±0.2) | 6.7 (±0.9) | 31.6 (±1.6) | | | | |

TABLE 8-5AWATER QUALITY PARAMETERS MEASURED DURING Cyprinodon
variegatus (SHEEPSHEAD MINNOW) ELUTRIATE TOXICITY TESTING ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| Test Elutrists | Mean (±Standard Deviation) | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|
| | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | |
| Inside Site 104 | 20.8 (±0.8) | 8.2 (±0.1) | 7.1 (±0.3) | 29.0 (±0.5) | | |
| Brewerton Eastern Extension | 20.2 (±0.7) | 8.2 (±0.1) | 6.9 (±0.4) | 28.9 (±0.5) | | |
| C&D Approaches - Cores | 20.5 (±0.5) | 8.2 (±0.2) | 7.0 (±0.2) | 28.9 (±0.5) | | |
| C&D Appr Surf. Grabs | 20.4 (±0.4) | 8.1 (±0.1) | 7.0 (±0.3) | 28.8 (±0.4) | | |
| Craighill | 20.6 (±0.6) | 8.1 (±0.1) | 6.9 (±0.4) | 29.3 (±0.8) | | |
| Craighill Angle East | 20.2 (±0.7) | 8.2 (±0.1) | 7.1 (±0.3) | 28.9 (±0.3) | | |
| Craighill Angle West | 20.1 (±0.6) | 8.2 (±0.1) | 7.2 (±0.2) | 29.0 (±0.3) | | |
| Craighill Entrance | 20.0 (±0.5) | 8.2 (±0.1) | 7.0 (±0.2) | 28.9 (±0.6) | | |
| Craighill Upper Range | 20.2 (±0.5) | 8.1 (±0.1) | 7.0 (±0.2) | 29.0 (±0.4) | | |
| Cutoff Angle | 20.5 (±0.5) | 8.2 (±0.1) | 7.0 (±0.2) | 29.0 (±0.6) | | |
| Swan Point | 20.3 (±0.3) | 8.2 (±0.1) | 7.1 (±0.2) | 29.6 (±0.7) | | |
| Tolchester North | 20.7 (±0.5) | 8.2 (±0.2) | 7.0 (±0.2) | 29.2 (±0.7) | | |
| Tolchester South | 20.3 (±0.6) | 8.1 (±0.1) | 7.1 (±0.3) | 29.4 (±0.9) | | |
| Tolchester Straightening | 20.8 (±0.5) | 8.1 (±0.2) | 6.9 (±0.4) | 29.1 (±0.4) | | |

 γ_{i}

TABLE 8-5BWATER QUALITY PARAMETERS MEASURED DURING Cyprinodon
variegatus (SHEEPSHEAD MINNOW) ELUTRIATE TOXICITY TESTING ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (±Standard Deviation) | | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Inside Site 104 | 19.6 (±1.0) | 8.3 (±0.1) | 7.2 (±0.3) | 30.7 (±1.9) | | | |
| Outside Site 104 | 20.0 (±0.9) | 8.3 (±0.1) | 7.1 (±0.2) | 30.8 (±1.6) | | | |
| Laboratory Control Sediment | 19.7 (±1.1) | 8.3 (±0.1) | 7.2 (±0.3) | 30.4 (±1.9) | | | |

TABLE 8-6A WATER QUALITY PARAMETERS MEASURED DURING Menidia beryllina
(INLAND SILVERSIDE) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT
SITE 104 (Round 2: Dec. 1999–Feb. 2000)

| | Mean (± Standard Deviation) | | | | | |
|---------------------------------------|-----------------------------|------------|------------------|-------------|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | |
| | (°C) | рН | (mg/L) | (ppt) | | |
| | , | <u>.</u> | | | | |
| Inside Site 104 | 20.2 (±0.8) | 8.2 (±0.1) | 5.7 (±1.3) | 33.2 (±2.9) | | |
| | | | | | | |
| Outside Site 104 | 20.6(+0.7) | 83(+01) | 57(+11) | 323(+27) | | |
| | 20.0 (10.7) | 0.5 (10.1) | 5.7 (11.1) | 52.5 (±2.7) | | |
| Decementary Francisco | 20.4 (1.1) | | | | | |
| Brewerton Eastern | 20.4 (±1.1) | 8.3 (±0.1) | 6.6 (±0.6) | 33.8 (±3.7) | | |
| Extension | | | | | | |
| | | | | | | |
| C&D Approaches - Cores | 20.9 (±0.8) | 8.1 (±0.1) | 6.3 (±1.0) | 29.8 (±1.1) | | |
| | | | ····· | | | |
| C&D Appr Surf. Grabs | 20.4 (±1.1) | 8.3 (±0.1) | 6.7 (±0.6) | 32.1 (±3.2) | | |
| | | | | | | |
| | | | | | | |
| Craighill | 20.2 (±0.5) | 8.2 (±0.1) | 6.6 (±0.5) | 32.1 (±1.6) | | |
| · · · · · · · · · · · · · · · · · · · | | | | | | |
| Craighill Angle East | 20.6 (±0.8) | 8.2 (±0.1) | 6.5 (±0.9) | 29.8 (±1.0) | | |
| | | | | | | |
| Craighill Angle West | 20.7 (+1.1) | 8.2 (+0.2) | 6.5 (+0.7) | 30.7 (+1.9) | | |
| 0 0 | | | | | | |
| Craighill Entrance | 100(17) | 83(.02) | 50(14) | 220(-20) | | |
| | 19.9 (±1.7) | 8.3 (±0.2) | J.9 (±1.4) | 33.0 (±3.0) | | |
| | | | | | | |
| Craighill Upper Range | 20.5 (±0.6) | 8.3 (±0.1) | 6.6 (±0.6) | 33.1 (±3.2) | | |
| | | | | | | |
| Cutoff Angle | 19.0 (±0.7) | 8.3 (±0.1) | 6.7 (±0.6) | 29.3 (±1.1) | | |
| | | | | | | |
| Swan Point | 20.9 (±0.9) | 8.2 (±0.1) | 5.8 (±1.3) | 35.1 (±5.6) | | |
| | | | | | | |
| Tolchester North | 193(+06) | 81(+01) | 66(+0.7) | 20.3 (+0.0) | | |
| Torenester North | 19.5 (±0.0) | 0.1 (±0.1) | 0.0 (±0.7) | 29.5 (±0.9) | | |
| | | | | | | |
| I olchester South | 21.1 (±0.8) | 8.2 (±0.1) | 6.7 (±0.5) | 29.5 (±1.7) | | |
| | | | | | | |
| Tolchester Straightening | 20.3 (±0.7) | 8.0 (±0.1) | 6.3 (±1.2) | 29.8 (±0.4) | | |
| | | | | | | |

TABLE 8-6BSUMMARY OF WATER QUALITY PARAMETERS MEASURED
DURING Menidia beryllina (INLAND SILVERSIDE) ELUTRIATE
TOXICITY TESTING ON SAMPLES FROM THE NORFOLK OCEAN
DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000)

| | Water Quality Parameters – Mean (±Standard Deviation) | | | | | | |
|--------------------|---|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Ocean Reference | 19.8 (±0.3) | 8.1 (±0.1) | 6.7 (±0.9) | 31.7 (±1.5) | | | |

TABLE 8-7AWATER QUALITY PARAMETERS MEASURED DURING Mytilus sp.
(BLUE MUSSEL) ELUTRIATE TOXICITY TESTING ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE
104 (Round 2: Dec. 1999-Feb. 2000) (a)

| | Mean (±Staudard Deviation) | | | | | | |
|---------------------------------------|----------------------------|---------------------------------------|------------------|---------------------------------------|--|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | | |
| | (°C) | pН | (mg/L) | (ppt) | | | |
| | | | | | | | |
| Inside Site 104 | 16.0 (±0.7) | 8.4 (±0.1) | 8.2 (±0.2) | 30.2 (±0.3) | | | |
| | | · · · · · · · · · · · · · · · · · · · | | | | | |
| Outside Site 104 | 16.0 (±0.4) | 8.4 (±0.1) | 8.2 (±0.2) | 30.2 (±0.2) | | | |
| | | | | | | | |
| Brewerton Eastern | 160(+07) | 84(+02) | 82(+0.2) | 29.8 (+0.7) | | | |
| Extension | 10.0 (10.7) | 0.1 (10.2) | 0.2 (10.2) | 20.0 (10.0) | | | |
| · · · · · · · · · · · · · · · · · · · | | | | | | | |
| C&D Annroaches - Cores | 163(+02) | 83(+02) | 82(+02) | 26.2 (+4.7) | | | |
| | 10.5 (10.2) | | 0.2 (10.2) | 20.2 (1) | | | |
| C&D Annu Suuf Cucha | 15.0 (.0.7) | 8 4 (, 0 2) | 82(.02) | 202(02) | | | |
| C&D Appr Suri. Grabs | 13.9 (±0.7) | 8.4 (±0.2) | 8.2 (±0.2) | 30.2 (±0.3) | | | |
| | | · · · · · · | ···· | | | | |
| Craighill | 16.2 (±0.6) | 8.1 (±0.1) | 8.2 (±0.2) | 30.1 (±0.2) | | | |
| | | | · · · | | | | |
| Craighill Angle East | 162(+07) | 82(+01) | 8.1 (+0.2) | 30.0 (+0.2) | | | |
| | 10.2 (10.7) | 0.2 (10.1) | 011 (1012) | | | | |
| Craighill Angle West | 157(.06) | 82(102) | 82(102) | 30.0 (. 0.2) | | | |
| Craiginii Angle West | 15.7 (±0.0) | 8.2 (±0.2) | 0.2 (±0.2) | 30.0 (±0.2) | | | |
| | | | | | | | |
| Craigniii Entrance | $16.3(\pm 0.5)$ | 8.2 (±0.2) | 8.1 (±0.1) | 30.0 (±0.2) | | | |
| | | | ·. | | | | |
| Craighill Upper Range | 16.1 (±0.7) | 8.2 (±0.1) | 8.1 (±0.1) | 30.0 (±0.4) | | | |
| | | | | | | | |
| Cutoff Angle | 16.1 (±0.7) | 8.3 (±0.2) | 8.2 (±0.2) | 30.2 (±0.3) | | | |
| | | | | | | | |
| Swan Point | 16.5 (±0.1) | 8.2 (±0.2) | 8.2 (±0.2) | 29.0 (±2.0) | | | |
| · · · · · · · · · · · · · · · · · · · | | | | · · · · · · · · · · · · · · · · · · · | | | |
| Tolchester North | 16.7 (±0.2) | 8.2 (±0.2) | 8.1 (±0.2) | 29.0 (±1.5) | | | |
| | | | · · · · · | <u> </u> | | | |
| Tolchester South | 160(+0.8) | 82(+01) | 8 2 (+0 2) | 30.0 (+0.2) | | | |
| | 10.0 (10.0) | | 0.2 (10.2) | JU.0 (±0.2) | | | |
| Talah satar Staalah tarima | 160(00) | 82(01) | 82(01) | 20.2 (. 0.2) | | | |
| i oicnester Straightening | 10.0 (±0.8) | 8.3 (±0.1) | 8.2 (±0.1) | 30.2 (±0.3) | | | |

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-7BWATER QUALITY PARAMETERS MEASURED DURING Mytilus sp.
(BLUE MUSSEL) ELUTRIATE TOXICITY TESTING ON SAMPLES
FROM NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000) ^(a)

| | Mean (±Standard Deviation) | | | | | | |
|--------------------|----------------------------|-----------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Ocean Reference | 15.8(±0.2) | 8.0(±0.1) | 8.1 (±0.2) | 30.9 (±0.3) | | | |

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-8AWATER QUALITY PARAMETERS MEASURED DURING 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Neanthes arenaceodentata
(ESTUARINE POLYCHAETE) ON SAMPLES FROM BALTIMORE HARBOR
APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | | | |
|---------------------------------------|---------------------------------------|------------|---------------------------------------|-----------------|--|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | | |
| | (°C) | pH | (mg/L) | (ppt) | | | |
| Laboratory Control | · · · · · · · · · · · · · · · · · · · | | | <u></u> | | | |
| Sediment | 102(.04) | 79(02) | (2(00) | 20.0 (1.0) | | | |
| | 19.2 (±0.4) | 7.8 (±0.3) | 6.2 (±0.6) | 30.9 (±1.0) | | | |
| Incide Site 104 | | | | | | | |
| Inside Site 104 | 19.2 (±0.6) | 7.9 (±0.2) | 5.6 (±0.4) | 31.0 (±0.7) | | | |
| | | | | | | | |
| | | | | | | | |
| Brewerton Eastern | 19.1 (±0.4) | 7.9 (+0.2) | 5.9 (+0.7) | 31.0(+0.8) | | | |
| Extension | | | | | | | |
| | · · · · · · · · · · · · · · · · · · | | | | | | |
| C&D Approaches - Cores | 19.0 (+0.5) | 7.9 (+0.2) | 6.2(+0.5) | 308(+08) | | | |
| · · · · · · · · · · · · · · · · · · · | 1710 (1010) | (10.2) | 0.2 (10.0) | 50.0 (10.0) | | | |
| C&D Appr Surf. Grabs | | 22400 | | | | | |
| | 19.0 (±0.6) | /./(±0.4) | 6.1 (±0.7) | 31.1 (±0.8) | | | |
| | | | • | | | | |
| Craighill | 191(+05) | 78(+02) | 62(+04) | 317(+13) | | | |
| · · · · · · · · · · · · · · · · · · · | 17.1 (10.5) | 7.0 (10.2) | 0.2 (10.4) | | | | |
| Craighill Angle East | | | | | | | |
| | 19.1 (±0.4) | 7.8 (±0.3) | 5.7 (±0.7) | 31.4 (±1.0) | | | |
| | | | | | | | |
| Craighill Angle West | 19.1 (±0.4) | 7.8 (±0.2) | 5.7 (±0.6) | $31.3(\pm 1.1)$ | | | |
| | | | · · · · · · · · · · · · · · · · · · · | | | | |
| Craighill Entrance | 101(.04) | 70(02) | 57(05) | 21.5 (1.0) | | | |
| | 19.1 (±0.4) | 7.9 (±0.3) | 5.7 (±0.5) | 31.5 (±1.0) | | | |
| Craighill Upper Range | | | | | | | |
| Craigini Opper Kange | 19.1 (±0.4) | 7.9 (±0.2) | 6.1 (±0.6) | 31.6 (±1.1) | | | |
| | | | | | | | |
| Cutoff Angle | 19.1 (+0.4) | 7.8(+0.2) | 57(+0.6) | 313(+10) | | | |
| | | | | 51.5 (11.0) | | | |
| Swan Point | | | | 1 | | | |
| | 19.1 (±0.4) | 7.9 (±0.3) | 5.9 (±0.7) | 31.1 (±0.7) | | | |
| | · | | | | | | |
| i oicnester North | 19.1 (±0.4) | 7.7 (±0.3) | 5.6 (±0.5) | 31.2 (±0.9) | | | |
| | | | · · · · · · · · · · · · · · · · · · · | `´´ | | | |
| Tolchester South | 101(05) | 79(103) | | 31.9 (1.0) | | | |
| | 19.1 (±0.5) | 7.8 (±0.3) | 6.0 (±0.8) | 31.2 (±1.0) | | | |
| Tolohastan Straightanin | | | | | | | |
| Loicnester Straightening | 19.1 (±0.5) | 7.8 (±0.3) | 6.3 (±0.7) | 31.3 (±1.0) | | | |

TABLE 8-8BWATER QUALITY PARAMETERS MEASURED DURING ADDITIONAL
10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Neanthes
arenaceodentata (ESTUARINE POLYCHAETE) ON SAMPLES FROM
BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT
SITE 104 (Round 2: Dec. 1999-Feb.2000)

| | Mean (±Standard Deviation) | | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Inside Site 104 | 20.4 (±0.6) | 8.3 (±0.1) | 7.2 (±0.3) | 29.2 (±0.4) | | | |
| Outside Site 104 | 20.3 (±0.6) | 8.2 (±0.2) | 7.2 (±0.3) | 28.8 (±0.7) | | | |
| Laboratory Control Sediment | 20.5 (±0.4) | 8.2 (±0.2) | 7.2 (±0.3) | 29.6 (±1.9) | | | |

TABLE 8-9AWATER QUALITY PARAMETERS MEASURED DURING 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Leptocheirus plumulosus (ESTUARINE
AMPHIPOD) ON SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | Mean (±Standard Deviation) | | | | | | |
|--------------------------------|----------------------------|------------|------------------|-------------|--|--|--|
| Test Elutriate | Temperature | | Dissolved Oxygen | Salinity | | | |
| | (°C) | pН | (mg/L) | (ppt) | | | |
| Laboratory Control Sediment | 19.5 (±0.5) | 7.6 (±0.2) | .6.2 (±0.7) | 20.9 (±0.6) | | | |
| Inside Site 104 | 19.4 (±0.4) | 7.7 (±0.3) | 5.8 (±0.8) | 21.6 (±0.8) | | | |
| Brewerton Eastern Extension | 19.4 (±0.4) | 7.7 (±0.2) | 6.0 (±0.8) | 21.3 (±0.7) | | | |
| C&D Approaches - Cores | 19.3 (±0.4) | 7.6 (±0.2) | 5.9 (±0.7) | 20.9 (±0.6) | | | |
| C&D Appr Surf. Grabs | 19.3 (±0.4) | 7.7 (±0.2) | 6.3 (±1.0) | 21.0 (±0.7) | | | |
| Craighill | 19.6 (±0.5) | 7.7 (±0.2) | 6.0 (±0.6) | 21.3 (±0.7) | | | |
| Craighill Angle East | 19.5 (±0.5) | 7.5 (±0.2) | 5.1 (±1.0) | 21.2 (±0.6) | | | |
| Craighill Angle West | 19.5 (±0.5) | 7.6 (±0.2) | 5.0 (±0.9) | 21.2 (±0.6) | | | |
| Craighill Entrance | 19.5 (±0.4) | 7.6 (±0.2) | 5.5 (±1.1) | 21.5 (±0.6) | | | |
| Craighill Upper Range | 19.4 (±0.5) | 7.6 (±0.2) | 6.3 (±0.7) | 21.5 (±0.7) | | | |
| Cutoff Angle | 19.4 (±0.5) | 7.6 (±0.2) | 5.7 (±0.9) | 21.3 (±0.8) | | | |
| Swan Point | 19.3 (±0.3) | 7.6 (±0.2) | 5.8 (±0.9) | 21.3 (±0.7) | | | |
| Tolchester North | 19.3 (±0.4) | 7.5 (±0.2) | 5.3 (±0.9) | 21.1 (±0.6) | | | |
| Tolchester South | 19.4 (±0.3) | 7.6 (±0.2) | 5.7 (±0.6) | 21.3 (±0.7) | | | |
| Tolchester Straightening | 19.4 (±0.4) | 7.7 (±0.2) | 6.2 (±0.7) | 21.1 (±0.5) | | | |

TABLE 8-9BWATER QUALITY PARAMETERS MEASURED DURING ADDITIONAL
10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Leptocheirus
plumulosus (ESTUARINE AMPHIPOD) ON SAMPLES FROM BALTIMORE
HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

.

| | Mean (±Standard Deviation) | | | | | | |
|--------------------------------|----------------------------|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Inside Site 104 | 21.8 (±3.2) | 8.0 (±0.1) | 5.5 (±1.0) | 22.6 (±1.8) | | | |
| Outside Site 104 | 22.1 (±2.8) | 8.0 (±0.1) | 3.9 (±1.1) | 21.0 (±1.6) | | | |
| Laboratory Control Sediment | 21.9 (±3.1) | 7.9 (±0.2) | 5.7 (±0.7) | 20.0 (±2.2) | | | |

TABLE 8-9CSUMMARY OF WATER QUALITY PARAMETERS FROM 10-DAY WHOLE
SEDIMENT TOXICITY TESTING WITH Leptocheirus plumulosus (ESTUARINE
AMPHIPOD) FOR THE NORFOLK OCEAN DISPOSAL SITE REFERENCE
AREA (Round 3: Feb. 2000)

| · · | Mean (±Standard Deviation) | | | | | | |
|--|----------------------------|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Lower Chesapeake Bay Control ^(a) | 20.3 (±0.4) | 8.1 (±0.1) | 5.2 (±0.6) | 21.5 (±1.2) | | | |

(a) = Tested in place of Ocean Reference sediment due to potential for adverse grain size effects.

TABLE 8-10SUMMARY OF WATER QUALITY PARAMETERS FROM 10-DAY
WHOLE SEDIMENT TOXICITY TESTING WITH Mysidopsis bahia FOR
THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000)

| | Mcan (±Standard Deviation) | | | | | | |
|-----------------|----------------------------|------------|----------------------------|-------------------|--|--|--|
| Test Elutriate | Temperature (°C) | рН | Dissolved Oxygen (mg/L) | Salinity (ppt) | | | |
| Ocean Reference | 20.5 (±0.6) | 8.0 (±0.2) | 6.2 (±0.5) | 31.5 (±1.3) | | | |

TABLE 8-11A RESULTS OF ELUTRIATE TOXICITY TESTING WITH Mysidopsis bahia (OPOSSUM SHRIMP) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | 96-Hour Survival (%)* | | | | 96-Hour | 96-Hour |
|-----------------------------------|-----------------------|------|-------------|-------------------|-----------------------------------|------------------------|
| Test Elutriate | Lab | Pe | rcent Eluti | ·iate | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 100 | 98 | 98 | 100 | > 100 | 100 |
| Brewerton Eastern Extension | 98 | 92 | 96 | 78 ^(a) | > 100 | 50 |
| C&D Approaches – Cores | 94 | 98 | 96 | 86 | > 100 | 100 |
| C&D Approaches – Surface Grabs | 100 | 92 | 98 | 84 ^(a) | > 100 | 50 |
| Craighill | 98 | .94 | 98 | 94 | > 100 | 100 |
| Craighill Angle East | 96 | 92 | 92 | 92 | > 100 | 100 |
| Craighill Angle West | 98 | 98 | 98 | 76 ^(a) | > 100 | 50 |
| Craighill Entrance | 98 | 98 | 98 | 54 ^(a) | > 100 | 50 |
| Craighill Upper Range | 100 | 98 | 98 | 98 | > 100 | 100 |
| Cutoff Angle | 98 | 90 | 98 | 92 | > 100 | 100 |
| Swan Point | 96 | 94 | 90 | 84 | > 100 | 100 |
| Tolchester North | 98 | 96 | 98 | 94 | > 100 | 100 |
| Tolchester South | 100 | 98 | 92 | 83 ^(a) | > 100 | 100 ^(b) |
| Tolchester Straightening | 96 | 92 | 98 | 94 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

(b) Multiple comparison (100, 50, 10 vs. control) shows no significance; single comparison (50 vs. control) shows significance (NOAEC = 10).

TABLE 8-11BRESULTS OF ADDITIONAL ELUTRIATE TOXICITY
TESTING WITH Mysidopsis bahia (OPOSSUM SHRIMP) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| | 96-Hour Survival (%)* | | | | 96-Hour | 96-Hour |
|--------------------------------|-----------------------|------|------|------------------|---------------|---------------|
| Test Elutriate | Lab Percent Elutriate | | | LC ₅₀ | NOAEC | |
| | Control | 10 % | 50 % | 100% | (% elutriate) | (% elutriate) |
| Inside 104 | 98 | 100 | 96 | 1 <u>00</u> | > 100 | 100 |
| Outside 104 | 96 | 98 | 94 | 94 | > 100 | 100 |
| Laboratory Control Sediment | 94 | • 92 | 90 | 88 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

TABLE 8-11C RESULTS OF ELUTRIATE TOXICITY TESTING WITH Mysidopsis bahia (OPOSSUM SHRIMP) ON SAMPLES FROM THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000)

| Test Elutriate | 9 | 6-Hour Survival (%)* | | | 96-Hour Survival (%)* | | | 96-Hour | 96-Hour |
|-----------------|----------------|----------------------|--------------------|--------------|-----------------------------------|------------------------|--|---------|---------|
| | Lab Control | Per 10 % | cent Elutr 50 % | iate 100% | LC ₅₀ (% elutriate) | NOAEC (% elutriate) | | | |
| Ocean Reference | 98 | 100 | 98 | 100 | > 100 | 100 | | | |

*Survival based on a mean of 5 replicate tests.

TABLE 8-12A RESULTS OF ELUTRIATE TOXICITY TESTING WITH Cyprinodon variegatus (SHEEPSHEAD MINNOW) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| | | 96-Hour S | urvival (% | 96-Hour | 96-Hour | |
|-----------------------------------|---------|-----------|------------|---------|-----------------------------------|------------------------|
| Test Elutriate | Lab | Pe | rcent Elut | riate | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 100 | 100 | 100 | 100 | > 100 | 100 |
| Brewerton Eastern Extension | 100 | 100 | 100 | 96 | > 100 | 100 |
| C&D Approaches – Cores | 100 | 100 | 96 | 98 | > 100 | 100 |
| C&D Approaches – Surface Grabs | 100 | 98 | 100 | 98 | > 100 | 100 |
| Craighill | 98 | 98 | 94 | 100 | > 100 | 100 |
| Craighill Angle East | 100 | 98 | 96 | 100 | > 100 | 100 |
| Craighill Angle West | 100 | 94 | 96 | 96 | > 100 | 100 |
| Craighill Entrance | 100 | 100 | 98 | 98 | > 100 | 100 |
| Craighill Upper Range | 100 | 98 | 98 | 98 | > 100 | 100 |
| Cutoff Angle | 100 | 92 | 98 . | 92 | > 100 | 100 |
| Swan Point | 100 | 98 | 100 | 94 | > 100 | 100 |
| Tolchester North | 100 | 100 | 100 | 98 | > 100 | 100 |
| Tolchester South | 98 | 98 | 98 | 98 | > 100 | 100 |
| Tolchester Straightening | 100 | 100 | 98 | 98 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

TABLE 8-12BRESULTS OF ADDITIONAL ELUTRIATE TOXICITY TESTING
WITH Cyprinodon variegatus (SHEEPSHEAD MINNOW) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH
CHANNELS AND PLACEMENT SITE 104
(Round 2: Dec. 1999-Feb. 2000)

| | 9 | 6-Hour Sı | ırvival (%) |)* | 96-Hour LC ₅₀ (% elutriate) | 96-Hour NOAEC (% elutriate) |
|--------------------------------|---------|-----------|-------------|-------|--|-----------------------------------|
| Test Elutriate | Lab | Pe | rcent Elutr | ·iate | | |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 100 | 100 | 100 | 100 | > 100 | 100 |
| Outside 104 | 98 | 100 | 100 | 96 | > 100 | 100 |
| Laboratory Control Sediment | 100 | 100 | 100 | 100 | > 100 | 100 |

*Survival based on a mean of 5 repliocate tests.

TABLE 8-13A RESULTS OF ELUTRIATE TOXICITY TESTING WITH *Menidia beryllina* (INLAND SILVERSIDE) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)

| | 96-Hour Survival (%)* | | | | 96-Hour | 96-Hour |
|-----------------------------------|-----------------------|------|------|-------------------|-----------------------------------|------------------------|
| Test Elutriate | Lab Percent Elutriate | | | riate | LC ₅₀ (% elutriate) | NOAEC (% elutriate) |
| | Control | 10 % | 50 % | 100% | - | |
| Inside 104 | 94 | 92 | 90 | 86 | > 100 | 100 |
| Outside 104 | 90 | 94 | 92 | 74 | > 100 | 100 |
| Brewerton Eastern Extension | 100 | 80 | 56 | 4 ^(a) | 53.4 | <10 |
| C&D Approaches – Cores | 91 | 98 | 90 | 76 ^(a) | > 100 | 50 |
| C&D Approaches – Surface Grabs | 94 | 78 | 54 | 10 ^(a) | 52.9 | 10 |
| Craighill | 96 | 98 | 60 | 38 ^(a) | 69.7 | 10 |
| Craighill Angle East | 92 | 97 | 60 | 0 ^(b) | 54.4 | 10 |
| Craighill Angle West | 92 | 97 | 77 | 0 ^(b) | 60.7 | 10 |
| Craighill Entrance | 92 | 96 | 72 | 22 ^(a) | 70.7 | 10 |
| Craighill Upper Range | 90 | 84 | 56 | 64 ^(a) | > 100 | 10 |
| Cutoff Angle | 92 | 94 | 50 | 0 ^(b) | 41.5 | 10 |
| Swan Point | 92 | 94 | 80 | 86 | > 100 | 100 |
| Tolchester North | 91 | 88 | 60 | 6 ^(a) | 50.8 | 10 |
| Tolchester South | 100 | 100 | 2 | 0 ^(b) | 23.8 | 10 |
| Tolchester Straightening | 91 | 90 | 92 | 80 | > 100 | -100 |

*Survival based on a mean of 5 replicate tests.

(a) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

TABLE 8-13BRESULTS OF ELUTRIATE TOXICITY TESTING WITH
Menidia beryllina (INLAND SILVERSIDE) ON SAMPLES FROM
THE NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA
(Round 3: Feb. 2000)

| Test Flutriate | 9 | 6-Hour Su | rvival (%) | 96-Hour | 96-Hour | |
|-----------------|---------|-------------------|------------|---------|---------------|---------------|
| I cot Enutrate | Lab | Percent Elutriate | | | LCso | NOAEC |
| | Control | 10 % | 50 % | 100% | (% elutriate) | (% elutriate) |
| Ocean Reference | 98 | 98 · | 100 | 98 | > 100 | 100 |

*Survival based on a mean of 5 replicate tests.

TABLE 8-14A RESULTS OF ELUTRIATE TOXICITY TESTING WITH *Mytilus* sp. (BLUE MUSSEL) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)⁽ⁿ⁾

| | 48-Hou | ır Normal | Developme | 48-Hour EC ₅₀ (% elutriate) | NOEC (% elutriate) | |
|-----------------------------------|-----------------|-----------|-------------|--|-----------------------|-----|
| Test Elutriate | Lab | Pe | rcent Eluti | | | |
| | Control | 10 % | 50 % | 100% | | |
| Inside 104 | 97 | 95 | 91 | 82 ^(b) | > 100 | 50 |
| Outside 104 | 96 | 92 | 85 | 2 ^(b) | 63.8 | 10 |
| Brewerton Eastern Extension | [.] 94 | 91 | 81 | 0.1 ^(b) | 60.9 | 10 |
| C&D Approaches – Cores | 98 | 92 | 0 | 0 ^(c) | 21.2 | <10 |
| C&D Approaches – Surface Grabs | 99 | 98 | 87 | 2 ^(b) | 64.2 | 10 |
| Craighill | 93 | 89 | 91 | 87 | > 100 | 100 |
| Craighill Angle East | 92 | 94 | 54 | 0 ^(c) | 43.5 | 10 |
| Craighill Angle West | 91 | 90 | 59 | 0 ^(c) | 46.8 | 10 |
| Craighill Entrance | 95 | 93 | 0 | 0 (c) | 22.0 | 10 |
| Craighill Upper Range | 90 | 89 | 86 | 87 | > 100 | 100 |
| Cutoff Angle | 96 | 98 | 92 | 27 ^(b) | 79.7 | 50 |
| Swan Point | 96 | 94 | 63 | 0 ^(c) | 47.8 | 10 |
| Tolchester North | 99 | 95 | 78 | 0.1 ^(b) | 56.4 | 10 |
| Tolchester South | 96 | 93 | 83 | 0 (c) | 61.7 | 10 |
| Tolchester Straightening | 95 | 95 | 84 | 0 ^(c) | 61.7 | 10 |

*Survival based on a mean of 5 replicate tests.

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

(b) Percent normal development in the 100 percent test concentration was statistically (p=0.05) lower than the laboratory control.

(c) A concentration which had no normally developed organisms was not statistically compared to the laboratory control.

TABLE 8-14B RESULTS OF ELUTRIATE TOXICITY TESTING WITH Mytilus sp.(BLUE MUSSEL) ON SAMPLES FROM THE NORFOLK OCEANDISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) ^(a)

| Test Elutriate | 48-Hou | ir Normal | Developme | 48-Hour | | |
|-----------------|-----------------------|-----------|-----------|------------------|-----------------------|-----|
| | Lab Percent Elutriate | | | EC ₅₀ | 96-Hour | |
| | Control 10 % 50 ° | 50 % | 100% | (% elutriate) | NOEC (% elutriate) | |
| Ocean Reference | 98.2 | 99.5 | 98.8 | 98.5 | > 100 | 100 |

*Survival based on a mcan of 5 replicate per tests.

(a) Embryo larval toxicity testing with Mytilus sp. was conducted by Northwestern Aquatic Sciences (NAS).

TABLE 8-15A RESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Neanthes arenaceodentata (ESTUARINE POLYCHAETE) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) |
|--------------------------------|----------------------------|-----------------------------------|
| Laboratory Control Sediment | 25 / 25 | 100 |
| Inside Site 104 | 24 / 25 | 96 |
| Brewerton Eastern Ext. | 25 / 25 | 100 |
| C&D Approaches - Cores | 24 / 25 | 96 |
| C&D Appr Surf. Grabs | 25 / 25 | 100 |
| Craighill | 24 / 25 | 96 |
| Craighill Angle East | 25 / 25 | 100 |
| Craighill Angle West | 23 / 25 | 92 |
| Craighill Entrance | 25 / 25 | 100 |
| Craighill Upper Range | 25 / 25 | 100 |
| Cutoff Angle | 24 / 25 | 96 |
| Swan Point | 25 / 25 | 100 |
| Tolchester North | 25 / 25 | 100 |
| Tolchester South | 24 / 25 | 96 |
| Tolchester Straightening | 24 / 25 | 96 |

*Based on 5 replicates of five animals each

TABLE 8-15BRESULTS OF ADDITIONAL 10-DAY WHOLE SEDIMENT TOXICITY
TESTING WITH Neanthes arenaceodentata (ESTUARINE POLYCHAETE) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | |
|--------------------------------|----------------------------|--------------------------------------|--|
| Inside Site 104 | 22 / 25 | 88 | |
| Outside Site 104 | 20 / 25 | 80 | |
| Laboratory Control Sediment | 24 / 25 | 96 | |

*Based on 5 replicates of five animals each

TABLE 8-16A RESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH Leptocheirus plumulosus (ESTUARINE AMPHIPOD) ON SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104 (Round 1: Sept.-Nov. 1999)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | |
|--------------------------------|----------------------------|--------------------------------------|--|
| Laboratory Control Sediment | 96 / 100 | .96 | |
| Inside Site 104 | 93 / 100 | 93 | |
| Brewerton Eastern Ext. | 88 / 100 | 88 | |
| C&D Approaches - Cores | 88 / 100 | 88 | |
| C&D Appr Surf. Grabs | 94 / 100 | 94 | |
| Craighill | 93 / 100 | 93 | |
| Craighill Angle East | 91 / 100 | 91 | |
| Craighill Angle West | 93 / 100 | 93 | |
| Craighill Entrance | 96 / 100 | 96 | |
| Craighill Upper Range | 86 / 100 | 86 | |
| Cutoff Angle | 92 / 100 | 92 | |
| Swan Point | 89/100 | 89 | |
| Tolchester North | 93 / 100 | 93 | |
| Tolchester South | 92 / 100 | 92 | |
| Tolchester Straightening | 89 / 100 | 89 | |

*Based on 5 replicates of 20 animals each
TABLE 8-16BRESULTS OF ADDITIONAL 10-DAY WHOLE SEDIMENT TOXICITY
TESTING WITH Leptocheirus plumulosus (ESTUARINE AMPHIPOD) ON
SAMPLES FROM BALTIMORE HARBOR APPROACH CHANNELS AND
PLACEMENT SITE 104 (Round 2: Dec. 1999-Feb. 2000)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | | |
|--------------------------------|----------------------------|--------------------------------------|--|--|
| Inside Site 104 | 87 / 100 | 87 ^(b) | | |
| Outside Site 104 | 96 / 100 | 96 | | |
| Laboratory Control Sediment | 98 / 100 | 98 | | |

(a) Survival in sample Inside Site 104 was statistically less (P=0.05) than the laboratory control sediment.

*Based on 5 replicates of 20 animals each

TABLE 8-16CRESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Leptocheirus plumulosus (ESTUARINE AMPHIPOD) FOR THE NORFOLK
OCEAN DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) ^(a)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | | |
|---------------------------------|----------------------------|--------------------------------------|--|--|
| Lower Chesapeake Bay Control | 94 / 100 | 94 | | |

(a) Lower Chesapeake Bay control sediment tested in place of Norfolk Ocean Disposal Site sediment to minimize potential grain size effect.

*Based on 5 replicates of 20 animals each

TABLE 8-17RESULTS OF 10-DAY WHOLE SEDIMENT TOXICITY TESTING WITH
Mysidopsis bahia (OPOSSUM SHRIMP) FOR THE NORFOLK OCEAN
DISPOSAL SITE REFERENCE AREA (Round 3: Feb. 2000) ^(a)

| Test Sediment | No. Alive/ No. Exposed* | 10-Day Percent Survival (mean) | | |
|-----------------|----------------------------|--------------------------------------|--|--|
| Ocean Reference | 47 / 50 | 94 | | |

(a) Mysidopsis bahia tested in place of Neanthes arenaceodentata for Ocean Reference testing.

*Based on 5 replicates of 10 animals each

TABLE 8-18A RESULTS OF REFERENCE TOXICANT TESTING – BALTIMORE HARBOR APPROACH CHANNELS AND PLACEMENT SITE 104

| Test Species | <u>Reference Toxicant</u> | Testing Date | <u>Endpoint</u> | Acceptable Control Chart <u>Limits</u> |
|--------------------------|--|---------------------|-----------------------------|---|
| Mysidopsis bahia | Potassium chloride | October 1999 | 48-hour LC50: 0.57 g/L KCl | 0.12-0.65 g/L KCl |
| | (KCI) | January 2000 | 48-1100F LC30. 0.71 g/L KCI | 0.32-0.86 g/L KCI |
| Cyprinodon variegatus | Potassium chloride | October 1999 | 48-hour LC50: 1.65 g/L KCl | 1.20-1.87 g/L KCl |
| | (KCl) | January 2000 | 48-hour LC50: 1.41 g/L KCl | 1.25-1.60 g/L KCl |
| Menidia beryllina | Potassium chloride (KCl) | January 2000 | 48-hour LC50: 1.18 g/L KCl | 1.10-1.49 g/L KCl |
| <i>Mytilus</i> sp. | Copper sulfate (CuSO ₄ •5H ₂ O) | December 1999 | 48-hour EC50: 10.3 μg/L Cu | 8.05-12.5 μg/L Cu |
| Neanthes arenaceodentata | Cadmium chloride | October 1999 | 96-hour LC50: 6.4 mg/L Cd | (a) |
| | $(CdCl_2)$ | December 1999 | 96-hour LC50: 5.7 mg/L Cd | (a) |
| Leptocheirus plumulosus | Cadmium chloride | October 1999 | 96-hour LC50: 0.22 mg/L Cd | 0.21-0.26 mg/L Cd |
| - • | $(CdCl_2)$ | January 2000 | 96-hour LC50: 0.24 mg/L Cd | 0.21-0.25 mg/L Cd |

(a) Control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50s of 6.4 and 5.7 mg/L Cd were similar to the LC50 of 5.7 mg/L Cd on a previous lot of *N. arenaceodentata* from September 1999.

TABLE 8-18B RESULTS OF REFERENCE TOXICANT TESTING - NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| Test Species Reference Toxican | | Endpoint | Acceptable Control Char Limits | | | |
|--------------------------------|--|-----------------------------|-----------------------------------|--|--|--|
| Mysidopsis bahia | Potassium chloride (KCl) | 96-hour LC50: 0.65 g/L KCl | 0.59-0.72 g/L KCl | | | |
| Menidia beryllina | Potassium chloride (KCl) | 96-hour LC50: 0.969 g/L KCl | 0.82-1.15 g/L KCl | | | |
| Leptocheirus plumulosus | Cadmium chloride (CdCl ₂) | 48-hour LC50: 5.9 mg/L Cd | (a) | | | |

(a) Control chart limits have not yet been established, because an insufficient number of reference toxicant tests (less than five) have been conducted. The LC50 of 5.9 mg/L Cd falls within the range of LC50's from previous *L. plumulosus* reference toxicant tests (2.2-9.0 mg/l Cd).

| | Opossum Shrimp | | Sheepshead Minnow | | | Blue Mussel | | | Inland Silverside | | | |
|---|------------------|----------------------------|-------------------|-----------------------|----------------------------|---------------|---------------|----------------------------|-------------------|-------------------|----------------------------|---------------|
| | Mysidopsis bahia | | | Cyprinodon variegatus | | | Mytilus sp. | | | Menidia beryllina | | |
| | | Statistical | | | Statistical | | | Statistical | | | Statistical | |
| | 96-hour LC50 | Difference 100% | 96-Hour NOAEC | 96-hour LC50 | Difference 100% | 96-Hour NOAEC | 48-hour EC50 | Difference 100% | 96-Hour NOEC | 96-hour LC 50 | Difference 100% | 96-Hour NOAEC |
| Sample Identification | (% clutriate) | vs. Control ^(a) | (% elutriate) | (% elutriate) | vs. Control ⁽ⁿ⁾ | (% elutriate) | (% elutriate) | vs. Control ^(a) | (% elutriate) | (% elutriate) | vs. Control ^(a) | (% elutriate) |
| | | | | | | | | | | | | |
| Inside Placement Site 104 | >100 | No | 100 | >100 | No | 100 | >100 | Yes | 50 | >100 | No | 100 |
| Brewerton Eastern Extension | >100 | Yes | 50 | ´ >100 | No | 100 | 60.9 | Yes | 10 | 53.4 | Yes | <10 |
| C&D Approaches (Cores) | >100 | No | 100 | >100 | No | 100 | 21.2 | Yes | <10 | >100 | Yes | 50 |
| C&D Approaches (Surficial) | >100 | Yes | 50 | >100 | No | 100 | 64.2 | Yes | 10 | 52.9 | Yes | 10 |
| Craighill | >100 | No | 100 | >100 | No | 100 | >100 | No | 100 | 69.7 | Yes | 10 |
| Craighill Entrance | >100 | Yes | 50 | >100 | No | 100 | 22.0 | Yes | 10 | 70.7 | Yes | 10 |
| Craighill Angle - East | >100 | No | 100 | >100 | No | 100 | 43.5 | Yes | 10 | 54.4 | Yes | 10 |
| Craighill Angle - West | >100 | Yes | 50 | >100 | No | 100 | 46.8 | Yes | 10 | 60.7 | Yes | 10 |
| Craighill Upper Range | >100 | No | 100 | >100 | No | 100 | >100 | No | 100 | >100 | Yes | 10 |
| Cutoff Angle | >100 | No | 100 | >100 | No | 100 | 79.7 | Yes | 50 | 41.5 | Yes | 10 |
| Swan Point | >100 | No | 100 | >100 | No | 100 | 47.8 | Yes | 10 | >100 | No | 100 |
| Tolchester Channel - North | >100 | No | 100 | >100 | No | 100 | 56.4 | Yes | 10 | 50.8 | Yes | 10 |
| Tolchester Channel - South | >100 | Yes | 100 | >100 | No | 100 | 61.7 | Yes | 10 | 23.8 | Yes | 10 |
| Tolchester Straightening | >100 | No | 100 | >100 | No | 100 | 61.7 | Ycs | 10 | >100 | No | 100 |
| | · | | | | | | | | | | | |
| Inside Placement Site 104 ^(b) | >100 | No | 100 | >100 | No | 100 | - | - | - | | - | - |
| Outside Placement Site 104 (*) | >100 | No | 100 | >100 | No | 100 | 63.8 | Yes | 10 | >100 | No | 100 |
| Norfolk Ocean Reference Site ⁽⁴⁾ | >100 | No | . 100 | NT | NT | NT | >100 | No | 100 | >100 | No | 100 |

TABLE 8-19A SUMMARY OF RESULTS FOR WATER COLUMN BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS, PLACEMENT SITE 104, OUTSIDE SITE 104 REFERENCE, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

NT = not tested: water column bioassays with Norfolk Ocean Reference elutriate conducted with opossum shrimp, hlue mussel, and inland silverside only

(a) Statistical significance analyzed at P=0.05; survival (I.C50) or normal development (EC50) in 100% elutriate significantly lower than the control.

(b) Inside Placement Site 104 elutriate re-tested concurrently with Outside Placement Site 104 water column bioassays (re-tested with opossum shrimp and sheepshead minnow only)

(c) Outside Placement Site 104 water column bioassays with opossum shrimp and sheepshead minnow conducted separately from channel elutriate hioassays

(d) Norfolk Ocean Reference testing conducted independently of Inside/Outside Placement Site 104 and channel elutriate tests

TABLE 8-19B SUMMARY OF RESULTS FOR WATER COLUMN BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS, PLACEMENT SITE 104, OUTSIDE SITE 104 REFERENCE, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

| | Opossum ShrImp Mysidopsis bahia | | Sheepshead Minnow Cyprinodon variegatus | | | Blue Mussel Mytilus sp. | | | Inland Silverside Menidia bervilina | | | |
|---|------------------------------------|--|---|-------------------------------|--|--|-------------------------------|--------------------------------|--|---------------|--------------------------------|--------------------------------------|
| Sample Identification | 96-hour LC50 (% elutriate) | Statistical Difference 100% vs. Control ^(a) | Mixing Factor Required to Achieve 0.01 LC50 | 96-hour LC50 (% elutriate) | Statistical Difference 100% vs. Control ^(a) | Mixing Factor Required to Achieve 0.01 LC50 | 48-hour EC50 (% elutriate) | Statistical Difference 100% | Mixing Factor Required to Achieve 0.01 EC50 | 96-hour LC50 | Statistical Difference 100% | Mixing Factor Required to Achieve |
| | | | | | | | | | Leso | (% eldulate) | vs. control | 0.01 LC 30 |
| Inside Placement Site 104 | >100 | No | NA | >100 | No | NA | >100 | Yes | <100 | >100 | No | NA |
| Brewerton Eastern Extension | >100 | Yes | <100 | >100 | No | NA | 60.9 | Yes | 164 | 53.4 | Yes | 188 |
| C&D Approaches (Cores) | >100 | No | NA | >100 | No | NA | 21.2 | Yes | 472 | >100 | Yes | <100 |
| C&D Approaches (Surficial) | >100 | Yes | <100 | >100 | No | NA | 64.2 | Yes | 156 | 52.9 | Yes | 189 |
| Craighill | >100 | No | NA | >100 | No | NA [·] | >100 | No | NA | 69.7 | Yes | 141 |
| Craighill Entrance | >100 | Yes | <100 | >100 | No | NA | 22.0 | Yes | 454 | 70.7 | Yes | 141 |
| Craighill Angle - East | >100 | No | NA | >100 | No | NA | 43.5 | Yes | 230 | 54.4 | Yes | 184 |
| Craighill Angle - West | >100 | Yes | <100 | >100 | No | NA | 46.8 | Yes | 214 | 60.7 | Yes | 165 |
| Craighill Upper Range | >100 | No | NA | >100 | No | NA | >100 | No | NA | >100 | Yes | NA |
| Cutoff Angle | >100 | No | NA | >100 | No | NA | 79.7 | Yes | 126 | 41.5 | Yes | 241 |
| Swan Point | >100 | No | NA | >100 | No | NA | 47.8 | Yes | 209 | >100 | No | NA |
| Tolchester Channel - North | - >100 | No | NA | >100 | No | NA | 56.4 | Yes | 177 | 50.8 | Yes | 197 |
| Tolchester Channel - South | >100 | Yes | <100 | >100 | No | NA | 61.7 | Yes | 162 | 23.8 | Yes | 420 |
| Tolchester Straightening | >100 | No | NA | >100 | No | NA | 61.7 | Yes | 162 | >100 | No | NA |
| Inside Placement Site 104 ^(b) | >100 | No | NA | >100 | No | NA | | | | | | |
| Outside Placement Site 104 ^(c) | >100 | No | NA | >100 | No | NA | 63.8 | Yes | 157 | >100 | No | NA |
| Norfolk Ocean Reference Site ^(d) | >100 | No | NA | NT | NT | - | >100 | . No | NA | >100 | No | NA |

NT = not tested; water column bioassays with Norfolk Ocean Reference elutriate conducted with opossum shrimp, blue mussel, and inland silverside only

(a) Statistical significance analyzed at P=0.05: survival (LC50) or normal development (EC50) in 100% elutriate significantly lower than the control.

(h) Inside Placement Site 104 elutriate re-tested concurrently with Outside Placement Site 104 water column bioassays (re-tested with opossum shrimp and sheepshead minnow only)

(c) Outside Placement Site 104 water column bioassays with opossum shrimp and sheepshead minnow conducted separately from channel elutriate bioassays

(d) Norfolk Ocean Reference testing conducted independently of Inside/Outside Placement Site 104 and channel elutriate tests

NA = mixing calculation is not applicable if mean survival in 100% elutriate is not statistically lower than the mean survival in the laboratory control.

| | | Estuarine Polychaet | e | Estuarine Amphipod | | | | |
|---|----------------------|---|--|-------------------------|---|--|--|--|
| | | Neanthes arenaceoda | ta | Leptocheirus plumulosus | | | | |
| Sample Identification | Mean Survival (%) | Statistical Difference vs. Inside Placement Site 104 ^(a) | 10% Difference vs. Inside Placement Site 104 | Mean Survival (%) | Statistical Difference vs. Inside Placement Site 104 ^(a) | 20% Difference vs. Inside Placement Site 104 | | |
| Laboratory Control Sediment ^(b) | 100 | NA | ΝA | | NA | NA | | |
| Inside Placement Site 104 | 96 | NA . | NA | 90 | NA | NA NA | | |
| Brewerton Eastern Extension | 100 | 'No | No | 88 | No | No | | |
| C&D Approaches (Cores) | 96 | No | No | 88 | No | No | | |
| C&D Approaches (Surficial) | 100 | No | No | 94 | No | No | | |
| Craighill | 96 | No | No | 93 | No | No | | |
| Craighill Entrance | 100 | No | No | 96 | No | No | | |
| Craighill Angle - East | 100 | No | No | . 91 | No | No | | |
| Craighill Angle - West | 92 | No | No | 93 | No | No | | |
| Craighill Upper Range | 100 | No | No | 86 | No | No | | |
| Cutoff Angle | · 96 | No | No | 92 | No | No | | |
| Swan Point | 100 | No | No | 89 | No | No | | |
| Tolchester Channel - North | 100 | No | No | 93 | · No | No | | |
| Tolchester Channel - South | 96 | No | No | 92 | No | No | | |
| Tolchester Straightening | 96 | No | No | 89 | No | No | | |
| Laboratory Control Sediment (b) | 96 | NA | NA | . 98 | NA | NA | | |
| Inside Placement Site 104 | 88 | NA | NA | 87 | . NA | NA | | |
| Outside Placement Site 104 | 80 . | NA | NA | 96 | NA | NA | | |
| Norfolk Ocean Reference Site (C)(d) | 94 | NA | NA | - | - | | | |
| Southern Chesapeake Bay Control Sediment (d)(e) | - | | - | 94 | NA | NA | | |

TABLE 8-20 SUMMARY OF RESULTS FOR WHOLE SEDIMENT BIOASSAYS BALTIMORE HARBOR APPROACH CHANNELS, INSIDE SITE 104, OUTSIDE SITE 104, AND NORFOLK OCEAN DISPOSAL SITE REFERENCE AREA

NA = not applicable; control and reference (Inside Site 104) survival not statistically compared as per USEPA/USACE guidelines (1998)

(a) Statistical significance analyzed at P=0.05; channel sediments statistically compared to Inside Placement Site 104

(b) Control serves as indicator for test acceptability/validity

(c) Mysidoposis bahia tested in place of Neanthes arenaceodentata for Norfolk Ocean Disposal Site reference area sediment

(d) Norfolk Ocean Reference tests conducted independently of Inside/Outside Placement Site 104 and channel whole-sediment tests

(e) Southern Chesapeake Bay Control Sediment tested with Leptocheirus in place of Norfolk Ocean Reference sediment as requested by EPA Region III-Philadelphia